

Portland State University

PDXScholar

Civil and Environmental Engineering Faculty
Publications and Presentations

Civil and Environmental Engineering

3-2007

Lake Roosevelt Water Quality and Hydrodynamic Model Calibration with Fish Bioenergetics

Michael Lee McKillip

Follow this and additional works at: https://pdxscholar.library.pdx.edu/cengin_fac



Part of the [Civil and Environmental Engineering Commons](#)

Let us know how access to this document benefits you.

Citation Details

McKillip, Mike, and Scott Wells. Lake Roosevelt Water Quality and Hydrodynamic Model Calibration with Fish Bioenergetics. Technical Report EWR-03-06, Department of Civil and Environmental Engineering, Portland State University. Portland, Oregon, 2007.

This Technical Report is brought to you for free and open access. It has been accepted for inclusion in Civil and Environmental Engineering Faculty Publications and Presentations by an authorized administrator of PDXScholar. Please contact us if we can make this document more accessible: pdxscholar@pdx.edu.

Lake Roosevelt Water Quality and Hydrodynamic Model Calibration with Fish Bioenergetics

By

Mike McKillip,

And

Scott Wells

Technical Report EWR-03-06

**PORTLAND STATE
UNIVERSITY**

Civil and Environmental Engineering Department
Maseeh College of Engineering and Computer Science
P.O. Box 751
Portland, Oregon 97201-0751

March 18, 2007

Table of Contents:

| | |
|--|------|
| Table of Contents:..... | i |
| Table of Figures | iv |
| Table of Tables | xiii |
| Acknowledgments..... | xiv |
| Introduction..... | 1 |
| Monitoring Sites..... | 3 |
| Hydrodynamic Calibration..... | 5 |
| Calibration Stations..... | 5 |
| Year 2000..... | 5 |
| Year 2001..... | 7 |
| Year 2002..... | 9 |
| Temperature Calibration | 11 |
| Calibration Stations..... | 12 |
| Grand Coulee Dam, Continuous Temperatures..... | 14 |
| Vertical Profile Stations, Periodic Sampling | 15 |
| Abiotic Water Quality Calibration..... | 18 |
| Calibration Stations..... | 18 |
| Constituent Calibration Discussion | 19 |
| Alkalinity and pH..... | 19 |
| Ammonium and Nitrate plus Nitrite | 19 |
| Dissolved Oxygen..... | 20 |
| Orthophosphate..... | 20 |
| Total Dissolved Solids | 20 |
| Grand Coulee Dam Outflow | 21 |
| Alkalinity | 22 |
| Ammonium | 22 |
| Dissolved Oxygen..... | 23 |
| Nitrate plus nitrite | 23 |
| pH | 24 |
| Orthophosphate..... | 25 |
| Total Dissolved Solids | 25 |
| Vertical Profile Stations, Periodic Sampling | 26 |
| Alternate Boundary Conditions | 33 |
| Spokane River Dissolved Oxygen | 33 |
| Total Inorganic Carbon (pH) | 34 |
| Biotic Modeling Approach and Calibration..... | 36 |
| Approach..... | 36 |
| Algae..... | 36 |
| Zooplankton | 37 |
| Fish Bioenergetics (kokanee)..... | 38 |
| Calibration | 40 |
| Algae..... | 40 |
| Zooplankton | 47 |

| | |
|---|-----|
| Fish Bioenergetics (kokanee)..... | 53 |
| Summary | 64 |
| References..... | 68 |
| Appendix A: Water quality boundary condition generation..... | 70 |
| Overview..... | 70 |
| Conductivity..... | 72 |
| Coliform bacteria | 73 |
| Alkalinity | 74 |
| Dissolved oxygen..... | 75 |
| Orthophosphate..... | 76 |
| Ammonium | 77 |
| Nitrate plus nitrite | 78 |
| Chlorophyll-a (algae)..... | 79 |
| pH | 80 |
| Total dissolved solids..... | 81 |
| Total inorganic carbon and total organic matter | 82 |
| Total inorganic carbon | 82 |
| Dissolved organic carbon..... | 82 |
| Total organic carbon | 83 |
| Total organic matter | 83 |
| Total suspended solids | 83 |
| Total inorganic suspended solids (ISS)..... | 83 |
| Dissolved organic matter | 83 |
| Particulate organic matter | 84 |
| Zooplankton | 85 |
| Data and Daily Boundary Condition Model Input Comparison..... | 86 |
| Mainstem Columbia River..... | 86 |
| Kettle River..... | 93 |
| Colville River..... | 96 |
| Spokane River..... | 98 |
| Sanpoil River | 101 |
| Banks Lake Return Flows..... | 102 |
| Appendix B: W2 control files..... | 105 |
| W2_con.npt..... | 105 |
| W2_bio_con.npt..... | 152 |
| Appendix C: Fish Bioenergetic Parameter Formulation..... | 153 |
| Growth | 153 |
| Fish energy density | 153 |
| Digestion..... | 154 |
| Digestion coefficient (function)..... | 155 |
| Stomach content and capacity..... | 155 |
| Consumption, Search Volume, and Reaction Distance | 156 |
| Egestion & Excretion..... | 157 |
| Specific Dynamic Action (SDA) | 158 |
| Respiration | 159 |
| Ancillary function values..... | 160 |

| | |
|---|-----|
| Appendix D: Bioenergetics control file explanation | 161 |
| Fish Computation..... | 162 |
| Bioenergetic time control..... | 162 |
| Fish temperature kinetics | 163 |
| Fish oxycaloric stoichiometric constant | 163 |
| Fish physical properties initial conditions | 163 |
| Zooplankton property cards | 164 |
| Foraging constants | 165 |
| Fish growth potential animation cards..... | 166 |
| Fish mass type..... | 167 |
| Fish mass function parameters..... | 167 |
| Diagnostic output controls | 168 |
| Single location controls | 168 |
| User defined input and output filenames | 169 |
| Appendix E: Plots of weighted model results and model-data comparisons..... | 170 |
| Station 0.0 | 171 |
| Station 1.0 | 175 |
| Station 2.0 | 179 |
| Station 3.0 | 183 |
| Station 4.0 | 187 |
| Station 5.5 | 191 |
| Station 6.0 | 194 |
| Station 6.5 | 198 |
| Station 7.0 | 200 |
| Station 8.0 | 203 |
| Station 8.5 | 207 |
| Station 9.0 | 211 |
| Appendix F: Vertical profile model-data comparison plots. | 215 |
| Total dissolved solids..... | 216 |
| Dissolved oxygen..... | 240 |
| Temperature | 264 |
| pH | 288 |
| Appendix G: Statistics Calculations | 312 |
| Appendix H: W2 Model Water Quality Parameters | 313 |

Table of Figures

| | |
|--|----|
| Figure 1. Lake Roosevelt model area. | 2 |
| Figure 2. Water quality sampling site locations..... | 4 |
| Figure 3. Model-data comparison, Grand Coulee Dam forebay stage, 2000. | 5 |
| Figure 4. Waterbalance flow magnitudes, 2000. | 6 |
| Figure 5. Waterbalance flows as percentage of downstream flows, 2000..... | 6 |
| Figure 6. Model-data comparison, Grand Coulee Dam forebay stage, 2001. | 7 |
| Figure 7. Waterbalance flow magnitudes, 2001. | 8 |
| Figure 8. Waterbalance flows as percentage of downstream flows, 2001..... | 8 |
| Figure 9. Model-data comparison, Grand Coulee Dam forebay stage, 2002. | 9 |
| Figure 10. Waterbalance flow magnitudes, 2002. | 10 |
| Figure 11. Waterbalance flows as percentage of downstream flows, 2002..... | 10 |
| Figure 12. The effects of wind sheltering coefficients on temperature calibration. The default value of WSC of 0.85 was not considered accurate but was used as a basis for comparison to the calibrated value. | 12 |
| Figure 13. Locations of the temperature calibration sites..... | 13 |
| Figure 14. Model-data temperature comparison, below Grand Coulee Dam, 2000..... | 15 |
| Figure 15. Selected model-data temperature profile comparisons at Porcupine Bay (LRFEP sta 4.0). | 16 |
| Figure 16. Selected model-data temperature profile comparisons at Spring Canyon (LRFEP sta 9.0). | 17 |
| Figure 17. Alkalinity time-series near Grand Coulee Dam. | 22 |
| Figure 18. Ammonium time-series near Grand Coulee Dam. | 22 |
| Figure 19. Dissolved oxygen time-series near Grand Coulee Dam..... | 23 |
| Figure 20. Nitrate time-series near Grand Coulee Dam. | 24 |
| Figure 21. pH time-series near Grand Coulee Dam..... | 24 |
| Figure 22. Orthophosphate time-series near Grand Coulee Dam. | 25 |
| Figure 23. Total dissolved solids time-series near Grand Coulee Dam..... | 26 |
| Figure 24. Selected model-data dissolved oxygen vertical profile comparisons at Porcupine Bay (LRFEP stat 4.0). | 27 |
| Figure 25. Selected model-data dissolved oxygen vertical profile comparisons at Spring Canyon (LRFEP sta 9.0). | 28 |
| Figure 26. Selected model-data total dissolved solids vertical profile comparisons at Porcupine Bay (LRFEP stat 4.0). | 29 |
| Figure 27. Selected model-data total dissolved solids vertical profile comparisons at Spring Canyon (LRFEP stat 9.0). | 30 |
| Figure 28. Selected model-data pH vertical profile comparisons at Porcupine Bay (LRFEP stat 4.0). | 31 |
| Figure 29. Selected model-data pH vertical profile comparisons at Spring Canyon (LRFEP stat 9.0). | 32 |
| Figure 30. Model-data comparison of dissolved oxygen at LRFEP station 4.0 under different boundary condition scenarios. | 34 |
| Figure 31. Model-data comparison of pH at LRFEP station 0.0 under different boundary condition scenarios..... | 35 |

| | |
|--|----|
| Figure 32. Model-data comparison of pH at LRFEP station 9.0 under different boundary condition scenarios..... | 35 |
| Figure 33. Conceptual diagram of the water quality, algae, and zooplankton interaction. | 36 |
| Figure 34. Model-data comparison of weighted total algae, LRFEP station 0.0..... | 40 |
| Figure 35. Model-data comparison of weighted total algae, LRFEP station 1.0..... | 41 |
| Figure 36. Model-data comparison of weighted total algae, LRFEP station 2.0..... | 41 |
| Figure 37. Model-data comparison of weighted total algae, LRFEP station 3.0..... | 42 |
| Figure 38. Model-data comparison of weighted total algae, LRFEP station 4.0..... | 42 |
| Figure 39. Model-data comparison of weighted total algae, LRFEP station 5.5..... | 43 |
| Figure 40. Model-data comparison of weighted total algae, LRFEP station 6.0..... | 43 |
| Figure 41. Model-data comparison of weighted total algae, LRFEP station 6.5..... | 44 |
| Figure 42. Model-data comparison of weighted total algae, LRFEP station 7.0..... | 44 |
| Figure 43. Model-data comparison of weighted total algae, LRFEP station 8.0..... | 45 |
| Figure 44. Model-data comparison of weighted total algae, LRFEP station 8.5..... | 45 |
| Figure 45. Model-data comparison of weighted total algae, LRFEP station 9.0..... | 46 |
| Figure 46. Model-data comparison of weighted total zooplankton, LRFEP station 0.0. | 47 |
| Figure 47. Model-data comparison of weighted total zooplankton, LRFEP station 1.0 | 47 |
| Figure 48. Model-data comparison of weighted total zooplankton, LRFEP station 2.0 | 48 |
| Figure 49. Model-data comparison of weighted total zooplankton, LRFEP station 3.0 | 48 |
| Figure 50. Model-data comparison of weighted total zooplankton, LRFEP station 4.0 | 49 |
| Figure 51. Model-data comparison of weighted total zooplankton, LRFEP station 5.5 | 49 |
| Figure 52. Model-data comparison of weighted total zooplankton, LRFEP station 6.0 | 50 |
| Figure 53. Model-data comparison of weighted total zooplankton, LRFEP station 6.5 | 50 |
| Figure 54. Model-data comparison of weighted total zooplankton, LRFEP station 7.0 | 51 |
| Figure 55. Model-data comparison of weighted total zooplankton, LRFEP station 8.0 | 51 |
| Figure 56. Model-data comparison of weighted total zooplankton, LRFEP station 8.5 | 52 |
| Figure 57. Model-data comparison of weighted total zooplankton, LRFEP station 9.0 | 52 |
| Figure 58. Base case temperature and prescribed fish mass function..... | 54 |
| Figure 59. Daily average consumption (includes nighttime) and prey density. | 54 |
| Figure 60. Daily growth and bioenergetic parameters..... | 55 |
| Figure 61. Daily maximum and actual consumption; p-values. | 55 |
| Figure 62. Daily growth and bioenergetic parameters, vertical foraging strategy..... | 57 |
| Figure 63. Comparison of fish location optimization strategies: best growth cell and vertical foraging. | 57 |
| Figure 64. Daily maximum and actual consumption for the foraging model. Comparison of best growth cell and vertical foraging p-values..... | 58 |
| Figure 65. Comparison of daily average consumption rates..... | 58 |
| Figure 66. Foraging depths at each time-step, best growth cell method..... | 59 |
| Figure 67. Foraging depths at each time-step, vertical foraging method..... | 59 |
| Figure 68. Comparison of water temperatures..... | 60 |
| Figure 69. Comparison of prey densities. | 61 |
| Figure 70. Comparison of consumption rates. | 61 |
| Figure 71. Comparison of p-values..... | 62 |
| Figure 72. Comparison of respiration and daily growth. | 62 |
| Figure 73. Growth and Consumption rate at C_{max} for a 10 g and 100g kokanee. | 63 |
| Figure 74. System wide daily fraction of total algal chl-a. | 80 |

| | |
|--|-----|
| Figure 75. TDS:Conductivity ratio at LRFEP station 0.0, 1999 to 2002. | 81 |
| Figure 76. Mainstem Columbia River conductivity boundary condition, 2000. | 86 |
| Figure 77. Mainstem Columbia River coliform boundary condition, 2000. | 87 |
| Figure 78. Mainstem Columbia River dissolved oxygen boundary condition, 2000. | 87 |
| Figure 79. Mainstem Columbia River alkalinity boundary condition, 2000. | 88 |
| Figure 80. Mainstem Columbia River orthophosphate boundary condition, 2000..... | 88 |
| Figure 81. Mainstem Columbia River ammonium boundary condition, 2000..... | 89 |
| Figure 82. Mainstem Columbia River nitrate plus nitrite boundary condition, 2000..... | 89 |
| Figure 83. Mainstem Columbia River algal dry weight boundary condition, 2000. | 90 |
| Figure 84. Mainstem Columbia River total suspended solids boundary condition, 2000..... | 90 |
| Figure 85. Mainstem Columbia River inorganic suspended solids boundary condition, 2000. .. | 90 |
| Figure 86. Mainstem Columbia River total organic carbon boundary condition, 2000. | 91 |
| Figure 87. Mainstem Columbia River total organic matter boundary condition, 2000..... | 91 |
| Figure 88. Mainstem Columbia River particulate organic matter boundary condition, 2000. | 92 |
| Figure 89. Mainstem Columbia River dissolved organic matter boundary condition, 2000..... | 92 |
| Figure 90. Kettle River conductivity boundary condition, 2000. | 93 |
| Figure 91. Kettle River coliform boundary condition, 2000. | 93 |
| Figure 92. Kettle River dissolved oxygen boundary condition, 2000. | 94 |
| Figure 93. Kettle River orthophosphate boundary condition, 2000..... | 94 |
| Figure 94. Kettle River ammonium boundary condition, 2000..... | 95 |
| Figure 95. Kettle River nitrate plus nitrite boundary condition, 2000..... | 95 |
| Figure 96. Colville River coliform boundary condition, 2000. | 96 |
| Figure 97. Colville River orthophosphate boundary condition, 2000. | 96 |
| Figure 98. Colville River ammonium boundary condition, 2000..... | 97 |
| Figure 99. Colville River nitrate plus nitrite boundary condition, 2000..... | 97 |
| Figure 100. Spokane River conductivity boundary condition, 2000. | 98 |
| Figure 101. Spokane River dissolved oxygen boundary condition, 2000. | 98 |
| Figure 102. Spokane River alkalinity boundary condition, 2000. | 99 |
| Figure 103. Spokane River orthophosphate boundary condition, 2000..... | 99 |
| Figure 104. Spokane River ammonium boundary condition, 2000..... | 100 |
| Figure 105. Spokane River nitrate plus nitrite boundary condition, 2000..... | 100 |
| Figure 106. Sanpoil River dissolved oxygen boundary condition, 2000..... | 101 |
| Figure 107. Banks Lake return flow conductivity boundary condition, 2000..... | 102 |
| Figure 108. Banks Lake return flow coliform boundary condition, 2000. | 102 |
| Figure 109. Banks Lake return flow dissolved oxygen boundary condition, 2000. | 103 |
| Figure 110. Banks Lake return flow orthophosphate boundary condition, 2000. | 103 |
| Figure 111. Banks Lake return flow ammonium boundary condition, 2000..... | 104 |
| Figure 112. Banks Lake return flow nitrate plus nitrite boundary condition, 2000..... | 104 |
| Figure 113. Plot of the Thornton-Lessem function..... | 160 |
| Figure 114. Model-data comparison of orthophosphate at LRFEP station 0.0. | 171 |
| Figure 115. Model-data comparison of ammonium at LRFEP station 0.0..... | 171 |
| Figure 116. Model-data comparison of nitrate plus nitrite at LRFEP station 0.0..... | 172 |
| Figure 117. Model-data comparison of dissolved oxygen at LRFEP station 0.0. | 172 |
| Figure 118. Model-data comparison of alkalinity at LRFEP station 0.0..... | 173 |
| Figure 119. Model-data comparison of pH at LRFEP station 0.0. | 173 |
| Figure 120. Model-data comparison of total dissolved solids at LRFEP station 0.0..... | 174 |

| | |
|--|-----|
| Figure 121. Model-data comparison of orthophosphate at LRFEP station 1.0. | 175 |
| Figure 122. Model-data comparison of ammonium at LRFEP station 1.0. | 175 |
| Figure 123. Model-data comparison of nitrate plus nitrite at LRFEP station 1.0. | 176 |
| Figure 124. Model-data comparison of dissolved oxygen at LRFEP station 1.0. | 176 |
| Figure 125. Model-data comparison of alkalinity at LRFEP station 1.0. | 177 |
| Figure 126. Model-data comparison of pH at LRFEP station 1.0. | 177 |
| Figure 127. Model-data comparison of total dissolved solids at LRFEP station 1.0. | 178 |
| Figure 128. Model-data comparison of orthophosphate at LRFEP station 2.0. | 179 |
| Figure 129. Model-data comparison of ammonium at LRFEP station 2.0. | 179 |
| Figure 130. Model-data comparison of nitrate plus nitrite at LRFEP station 2.0. | 180 |
| Figure 131. Model-data comparison of dissolved oxygen at LRFEP station 2.0. | 180 |
| Figure 132. Model-data comparison of alkalinity at LRFEP station 2.0. | 181 |
| Figure 133. Model-data comparison of pH at LRFEP station 2.0. | 181 |
| Figure 134. Model-data comparison of total dissolved solids at LRFEP station 2.0. | 182 |
| Figure 135. Model-data comparison of orthophosphate at LRFEP station 3.0. | 183 |
| Figure 136. Model-data comparison of ammonium at LRFEP station 3.0. | 183 |
| Figure 137. Model-data comparison of nitrate plus nitrite at LRFEP station 3.0. | 184 |
| Figure 138. Model-data comparison of dissolved oxygen at LRFEP station 3.0. | 184 |
| Figure 139. Model-data comparison of alkalinity at LRFEP station 3.0. | 185 |
| Figure 140. Model-data comparison of pH at LRFEP station 3.0. | 185 |
| Figure 141. Model-data comparison of total dissolved solids at LRFEP station 3.0. | 186 |
| Figure 142. Model-data comparison of orthophosphate at LRFEP station 4.0. | 187 |
| Figure 143. Model-data comparison of ammonium at LRFEP station 4.0. | 187 |
| Figure 144. Model-data comparison of nitrate plus nitrite at LRFEP station 4.0. | 188 |
| Figure 145. Model-data comparison of dissolved oxygen at LRFEP station 4.0. | 188 |
| Figure 146. Model-data comparison of alkalinity at LRFEP station 4.0. | 189 |
| Figure 147. Model-data comparison of pH at LRFEP station 4.0. | 189 |
| Figure 148. Model-data comparison of total dissolved solids at LRFEP station 4.0. | 190 |
| Figure 149. Model-data comparison of orthophosphate at LRFEP station 5.5. | 191 |
| Figure 150. Model-data comparison of ammonium at LRFEP station 5.5. | 191 |
| Figure 151. Model-data comparison of nitrate plus nitrite at LRFEP station 5.5. | 192 |
| Figure 152. Model-data comparison of dissolved oxygen at LRFEP station 5.5. | 192 |
| Figure 153. Model-data comparison of pH at LRFEP station 5.5. | 193 |
| Figure 154. Model-data comparison of total dissolved solids at LRFEP station 5.5. | 193 |
| Figure 155. Model-data comparison of orthophosphate at LRFEP station 6.0. | 194 |
| Figure 156. Model-data comparison of ammonium at LRFEP station 6.0. | 194 |
| Figure 157. Model-data comparison of nitrate plus nitrite at LRFEP station 6.0. | 195 |
| Figure 158. Model-data comparison of dissolved oxygen at LRFEP station 6.0. | 195 |
| Figure 159. Model-data comparison of alkalinity at LRFEP station 6.0. | 196 |
| Figure 160. Model-data comparison of pH at LRFEP station 6.0. | 196 |
| Figure 161. Model-data comparison of total dissolved solids at LRFEP station 6.0. | 197 |
| Figure 162. Model-data comparison of dissolved oxygen at LRFEP station 6.5. | 198 |
| Figure 163. Model-data comparison of pH at LRFEP station 6.5. | 198 |
| Figure 164. Model-data comparison of total dissolved solids at LRFEP station 6.5. | 199 |
| Figure 165. Model-data comparison of orthophosphate at LRFEP station 7.0. | 200 |
| Figure 166. Model-data comparison of ammonium at LRFEP station 7.0. | 200 |

| | |
|---|-----|
| Figure 167. Model-data comparison of nitrate plus nitrite at LRFEP station 7.0..... | 201 |
| Figure 168. Model-data comparison of dissolved oxygen at LRFEP station 7.0. | 201 |
| Figure 169. Model-data comparison of alkalinity at LRFEP station 7.0. | 202 |
| Figure 170. Model-data comparison of pH at LRFEP station 7.0. | 202 |
| Figure 171. Model-data comparison of total dissolved solids at LRFEP station 7.0..... | 203 |
| Figure 172. Model-data comparison of orthophosphate at LRFEP station 8.0. | 203 |
| Figure 173. Model-data comparison of ammonium at LRFEP station 8.0..... | 204 |
| Figure 174. Model-data comparison of nitrate plus nitrite at LRFEP station 8.0..... | 204 |
| Figure 175. Model-data comparison of dissolved oxygen at LRFEP station 8.0. | 205 |
| Figure 176. Model-data comparison of alkalinity at LRFEP station 8.0..... | 205 |
| Figure 177. Model-data comparison of pH at LRFEP station 8.0. | 206 |
| Figure 178. Model-data comparison of total dissolved solids at LRFEP station 8.0..... | 206 |
| Figure 179. Model-data comparison of orthophosphate at LRFEP station 8.5. | 207 |
| Figure 180. Model-data comparison of ammonium at LRFEP station 8.5..... | 207 |
| Figure 181. Model-data comparison of nitrate plus nitrite at LRFEP station 8.5..... | 208 |
| Figure 182. Model-data comparison of dissolved oxygen at LRFEP station 8.5. | 208 |
| Figure 183. Model-data comparison of alkalinity at LRFEP station 8.5..... | 209 |
| Figure 184. Model-data comparison of pH at LRFEP station 8.5. | 209 |
| Figure 185. Model-data comparison of total dissolved solids at LRFEP station 8.5..... | 210 |
| Figure 186. Model-data comparison of orthophosphate at LRFEP station 9.0. | 211 |
| Figure 187. Model-data comparison of ammonium at LRFEP station 9.0..... | 211 |
| Figure 188. Model-data comparison of nitrate plus nitrite at LRFEP station 9.0..... | 212 |
| Figure 189. Model-data comparison of dissolved oxygen at LRFEP station 9.0. | 212 |
| Figure 190. Model-data comparison of alkalinity at LRFEP station 9.0..... | 213 |
| Figure 191. Model-data comparison of pH at LRFEP station 9.0. | 213 |
| Figure 192. Model-data comparison of total dissolved solids at LRFEP station 9.0..... | 214 |
| Figure 193. Vertical total dissolved solids model-data comparison, J384. | 216 |
| Figure 194. Vertical total dissolved solids model-data comparison, J385. | 217 |
| Figure 195. Vertical total dissolved solids model-data comparison, J410. | 217 |
| Figure 196. Vertical total dissolved solids model-data comparison, J411. | 218 |
| Figure 197. Vertical total dissolved solids model-data comparison, J412. | 218 |
| Figure 198. Vertical total dissolved solids model-data comparison, J445. | 219 |
| Figure 199. Vertical total dissolved solids model-data comparison, J446. | 219 |
| Figure 200. Vertical total dissolved solids model-data comparison, J447. | 220 |
| Figure 201. Vertical total dissolved solids model-data comparison, J448. | 220 |
| Figure 202. Vertical total dissolved solids model-data comparison, J473. | 221 |
| Figure 203. Vertical total dissolved solids model-data comparison, J474. | 221 |
| Figure 204. Vertical total dissolved solids model-data comparison, J475. | 222 |
| Figure 205. Vertical total dissolved solids model-data comparison, J489. | 222 |
| Figure 206. Vertical total dissolved solids model-data comparison, J490. | 223 |
| Figure 207. Vertical total dissolved solids model-data comparison, J491. | 223 |
| Figure 208. Vertical total dissolved solids model-data comparison, J502. | 224 |
| Figure 209. Vertical total dissolved solids model-data comparison, J503. | 224 |
| Figure 210. Vertical total dissolved solids model-data comparison, J522. | 225 |
| Figure 211. Vertical total dissolved solids model-data comparison, J523. | 225 |
| Figure 212. Vertical total dissolved solids model-data comparison, J524. | 226 |

| | |
|---|-----|
| Figure 259. Vertical dissolved oxygen profile model-data comparison, J524. | 250 |
| Figure 260. Vertical dissolved oxygen profile model-data comparison, J536. | 250 |
| Figure 261. Vertical dissolved oxygen profile model-data comparison, J537. | 251 |
| Figure 262. Vertical dissolved oxygen profile model-data comparison, J552. | 251 |
| Figure 263. Vertical dissolved oxygen profile model-data comparison, J553. | 252 |
| Figure 264. Vertical dissolved oxygen profile model-data comparison, J554. | 252 |
| Figure 265. Vertical dissolved oxygen profile model-data comparison, J572. | 253 |
| Figure 266. Vertical dissolved oxygen profile model-data comparison, J573. | 253 |
| Figure 267. Vertical dissolved oxygen profile model-data comparison, J585. | 254 |
| Figure 268. Vertical dissolved oxygen profile model-data comparison, J586. | 254 |
| Figure 269. Vertical dissolved oxygen profile model-data comparison, J587. | 255 |
| Figure 270. Vertical dissolved oxygen profile model-data comparison, J599. | 255 |
| Figure 271. Vertical dissolved oxygen profile model-data comparison, J600. | 256 |
| Figure 272. Vertical dissolved oxygen profile model-data comparison, J620. | 256 |
| Figure 273. Vertical dissolved oxygen profile model-data comparison, J621. | 257 |
| Figure 274. Vertical dissolved oxygen profile model-data comparison, J622. | 257 |
| Figure 275. Vertical dissolved oxygen profile model-data comparison, J634. | 258 |
| Figure 276. Vertical dissolved oxygen profile model-data comparison, J635. | 258 |
| Figure 277. Vertical dissolved oxygen profile model-data comparison, J648. | 259 |
| Figure 278. Vertical dissolved oxygen profile model-data comparison, J650. | 259 |
| Figure 279. Vertical dissolved oxygen profile model-data comparison, J651. | 260 |
| Figure 280. Vertical dissolved oxygen profile model-data comparison, J663. | 260 |
| Figure 281. Vertical dissolved oxygen profile model-data comparison, J664. | 261 |
| Figure 282. Vertical dissolved oxygen profile model-data comparison, J685. | 261 |
| Figure 283. Vertical dissolved oxygen profile model-data comparison, J686. | 262 |
| Figure 284. Vertical dissolved oxygen profile model-data comparison, J712. | 262 |
| Figure 285. Vertical dissolved oxygen profile model-data comparison, J714. | 263 |
| Figure 286. Vertical dissolved oxygen profile model-data comparison, J720. | 263 |
| Figure 287. Vertical temperature profile model-data comparison, J384. | 264 |
| Figure 288. Vertical temperature profile model-data comparison, J385. | 265 |
| Figure 289. Vertical temperature profile model-data comparison, J410. | 265 |
| Figure 290. Vertical temperature profile model-data comparison, J411. | 266 |
| Figure 291. Vertical temperature profile model-data comparison, J412. | 266 |
| Figure 292. Vertical temperature profile model-data comparison, J445. | 267 |
| Figure 293. Vertical temperature profile model-data comparison, J446. | 267 |
| Figure 294. Vertical temperature profile model-data comparison, J447. | 268 |
| Figure 295. Vertical temperature profile model-data comparison, J448. | 268 |
| Figure 296. Vertical temperature profile model-data comparison, J473. | 269 |
| Figure 297. Vertical temperature profile model-data comparison, J474. | 269 |
| Figure 298. Vertical temperature profile model-data comparison, J475. | 270 |
| Figure 299. Vertical temperature profile model-data comparison, J489. | 270 |
| Figure 300. Vertical temperature profile model-data comparison, J490. | 271 |
| Figure 301. Vertical temperature profile model-data comparison, J491. | 271 |
| Figure 302. Vertical temperature profile model-data comparison, J502. | 272 |
| Figure 303. Vertical temperature profile model-data comparison, J503. | 272 |
| Figure 304. Vertical temperature profile model-data comparison, J522. | 273 |

| | |
|--|-----|
| Figure 305. Vertical temperature profile model-data comparison, J523. | 273 |
| Figure 306. Vertical temperature profile model-data comparison, J524. | 274 |
| Figure 307. Vertical temperature profile model-data comparison, J536. | 274 |
| Figure 308. Vertical temperature profile model-data comparison, J537. | 275 |
| Figure 309. Vertical temperature profile model-data comparison, J552. | 275 |
| Figure 310. Vertical temperature profile model-data comparison, J553. | 276 |
| Figure 311. Vertical temperature profile model-data comparison, J554. | 276 |
| Figure 312. Vertical temperature profile model-data comparison, J572. | 277 |
| Figure 313. Vertical temperature profile model-data comparison, J573. | 277 |
| Figure 314. Vertical temperature profile model-data comparison, J585. | 278 |
| Figure 315. Vertical temperature profile model-data comparison, J586. | 278 |
| Figure 316. Vertical temperature profile model-data comparison, J587. | 279 |
| Figure 317. Vertical temperature profile model-data comparison, J599. | 279 |
| Figure 318. Vertical temperature profile model-data comparison, J600. | 280 |
| Figure 319. Vertical temperature profile model-data comparison, J620. | 280 |
| Figure 320. Vertical temperature profile model-data comparison, J621. | 281 |
| Figure 321. Vertical temperature profile model-data comparison, J622. | 281 |
| Figure 322. Vertical temperature profile model-data comparison, J634. | 282 |
| Figure 323. Vertical temperature profile model-data comparison, J635. | 282 |
| Figure 324. Vertical temperature profile model-data comparison, J648. | 283 |
| Figure 325. Vertical temperature profile model-data comparison, J650. | 283 |
| Figure 326. Vertical temperature profile model-data comparison, J651. | 284 |
| Figure 327. Vertical temperature profile model-data comparison, J663. | 284 |
| Figure 328. Vertical temperature profile model-data comparison, J664. | 285 |
| Figure 329. Vertical temperature profile model-data comparison, J685. | 285 |
| Figure 330. Vertical temperature profile model-data comparison, J686. | 286 |
| Figure 331. Vertical temperature profile model-data comparison, J712. | 286 |
| Figure 332. Vertical temperature profile model-data comparison, J714. | 287 |
| Figure 333. Vertical temperature profile model-data comparison, J720. | 287 |
| Figure 334. Vertical pH model-data comparison, J384. | 288 |
| Figure 335. Vertical pH model-data comparison, J385. | 289 |
| Figure 336. Vertical pH model-data comparison, J410. | 289 |
| Figure 337. Vertical pH model-data comparison, J411. | 290 |
| Figure 338. Vertical pH model-data comparison, J412. | 290 |
| Figure 339. Vertical pH model-data comparison, J445. | 291 |
| Figure 340. Vertical pH model-data comparison, J446. | 291 |
| Figure 341. Vertical pH model-data comparison, J447. | 292 |
| Figure 342. Vertical pH model-data comparison, J448. | 292 |
| Figure 343. Vertical pH model-data comparison, J473. | 293 |
| Figure 344. Vertical pH model-data comparison, J474. | 293 |
| Figure 345. Vertical pH model-data comparison, J475. | 294 |
| Figure 346. Vertical pH model-data comparison, J489. | 294 |
| Figure 347. Vertical pH model-data comparison, J490. | 295 |
| Figure 348. Vertical pH model-data comparison, J491. | 295 |
| Figure 349. Vertical pH model-data comparison, J502. | 296 |
| Figure 350. Vertical pH model-data comparison, J503. | 296 |

| | |
|--|-----|
| Figure 351. Vertical pH model-data comparison, J522..... | 297 |
| Figure 352. Vertical pH model-data comparison, J523..... | 297 |
| Figure 353. Vertical pH model-data comparison, J524..... | 298 |
| Figure 354. Vertical pH model-data comparison, J536..... | 298 |
| Figure 355. Vertical pH model-data comparison, J537..... | 299 |
| Figure 356. Vertical pH model-data comparison, J552..... | 299 |
| Figure 357. Vertical pH model-data comparison, J553..... | 300 |
| Figure 358. Vertical pH model-data comparison, J554..... | 300 |
| Figure 359. Vertical pH model-data comparison, J572..... | 301 |
| Figure 360. Vertical pH model-data comparison, J573..... | 301 |
| Figure 361. Vertical pH model-data comparison, J585..... | 302 |
| Figure 362. Vertical pH model-data comparison, J586..... | 302 |
| Figure 363. Vertical pH model-data comparison, J587..... | 303 |
| Figure 364. Vertical pH model-data comparison, J599..... | 303 |
| Figure 365. Vertical pH model-data comparison, J600..... | 304 |
| Figure 366. Vertical pH model-data comparison, J620..... | 304 |
| Figure 367. Vertical pH model-data comparison, J621..... | 305 |
| Figure 368. Vertical pH model-data comparison, J622..... | 305 |
| Figure 369. Vertical pH model-data comparison, J634..... | 306 |
| Figure 370. Vertical pH model-data comparison, J635..... | 306 |
| Figure 371. Vertical pH model-data comparison, J648..... | 307 |
| Figure 372. Vertical pH model-data comparison, J650..... | 307 |
| Figure 373. Vertical pH model-data comparison, J651..... | 308 |
| Figure 374. Vertical pH model-data comparison, J663..... | 308 |
| Figure 375. Vertical pH model-data comparison, J664..... | 309 |
| Figure 376. Vertical pH model-data comparison, J685..... | 309 |
| Figure 377. Vertical pH model-data comparison, J686..... | 310 |
| Figure 378. Vertical pH model-data comparison, J712..... | 310 |
| Figure 379. Vertical pH model-data comparison, J714..... | 311 |
| Figure 380. Vertical pH model-data comparison, J720..... | 311 |

Table of Tables

| | |
|---|-----|
| Table 1. Model calibration periods..... | 3 |
| Table 2. Grand Coulee Dam forebay stage statistics, 2000..... | 6 |
| Table 3. Grand Coulee Dam forebay stage statistics, 2001..... | 7 |
| Table 4. Grand Coulee Dam forebay stage statistics, 2002..... | 9 |
| Table 5. LFREP vertical profile stations..... | 14 |
| Table 6. Grand Coulee Dam temperature statistics, 2000..... | 15 |
| Table 7. Organization of the Grand Coulee Dam outflow constituent model-data comparison statistics..... | 18 |
| Table 8. Discrete constituent model-data comparison statistics..... | 21 |
| Table 9. Calibration statistics summary, 2000..... | 64 |
| Table 10. Conductivity boundary condition generation summary..... | 72 |
| Table 11. Coliform bacteria boundary condition generation summary..... | 73 |
| Table 12. Alkalinity boundary condition generation summary..... | 74 |
| Table 13. Dissolved oxygen boundary condition generation summary..... | 75 |
| Table 14. Orthophosphate boundary condition generation summary..... | 76 |
| Table 15. Ammonium boundary condition generation summary..... | 77 |
| Table 16. Nitrate plus nitrite boundary condition generation summary..... | 78 |
| Table 17. Chlorophyll-a (algae) boundary condition generation summary..... | 79 |
| Table 18. pH boundary condition generation summary..... | 80 |
| Table 19. Dissolved organic carbon boundary condition generation summary..... | 82 |
| Table 20. Total suspended solids boundary condition generation summary..... | 83 |
| Table 21. Zooplankton boundary condition generation summary..... | 85 |
| Table 22. Bioenergetics parameters summary..... | 153 |
| Table 23. Digestion parameter variables and units..... | 154 |
| Table 24. Digestion coefficient formulations..... | 155 |
| Table 25. Consumption parameter variables and units..... | 156 |
| Table 26. Egestion parameter variables and units..... | 157 |
| Table 27. Excretion parameter variables and units..... | 157 |
| Table 28. Specific dynamic action parameter variables and units..... | 158 |
| Table 29. Respiration parameter variables and units..... | 159 |
| Table 30. Representative ancillary function values for 76 and 125 g kokanee at 4 and 20 °C..... | 160 |
| Table 31. Vertical profile calibration statistics, 2000..... | 215 |
| Table 32. W2 Model Water Quality Parameters Summary..... | 313 |

Acknowledgments

Funds were provided by the Bonneville Power Administration to the Spokane Tribe of Indians to conduct this project for the Lake Roosevelt Fisheries Evaluation Program. Deanne-Pavlik-Kunkel and Ben Scofield (Lake Roosevelt Fisheries Evaluation Program) provided data, system knowledge, and data interpretation. Dr. Mike Mazur and Dr. Dave Beauchamp (University of Washington, School of Aquatic and Fishery Sciences) collaborated on the fish bioenergetics modeling. Dr. Robert Annear and Dr. Chris Berger (Portland State University, Water Resources Research Group) provided technical assistance and advice for the hydrodynamic and water quality model.

Introduction

An understanding of the effects of hydrodynamics and reservoir operations on the Franklin D. Roosevelt Lake (Lake Roosevelt) aquatic food web allows for better management of the reservoir. A CE-QUAL-W2, v.3.5, hydrodynamic and water quality model (Cole and Wells, 2006¹) is being applied to the reservoir. The models zooplankton algorithms are expanded and a fish bioenergetics model is incorporated. The Lake Roosevelt model extent is shown in Figure 1. The model includes the lacustrine arms up to full pool on the Sanpoil, Kettle, and Colville Rivers; the Spokane River arm up to Little Falls Dam; and the Columbia River from Grand Coulee Dam to the U.S.-Canadian border.

The previous companion report, “Boundary Conditions and Set-up” (McKillip, Annear, and Wells, 2006), covered

- A limnological overview
- Hydrodynamic boundary condition data and model inputs
- Grand Coulee Dam structures (powerhouse and spillway characteristics)
- Water temperature boundary condition data and model inputs
- Meteorological data and model inputs
- Water quality boundary condition data
- Model bathymetry data and model grid development
- Topographic shading
- Primary and secondary production data
- Kokanee hatchery release data

This report discusses the model calibration and issues related to the calibration. This report discusses the topics of

- 1) Hydrodynamic calibration: Hydrodynamic calibration focuses on matching the water surface elevation at Grand Coulee Dam.
- 2) Temperature calibration: Temperature calibration focuses on matching temperature profiles throughout the reservoir and continuous data below Grand Coulee Dam. Many of the calibration issues centered on properly characterizing the localized wind and powerhouse withdrawals.
- 3) Abiotic water quality calibration: Abiotic water quality calibration focused on matching water quality profile data in the reservoir. The selection of proper rate kinetics and understanding the impact of hydrodynamics on water quality state variables was critical to proper calibration

¹ The CE-QUAL-W2 version was upgraded from v.3.2 to v.3.5 since release of the Data Report (McKillip, Annear, and Wells, 2006 (EWR-01-05)). References to the User Manual also changed from Cole and Wells, 2004 (v.3.2) to Cole and Wells, 2006 (v.3.5.)

4) Bioenergetic (algae, zooplankton, kokanee) modeling approach and calibration

5) Sensitivity analyses

The hydrodynamic calibration was performed for three independent years: 2000, 2001, and 2002. Temperature, water quality and bioenergetic calibrations were performed for 2000. Table 1 shows the model calibration periods.

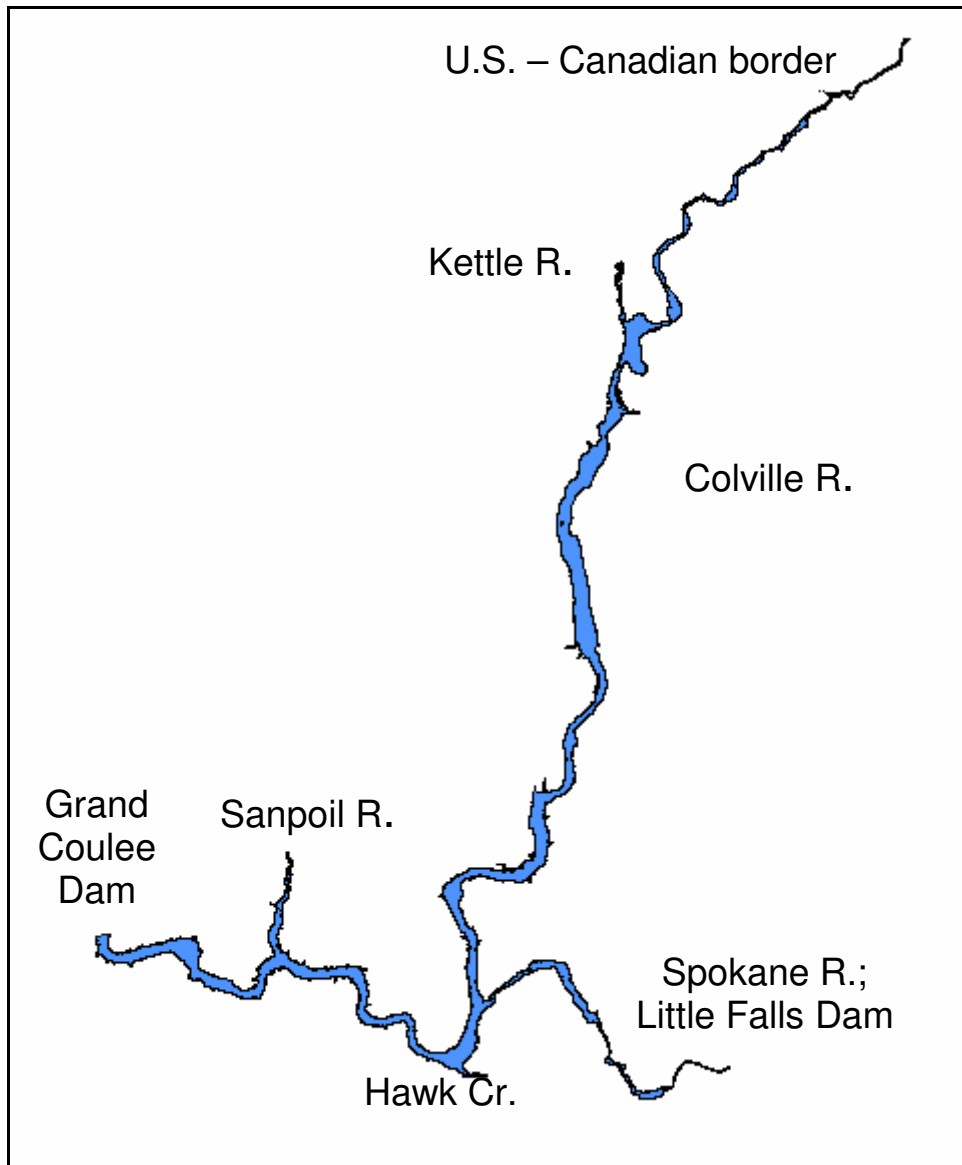


Figure 1. Lake Roosevelt model area.

Table 1. Model calibration periods.

| Calibration | Start date | End date |
|--------------------|-------------------|-------------------|
| Hydrodynamic | January 1, 2000 | December 31, 2000 |
| | January 1, 2001 | December 31, 2001 |
| | January 1, 2002 | December 31, 2002 |
| Water temperature | January 1, 2000 | December 31, 2000 |
| Water quality | January 1, 2000 | December 31, 2000 |
| Bioenergetics | January 1, 2000 | December 31, 2000 |

Monitoring Sites

The water quality monitoring sites are discussed in McKillip, Annear, and Wells (2005). Figure 2 shows the locations of the water quality monitoring stations. Hydrodynamic calibration sites include USACOE FDRW at Grand Coulee Dam and GCGW downstream of the dam. These sites were used for boundary conditions and for model-data comparisons during calibration within Lake Roosevelt.

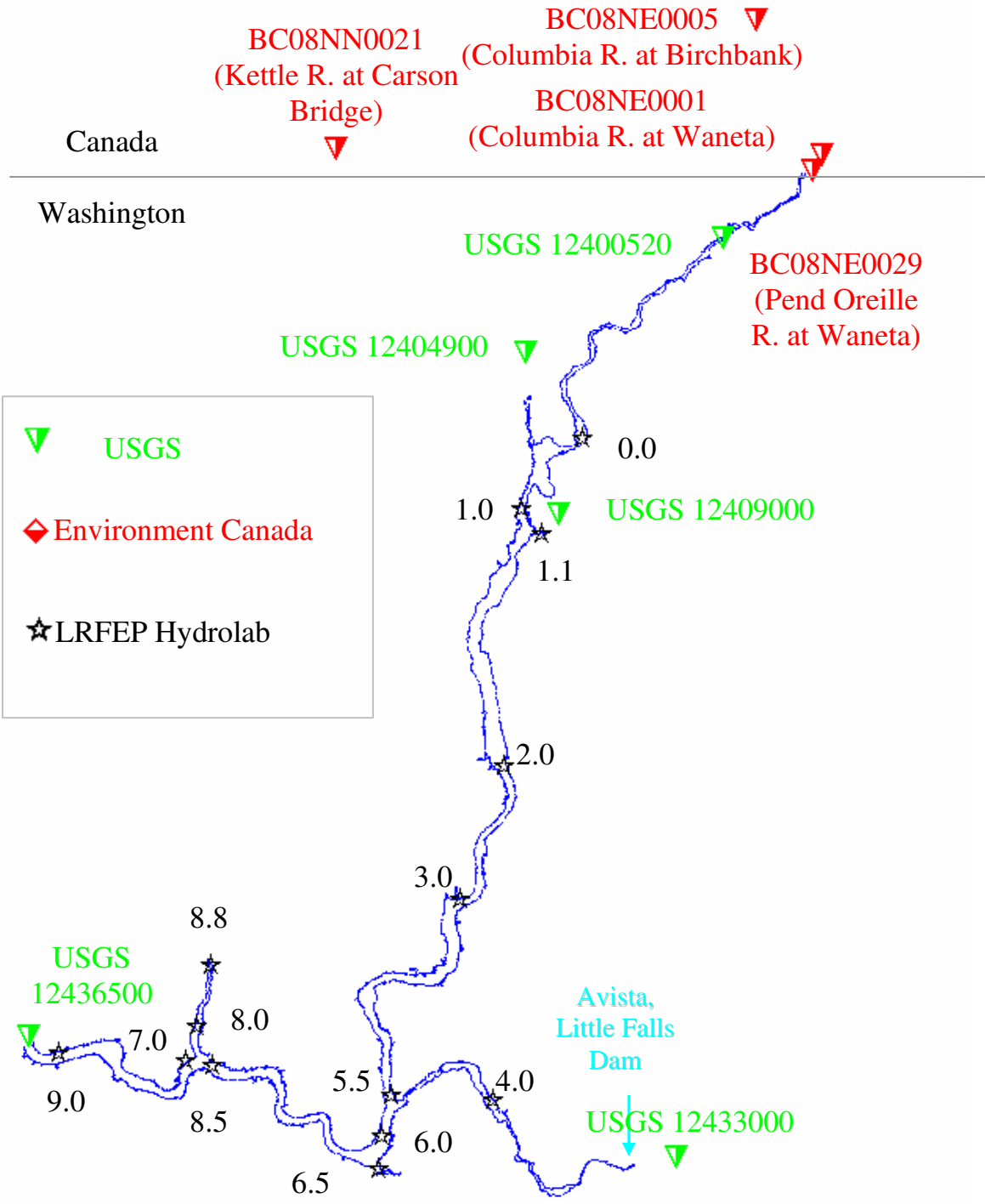


Figure 2. Water quality sampling site locations.

Hydrodynamic Calibration

Hydrodynamic calibration was performed by balancing all of the sources and sinks with a waterbalance flow to match the dam forebay water surface elevation data. Sources include tributary and mainstem inflows, precipitation, and return flows from Banks Lake (cogeneration flows). Sinks include outflows at the dam (powerhouse flows, outlet tubes, and spillway flows) irrigation withdrawals to Banks Lake, and evaporation.

Calibration Stations

The sole hydrodynamic calibration station is at the forebay of Grand Coulee Dam, USACOE gage (GCL). The station reports hourly water surface elevation (stage).

Year 2000

Figure 3 shows the model-data comparison of forebay stage. Table 2 reports the model-data comparison statistics. The magnitude of the waterbalance flows are shown in Figure 4. Figure 5 shows the percent of the waterbalance flows compared to the total flow through the dam. The magnitude of the waterbalance flows is largely within the flow gage measurement error range of 5 to 10%.

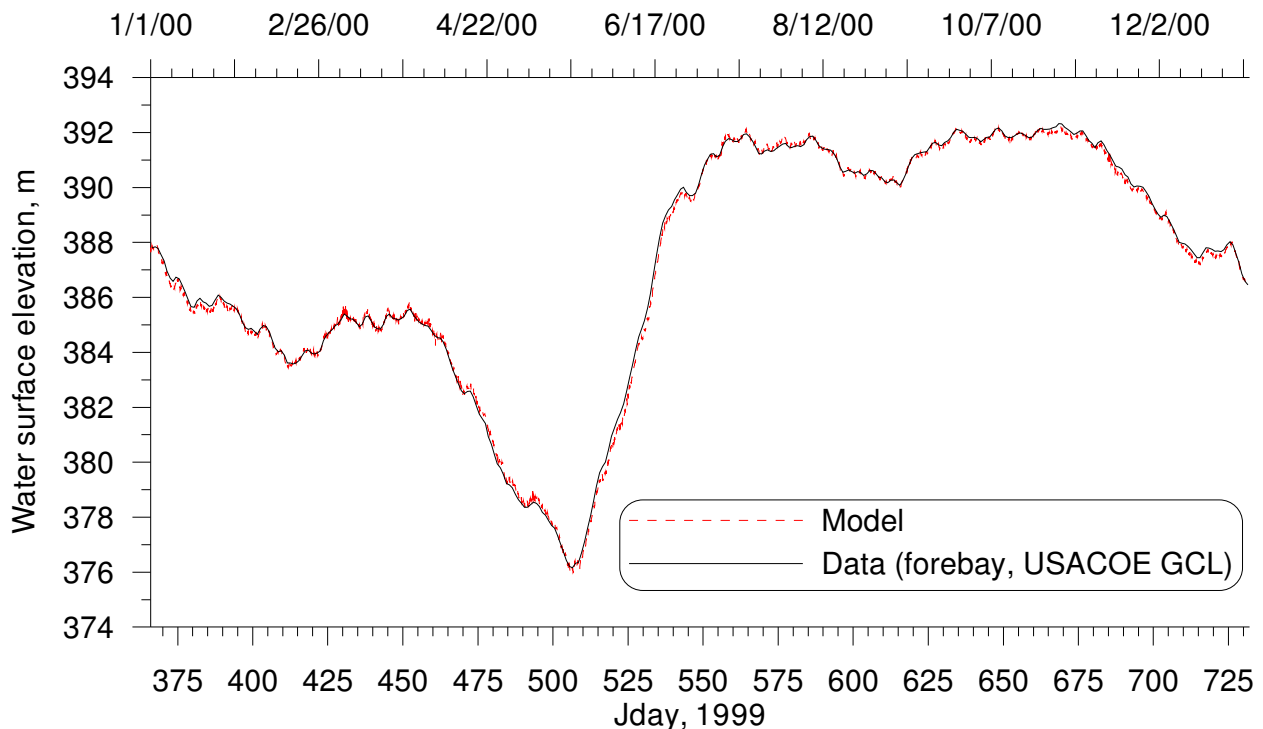


Figure 3. Model-data comparison, Grand Coulee Dam forebay stage, 2000.

Table 2. Grand Coulee Dam forebay stage statistics, 2000.

| Statistic (m) | Count | ME* | AME* | RMS* |
|-----------------------|-------|-------|------|------|
| Daily-average values | 366 | 0.01 | 0.01 | 0.01 |
| Hourly-average values | 8762 | -0.03 | 0.17 | 0.17 |

* ME=mean error, AME=absolute mean error, RMS=root mean square error, see Appendix G.

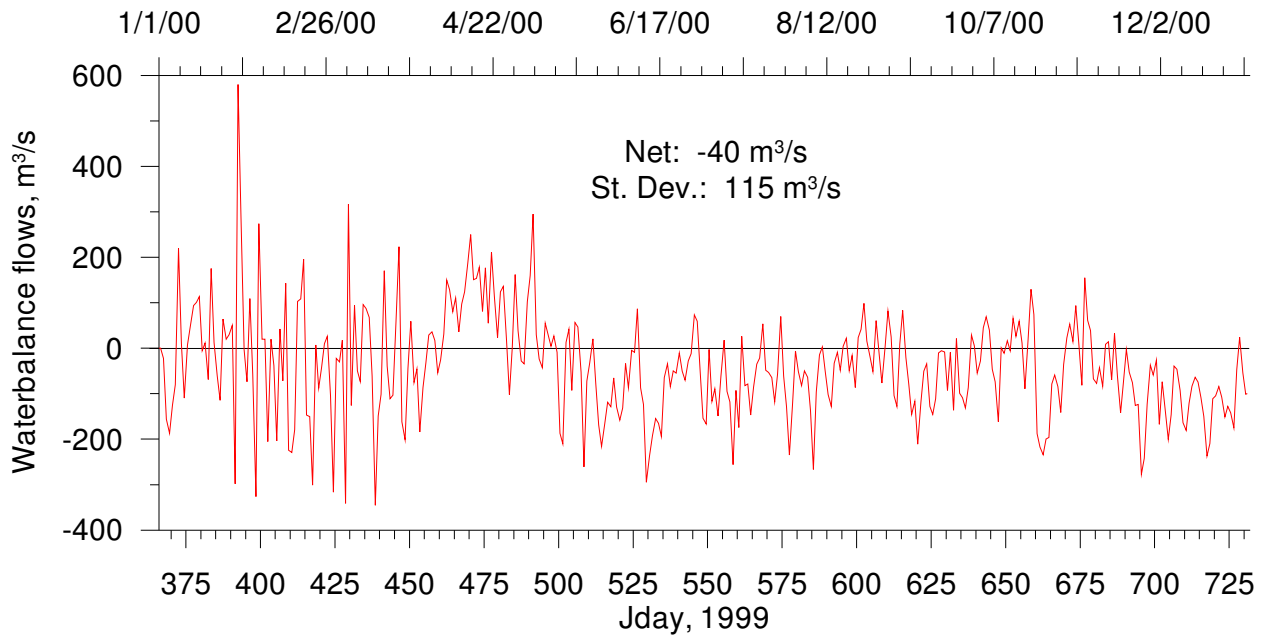


Figure 4. Waterbalance flow magnitudes, 2000.

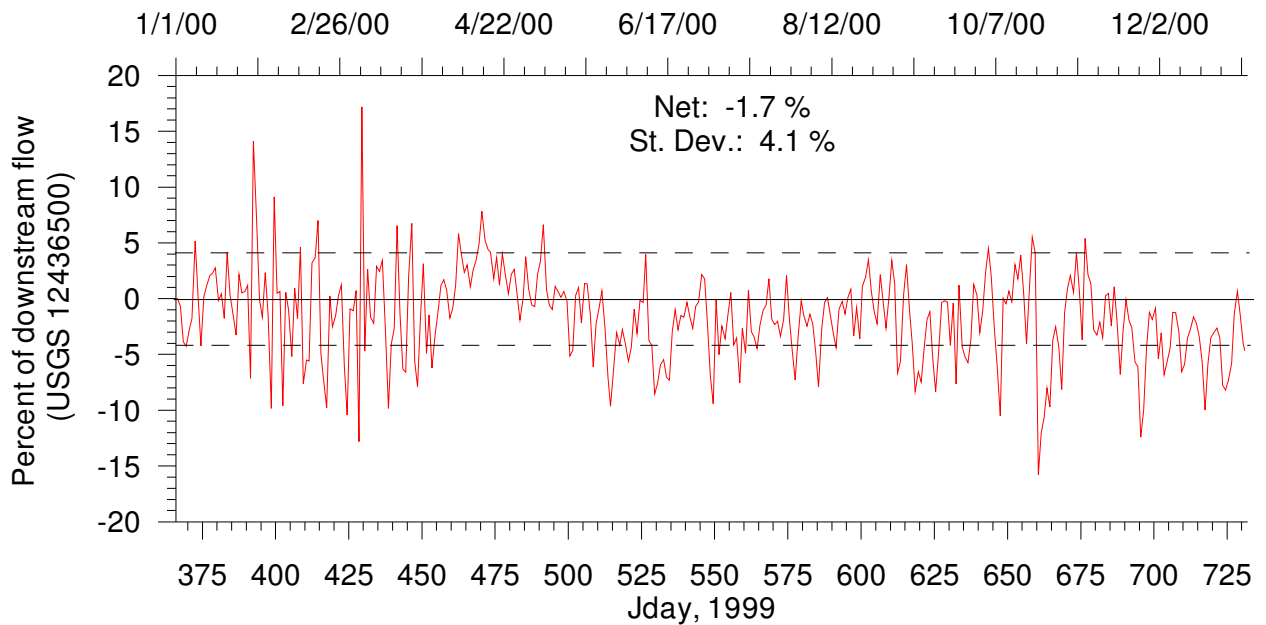


Figure 5. Waterbalance flows as percentage of downstream flows, 2000.

Year 2001

Figure 6 shows the model-data comparison of forebay stage. Table 3 reports the model-data comparison statistics. The waterbalance flows are shown terms of magnitude (Figure 7) and percent of total flow through the dam (Figure 8). Unlike the water balance for 2000, the 2001 water balance shows a bias toward negative flows (water being removed from the river).

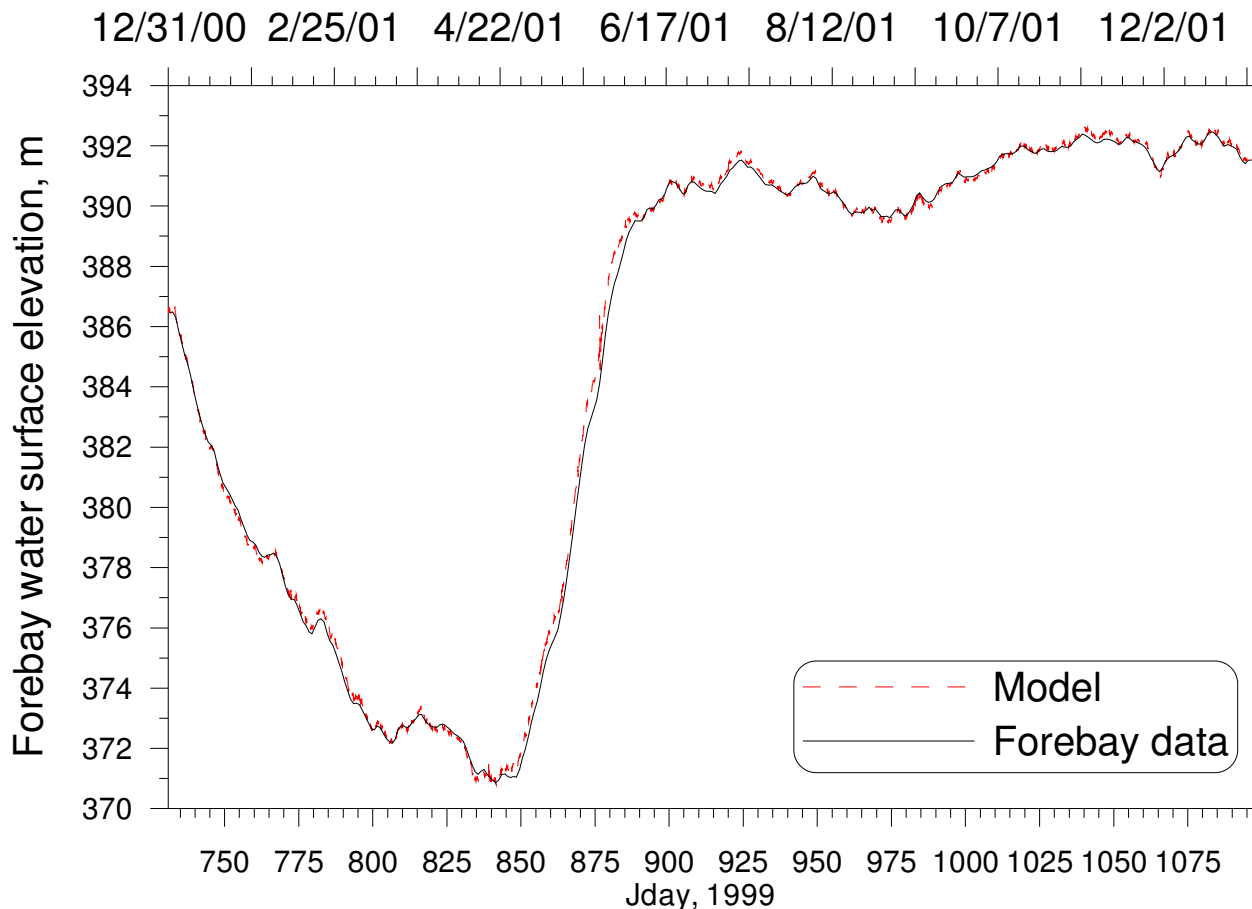


Figure 6. Model-data comparison, Grand Coulee Dam forebay stage, 2001.

Table 3. Grand Coulee Dam forebay stage statistics, 2001.

| Statistic (m) | Count | ME* | AME* | RMS* |
|-----------------------|-------|------|------|------|
| Daily-average values | 366 | 0.13 | 0.17 | 0.27 |
| Hourly-average values | 8762 | 0.01 | 0.12 | 0.12 |

* ME=mean error, AME=absolute mean error, RMS=root mean square error, see Appendix G.

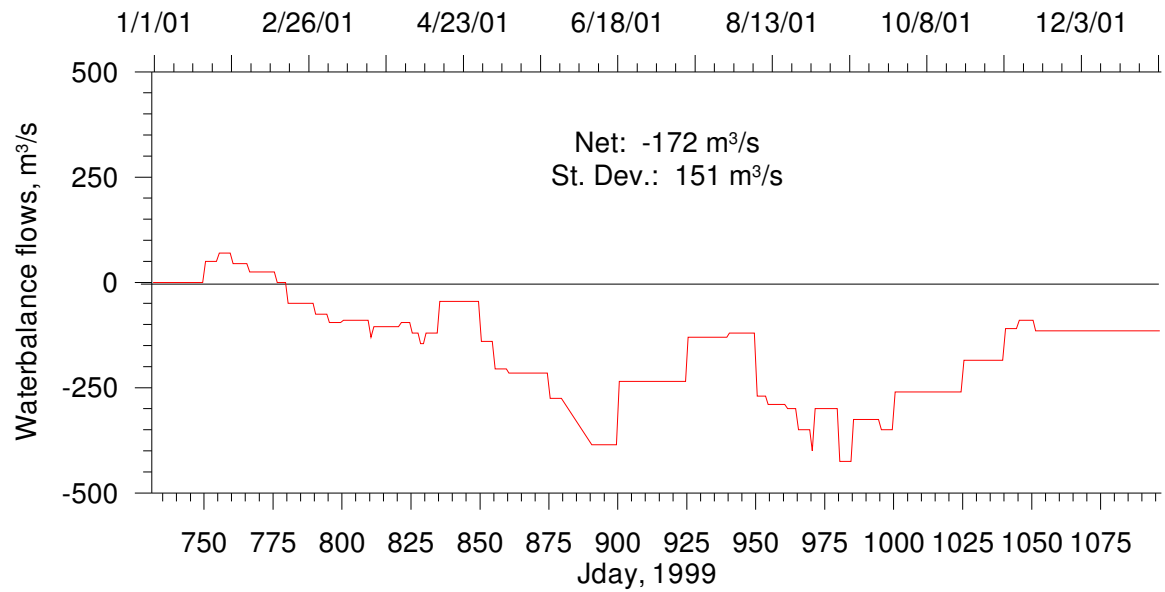


Figure 7. Waterbalance flow magnitudes, 2001.

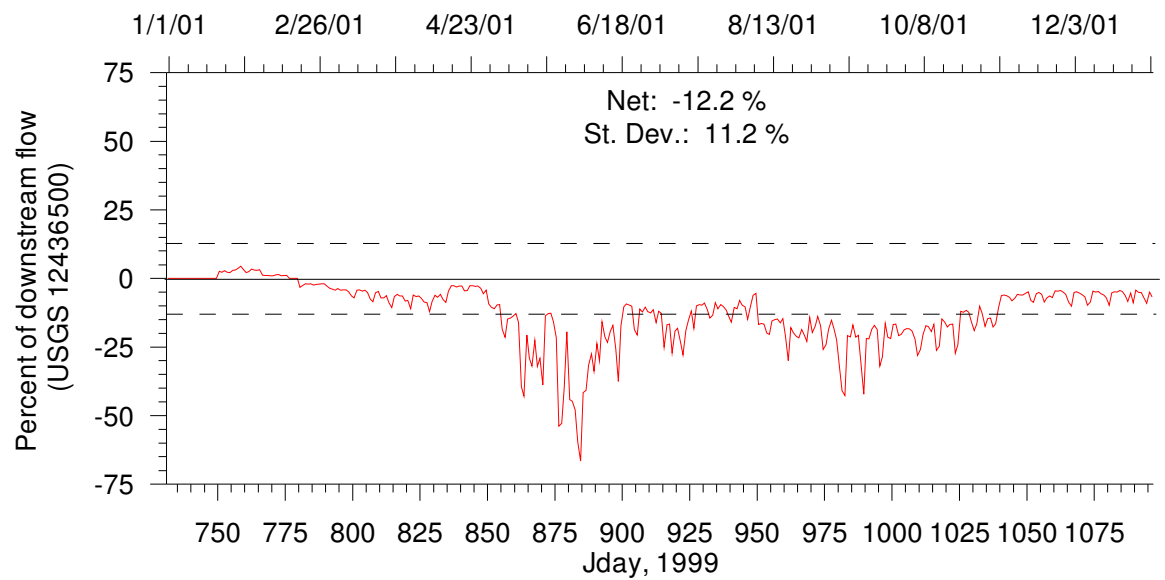


Figure 8. Waterbalance flows as percentage of downstream flows, 2001.

Year 2002

Figure 9 shows the model-data comparison of forebay stage. Table 4 reports the model-data comparison statistics. The waterbalance flows are shown terms of magnitude (Figure 10) and percent of total flow through the dam (Figure 11). Unlike the water balance for 2000, the 2002 water balance shows a bias toward negative flows (water being removed from the river).

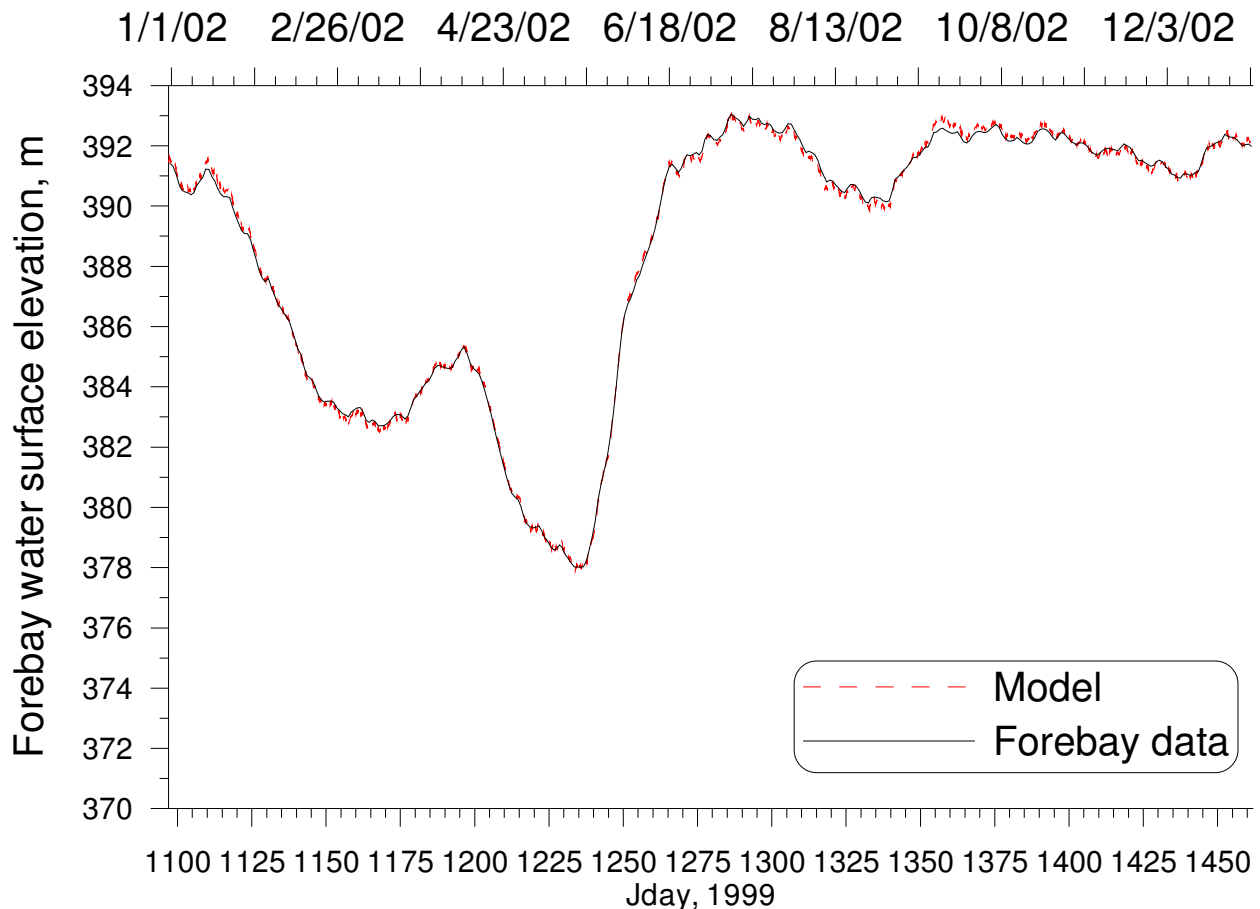


Figure 9. Model-data comparison, Grand Coulee Dam forebay stage, 2002.

Table 4. Grand Coulee Dam forebay stage statistics, 2002.

| Statistic (m) | Count | ME* | AME* | RMS* |
|-----------------------|-------|------|------|------|
| Daily-average values | 366 | 0.02 | 0.02 | 0.02 |
| Hourly-average values | 8763 | 0.01 | 0.10 | 0.10 |

* ME=mean error, AME=absolute mean error, RMS=root mean square error, see Appendix G.

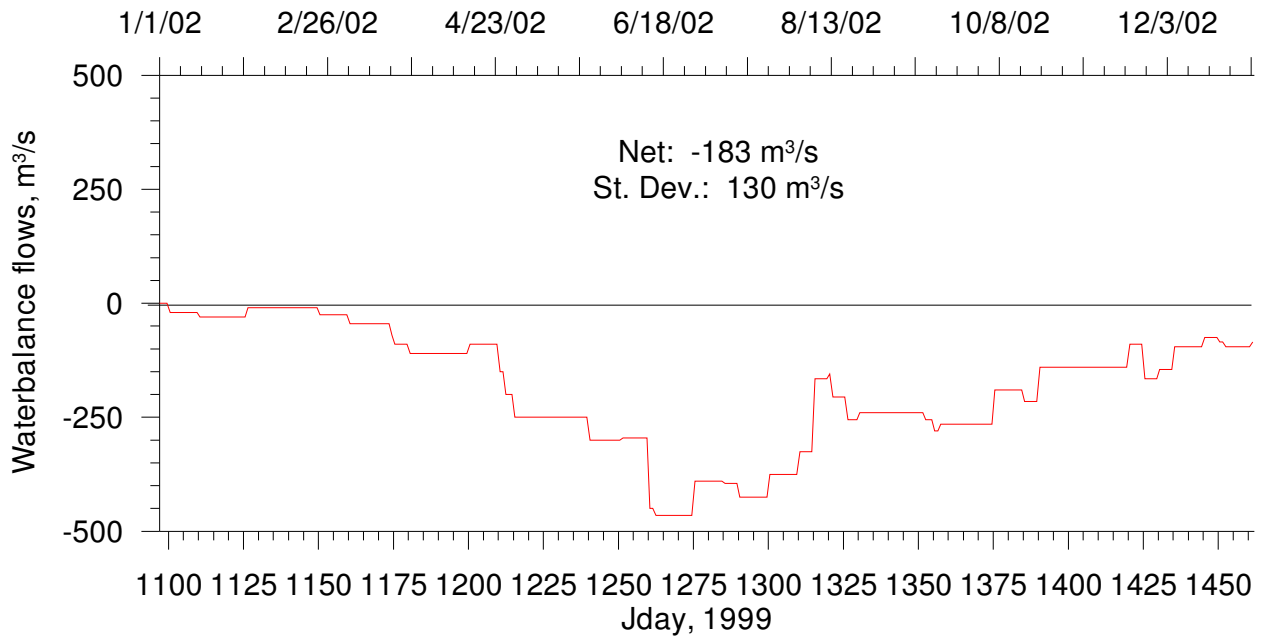


Figure 10. Waterbalance flow magnitudes, 2002.

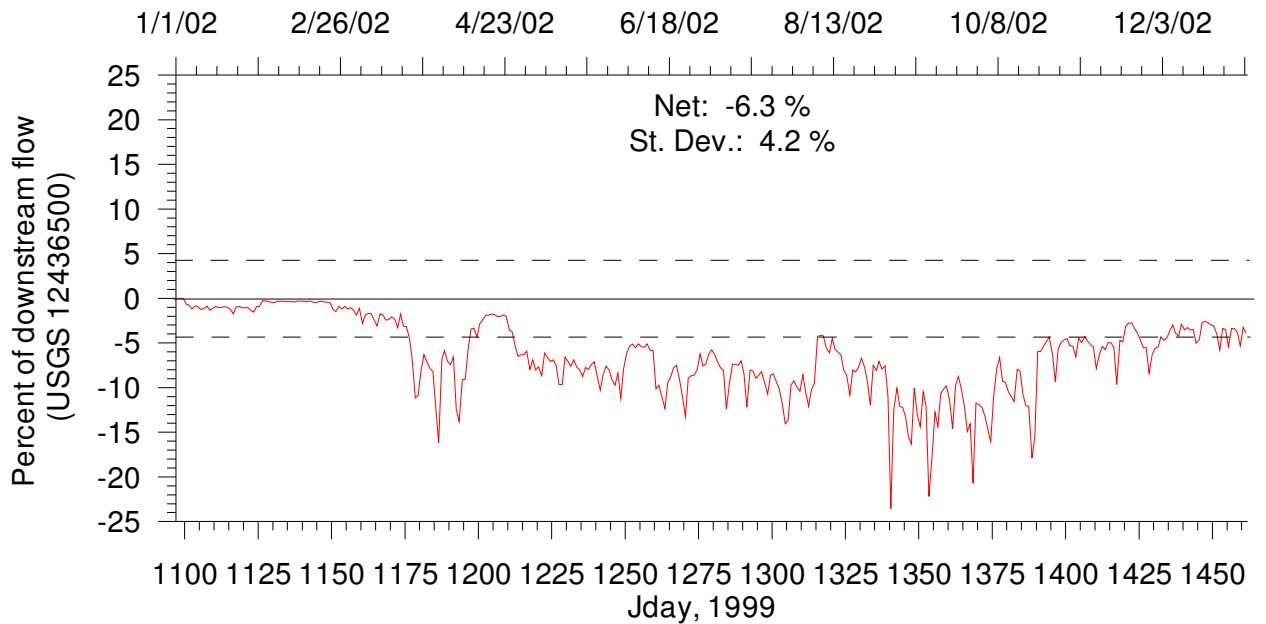


Figure 11. Waterbalance flows as percentage of downstream flows, 2002.

Temperature Calibration

The temperature calibration focused on matching the vertical stratification by adjusting the local wind sheltering coefficient and properly characterizing the powerhouse flows. The selective withdrawal elevation for flow through the third powerhouse had a lower bounds set to allow for more of the warmer surface water characteristic during stratification to be withdrawn. Meteorological inputs were adjusted to allow for the proper level of mean heating in the vertical profile sampling stations and continuous data downstream of Grand Coulee Dam.

Figure 12 shows a comparison of the vertical temperature profile at LRFEP station 9.0 (upstream of Grand Coulee Dam) under the calibrated wind sheltering coefficients [WSC] and with the default values [1.0]. Areas upstream of the dam had decreased WSC values (this in a decrease in wind speed, and hence mixing) which helped to allow greater stratification in the epilimnion.

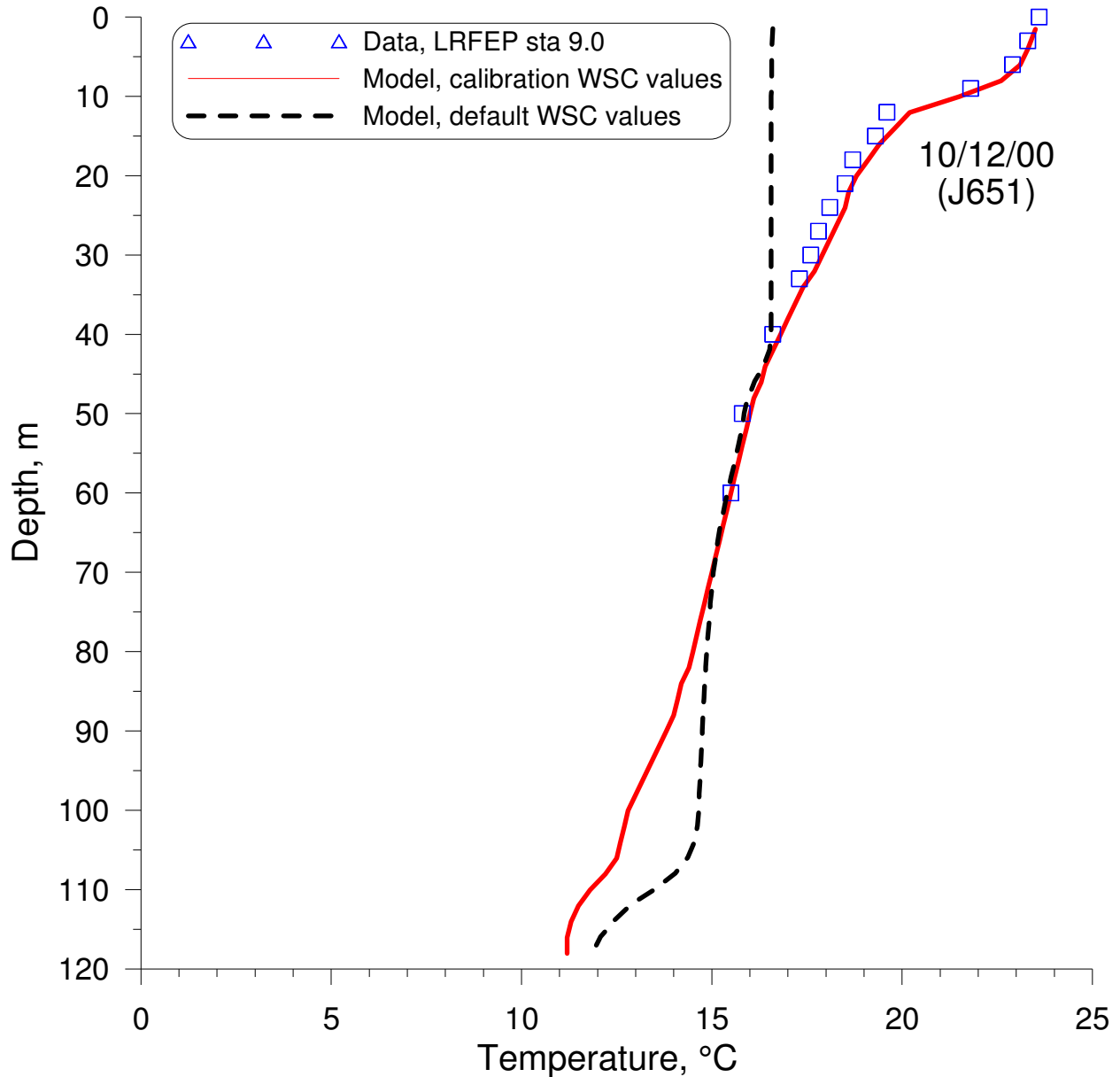


Figure 12. The effects of wind sheltering coefficients on temperature calibration. The default value of WSC of 0.85 was not considered accurate but was used as a basis for comparison to the calibrated value.

Calibration Stations

Two temperature calibration data types were available. Periodic vertical profile data were available at some or all of the 11 LRFEP stations shown in Figure 13. Table 5 lists the gage locations, numbers, and names. Sampling general occurred at a temporal frequency up to monthly at a typical vertical resolution of 3 m over the bulk of the vertical range. Roughly 10 km (6 mi) downstream of Grand Coulee Dam is the USACOE gage (GCGW) which records hourly temperatures. The Columbia River at the gage is riverine, and the temperatures were taken to be representative of the mean temperature.

Two additional temperature data sources were not used for calibration. The USGS gage at Northport (12400520) reported low frequency samples. The values reported were generally much colder than the nearby upstream temperatures reported at the International Boundary (USACOE CIBW) used for the Columbia River temperature boundary condition. The upstream gage, CIBW, agreed well with the most upstream vertical profile station (LRFEP 0.0).

The USBR collects temperatures from a station at the left side of Grand Coulee Dam that floats 60 ft (18.3 m) below the water surface (reported as USACOE: FDRW). These data were likely unrepresentative of the temperatures in the last model segment (which is 1000 m in length) as the instrument was attached to a trash rack near the dam face. Refer to McKillip, Annear, and Wells (2005) for further discussion of the instrument and data quality.

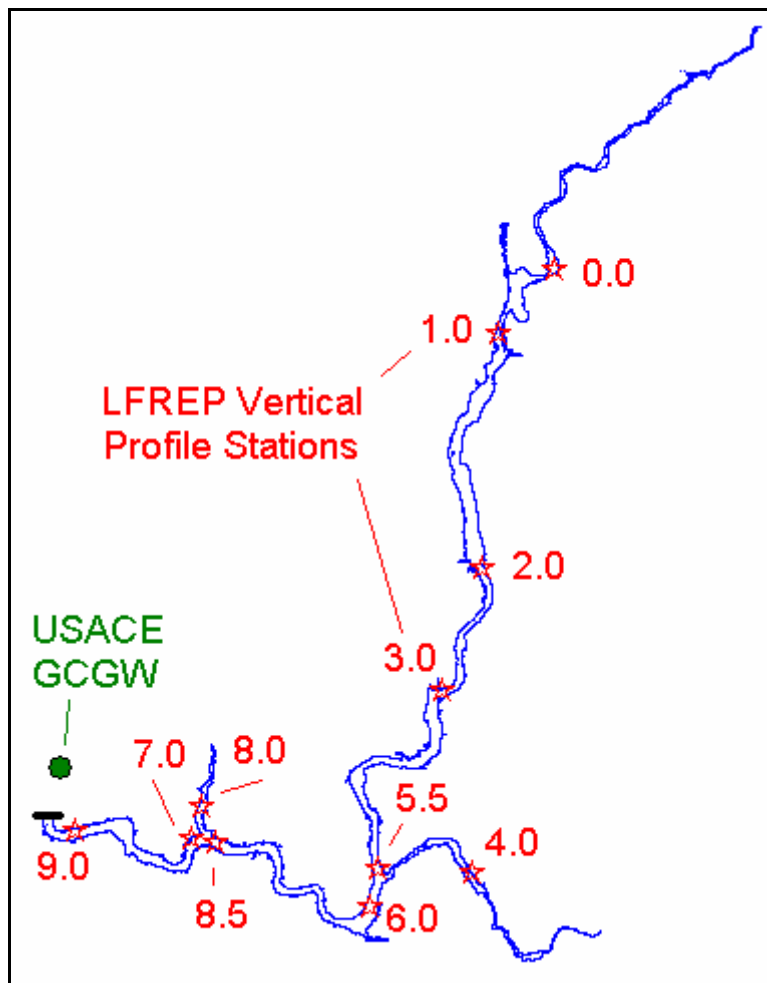


Figure 13. Locations of the temperature calibration sites.

Table 5. LRFEP vertical profile stations.

| Gage/Station | Location Name | Latitude | Longitude |
|---------------------|-----------------------|-----------------|------------------|
| 0.0 | Evan's Landing | 48.6830 | 118.0216 |
| 1.0 | Kettle Falls | 48.5992 | 118.1310 |
| 2.0 | Gifford | 48.2944 | 118.1540 |
| 3.0 | Hunters | 48.1371 | 118.2261 |
| 4.0 | Porcupine Bay | 47.9018 | 118.1651 |
| 5.5 | Spokane R. Confluence | 47.9043 | 118.3431 |
| 6.0 | Seven Bays | 47.8566 | 118.3571 |
| 7.0 | Keller Ferry | 47.9398 | 118.7046 |
| 8.0 | Sanpoil R. | 47.9814 | 118.6859 |
| 8.5 | Sanpoil R. Confluence | 48.0545 | 118.6643 |
| 9.0 | Spring Canyon | 47.9462 | 118.9285 |

Grand Coulee Dam, Continuous Temperatures

Temperature calibration focused on adjusting the wind magnitude (via the wind sheltering coefficient) temporally and spatially. Because actual wind speed and direction are highly variable around the lake, in order to monitor the wind with sufficient accuracy for calibration, one or a couple of wind monitoring locations may be inadequate to account for the full spatial and temporal variability of the wind field. The temperature data provide a good measure of the level of wind-driven mixing available between stations over the sampling time intervals. Thus, by adjusting the wind magnitude over reaches and time periods where the level of mixing is known (i.e., where temperature data are present), the mean wind speed can be better approximated.

The first target was to approximate the mean downstream temperature data (USACOE GCGW). The second target was to match the vertical temperature profiles (refer to Appendix F), which are a good indicator of the appropriate wind magnitude. In matching the temperatures near Grand Coulee Dam, the vertical profile data at LRFEP station 9.0 were given greater weight than the downstream river temperatures at USACOE GCGW. There is some uncertainty in how representative the downstream data are of the combined dam outflow temperature. This comparison is shown in Figure 14. Statistics are reported in Table 6. The depth of the powerhouse intakes, when compared to the vertical temperature profiles of the data at LRFEP station 9.0 and the model at the last Columbia River segment, suggests that the outflow temperature should be much colder during the summer than the downstream, riverine data. In order to 1) match the downstream, riverine data and 2) obtain the shape of the vertical profile data, the bottom of the selective withdrawal algorithm for the third powerhouse was limited to a minimum elevation of 353 m. This is higher than the centerline intake elevation of 347.5 m. Given the narrow inlet length of the third powerhouse, the surface waters appear to be preferentially withdrawn from the surface. This model characterization is a simplification of the more complicated three-dimensional nature of the flow within the third powerhouse inlet.

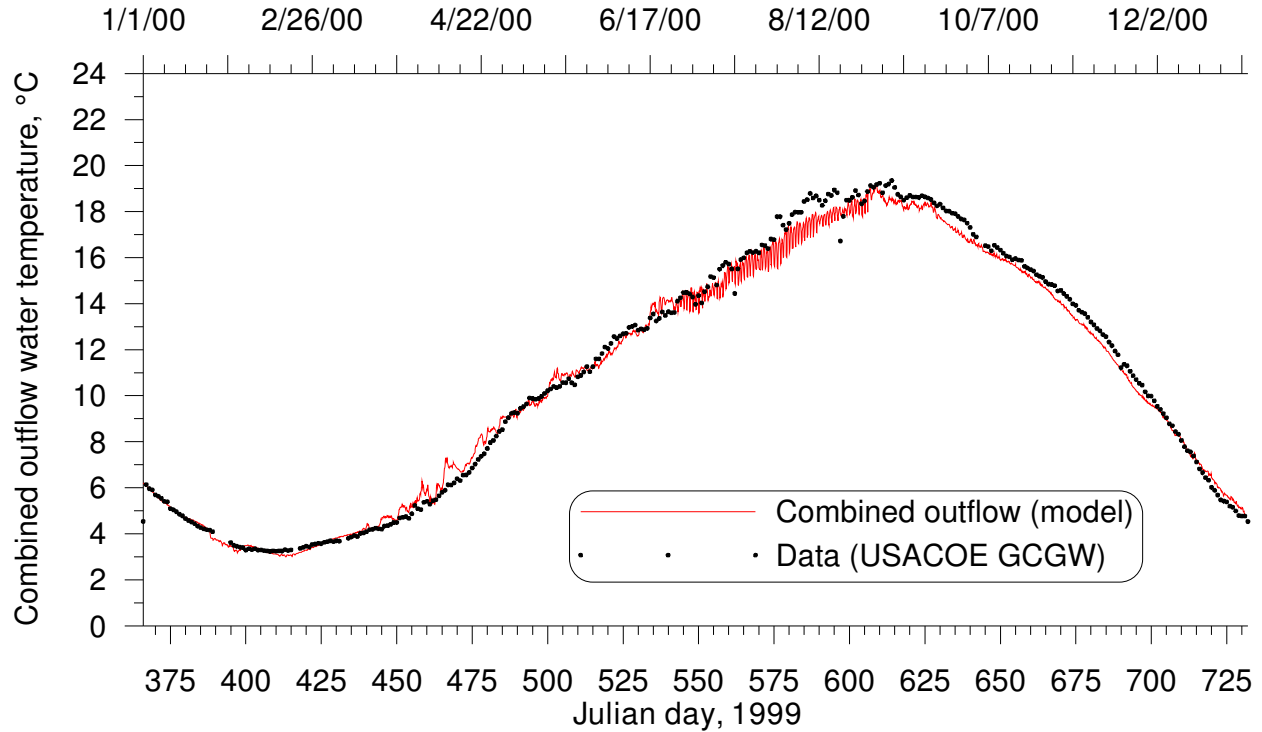


Figure 14. Model-data temperature comparison, below Grand Coulee Dam, 2000.

Table 6. Grand Coulee Dam temperature statistics, 2000.

| Statistic (°C) | Count | ME* | AME* | RMS* |
|----------------------------------|-------|-------|------|------|
| Daily-average values | 356 | -0.12 | 0.36 | 0.36 |
| 15-min and 60-min average values | 24428 | -0.24 | 0.45 | 0.45 |

* ME=mean error, AME=absolute mean error, RMS=root mean square error, see Appendix G.

Vertical Profile Stations, Periodic Sampling

The temperature calibration plots and statistics from the LRFEP vertical profile station data are reported in Appendix F. Selected profiles are shown for station 4.0 (Figure 15) and station 9.0 (Figure 16) in this section. Three sampling periods are illustrated for the reservoir conditions near minimum spring pool, near peak thermal stratification, and after fall turnover. Overall temperature model-data profile errors were -0.14°C , 0.49°C , and 0.50°C , for the mean error, absolute mean error, and root-mean-square error, respectively, comparing 114 vertical temperature profiles. Detailed error statistics are shown in Appendix F Table 31.

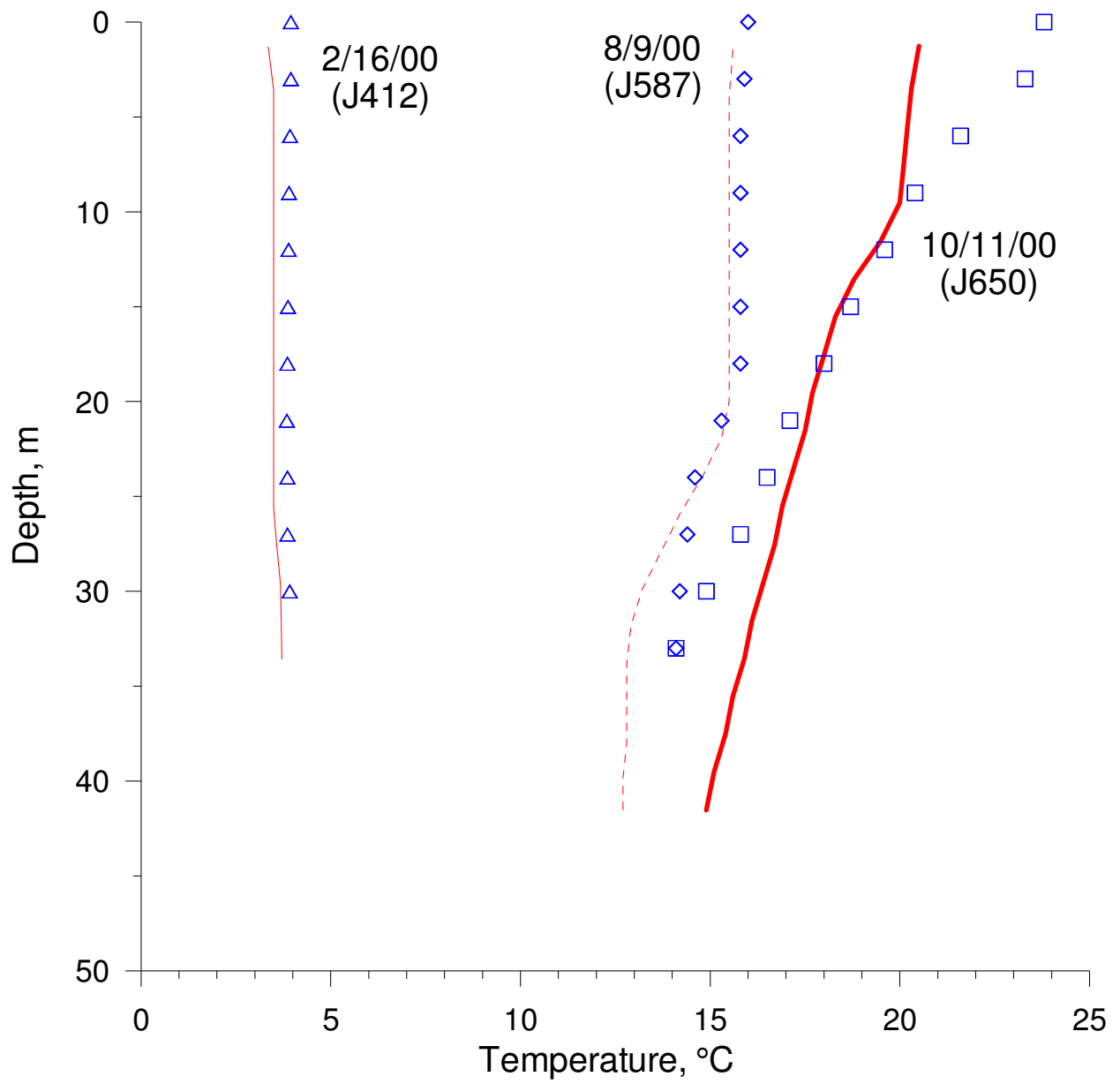


Figure 15. Selected model-data temperature profile comparisons at Porcupine Bay (LRFEP sta 4.0).

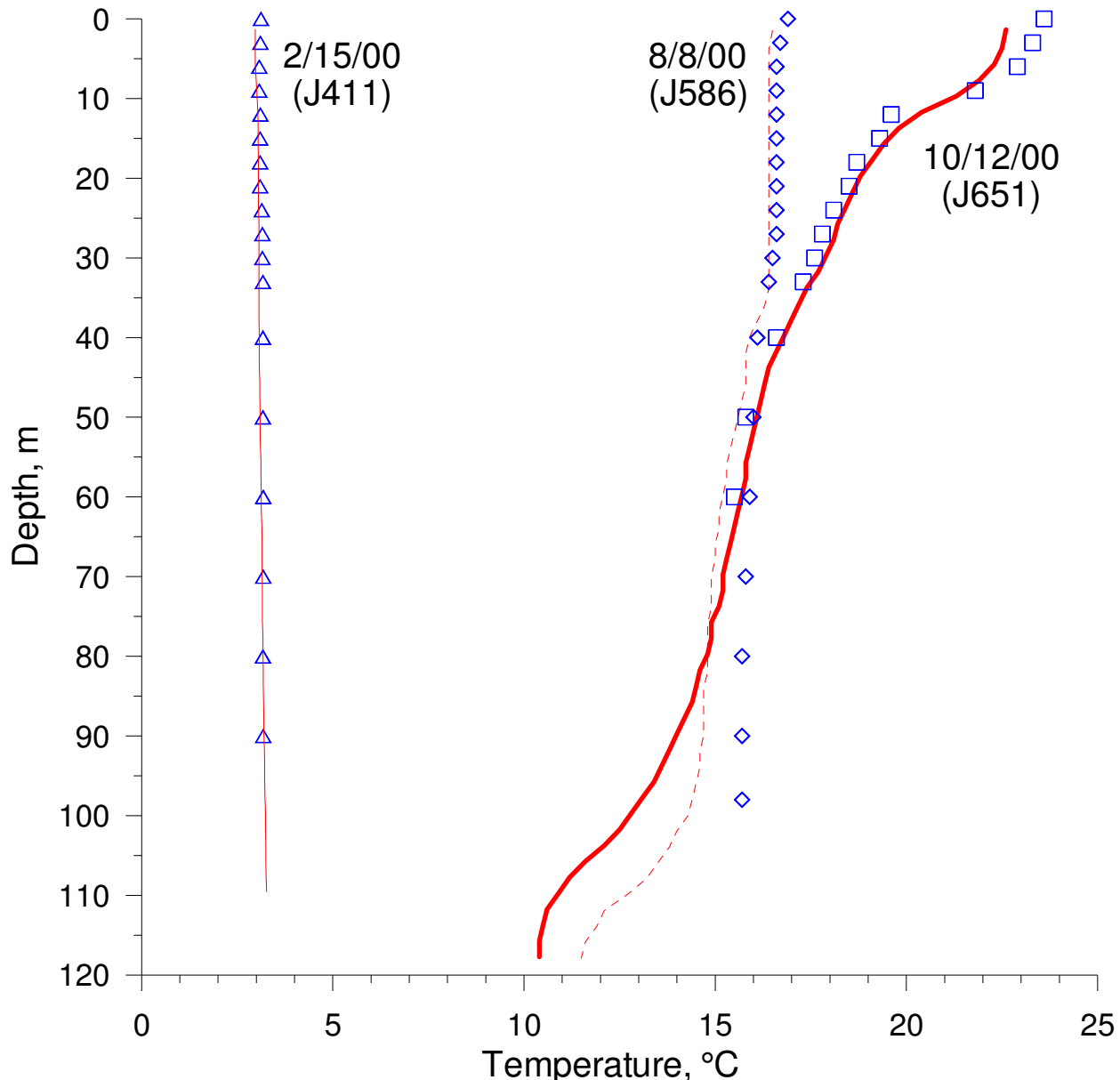


Figure 16. Selected model-data temperature profile comparisons at Spring Canyon (LRFEP sta 9.0).

Abiotic Water Quality Calibration

Calibration Stations

The abiotic water quality calibration stations included the LRFEP stations shown in Figure 13 (in the temperature calibration section) and the USGS gage below Grand Coulee Dam (12436500).

The abiotic water quality calibration plots were grouped into three areas of the report to take advantage of the structure of the data.

- The flow weighted constituent values of the outflow at Grand Coulee Dam are reported in the next section (Grand Coulee Dam Outflow). The time-series results are compared with the vertical data upstream of the dam (LRFEP station 9.0) and the grab sampling in the riverine reach below the dam (USGS 12436500). Data below the dam are not available for all constituents.
- The LRFEP monitoring stations report several constituents associated over specific depths. These constituents are total dissolved solids, dissolved oxygen, and pH. Plots showing the vertical structure are reported in Appendix F.
- Many of the constituents sampled by the LRFEP had a depth loosely associated with the sample. The depths were reported as being in the photic or aphotic zone. The model results were volume-weighted over the upper 10 m to provide a rough estimate of the photic zone constituent value. Plots showing the weighted model results are reported in Appendix E.

Some of the constituents results are illustrated as both continuous outflow concentrations (next section: Grand Coulee Dam, Continuous Temperatures) and as vertical profiles (Appendix F). Model-data comparison statistics are reported based on the structure of the data as shown in Table 7.

Table 7. Organization of the Grand Coulee Dam outflow constituent model-data comparison statistics.

| Constituent | Data | | Statistics Table |
|------------------------|-----------------|------------------|------------------|
| | structure | source | |
| Alkalinity | discrete (grab) | STOI Lab sta 9.0 | Table 8 |
| Ammonium | discrete (grab) | STOI Lab sta 9.0 | Table 8 |
| Dissolved oxygen | profile | LRFEP sta 9.0 | Table 31 |
| Nitrate + nitrite | discrete (grab) | STOI Lab sta 9.0 | Table 8 |
| pH | profile | LRFEP sta 9.0 | Table 31 |
| Orthophosphate | discrete (grab) | STOI Lab sta 9.0 | Table 8 |
| Total dissolved solids | profile | LRFEP sta 9.0 | Table 31 |

Constituent Calibration Discussion

Water quality calibration is discussed in several sections:

- Continuous abiotic data near Grand Coulee Dam
- Instantaneous vertical profile of abiotic data at LRFEP stations
- Model-data vertical profile comparisons in Appendix E
- Alternative boundary conditions for DO and pH
- Boundary condition generation in Appendix A

Calibration to temperature, hydrodynamic, abiotic, and biotic data is to a varying degree simultaneous. A discussion of the abiotic calibration is presented in this section in an attempt to present the calibration adjustments in a single section.

In general, the approach to calibration is to calibrate to hydrodynamic data, and then to the temperature data, which often requires adjustment of the hydrodynamic calibration. The most upstream location is calibrated first, for each calibration target, and the earliest targets (in time) are also calibrated first. The downstream stations, and the later time periods, are heavily influenced by the upstream and earlier periods. Thus, a continuous reevaluation of the upstream calibration and boundary conditions is made during calibration.

While considering this general approach, calibration of specific constituents is discussed below.

Alkalinity and pH

Since there is little carbonate chemistry activity in the system, alkalinity behaves like a conservative constituent. The calibration focused on ensuring that all of the carbon chemistry initial conditions were appropriate. The W2 model uses alkalinity, total inorganic carbon (TIC), and bicarbonate (HCO_3) to model the carbon system. pH is a derived constituent, meaning its value is determined from other constituents. Overall, pH model-data profile errors were -0.42, 0.43, and 0.43, for the mean error, absolute mean error, and root-mean-square error, respectively, comparing 114 vertical pH profiles. Detailed error statistics are shown in Appendix F Table 31.

There is an apparent data conflict between the USGS and LRFEP data. The sub-section Total Inorganic Carbon (pH) of the Alternate Boundary Conditions section examines this conflict.

Ammonium and Nitrate plus Nitrite

Calibration of the nitrogen budget required only small changes from the default conditions. The major change was to reduce the amount of nitrogen in the decaying organic matter [ORGN] to reduce the ammonium concentrations.

Dissolved Oxygen

Dissolved oxygen calibration involved a review of the boundary conditions, inclusion of sediment oxygen demand (SOD), and an investigation of the reaeration equations and algal populations. As might be expected from low algal concentrations, the mainstem Columbia and Spokane River DO concentrations were not sensitive to algal concentration. Similarly, such a large volume system was not sensitive to the reaeration formulation. Inclusion of SOD allowed the model to capture some of the vertical gradients near the bottom of the reservoir. However, the model was not able to fully capture all of the observed vertical gradients—the model predicted less vertical variation. Also, the data exhibit significant shifts in concentrations from month to month and site to site that are not readily explained by known physical processes. For example, the data showed monthly changes of 1 mg/l at a site that would not occur at the next downstream site. Overall, dissolved oxygen model-data profile errors were -0.01 mg/l, 0.14 mg/l, and 0.16 mg/l, for the mean error, absolute mean error, and root-mean-square error, respectively, comparing 114 vertical dissolved oxygen profiles. Detailed error statistics are shown in Appendix F Table 31.

Orthophosphate

Phosphorous calibration was problematic given the large spring spike over the entire system, followed by very low concentrations. The physical cause of this phenomenon is not clear. The phenomenon was characterized as early spring inflows which were then sorbed onto inorganic suspended solids. This approach allowed some of the spring spike to be captured by the model, but the summer and fall concentrations were elevated. Additionally, the organic matter stoichiometry [ORGP] was reduced to lower concentrations. The orthophosphate concentrations were not sensitive to algal stoichiometry or algal concentrations within the range of the system.

Total Dissolved Solids

Total dissolved solids are largely conservative, so calibration focused on ensuring good boundary conditions. The calibration was acceptable for the Columbia River, but the Spokane River showed some problems. While the shape of the vertical gradient was typically captured, there were errors (shifts) in the concentrations. Overall, Total dissolved solids model-data profile errors were -0.18 mg/l, 0.39 mg/l, and 0.44 mg/l, for the mean error, absolute mean error, and root-mean-square error, respectively, comparing 114 vertical TDS profiles. Detailed error statistics are shown in Appendix F Table 31.

Grand Coulee Dam Outflow

The water quality constituents of the modeled outflow from Grand Coulee Dam are compared to data both upstream at LRFEP station 9.0 and downstream at USGS 12436500. The data at these two points do not always agree. The model uses flow weighting from each outlet structure in reporting the model constituent concentrations. The downstream data are from grab samples at a 2riverine reach. The upstream samples are a mix of vertical profile points and grab samples in both the aphotic and euphotic zones. The vertical profile data points are shown to give a sense of the range of the constituent values at the sampling point.

The model-data comparison statistics for the discrete sampling constituents are shown in Table 8. The statistics for the vertical profile sampling constituents are reported in Appendix F (Table 31).

Table 8. Discrete constituent model-data comparison statistics.

| Constituent | Orthophosphate | Ammonium | Nitrate plus nitrite | Alkalinity |
|--------------------|-----------------------|-----------------|-----------------------------|------------------------|
| Unit | mg/L-P | mg/L-N | mg/L-N | mg/L-CaCO ₃ |
| Count | 25 | 25 | 25 | 25 |
| ME | 0.00072 | -0.0050 | -0.017 | 0.89 |
| AME | 0.00097 | 0.0067 | 0.036 | 4.18 |
| RMS | 0.00097 | 0.0067 | 0.036 | 4.18 |

Alkalinity

The alkalinity model-data comparison near Grand Coulee Dam is shown in Figure 17. After ensuring good boundary conditions and hydrodynamic calibration, no further adjustments were made.

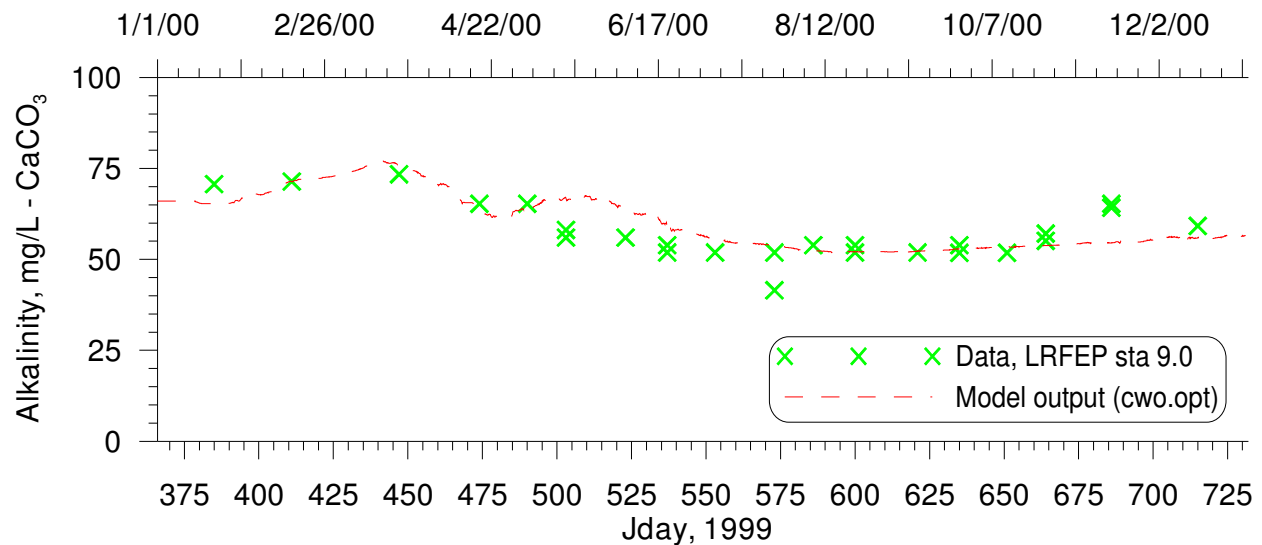


Figure 17. Alkalinity time-series near Grand Coulee Dam.

Ammonium

The ammonium model-data comparison near Grand Coulee Dam is shown in Figure 18. After ensuring good boundary conditions and hydrodynamic calibration, no further adjustments were made. The USGS sampling had a higher detection limit than the LRFEP sampling. The model predicts ammonium concentrations below the USGS sampling detection limit.

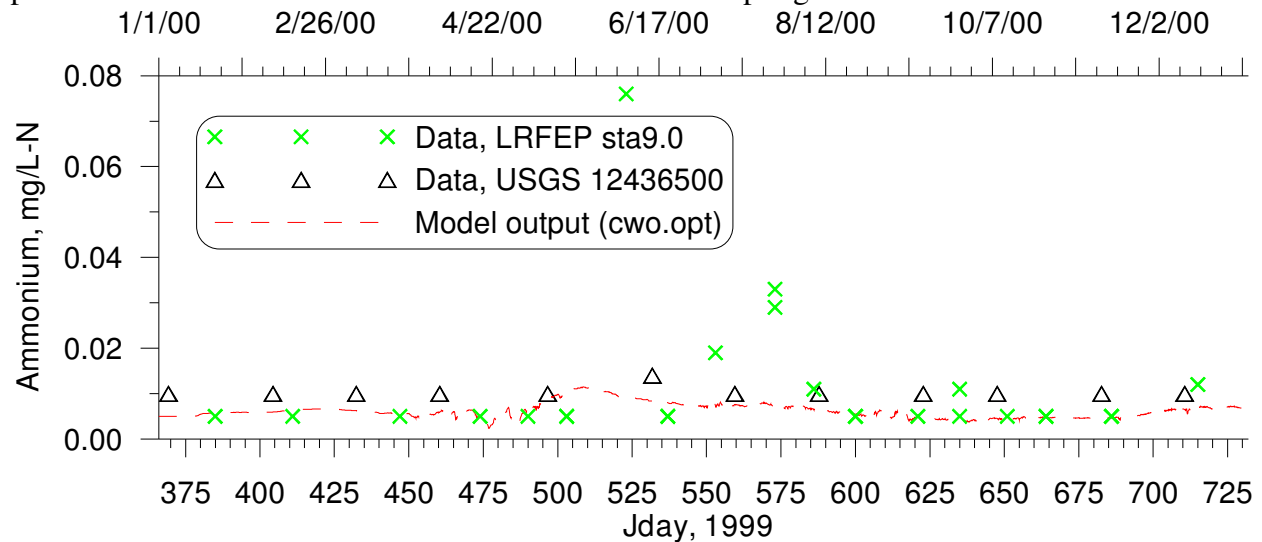


Figure 18. Ammonium time-series near Grand Coulee Dam.

Dissolved Oxygen

The dissolved oxygen model-data comparison near Grand Coulee Dam is shown in Figure 19. After ensuring good boundary conditions and hydrodynamic calibration, no further adjustments were made. The model values are typically near saturation; however, the data are above saturation in the early spring and below saturation in the fall. The source of this discrepancy is unclear. Model results were not sensitive to the algal concentration or the air-water reaeration formulation.

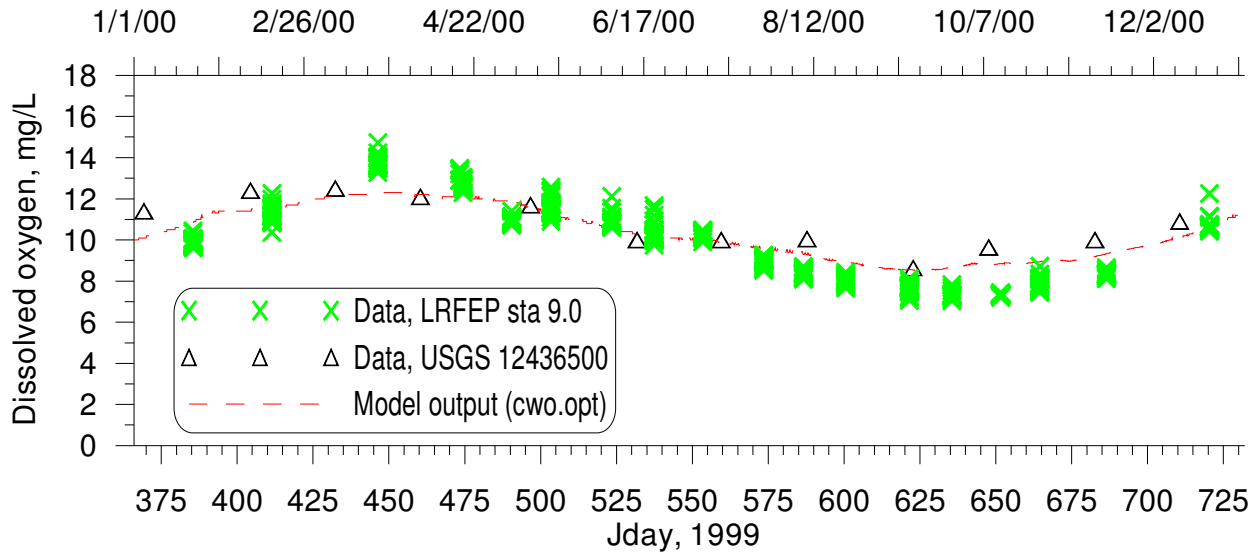


Figure 19. Dissolved oxygen time-series near Grand Coulee Dam.

Nitrate plus nitrite

The nitrate plus nitrite model-data comparison near Grand Coulee Dam is shown in Figure 20. After ensuring good boundary conditions and hydrodynamic calibration, no further adjustments were made.

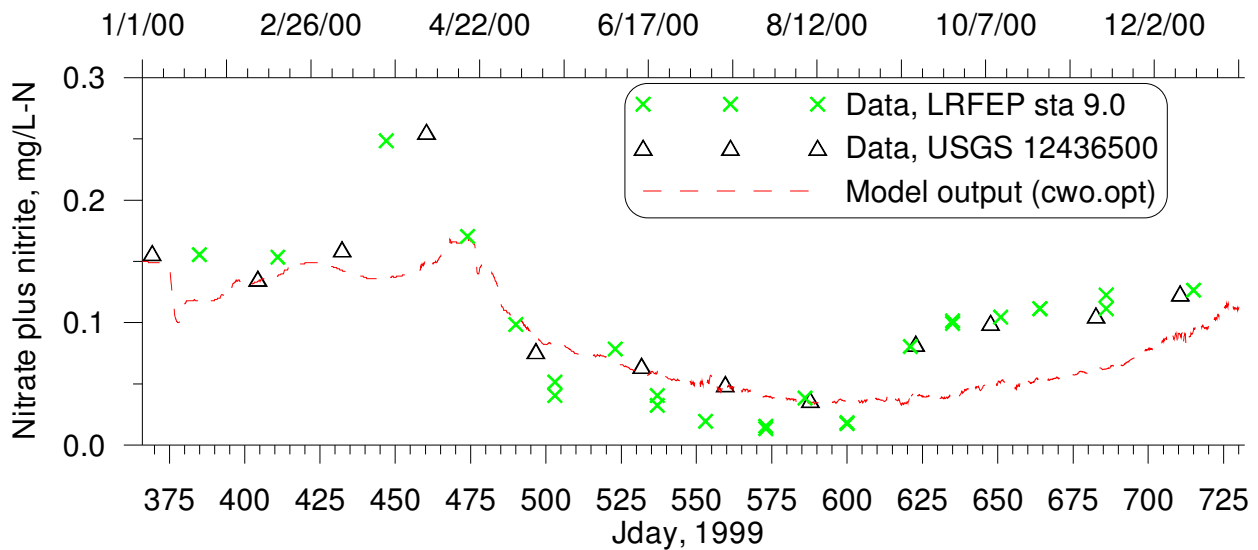


Figure 20. Nitrate time-series near Grand Coulee Dam.

pH

The pH model-data comparison near Grand Coulee Dam is shown in Figure 21. Calibration included ensuring good boundary conditions and good hydrodynamic calibration. While not clearly illustrated in Figure 21, the model does capture the strength of the vertical gradient seen during the summer. Refer to the vertical profile plots in Appendix F. During October and November of 2000, there was a known problem with the sensor probe.

The section Total Inorganic Carbon (pH) in the Alternate Boundary Conditions section of this report discussed the effects of alternate pH boundary conditions.

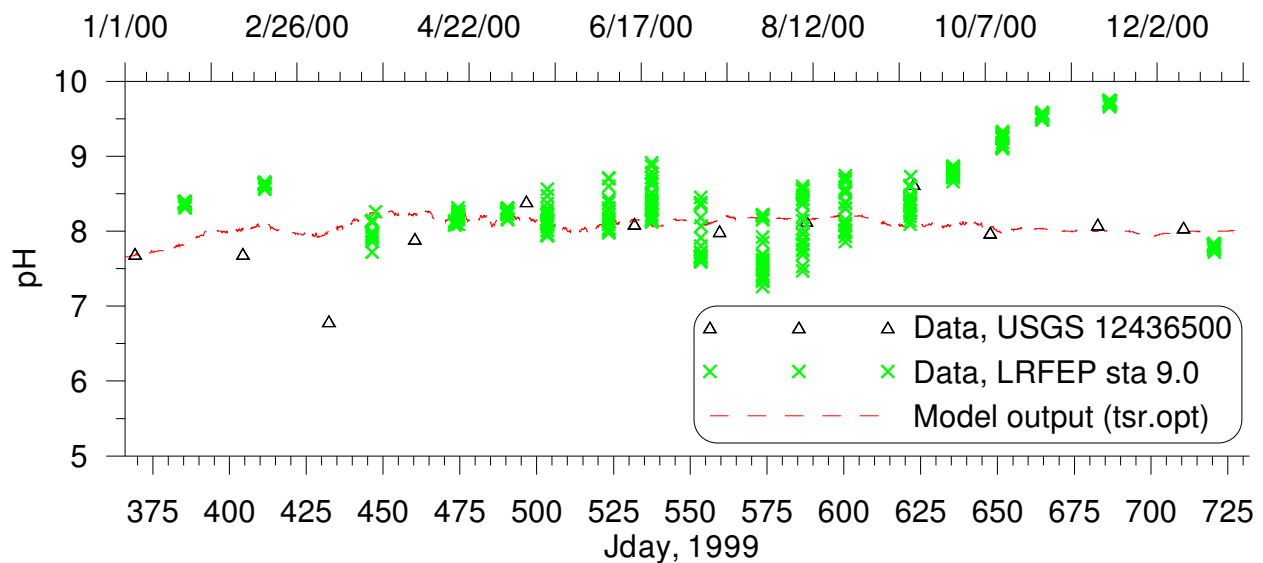


Figure 21. pH time-series near Grand Coulee Dam.

Orthophosphate

The orthophosphate model-data comparison near Grand Coulee Dam is shown in Figure 22. Calibration included ensuring good boundary conditions and good hydrodynamic calibration. The organic matter decay rate was decreased, sorption onto inorganic suspend solids was allowed, and the stoichiometry of organic matter was adjusted to calibrate orthophosphate.

There are several possible sources of the model-data discrepancy. The very low phosphorous concentration makes accurate measurements difficult. The attached algae data are sparse and may not be representative. The relationship between sorbed phosphorous and the stoichiometry of the inflow organic matter is likely neither invariant in time or space. The transition from winter periphyton to spring phytoplankton as the dominant primary production and the heterogeneity in boundary conditions are likely to explain much of the systems behavior.

The USGS sampling had a higher detection limit than the LRFEP sampling. The model predicts orthophosphate concentrations below the USGS sampling detection limit.

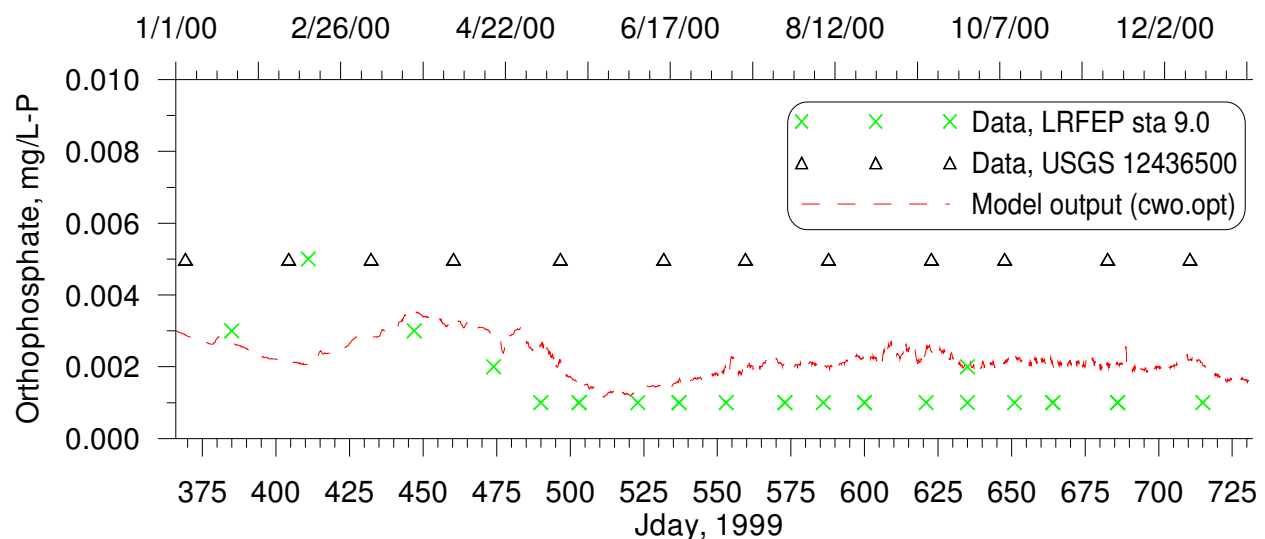


Figure 22. Orthophosphate time-series near Grand Coulee Dam.

Total Dissolved Solids

The total dissolved solids model-data comparison near Grand Coulee Dam is shown in Figure 23. After ensuring good boundary conditions and hydrodynamic calibration, no further adjustments were made.

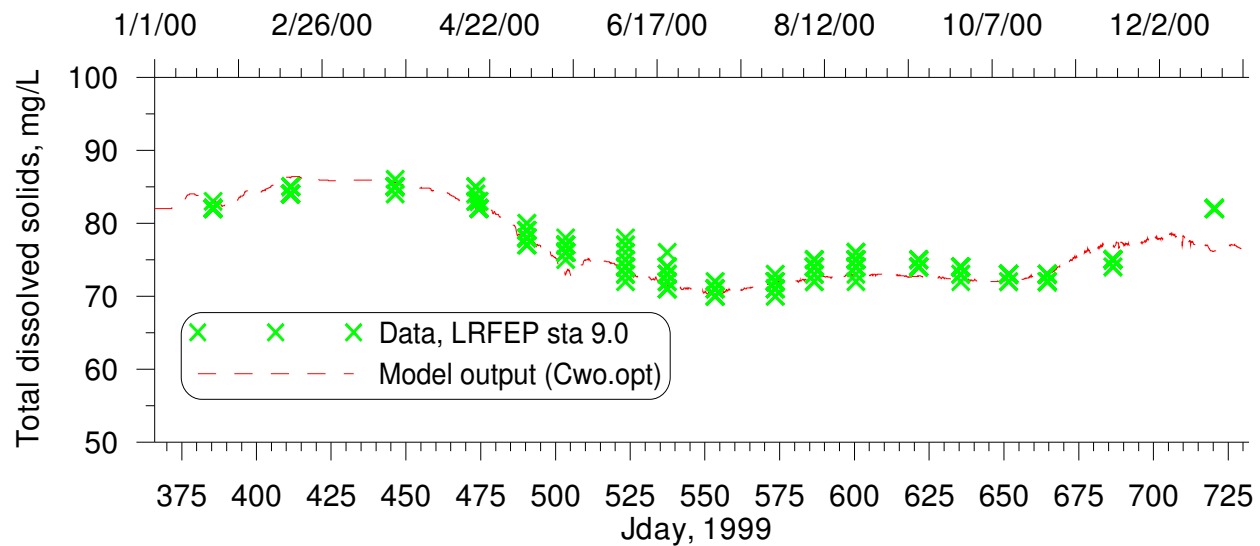


Figure 23. Total dissolved solids time-series near Grand Coulee Dam.

Vertical Profile Stations, Periodic Sampling

Dissolved oxygen, total dissolved solids, and pH calibration plots and statistics from the LRFEP vertical profile station data are reported in Appendix F. Selected profiles are shown for station 4.0 and station 9.0 in this section. Three sampling periods are illustrated for the reservoir conditions near minimum spring pool, near peak thermal stratification, and after fall turnover.

Ensuring hydrodynamic and temperature calibration was the first step in water quality constituent calibration.

Dissolved oxygen (Figure 24 and Figure 25) calibration also included sediment oxygen demand (SOD). Total dissolved solids (Figure 26 and Figure 27) are a conservative constituent, so no additional calibration was made. Calibration for pH (Figure 28 and Figure 29) focused on the selection of boundary conditions.

An investigation of alternative boundary conditions is presented in the subsequent section, Alternate Boundary Conditions.

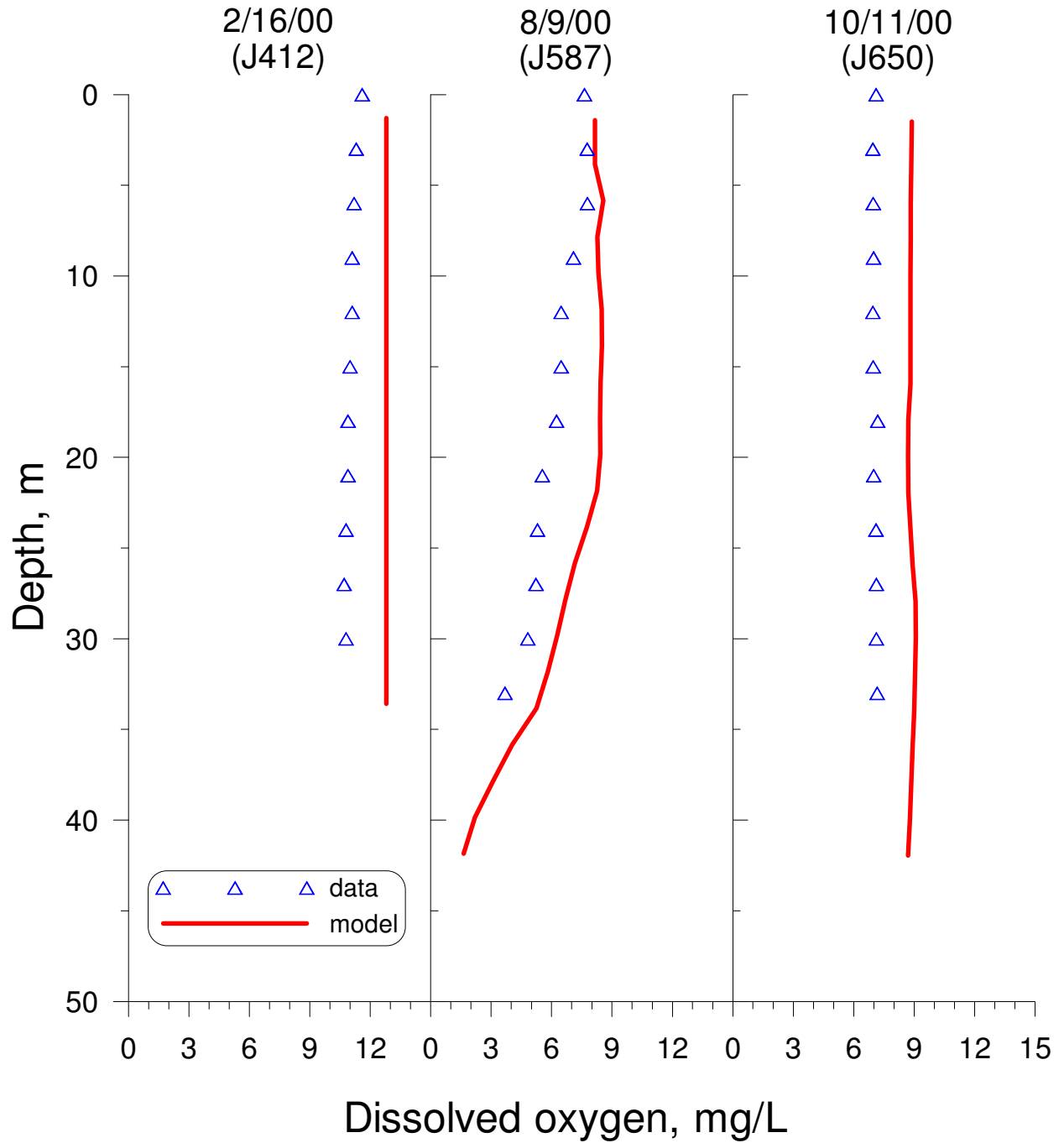


Figure 24. Selected model-data dissolved oxygen vertical profile comparisons at Porcupine Bay (LRFEP stat 4.0).

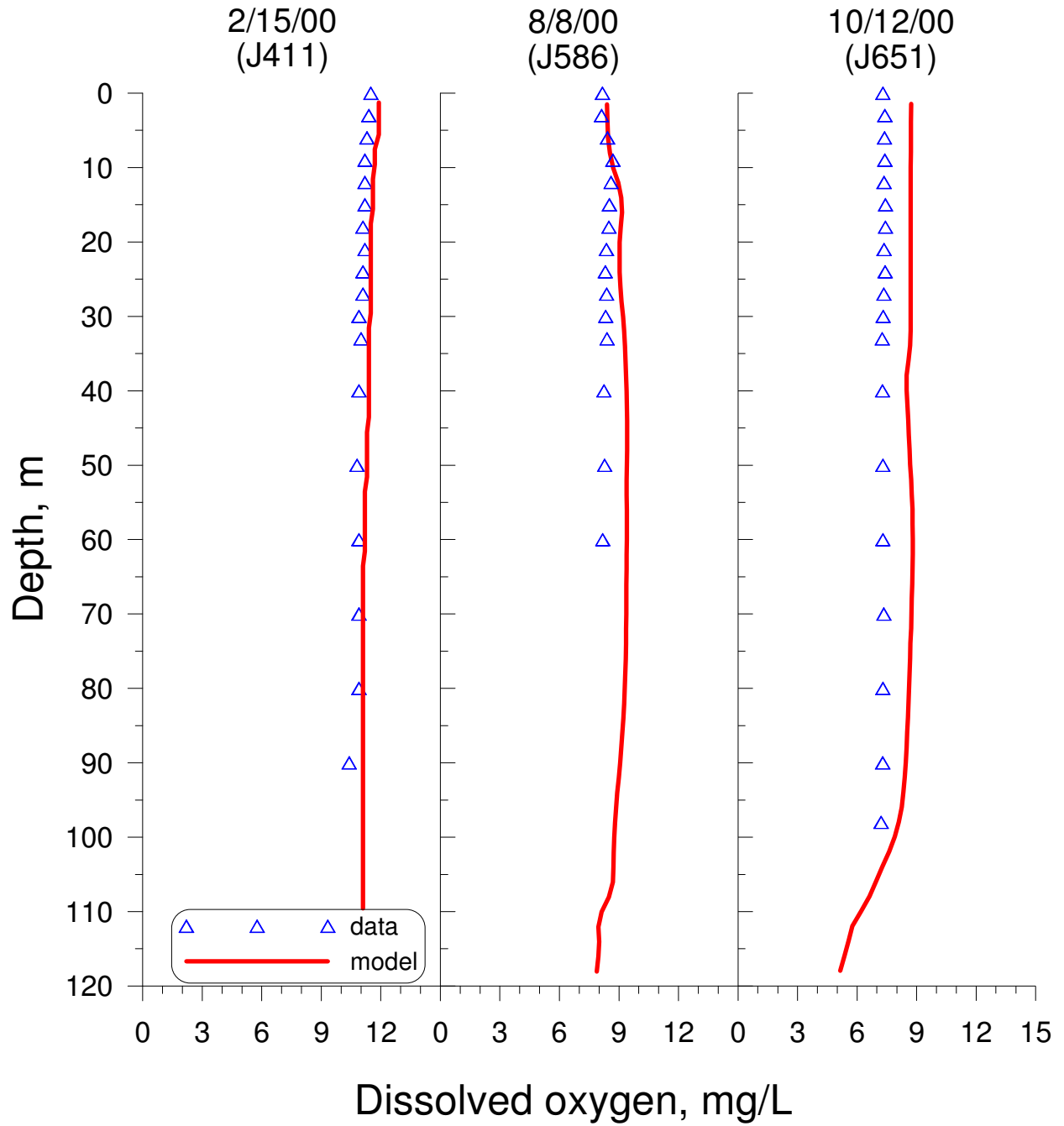


Figure 25. Selected model-data dissolved oxygen vertical profile comparisons at Spring Canyon (LRFEP sta 9.0).

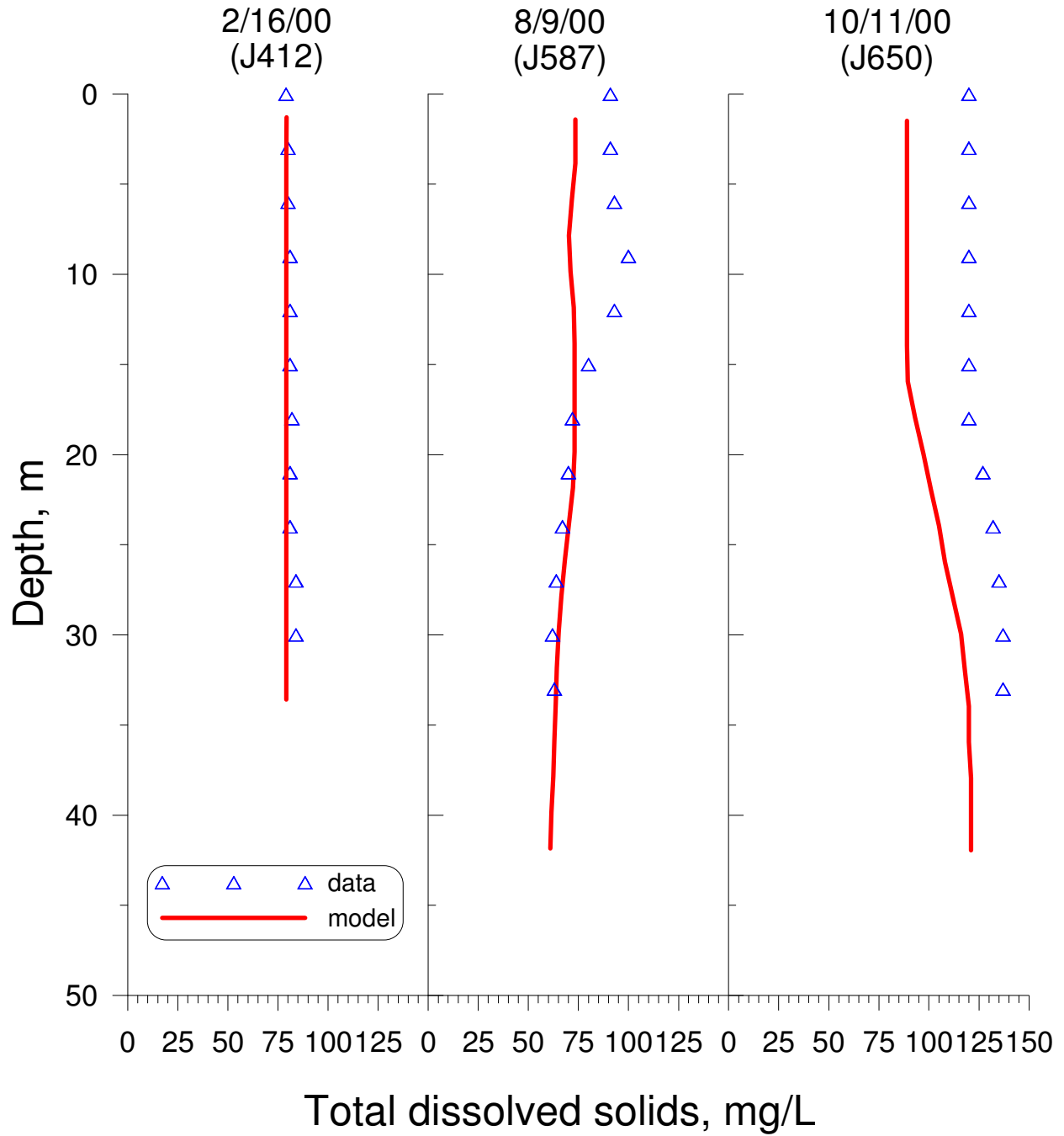


Figure 26. Selected model-data total dissolved solids vertical profile comparisons at Porcupine Bay (LRFEP stat 4.0).

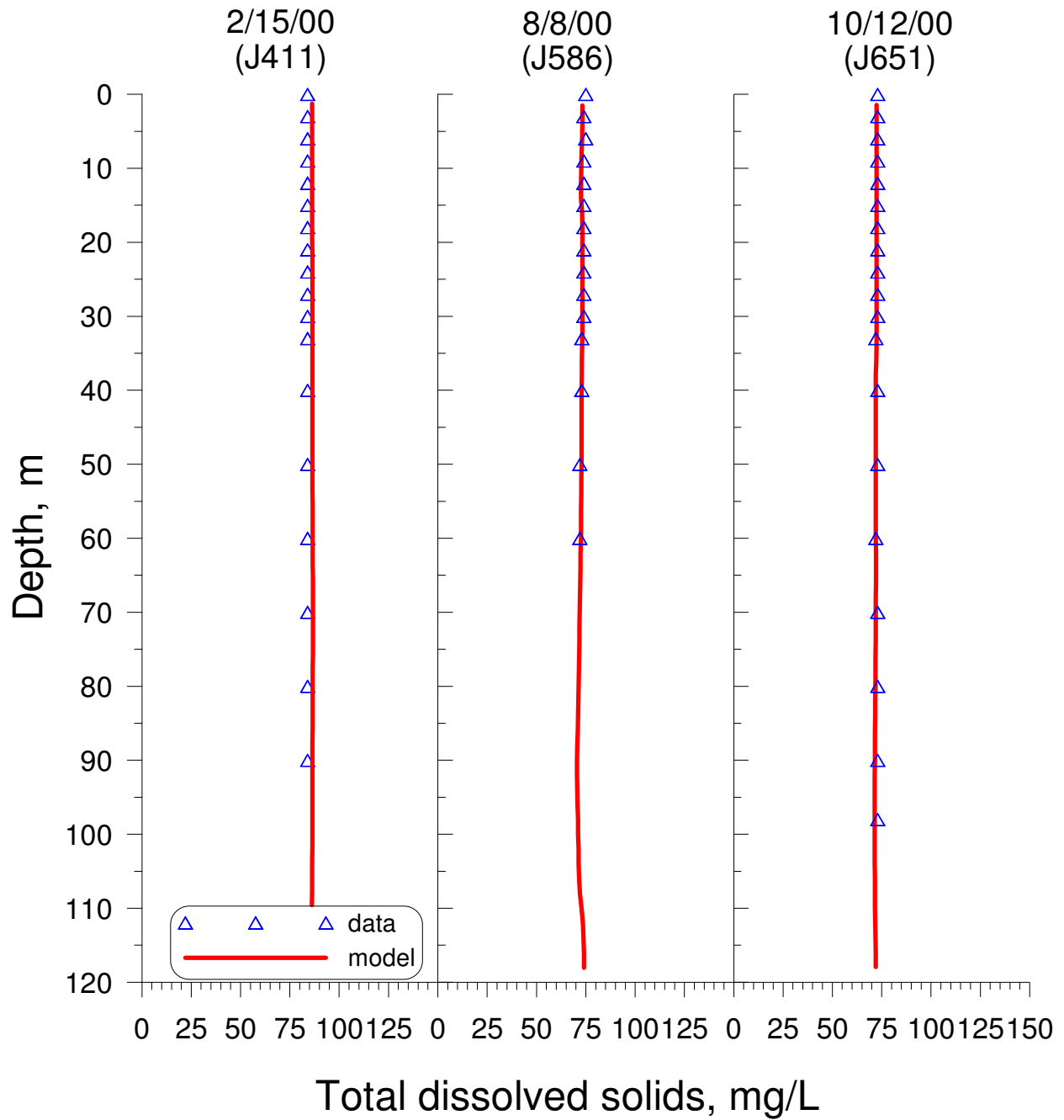


Figure 27. Selected model-data total dissolved solids vertical profile comparisons at Spring Canyon (LRFEP stat 9.0).

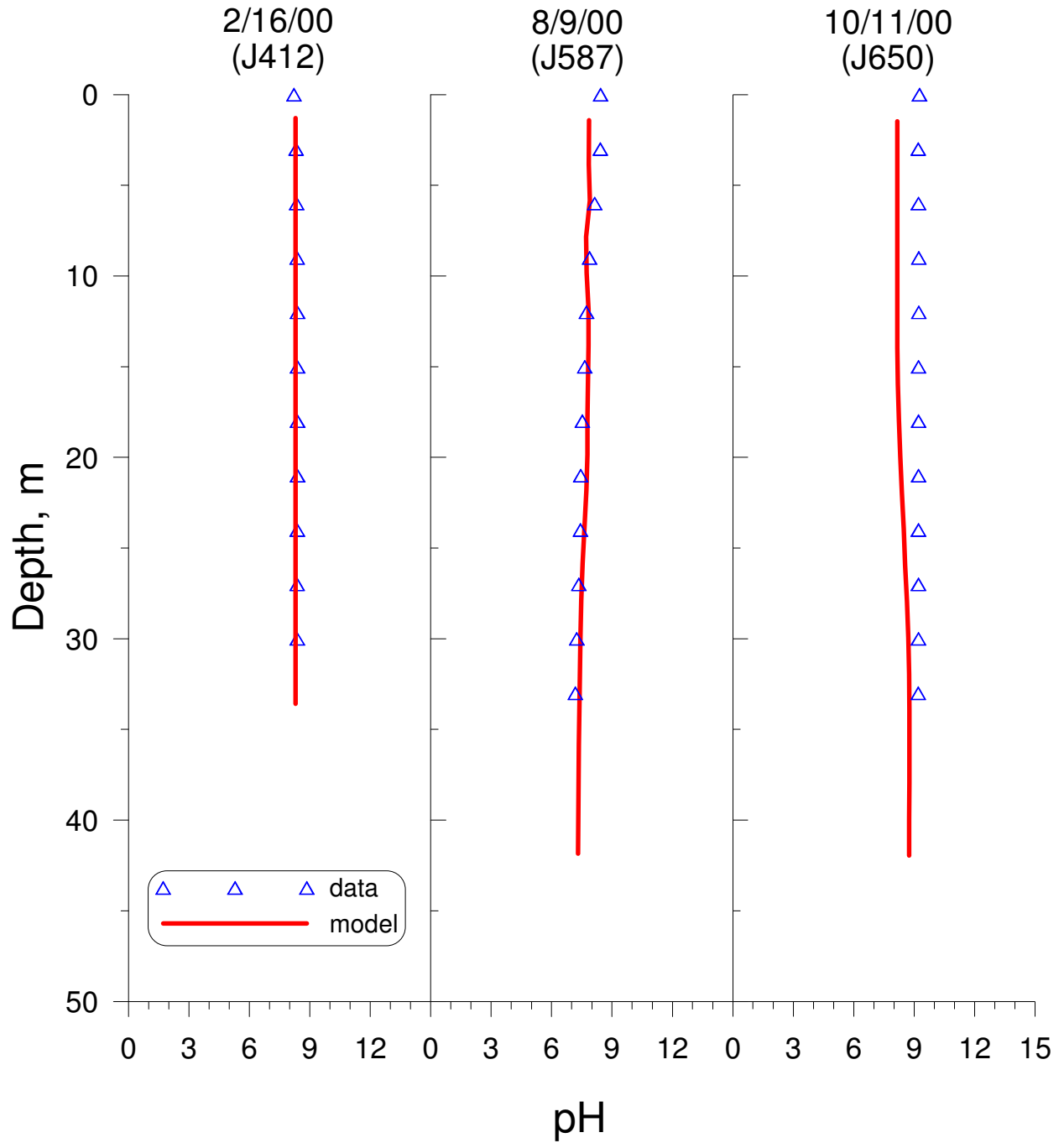


Figure 28. Selected model-data pH vertical profile comparisons at Porcupine Bay (LRFEP stat 4.0).

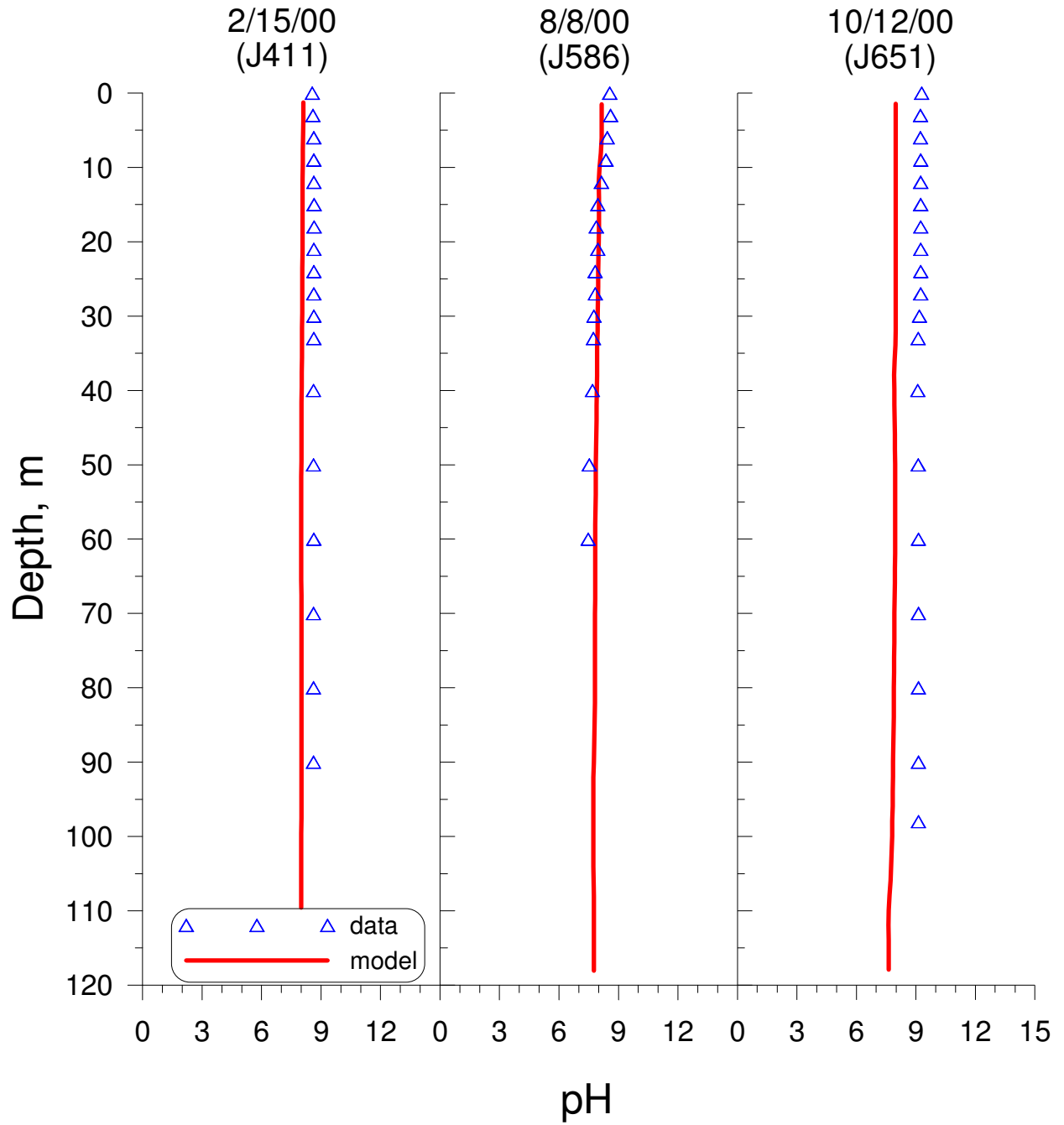


Figure 29. Selected model-data pH vertical profile comparisons at Spring Canyon (LRFEP stat 9.0).

Alternate Boundary Conditions

Two alternative boundary conditions were examined as a tool for examining the use of difference data sources as water quality constituent boundary conditions. These examinations were conducted to show the sensitivity of model results to apparent data conflicts.

Spokane River Dissolved Oxygen

The LRFEP monitoring station 4.0 on the Spokane River is located at approximately just downstream of the riverine to lacustrine system transition point. Thus, the model predictions, especially as they regard to the stratified condition, are sensitive to flow magnitudes and wind driven mixing. Additionally, the monitoring station is located at a bend in the pre-dam river channel which may influence the near-bank system. Attempts to account for these effects did not resolve the apparent data conflict between the choices of Spokane River upstream boundary conditions for dissolved oxygen.

There are hourly-averaged dissolved oxygen data available from Avista Utilities at the Little Falls Dam tailrace. These are the preferred boundary condition data due to their sampling frequency and proximity to the model boundary. There are instantaneous in-stream vertical profile data available from the LRFEP at station 4.0. Ideally, the model should predict the in-stream data using the upstream data as model input. However, the model does not match the data at station 4.0, even when the in-stream data are used as the upstream boundary condition. The model-data comparison under both boundary condition scenarios are illustrated in Figure 30.

The model values at station 4.0 during January and February are largely an echo of the upstream boundary condition. During the early and late summer, when stratification occurs, both data boundary condition scenarios under predict the data. After fall turnover, the model over predicts the data under both scenarios. This suggests that the actual behavior at station 4.0 is significantly influenced by the lacustrine arm of the Spokane River and its interaction with the Columbia River.

The LRFEP conducted a study along the Spokane River arm in 2006 which showed that the Columbia River water rides upstream over trapped Spokane River water during the summer. This results in two to three months of segregation where the hypolimnion suffers from lack of DO replenishment and potential hypoxia.

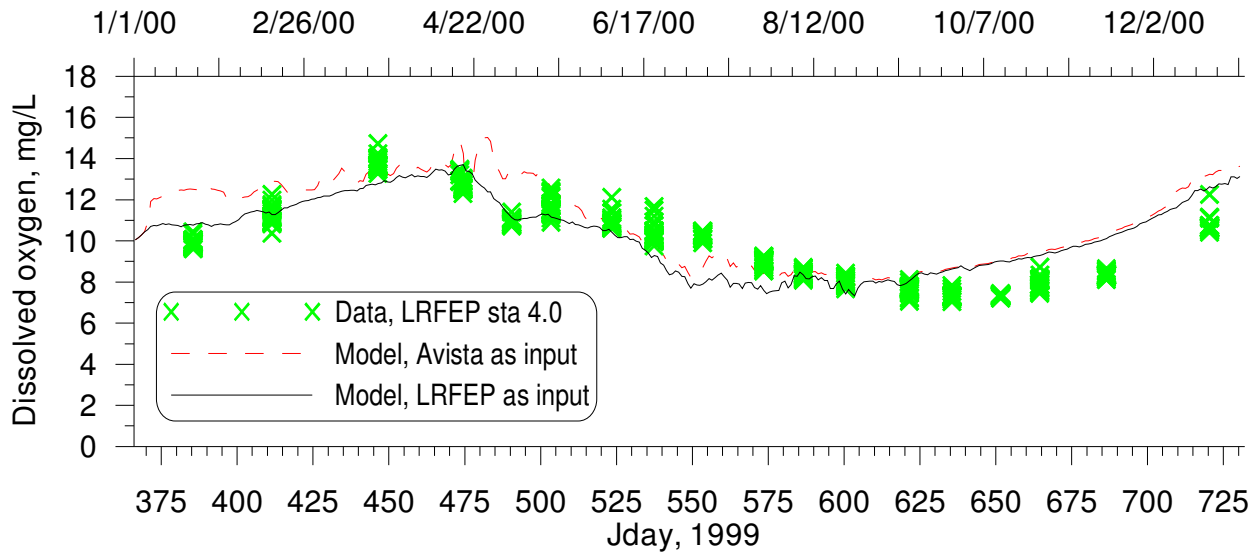


Figure 30. Model-data comparison of dissolved oxygen at LRFEP station 4.0 under different boundary condition scenarios.

Total Inorganic Carbon (pH)

Within the Lake Roosevelt system, there are few sources and sinks of carbon. Plant and phytoplankton concentrations are low and there is little geochemical activity, so pH is largely dictated by boundary conditions and atmospheric interaction. The pH data available from Environment Canada near the confluence of the mainstem Columbia River and the Pend Oreille River form the preferred upstream boundary condition data due to their proximity and roughly bi-weekly sampling frequency. In-stream data are available at LRFEP station 0.0, with a typical travel time of 1 day.

Given the proximity of station 0.0 to the upstream boundary (the U.S.-Canadian border) and the relatively conservative-constituent behavior of pH within the system, it is expected that the model prediction at station 0.0 would echo the upstream carbon boundary condition. Figure 31 illustrates that when the LRFEP data are used as the boundary condition, the model does echo the input. However, the Environment Canada data, when used as the boundary condition, do not show good model-data agreement at station 0.0. This suggests that the LRFEP and Environment Canada data are not in agreement and one or both sources may have some measurement error.

The LRFEP pH data values are generally between 8 and 9. Much of the system shows pH over 9 in the late fall, over multiple years. The cause of the elevated pH is unclear. When the LRFEP station 0.0 and 4.0 pH data are used as model input for the Columbia and Spokane Rivers, respectively, the mean pH values are conserved downstream at LRFEP station 9.0 upstream of Grand Coulee Dam, but the data at station 9.0 does not match the model predictions. The data show vertical gradients and values much larger than the model predicts. Figure 32 illustrates the model-data comparison.

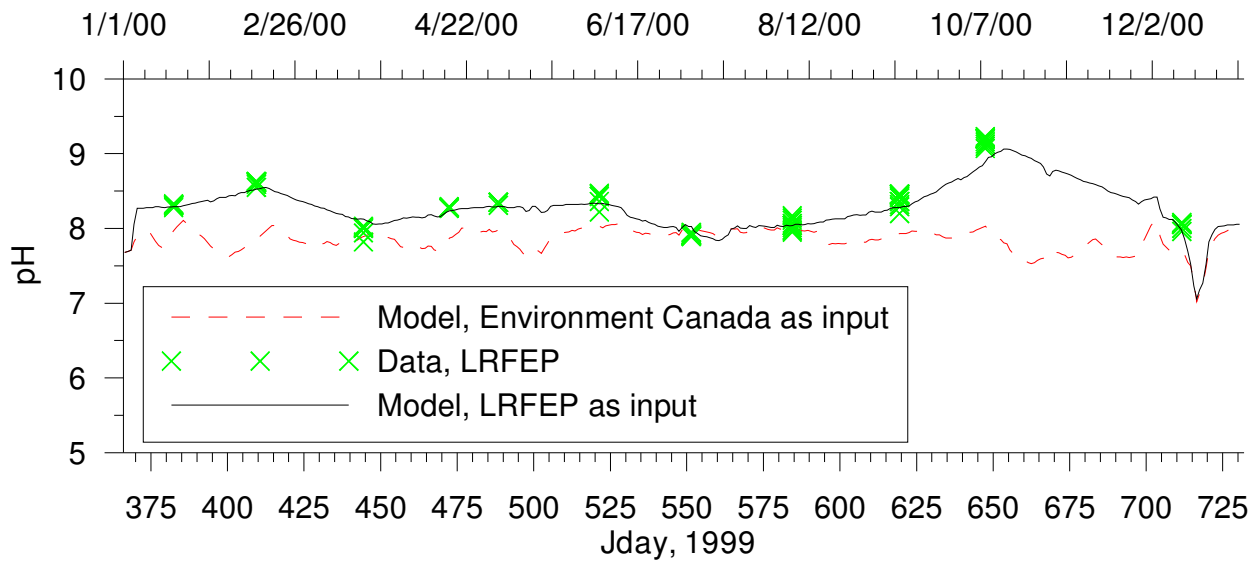


Figure 31. Model-data comparison of pH at LRFEP station 0.0 under different boundary condition scenarios.

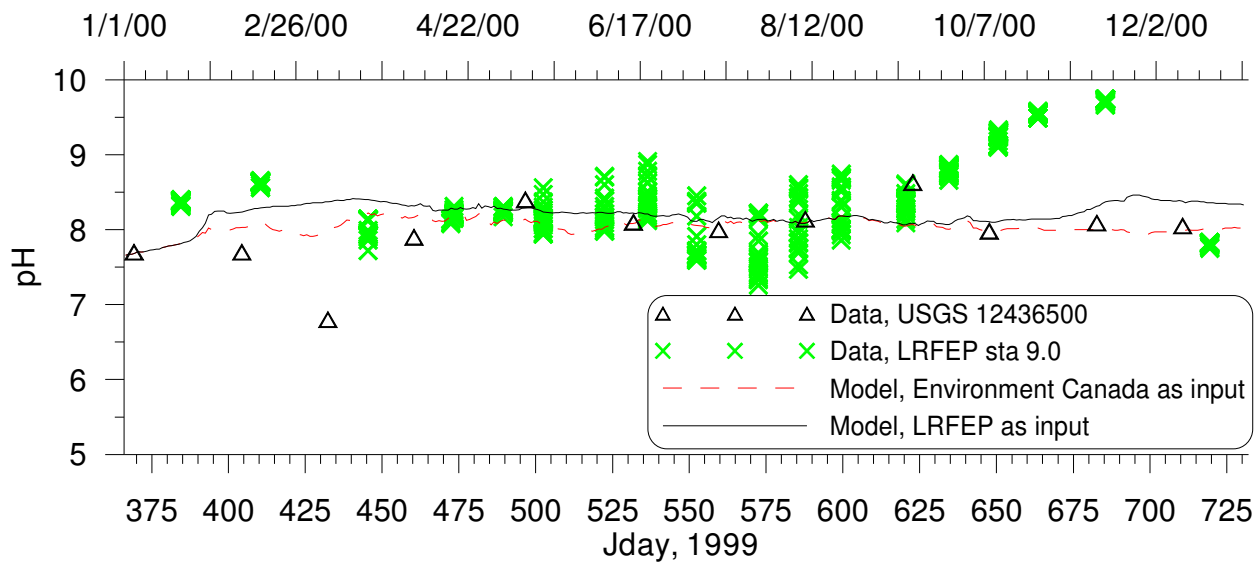


Figure 32. Model-data comparison of pH at LRFEP station 9.0 under different boundary condition scenarios.

Biotic Modeling Approach and Calibration

Approach

The biotic modeling approaches used the conservation of mass as the governing equation. For a more detailed discussion of the algae and zooplankton algorithms, refer to Cole and Wells (2006). Details of the fish bioenergetics model are covered in Appendix C. A conceptual schematic of the interaction among the water quality constituents, algal groups, and zooplankton groups is shown as Figure 33. The fish model uses the zooplankton densities and water temperatures as input, and does not influence the model results.

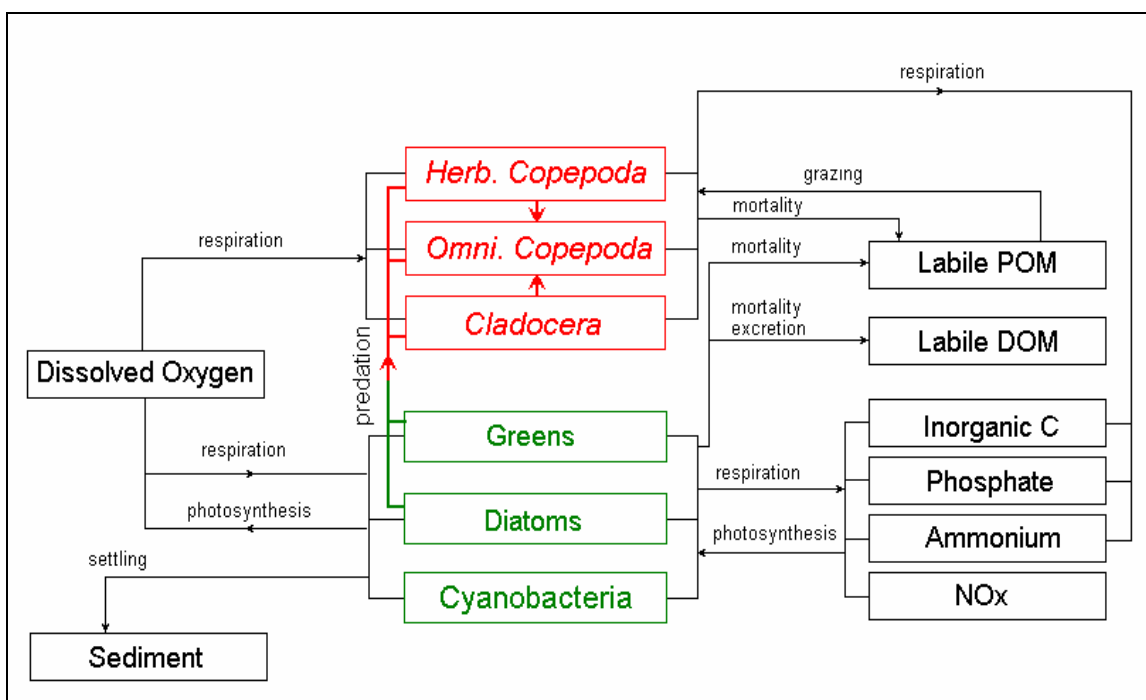


Figure 33. Conceptual diagram of the water quality, algae, and zooplankton interaction.

Algae

Algae biomass was divided into three functional groups based on the taxa data: diatoms, mixed green algae, and cyanobacteria.

The diatom group differs from the mixed green algae in several ways. Diatoms use silica as a nutrient. Since silica was not a limiting nutrient in Lake Roosevelt, silica was not included in the model; however, diatoms are heavier than green algae and rely more on turbulent mixing to remain in the epilimnion due to their larger settling velocity. Diatoms are known to bloom in the spring in Lake Roosevelt, and provide a large portion of the primary production available to zooplankton. *Asterionella formosa* comprised the largest single species within the diatom group

(~50% of the *Bacillariophyceae* biomass) and *Asterionella* was used as the basis for the diatom group kinetics.

The mixed green algae group (or simply the green) was comprised of mostly brown-green algae (*Chryptophyta*) and then green algae (*Chlorophyta*). The green algae are the dominant perennial food supply for zooplankton. Brown-green kinetic values were selected from the literature.

The cyanobacteria group (or blue-green algae), while a small portion of the total algal biomass, was included separately as zooplankton do not appreciate prey up this group. While *Oscillatoria sp.* dominated the group, general cyanobacteria literature values were used for the kinetics. Values for *Oscillatoria* kinetics were scarce and variable in the literature.

Zooplankton

The zooplankton biomass was divided into three functional groups: herbivorous copepoda, omnivorous copepoda, and cladocera.

While all zooplankton biomass decreases during the winter, the copepoda maintain larger numbers than the cladocera in Lake Roosevelt. *Leptodiptomus ashlandi* and *Diacyclops bicuspidatus thomasi* comprised the main copepoda species. Omnivorous copepoda were assumed to comprise 10% of the total copepoda biomass (personal communication, Mike Mazur). Omnivorous copepoda were assumed to prefer prey equally between the herbivorous copepoda and cladocera.

The cladocera biomass is dominated by *Daphnia sp.* Other cladocera make up roughly 5% of the total cladocera biomass. *Daphnia pulex* dominate the cladocera biomass (~50%). *Daphnia* differ from copepoda in that they are much larger. Diet studies show that cladocera are the preferred prey item for kokanee in Lake Roosevelt (refer to LRFEP annual reports). While copepoda provide for a perennial food supply for zooplanktivorous fish, sampling has shown that *Daphnia* spp. make up the majority of the kokanee prey biomass, especially over the spring and summer. Winter diets also show a largely *Daphnia* diet, despite the very low *Daphnia* concentrations.

Fish Bioenergetics (kokanee)

The bioenergetics growth model is based on Stockwell and Johnson (1999). A more detailed discussion is presented in Appendix C. The model was used to predict fish growth potential, represented as daily change in fish mass. Fish growth was a function of fish mass, prey density, fish and prey energy densities, water temperature, and ambient light. Fish growth potential was determined for each model cell in 30 minute time steps. The results were used in three different *growth prediction* approaches with four fish *growth application* approaches. *Growth prediction* refers to how the daily growth is determined; *growth application* refers to how the daily fish mass is formulated.

Four fish mass time series formulations (or *growth applications*) were allowed, and could be applied to each of the three *growth predictions*.

- 1) Fish mass could remain constant (steady).
- 2) Fish mass could grow at a user specified exponential rate. This facilitates comparisons among locations and foraging methods.
- 3) Fish mass could be set to daily user specified values. This allowed model output to be returned as model input.
- 4) Daily fish growth could be added to the starting fish mass. Fish mass then becomes model predicted.

General approach

The general approach is to simply determine fish growth potential in each model cell. If fish growth is then applied to the cell, then each cell will then have its own fish mass. Cells located at the bottom of the lake are unlikely to show positive growth, and would be unlikely to represent foraging strategies for the pelagic visual zooplanktivore. A more representative strategy is approximated by the “best growth” approach.

Best growth approach

The “best growth” approach is similar to the general approach. However, instead of simulating growth in each cell, growth in the segment is simulated. At each time step, the cell of most positive growth is determined. The stomach contents are then passed to all of the cells in the segment for the next time step. At the end of the day, growth is determined. The daily *growth application* formulation is applied to all cells. Two assumptions limit this approach. Fish are not allowed to migrate to other segments, and fish growth over a single 30 minute time step does not directly consider the future growth from consumption. Since growth only occurs after digestion, which typically takes a few to several hours, when growth potential becomes negative due to large respiration costs associated with warm water, the approach can result in a non-foraging fish. This is unrepresentative of fish foraging. An improvement to the best growth approach is the “vertical foraging” approach.

Vertical foraging approach

The “vertical foraging” approach makes two improvements to the “best growth” approach. First, in determining the most favorable cell for growth, 20% of the energy from consumption is added to the energy available for growth. This value was based on the general energy balance of Brett and Groves (1979). The determination of growth did not include this amount; it was used only to give weight to the potential growth from consumption over a time step. The second improvement attempted to simulate the vertical foraging forays where kokanee move to the warmer high prey density water to feed and then move to deeper, colder water to digest and benefit from the lower metabolic costs. To simulate these forays, during periods of daylight, if growth potential is found to be negative, then the cell selected at each time step alternates between that of greatest consumption and that of minimum respiration costs. Daylight is user defined as some light threshold in the surface layer. The decision process at each timestep is:

- 1) Use the cell of best growth if
 - a. growth is positive
 - b. during nighttime (selects least negative)
- 2) If daylight (surface lux >1), alternate foraging and “digesting”
 - a. Forage at the cell of greatest consumption, C
 - b. “Digest” or “rest” at the cell of minimum respiration, R
 - c. Alternate foraging and digesting timesteps during daylight.

Calibration

Algae and zooplankton data are available at the LRFEP stations reported in the temperature calibration section (see Figure 13). Data were collected over a variety of depths. Algae were sampled over the euphotic zone. Zooplankton were collected at 17, 33, and 66 m depths. Duplicate samples at the same depth showed large variability.

All data for a given day were arithmetically averaged. Model results, which vary from layer to layer, were averaged over the first 10 meters of depth for comparison to the averaged data.

Algae

Weighted total algae model-data comparisons are shown in Figure 34 through Figure 45.

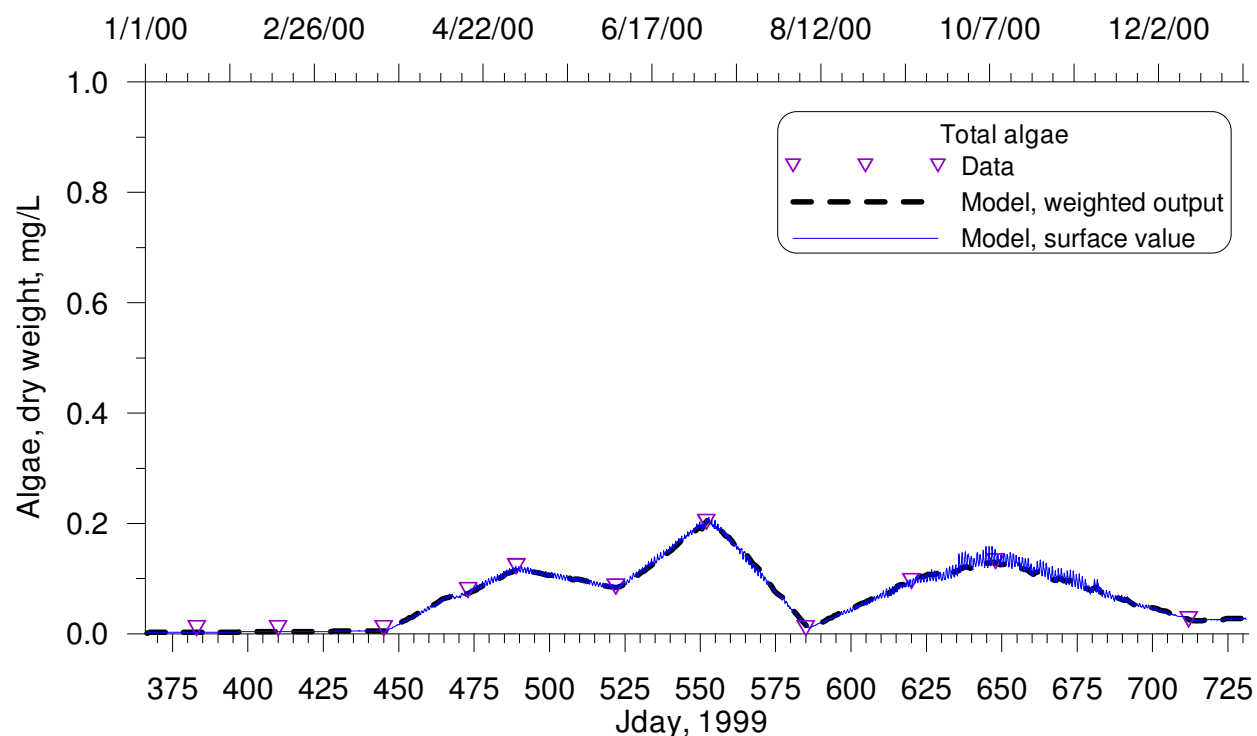


Figure 34. Model-data comparison of weighted total algae, LRFEP station 0.0.

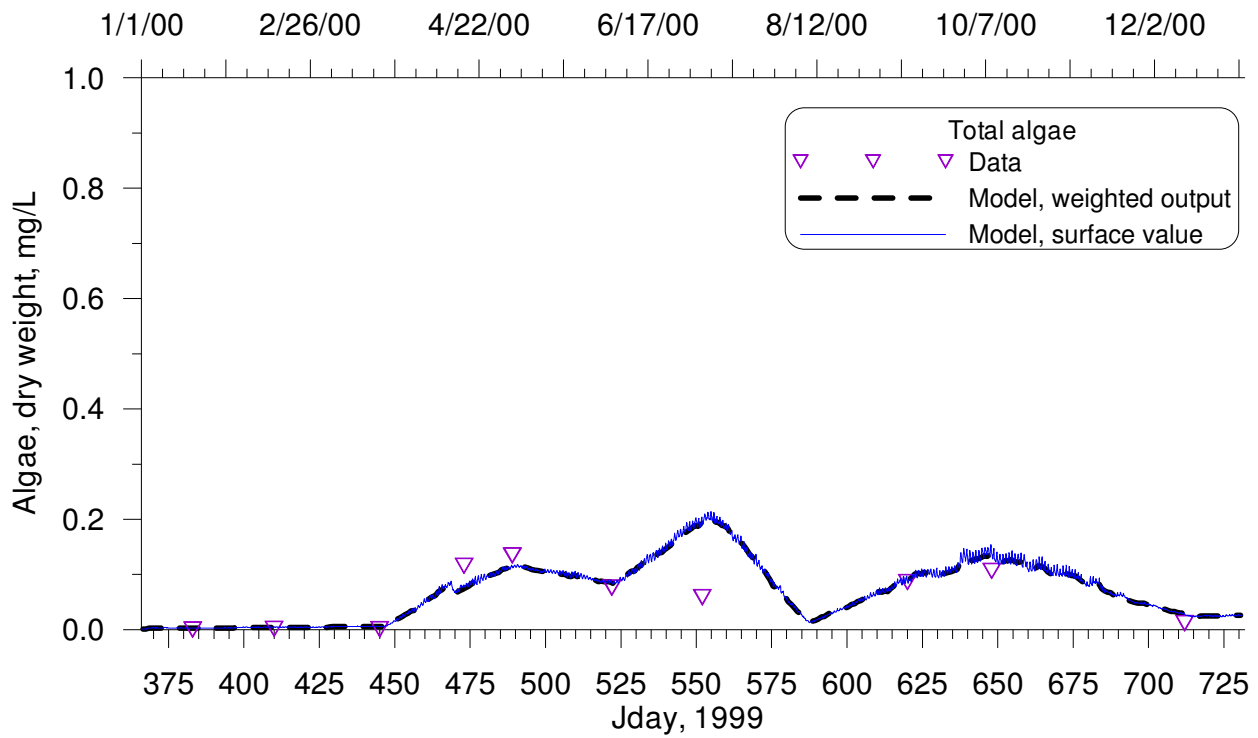


Figure 35. Model-data comparison of weighted total algae, LRFEP station 1.0

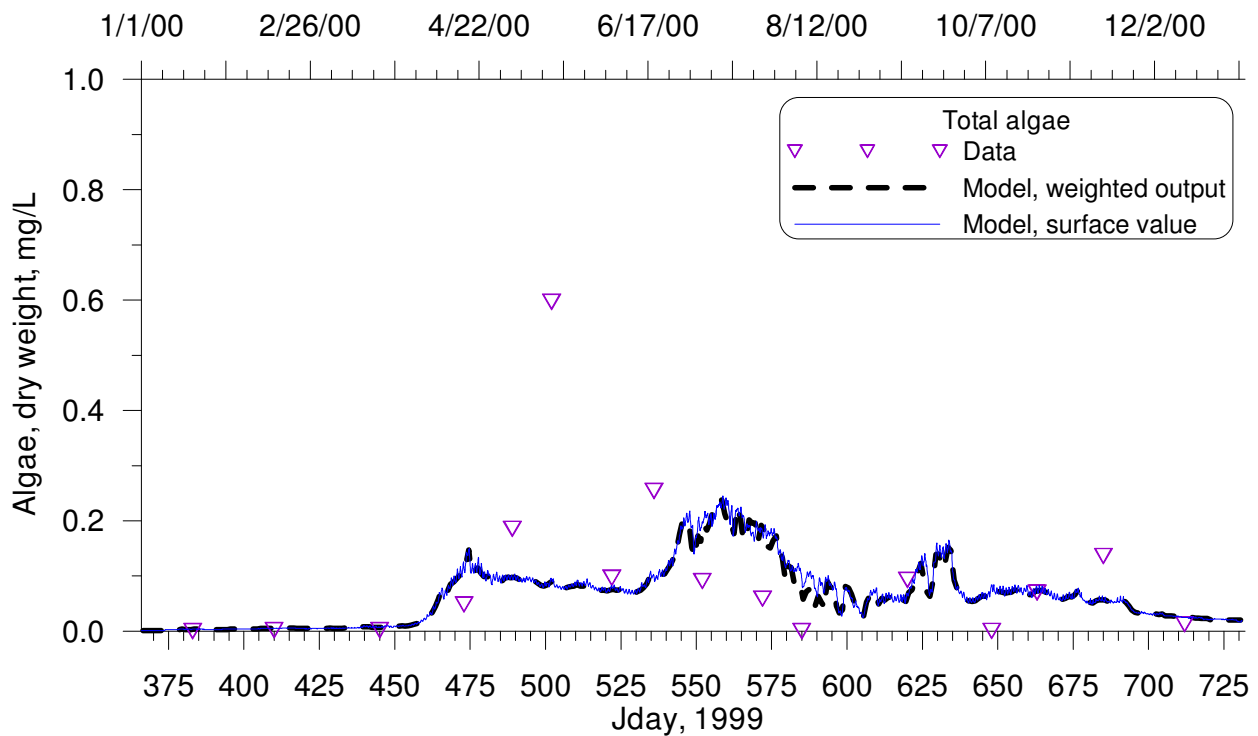


Figure 36. Model-data comparison of weighted total algae, LRFEP station 2.0

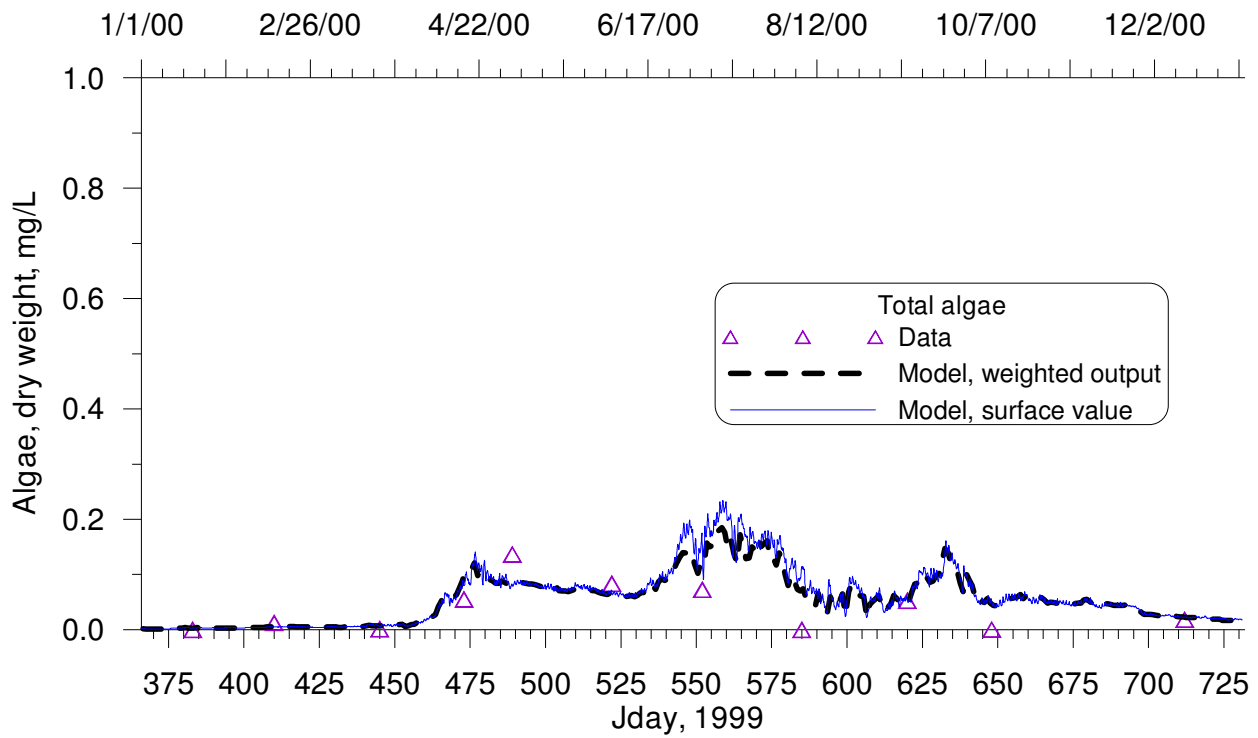


Figure 37. Model-data comparison of weighted total algae, LRFEP station 3.0

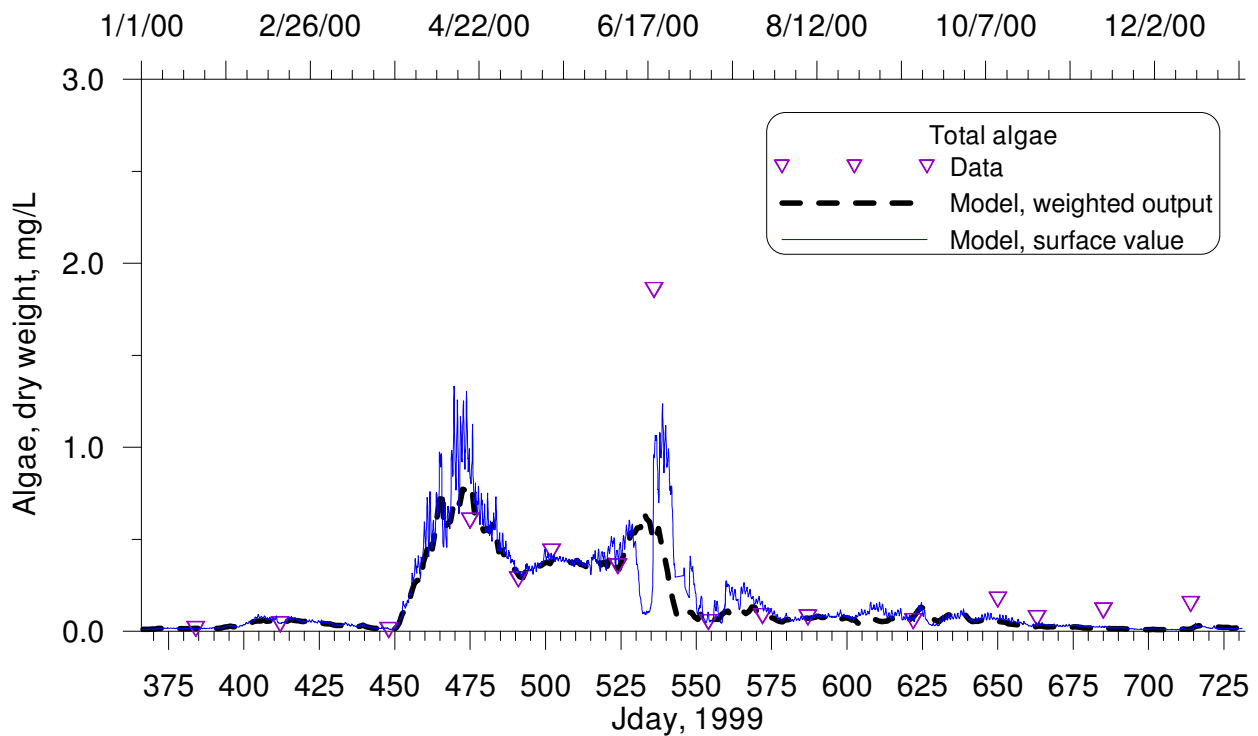


Figure 38. Model-data comparison of weighted total algae, LRFEP station 4.0

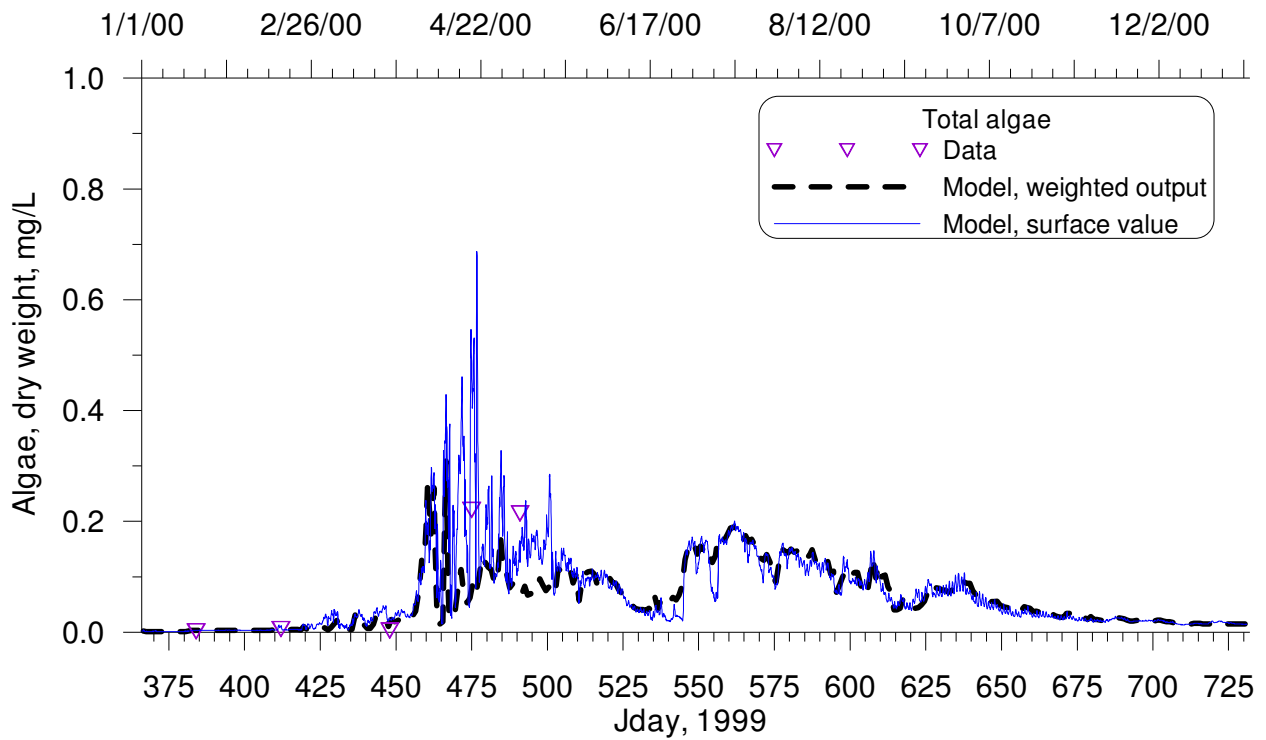


Figure 39. Model-data comparison of weighted total algae, LRFEP station 5.5

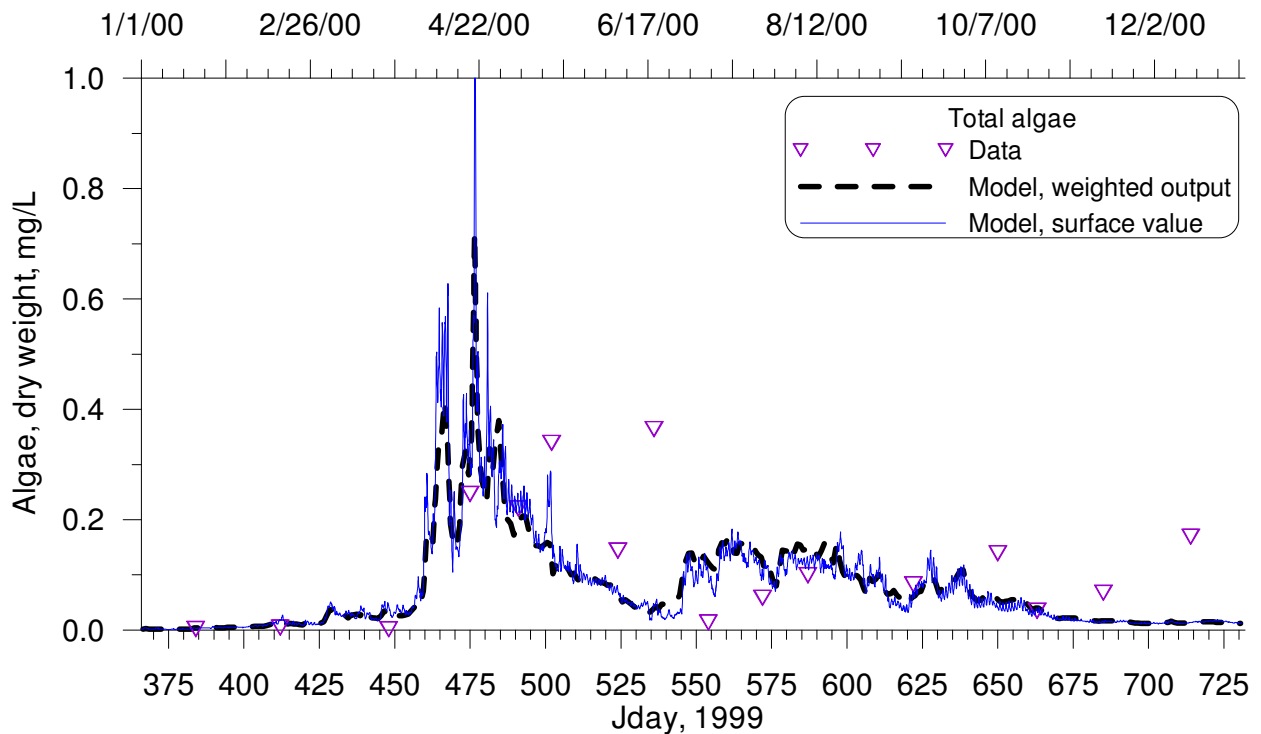


Figure 40. Model-data comparison of weighted total algae, LRFEP station 6.0

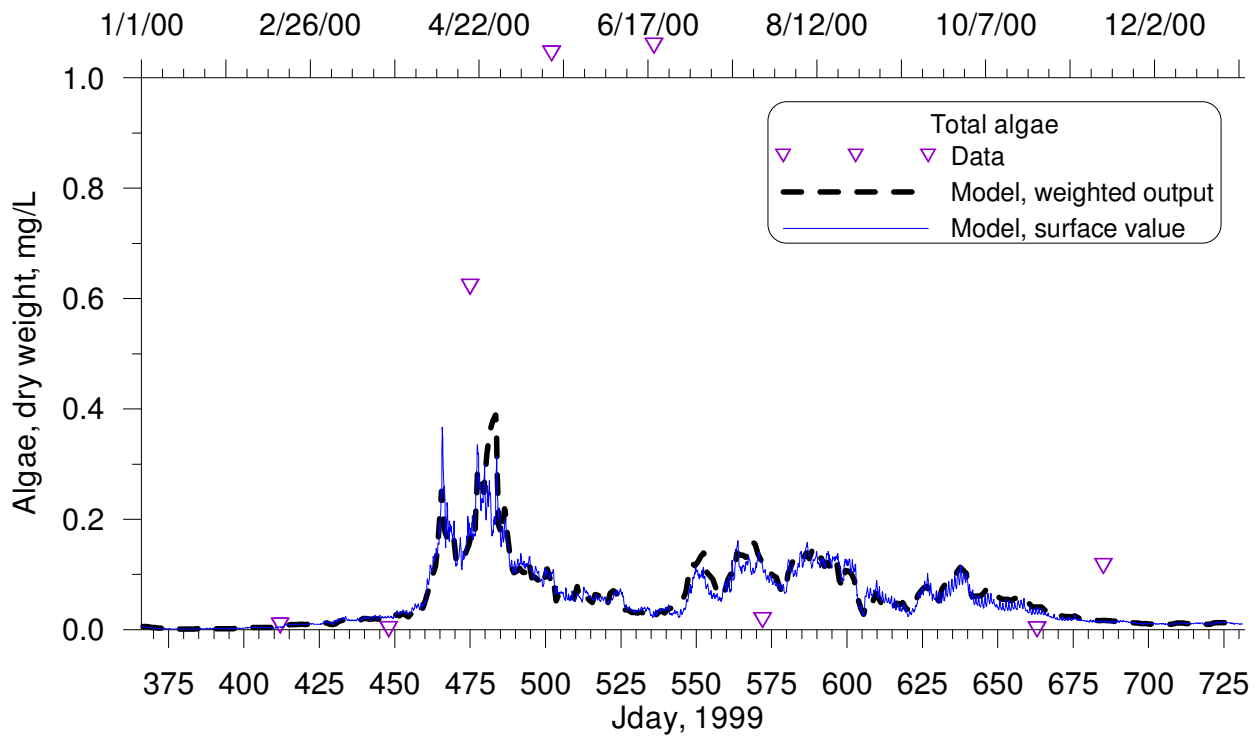


Figure 41. Model-data comparison of weighted total algae, LRFEP station 6.5

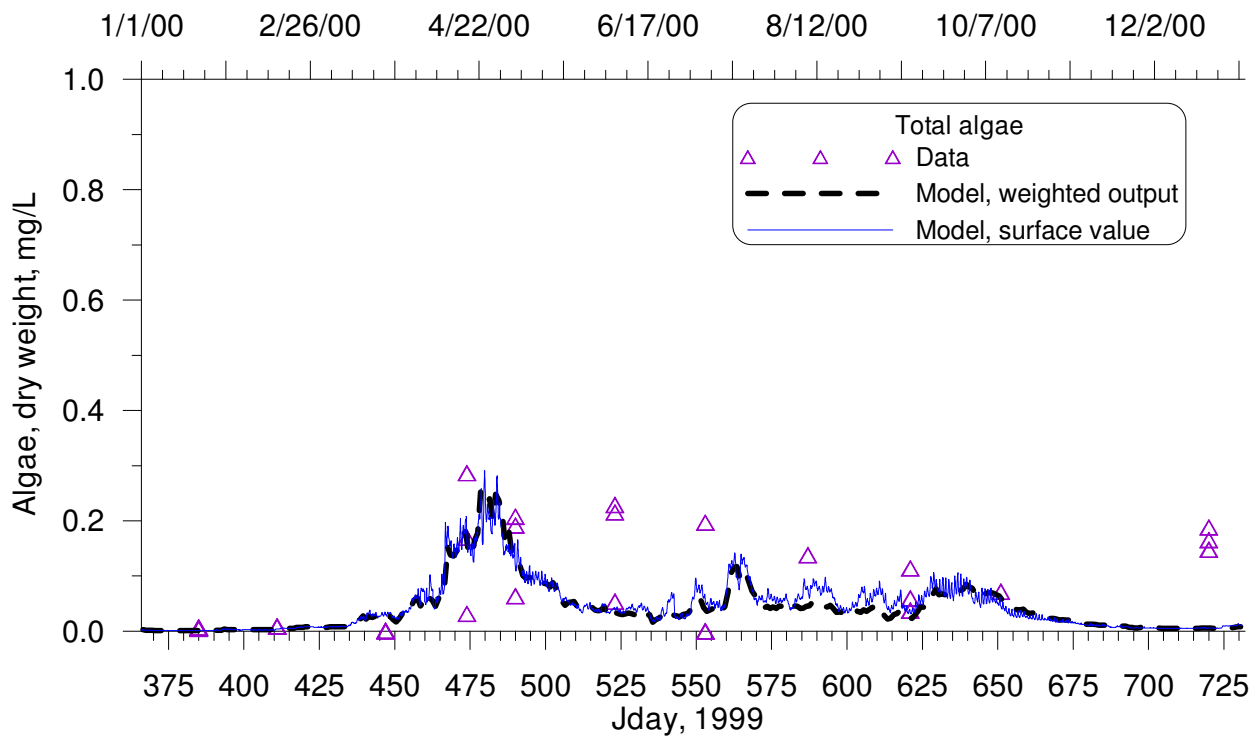


Figure 42. Model-data comparison of weighted total algae, LRFEP station 7.0

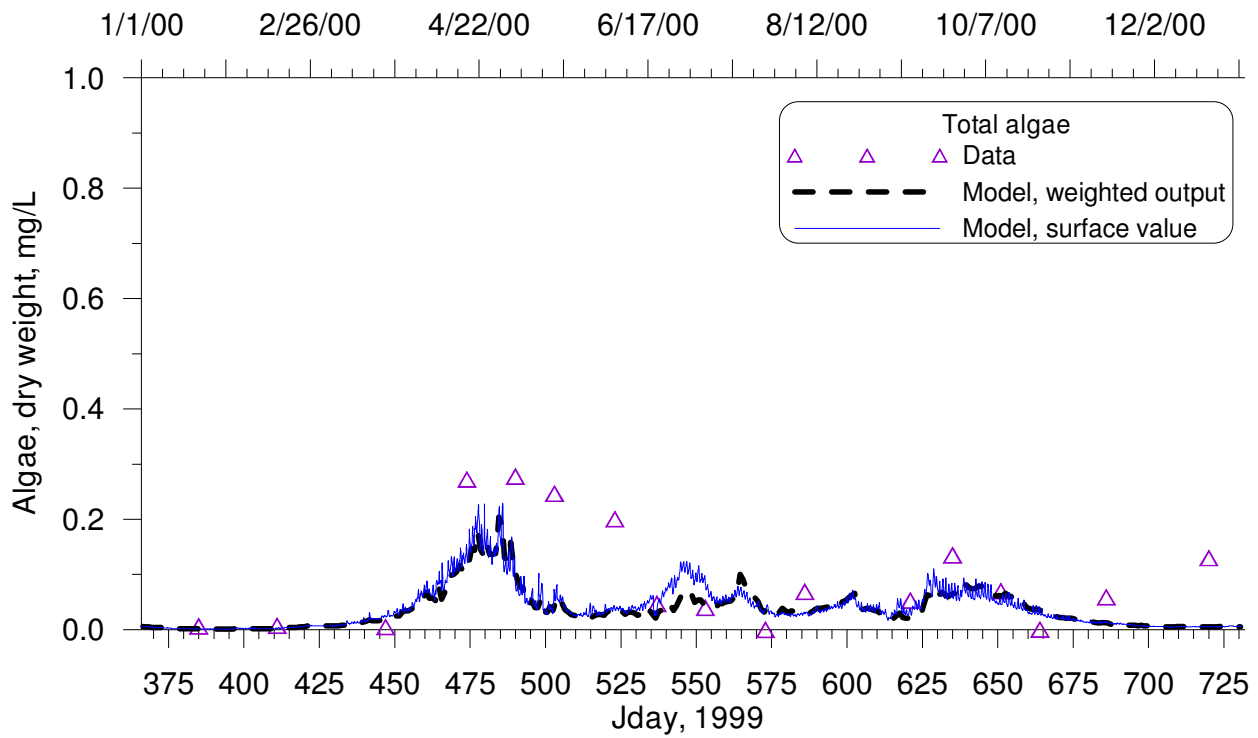


Figure 43. Model-data comparison of weighted total algae, LRFEP station 8.0

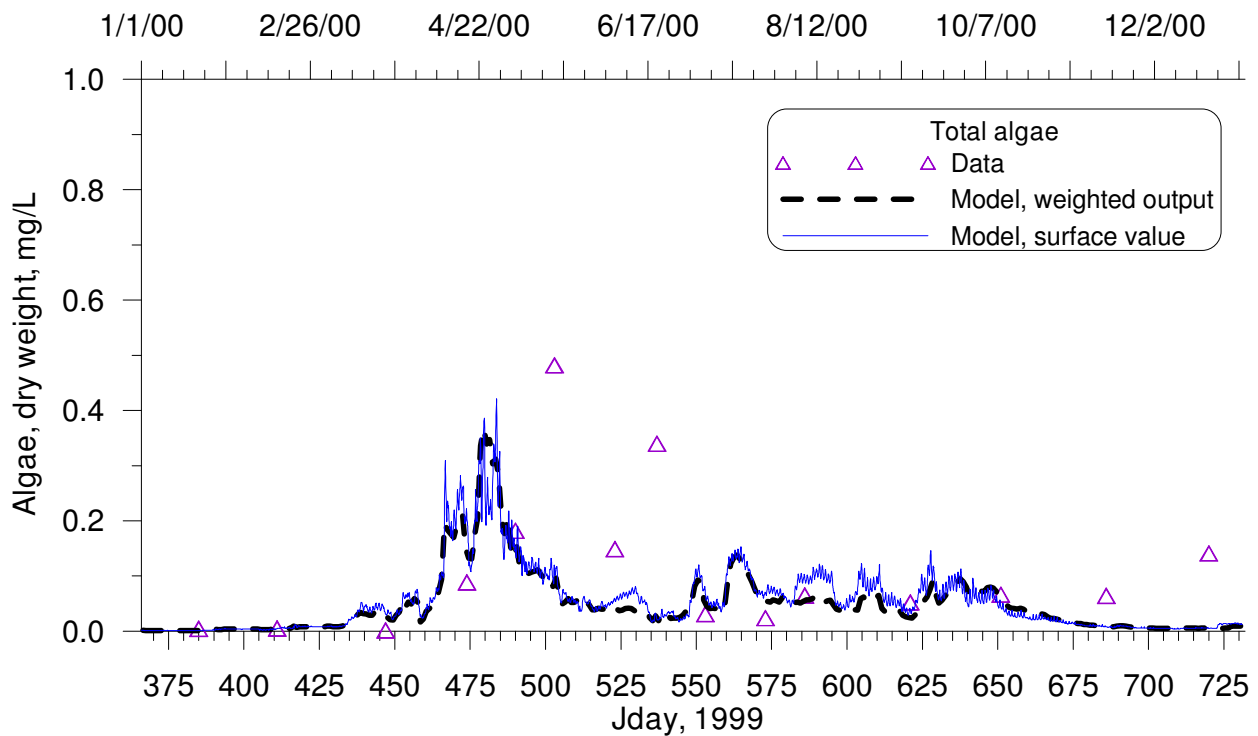


Figure 44. Model-data comparison of weighted total algae, LRFEP station 8.5

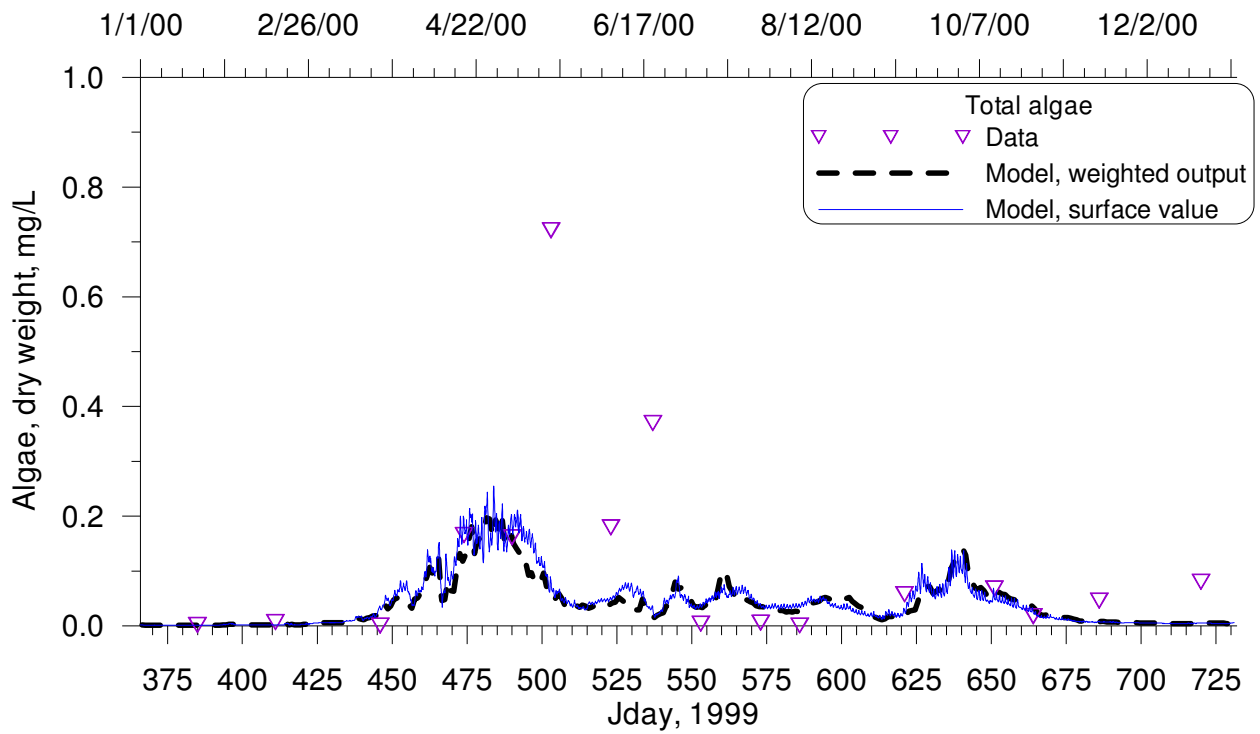


Figure 45. Model-data comparison of weighted total algae, LRFEP station 9.0

Zooplankton

Weighted total zooplankton model-data comparisons are shown in Figure 46 through Figure 57.

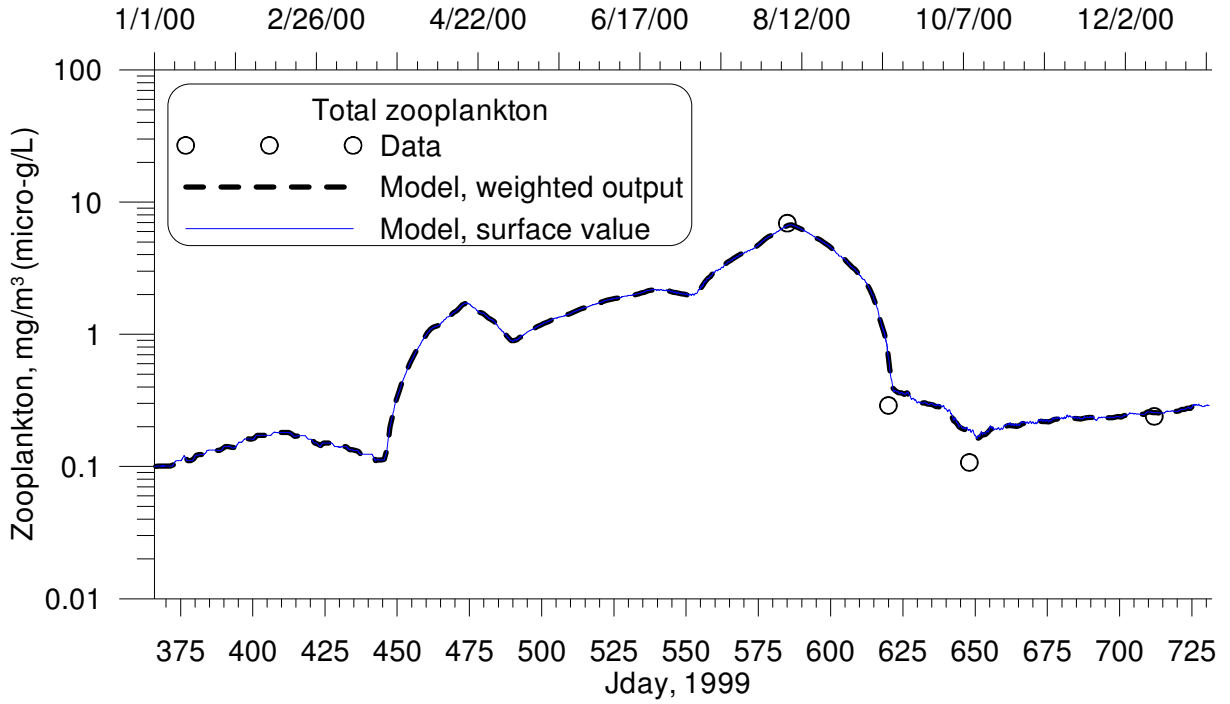


Figure 46. Model-data comparison of weighted total zooplankton, LRFEP station 0.0.

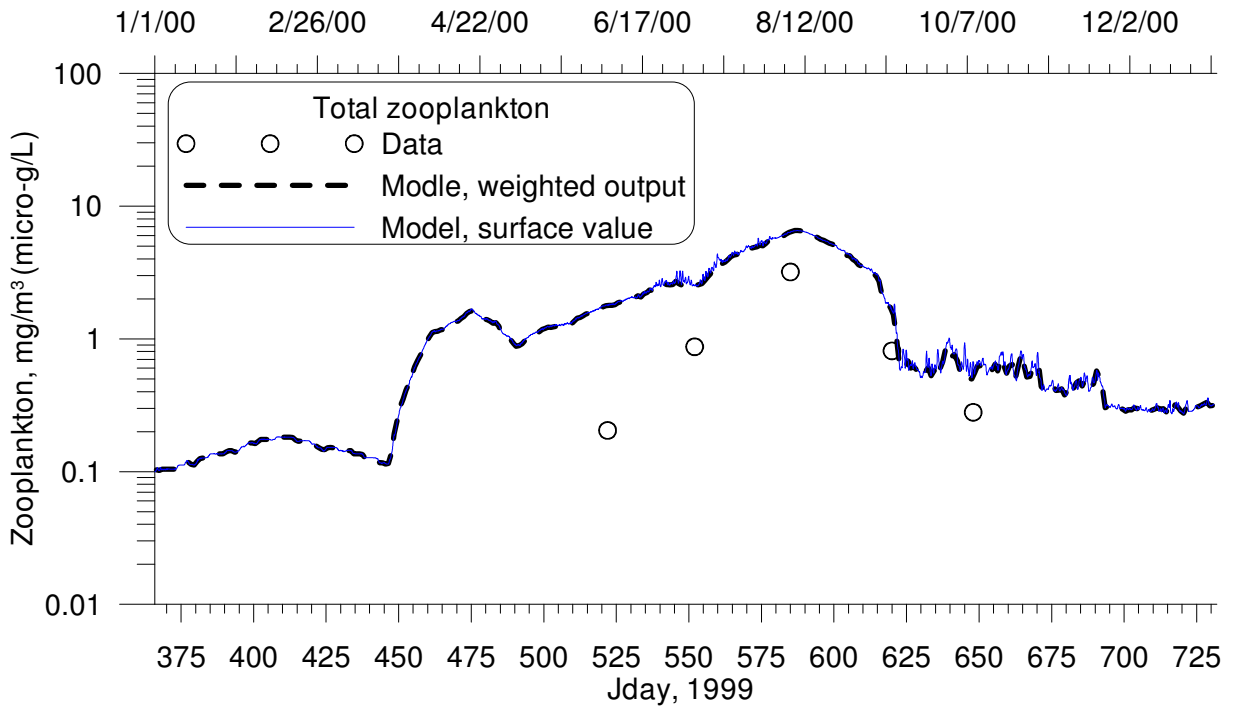


Figure 47. Model-data comparison of weighted total zooplankton, LRFEP station 1.0

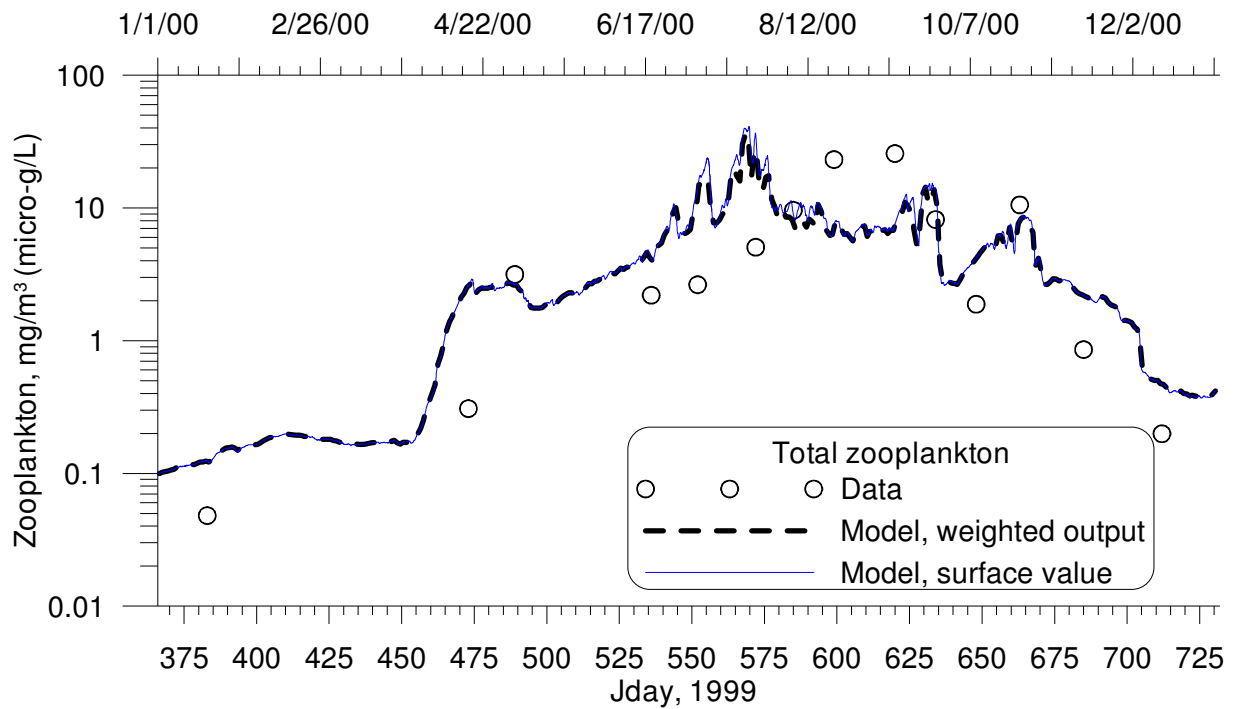


Figure 48. Model-data comparison of weighted total zooplankton, LRFEP station 2.0

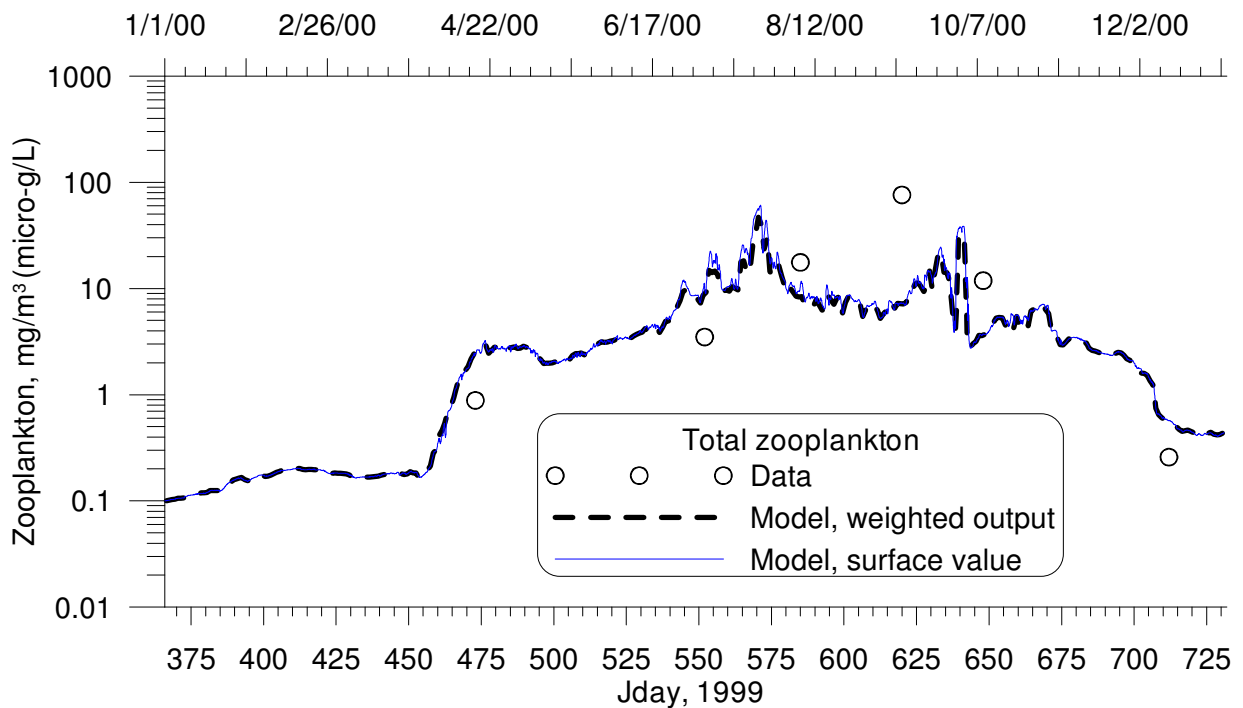


Figure 49. Model-data comparison of weighted total zooplankton, LRFEP station 3.0

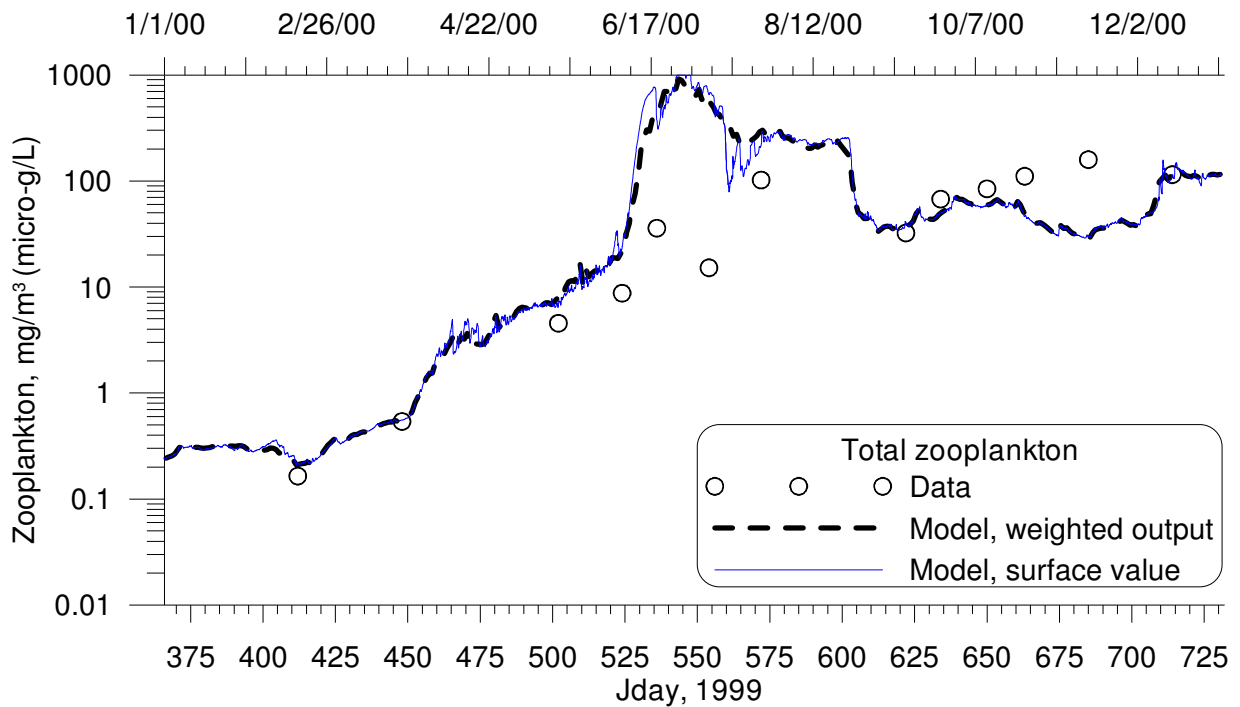


Figure 50. Model-data comparison of weighted total zooplankton, LRFEP station 4.0

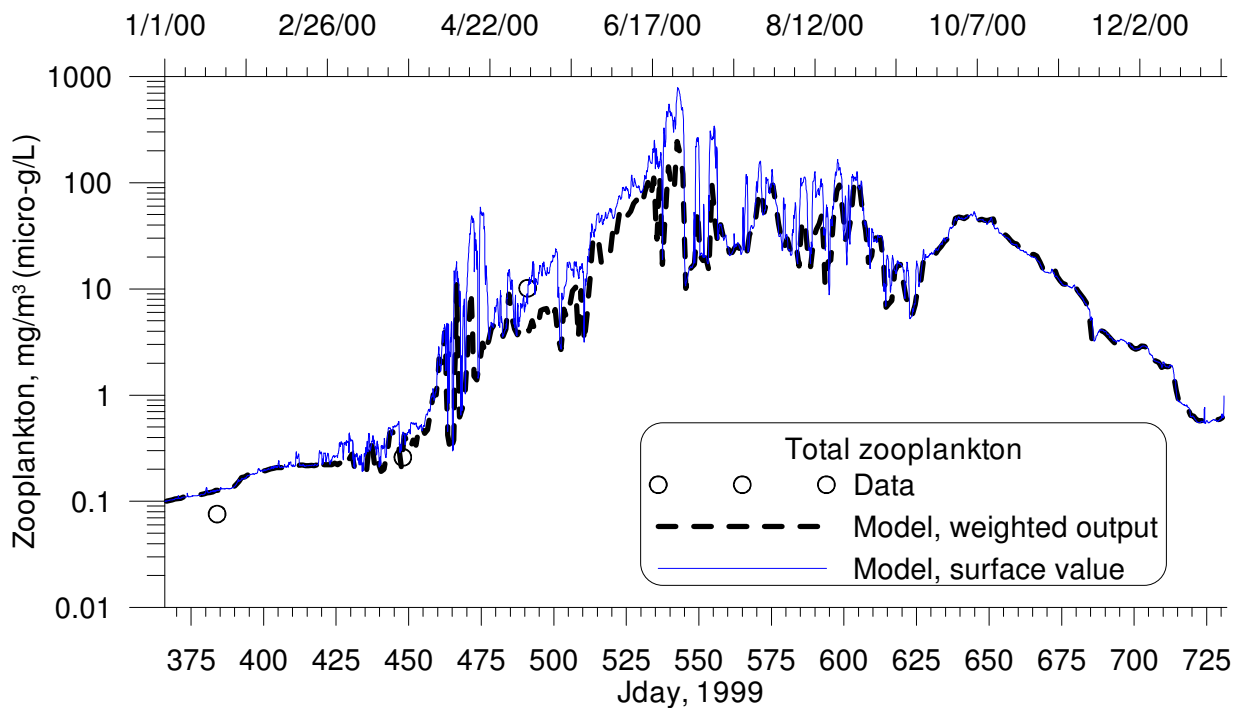


Figure 51. Model-data comparison of weighted total zooplankton, LRFEP station 5.5

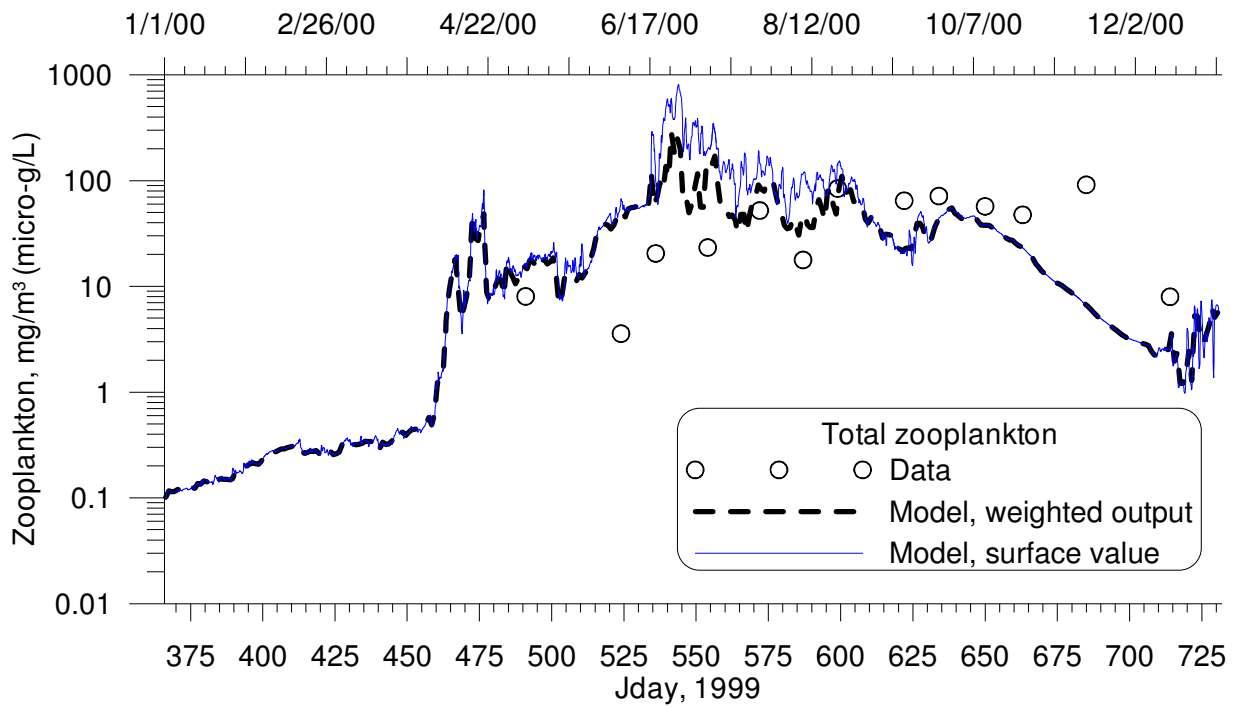


Figure 52. Model-data comparison of weighted total zooplankton, LRFEP station 6.0

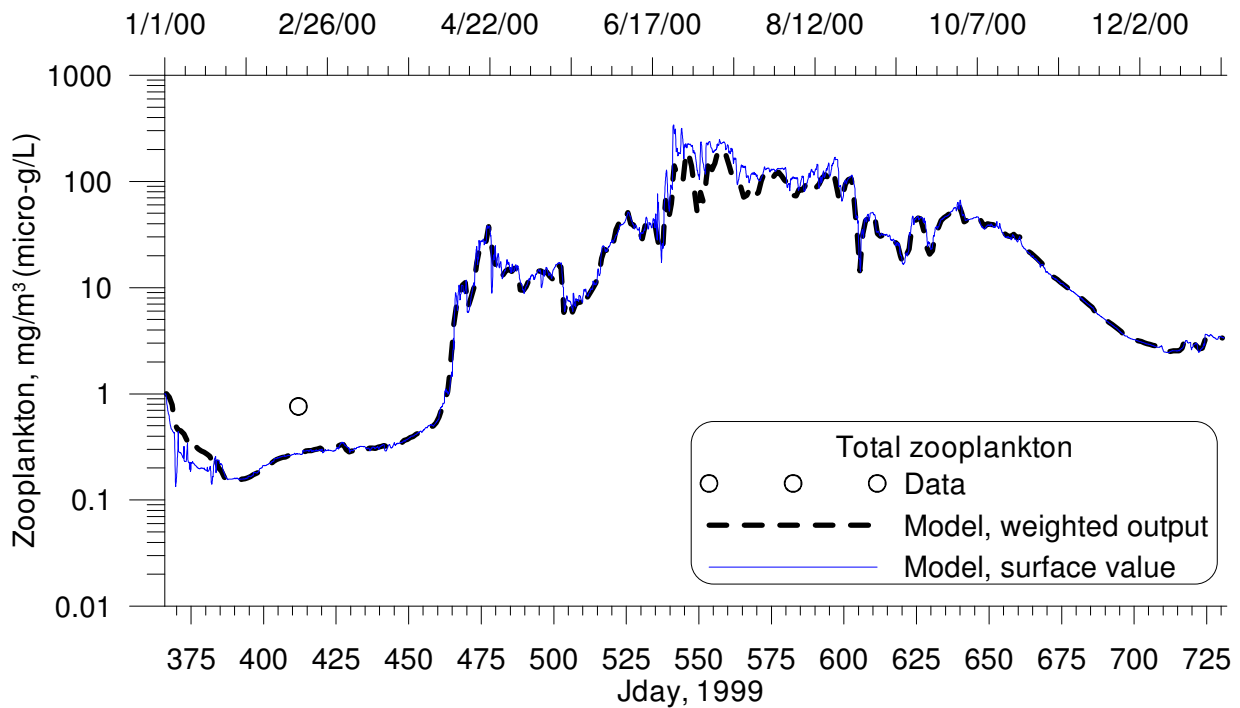


Figure 53. Model-data comparison of weighted total zooplankton, LRFEP station 6.5

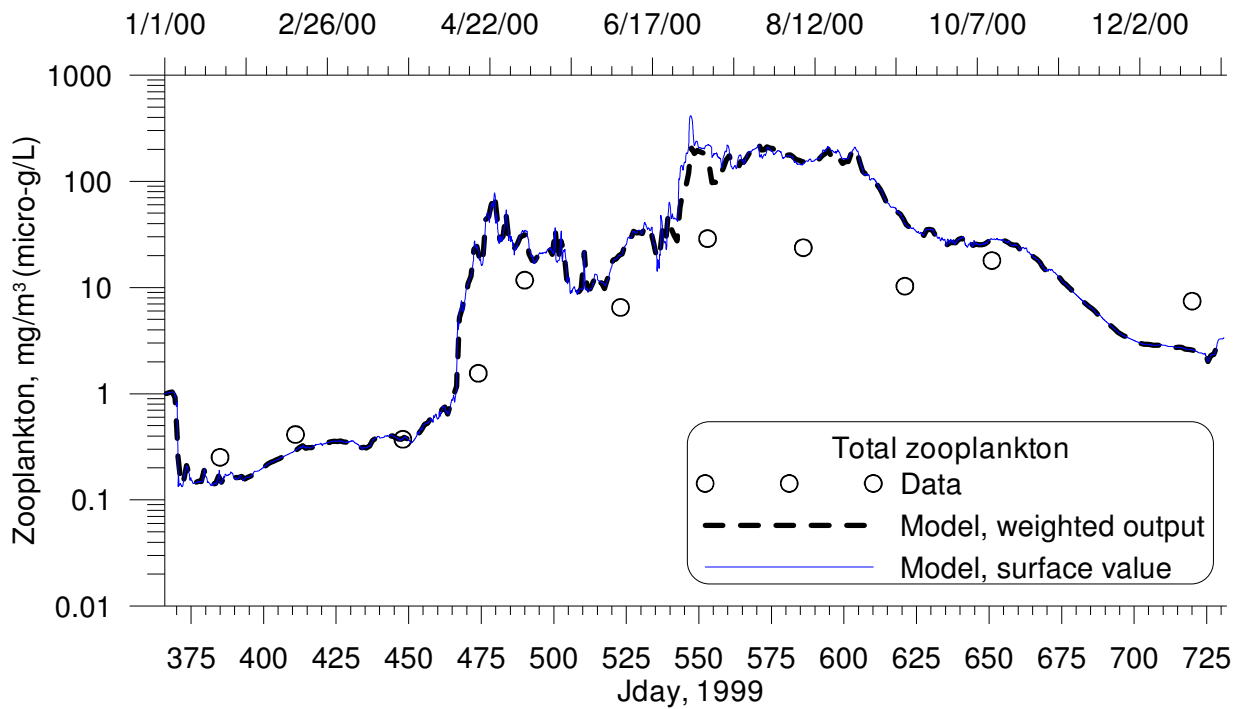


Figure 54. Model-data comparison of weighted total zooplankton, LRFEP station 7.0

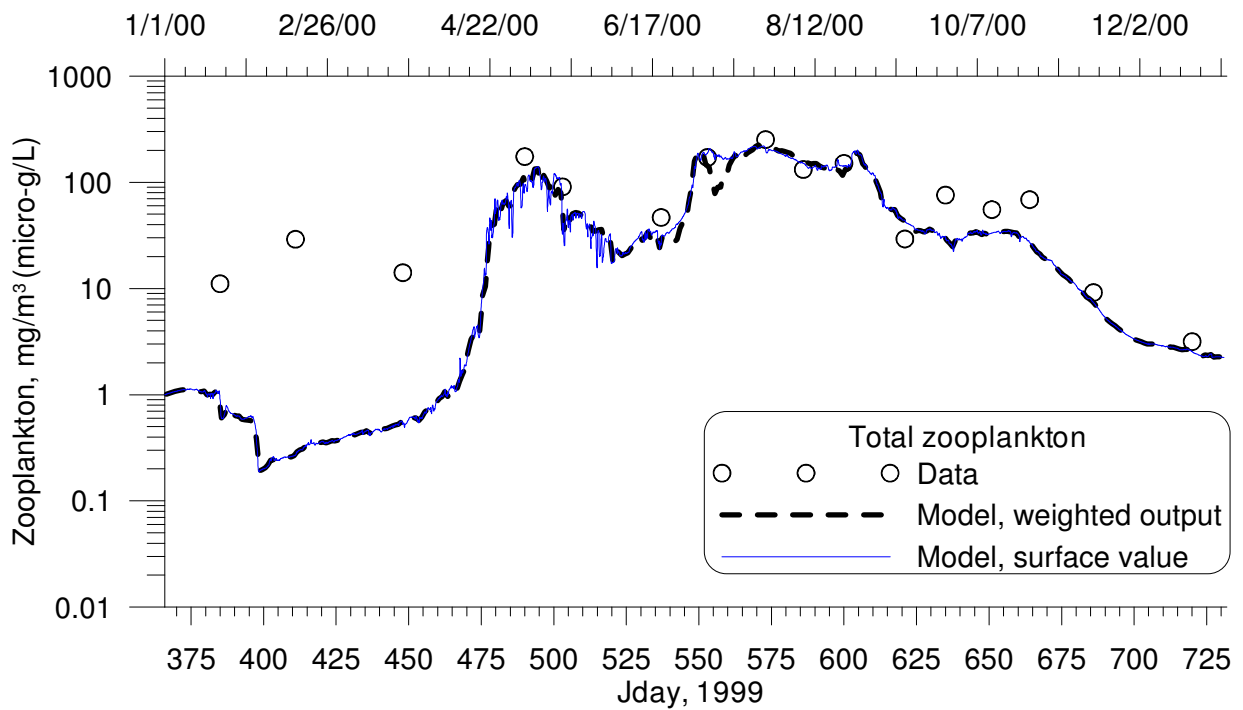


Figure 55. Model-data comparison of weighted total zooplankton, LRFEP station 8.0

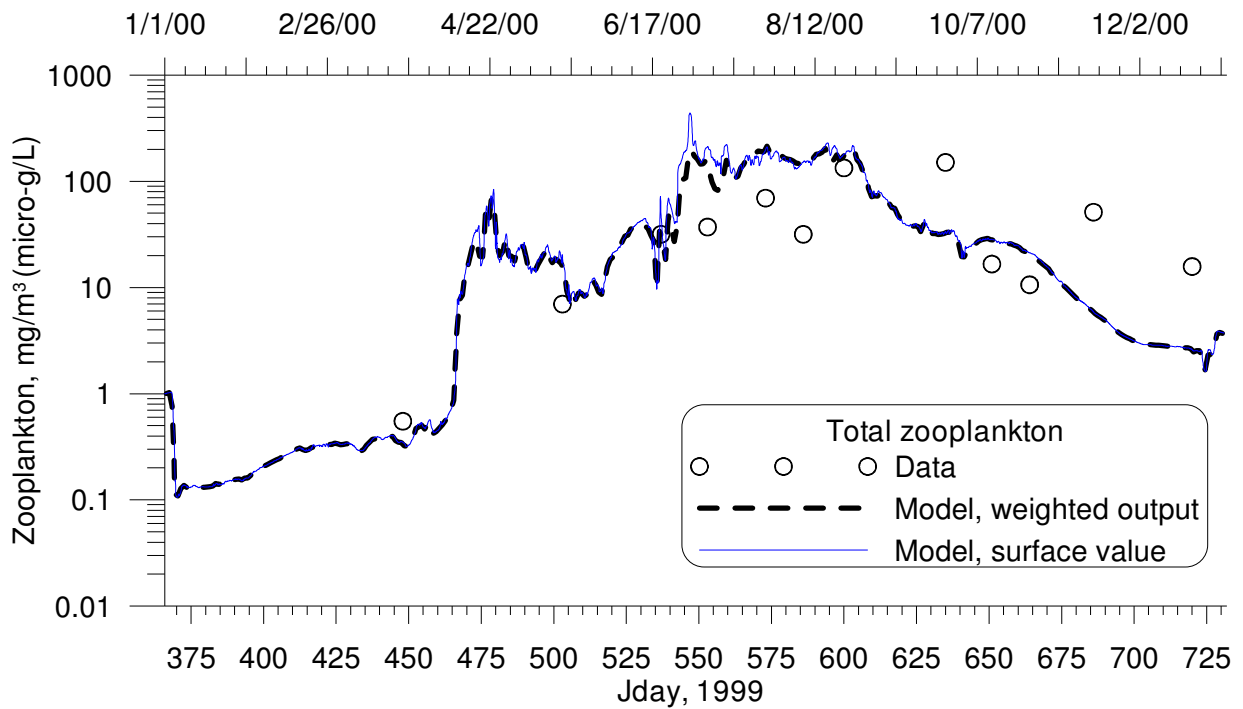


Figure 56. Model-data comparison of weighted total zooplankton, LRFEP station 8.5

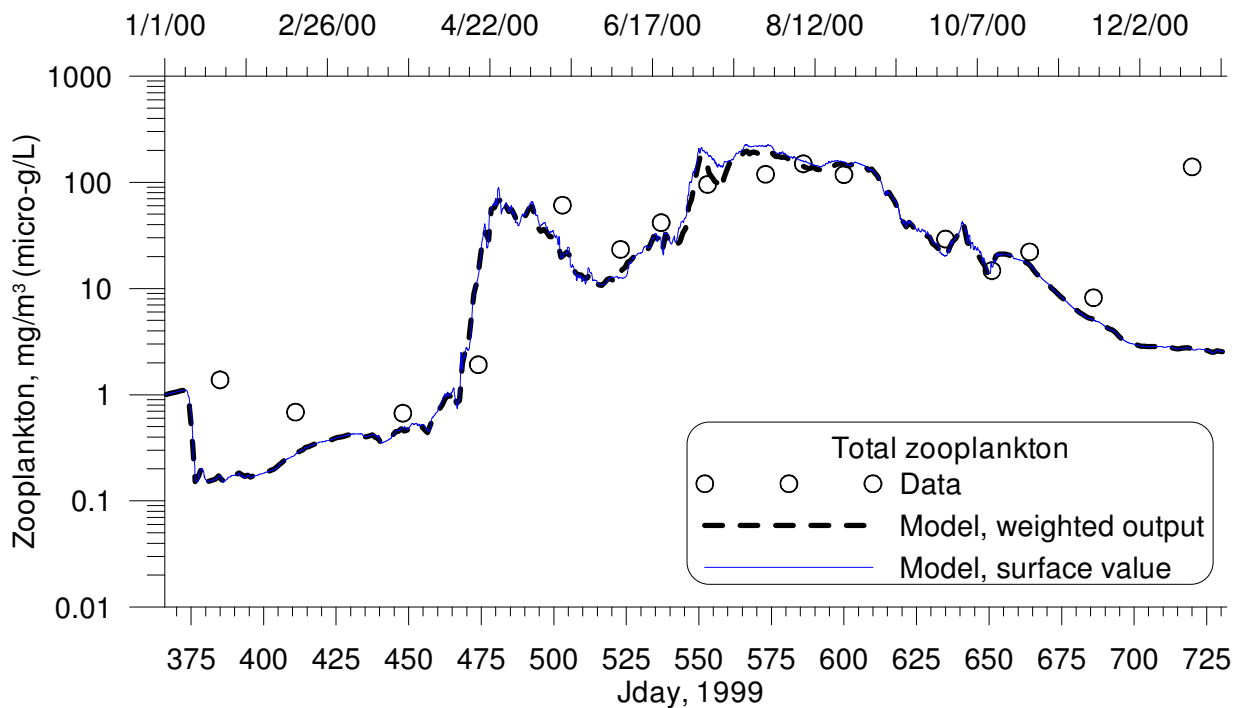


Figure 57. Model-data comparison of weighted total zooplankton, LRFEP station 9.0

Fish Bioenergetics (kokanee)

Results for a variety of cases are reported and show that the model is behaving appropriately, and that some diagnostic conclusions can be drawn.

A prescribed mass case, based on fish growth data, is examined at the Spring Canyon monitoring location (upstream of Grand Coulee Dam). The cell of best growth was selected at each time-step, and stomach content was passed on to the other vertical cells. At the end of the day (taken to be midnight), growth was determined and passed on. A similar case was examined at the Porcupine Bay monitoring location (on the Spokane River arm). The Spokane arm is known to have different fish habitat characteristics, and provides a contrast with the Columbia River site.

A simple vertical foraging strategy was employed to try to capture some realistic foraging behavior; the best growth strategy predicts a period of minimal consumption in the late summer and fall when prey is abundant. The decision process at each time-step is:

- 3) Use the cell of best growth if
 - a. growth is positive
 - b. during nighttime (selects least negative)
- 4) If daylight (surface lux >1), alternate foraging and “digesting”
 - a. Forage at the cell of greatest consumption, C
 - b. “Digest” or “rest” at the cell of minimum respiration, R
 - c. alternate foraging and digesting time-steps during daylight.

These cases show that fish growth can be reasonably modeled. They also predict that there are periods when foraging at some locations in the reservoir are not practical due to the large metabolic costs associated with warm water. Specifically, near the time of fall turnover, the smallest respiration costs at Spring Canyon exceed the best energy gain from consumption. This indicates that fish are unlikely to a) inhabit that part of the reservoir, and/or b) are utilizing a strategy not modeled. Possibilities include finding cold water inflows, or foraging under more successful techniques such as high prey density littoral regions. The comparison between the two sites suggests the possibility of horizontal migration. When the bioenergetics are poor at Spring Canyon in the late fall and winter, they are more favorable at Porcupine Bay.

The shown runs used a handling time of 0.5 sec. Runs with a 0.33 sec handling time do not greatly influence the results: the positive and negative growth periods are the same, but the magnitude is slightly improved.

The results of the fish bioenergetics model show general agreement with literature values. There is a little difference for a 10-g kokanee (see Figure 73 and subsequent comments and figure).

Base Lake Roosevelt Results at Spring Canyon (LRFEP sta 9.0)

Energy content of prey: a constant of 2420 J/g

Growth method: prescribed function based on data.

Feeding: model output prey densities are used.

Foraging: Best cell at each time-step; growth calculated and passed on to all cells in the segment daily.

Comments: During the warm water periods in roughly July through October, the best place to be for a single time-step is at the bottom of the lake. However, only negligible consumption is then possible. This is unlikely the actual fish behavior to these conditions. Compare with the next section.

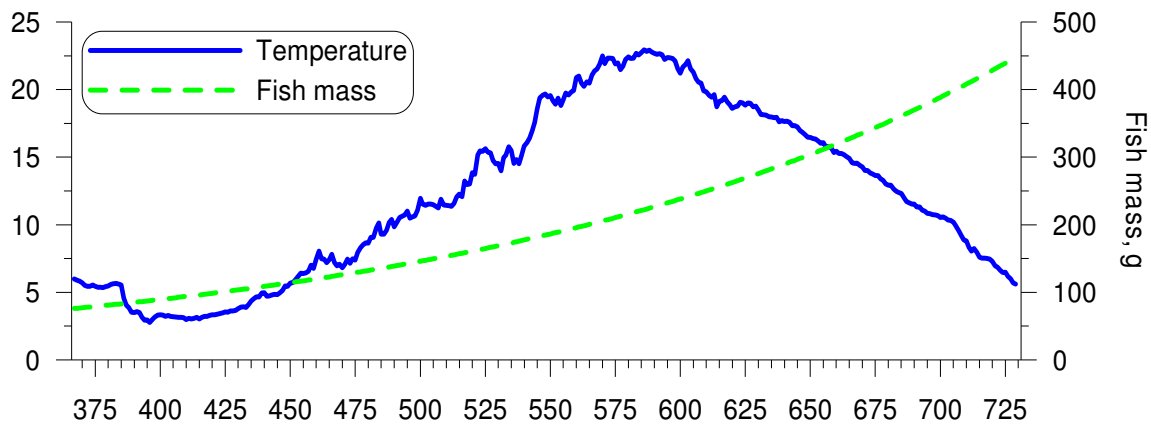


Figure 58. Base case temperature and prescribed fish mass function.

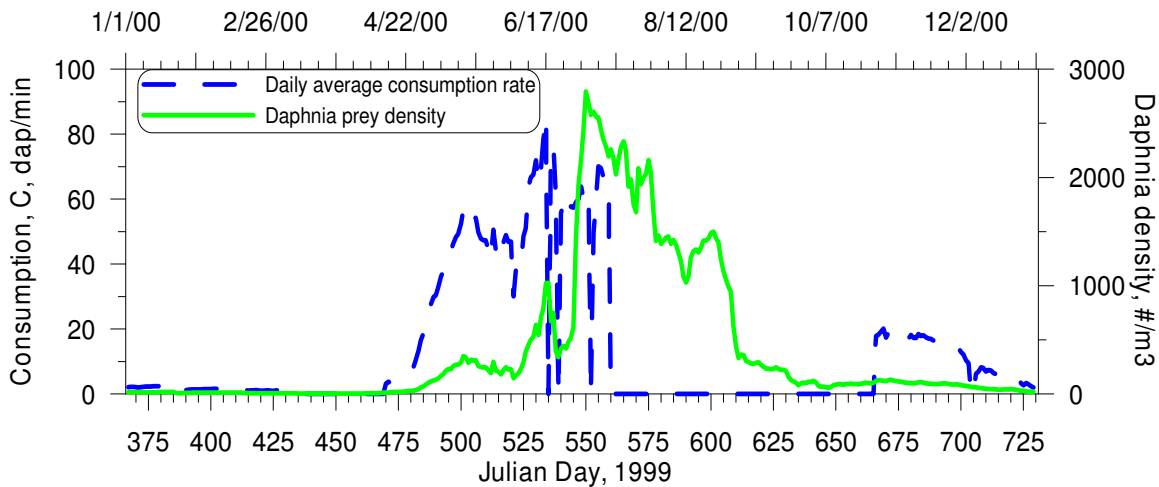


Figure 59. Daily average consumption (includes nighttime) and prey density.

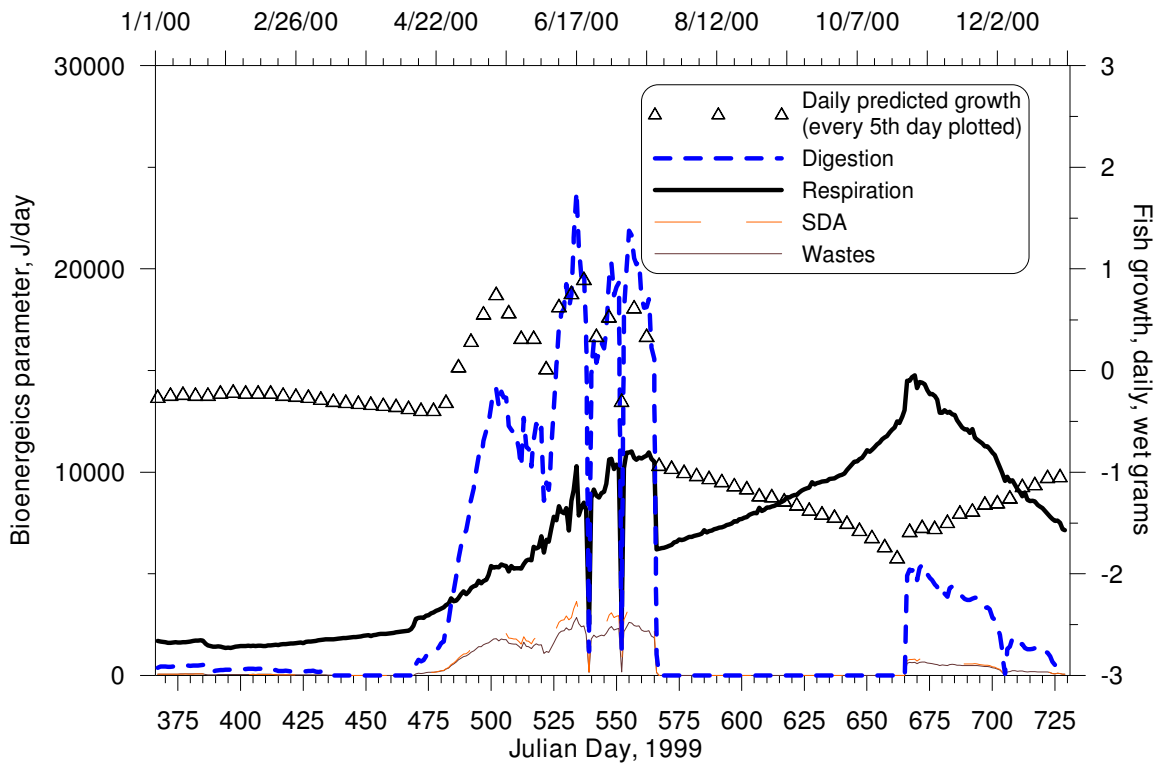


Figure 60. Daily growth and bioenergetic parameters.

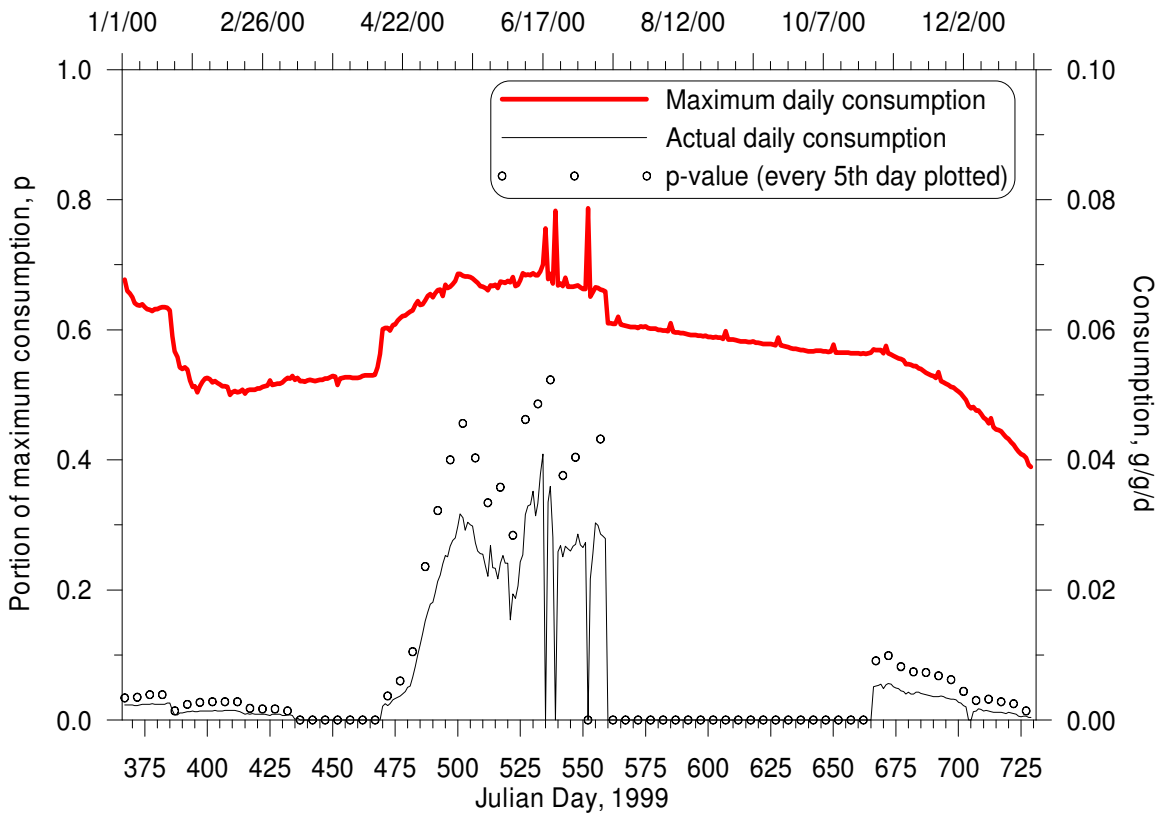


Figure 61. Daily maximum and actual consumption; p-values.

Base Lake Roosevelt Results at Spring Canyon (LRFEP sta 9.0);

Simple foraging algorithm

Energy content of prey: a constant of 2420 J/g

Growth method: prescribed function based on data.

Feeding: model output prey densities are used.

Foraging: “Best growth” with conditional vertical foraging during the day. The decision process at each time-step:

- 5) Use the cell of best growth if
 - a. growth is positive
 - b. during nighttime (selects least negative)
- 6) If daylight (surface lux >1), alternate foraging and “digesting”
 - a. Forage at the cell of greatest consumption, C
 - b. “Digest” or “rest” at the cell of minimum respiration, R
 - c. alternate foraging and digesting time-steps during daylight.

Comments: This is an attempt at simple vertical foraging in an attempt to optimize consumption and respiration. Also, when assessing which cell has the best growth, 20% of the consumption is directly added to the digestion parameter to allow for potential digestion (Brett & Groves, 1979). Without this estimate (only used to estimate the best cell, and not for any parameter calculations), the model will pick the most favorable growth locations using the stomach content at the start of the time-step. Since only a small portion of the prey consumed are digested in the same time-step, the model tends to pick the locations of least respiration cost when the waters become warm which in turn leads to empty stomachs.

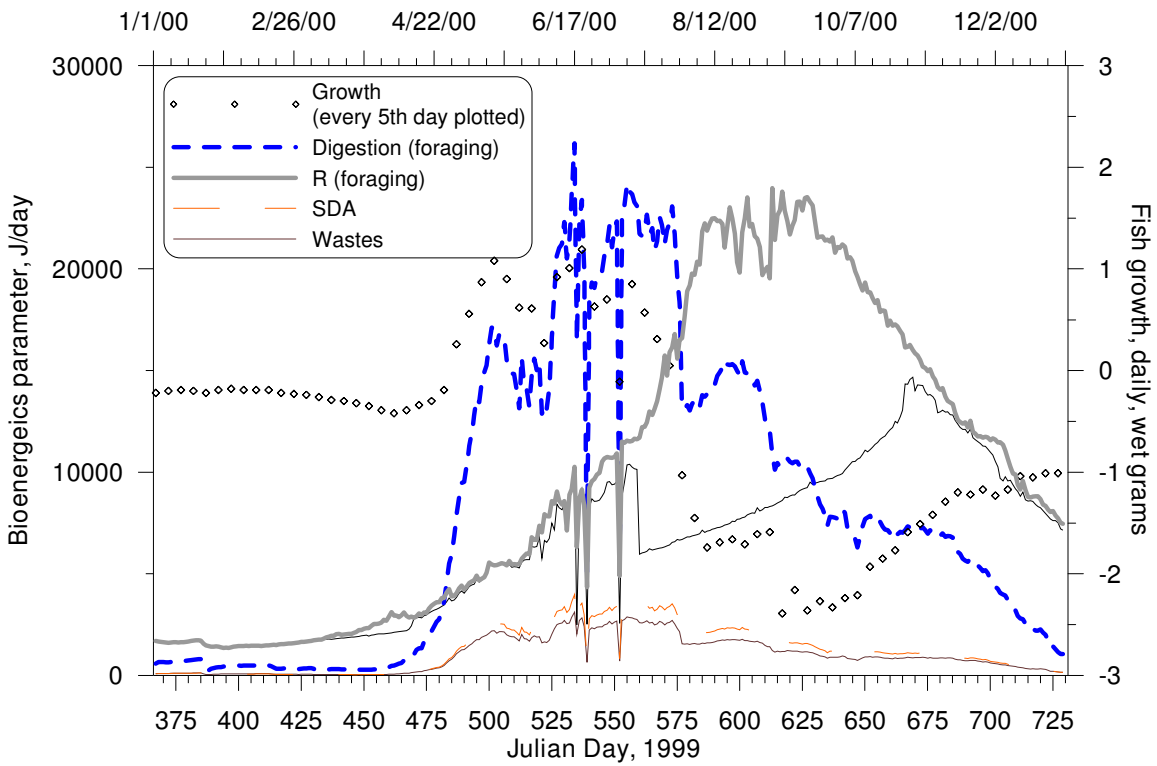


Figure 62. Daily growth and bioenergetic parameters, vertical foraging strategy.

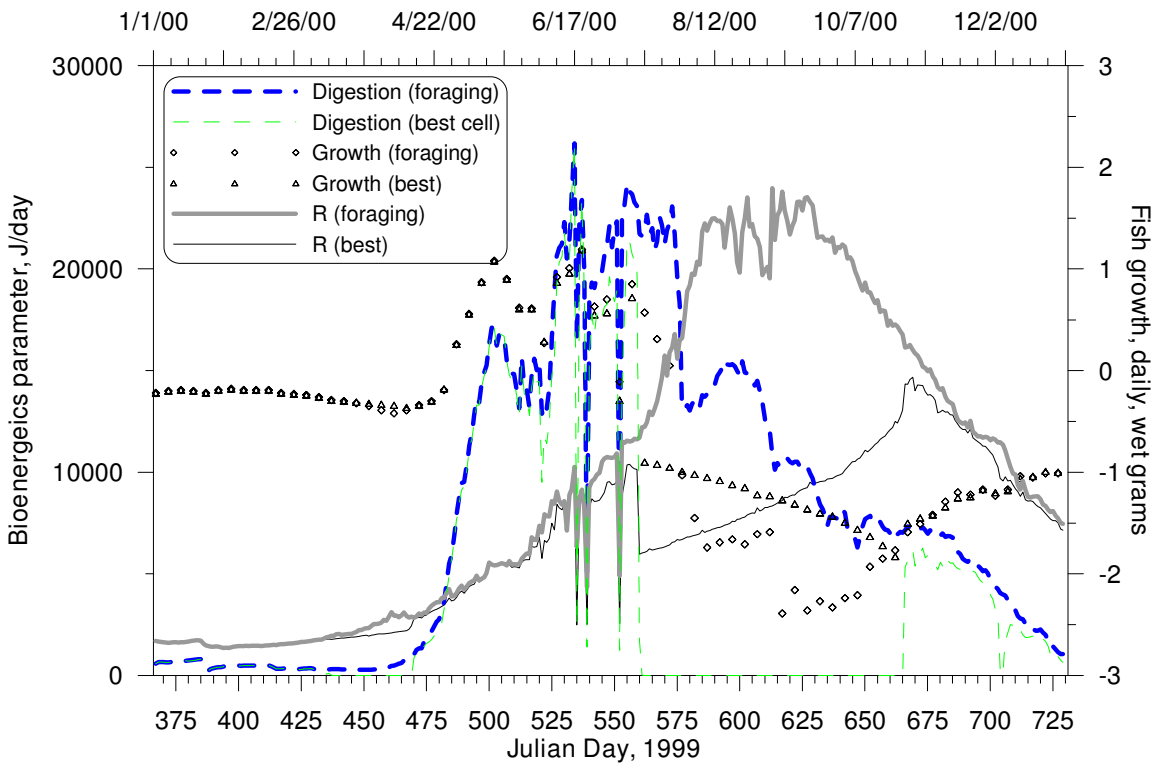


Figure 63. Comparison of fish location optimization strategies: best growth cell and vertical foraging.

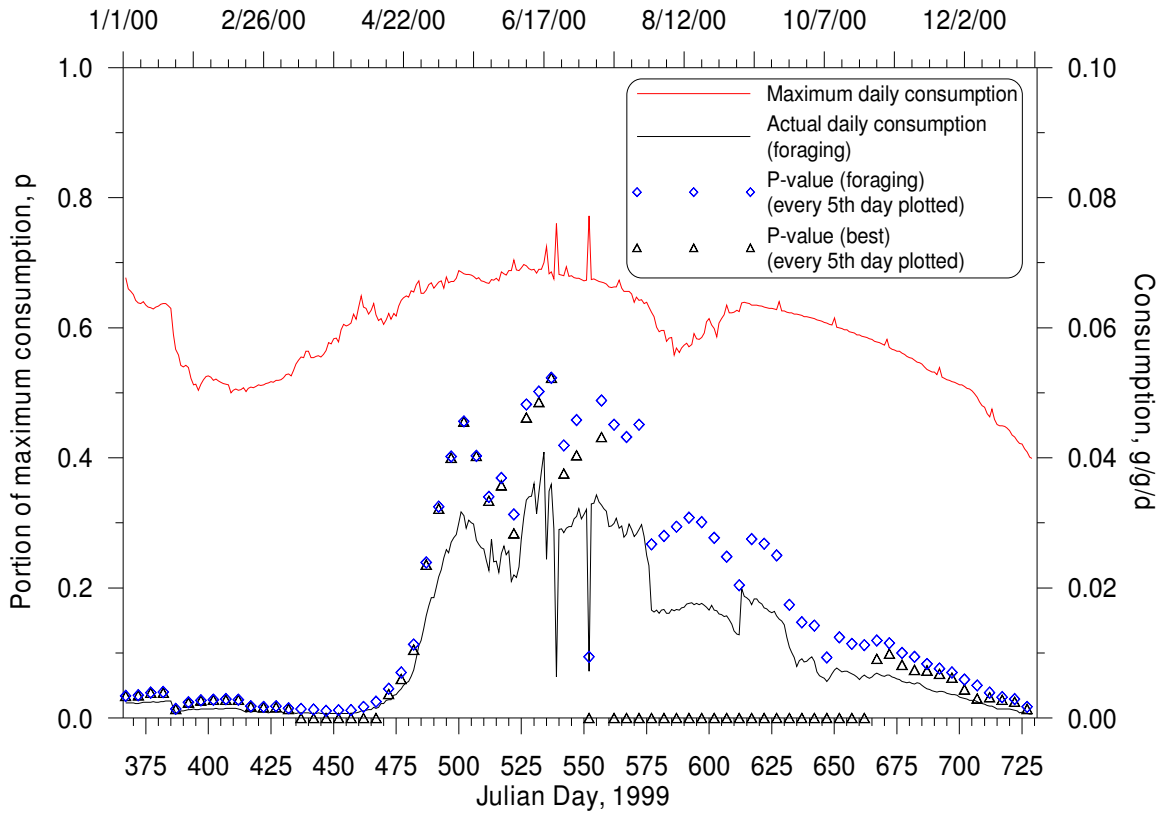


Figure 64. Daily maximum and actual consumption for the foraging model. Comparison of best growth cell and vertical foraging p-values.

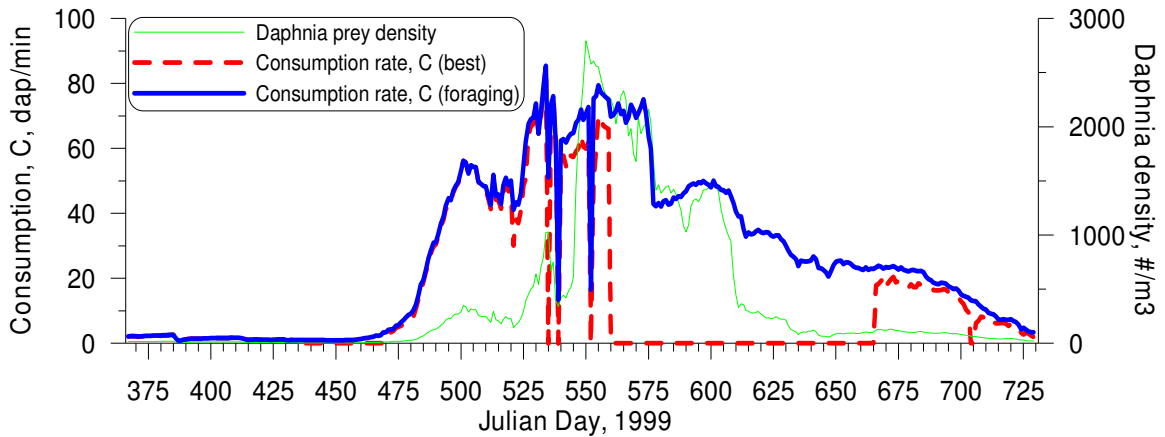


Figure 65. Comparison of daily average consumption rates.

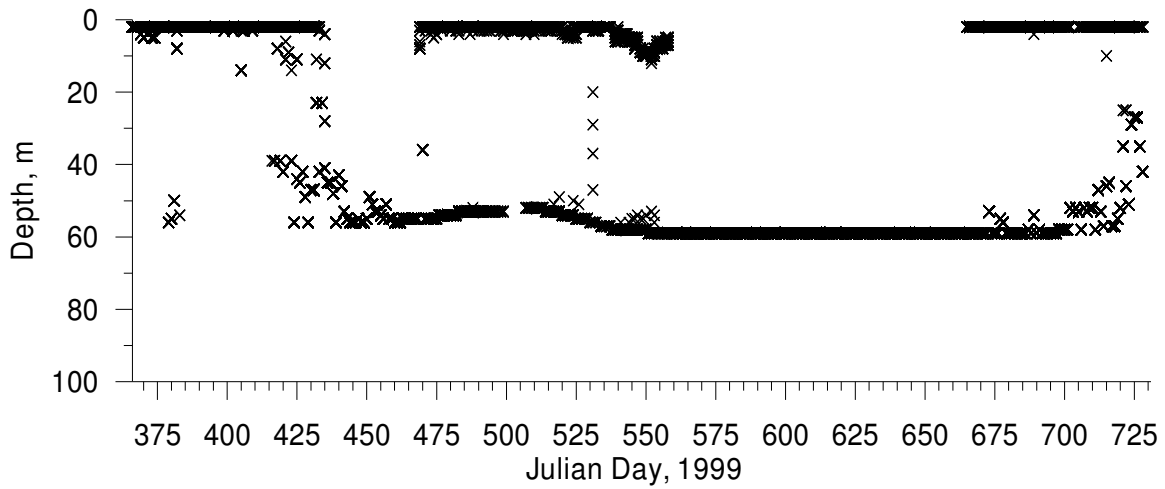


Figure 66. Foraging depths at each time-step, best growth cell method.

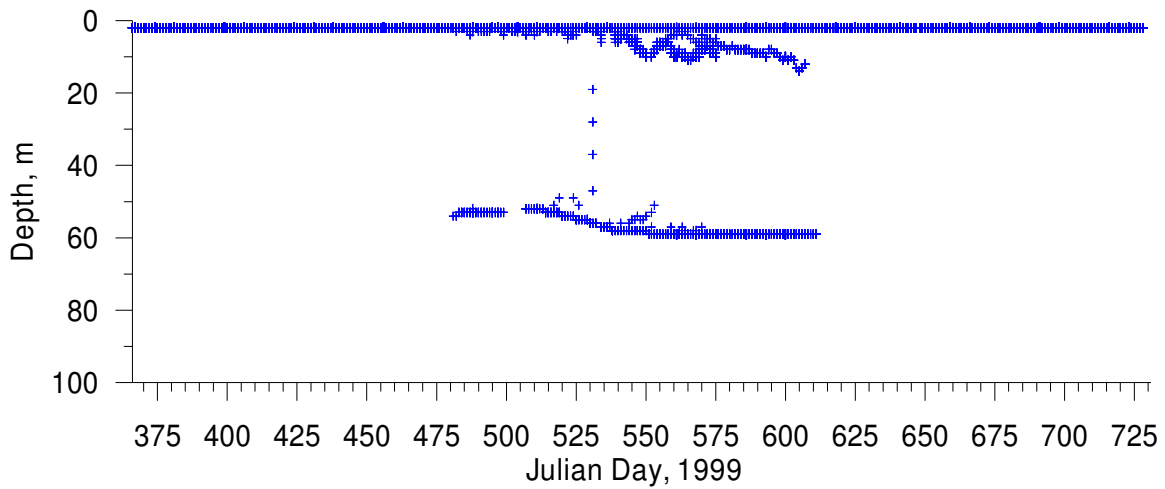


Figure 67. Foraging depths at each time-step, vertical foraging method.

[Figure 66 and Figure 67 were hard to distinguish on the same plot; individual points which look to be the same generally are.]

Comparison of Base Lake Roosevelt Results at Porcupine Bay (Spokane River, LRFEP sta 4.0) with Spring Canyon (Columbia River, LRFEP sta 9.0)

Energy content of prey: a constant of 2420 J/g

Growth method: prescribed function based on data.

Feeding: model output prey densities are used.

Foraging: “Best growth” with conditional vertical foraging during the day. The decision process at each time-step:

- 7) Use the cell of best growth if
 - a. growth is positive
 - b. during nighttime (selects least negative)
- 8) If daylight (surface lux >1), alternate foraging and “digesting”
 - a. Forage at the cell of greatest consumption, C
 - b. “Digest” or “rest” at the cell of minimum respiration, R
 - c. alternate foraging and digesting time-steps during daylight.

Comments: This is the same run as previously reported; a new location is added.

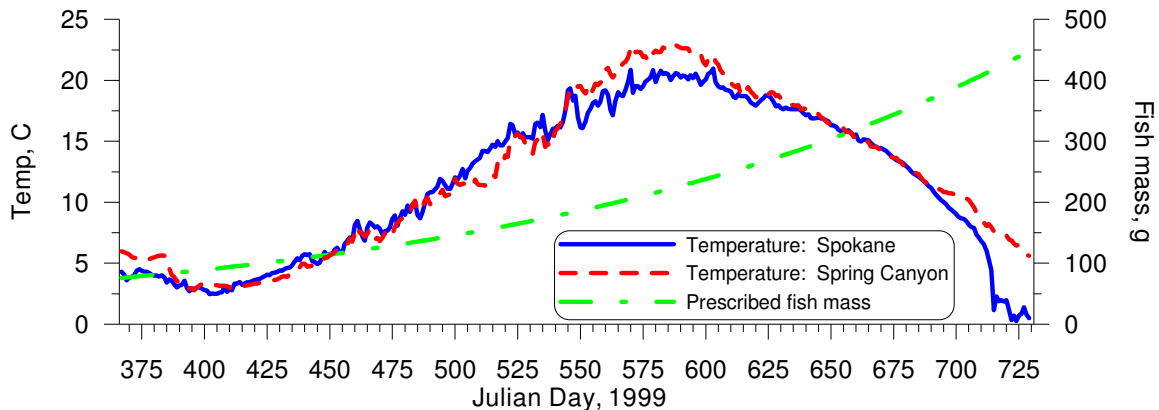


Figure 68. Comparison of water temperatures.

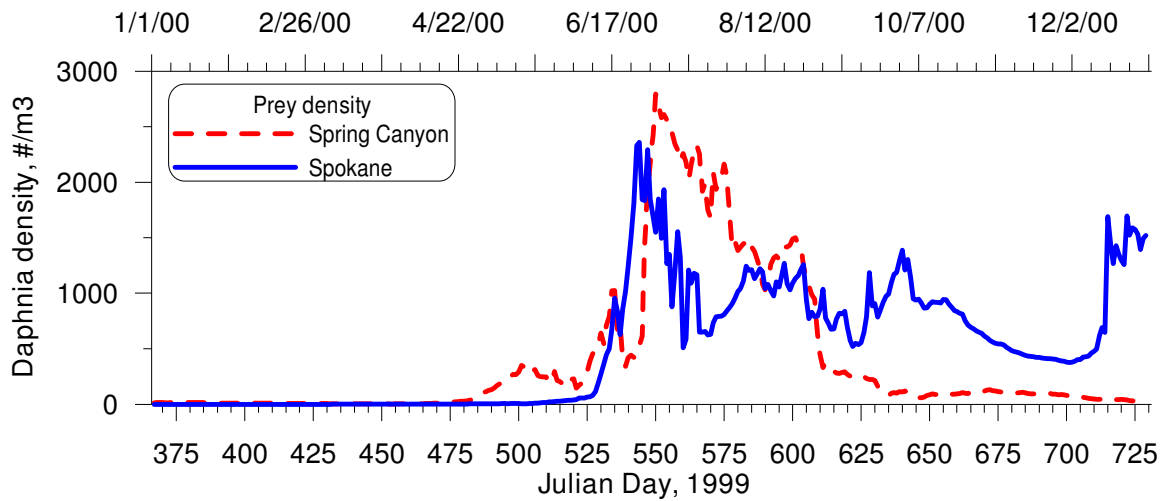


Figure 69. Comparison of prey densities.

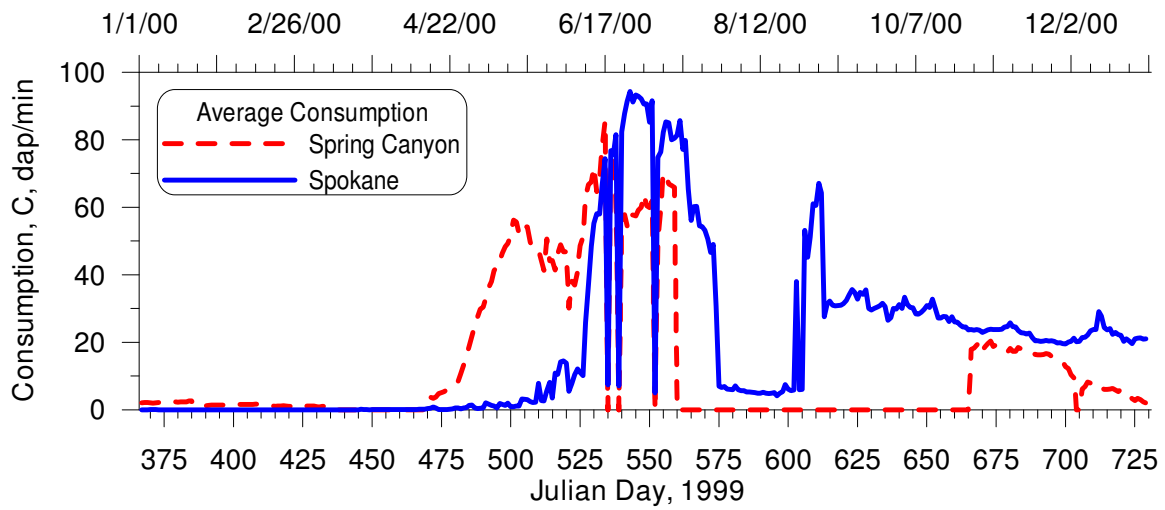


Figure 70. Comparison of consumption rates.

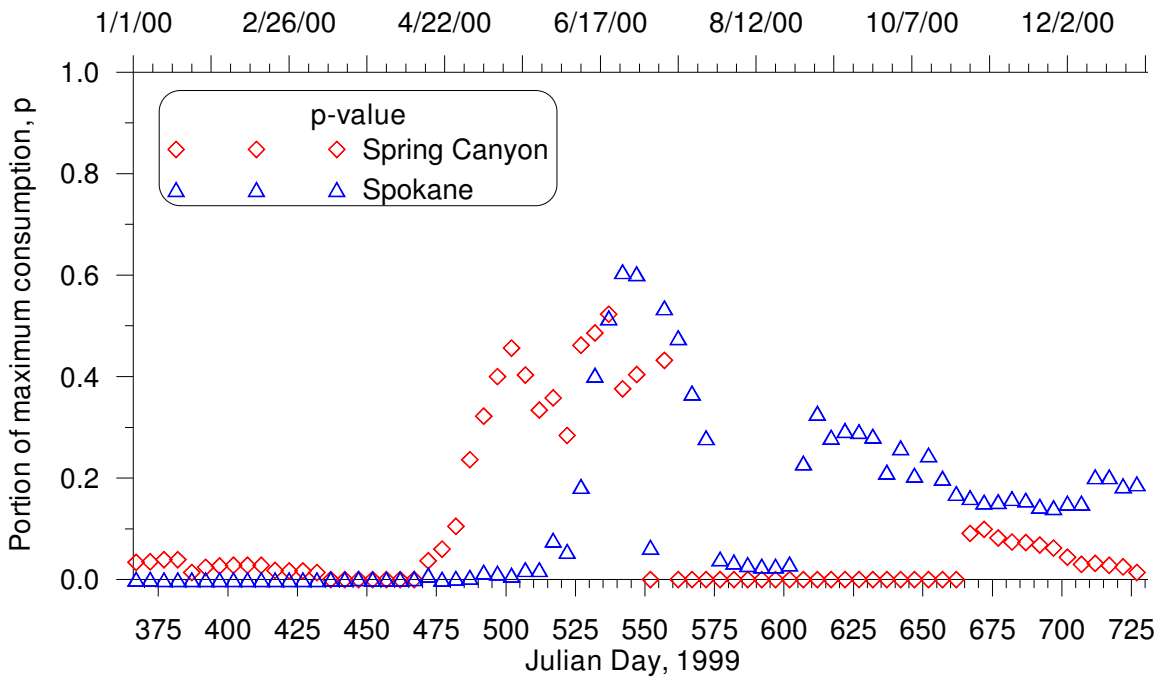


Figure 71. Comparison of p-values.

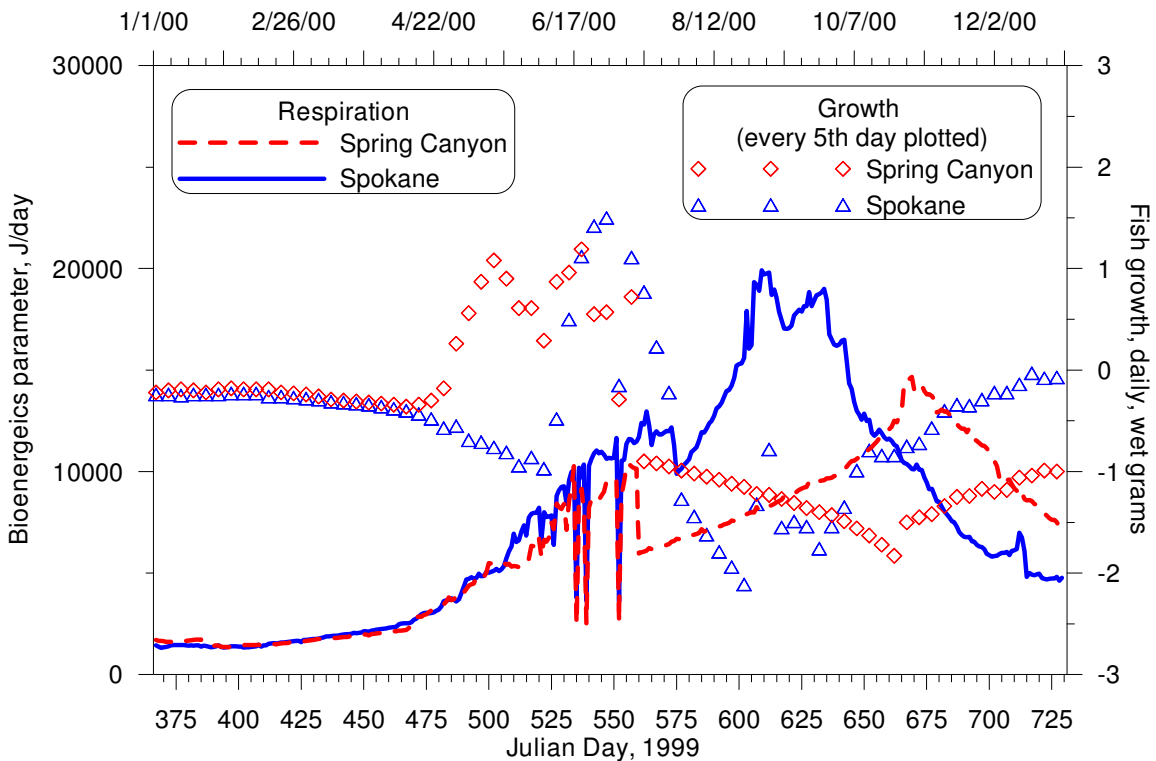


Figure 72. Comparison of respiration and daily growth.

Fixed mass kokanee, diagnostic run at 100% C_{max} .

Energy content of prey: a constant of 2800 J/g

Growth method: no growth: fixed masses of 10 and 100 grams.

Feeding: consumption rate set by C_{max} = function (T, mass)

Foraging: Best cell at each time-step.

Comments: Diagnostic run to compare model results to literature results.

C_{max} is formulated using Beauchamp, et al. (1989):

$$C_{max} = 0.303 \cdot M^{-0.275} \cdot TL(temp) \quad \text{in g/g/d}$$

$$C_{max} = 0.303 \cdot M^{-0.275} \cdot TL(temp) \cdot M \cdot E_{daphnia} \quad \text{in J/d}$$

Foraging rate (C, in daphnia/minute) was determined by dividing C_{max} by 1440; thus feeding occurred continuously and was only a function of temperature (mass was constant).

The allometric function reported by Hewitt & Johnson (1992) is used for the Kokanee energy density.

$$E_{fish} \left[\frac{J}{g} \right] = \begin{cases} 4.1868 \cdot (1.8510 \cdot M + 1250) & \text{for } M \leq 196 \text{ g} \\ 4.1868 \cdot (1.1254 \cdot M + 1588) & \text{for } M > 196 \text{ g} \end{cases}$$

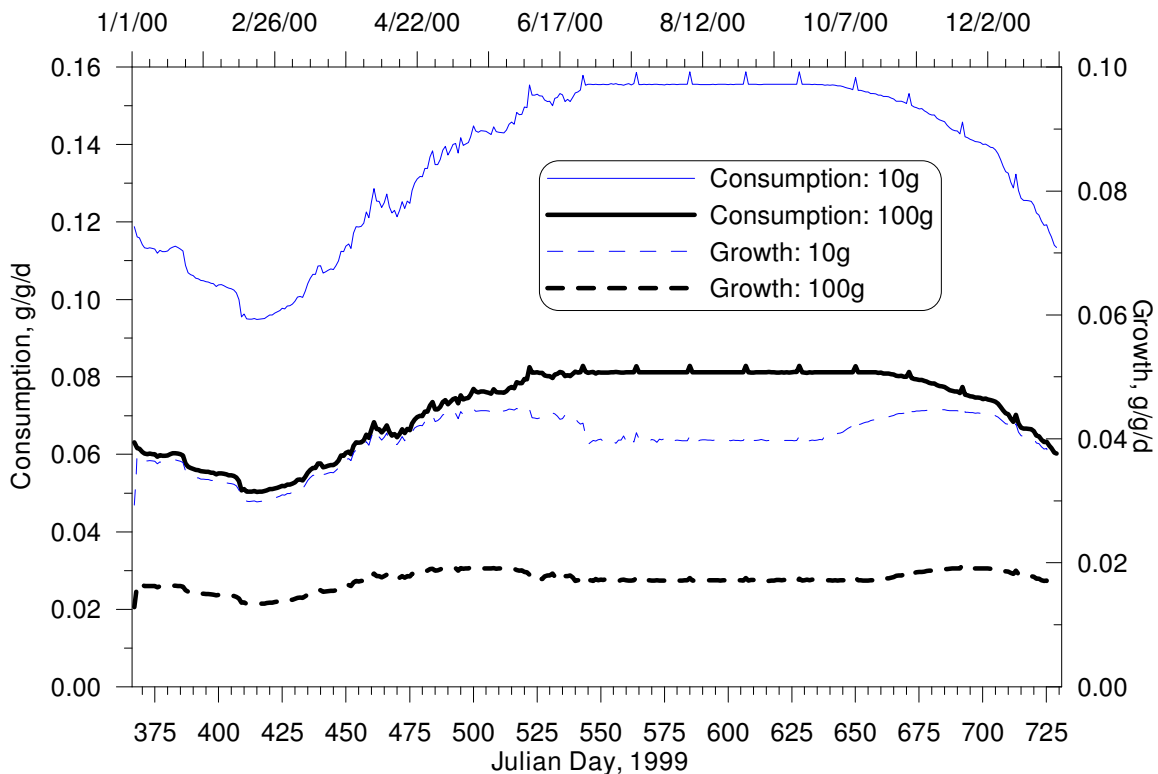


Figure 73. Growth and Consumption rate at C_{max} for a 10 g and 100g kokanee.

Summary

A CE-QUAL-W2, version 3.5, hydrodynamic, temperature, and water quality model was developed and calibrated for the Franklin D. Roosevelt Lake from Grand Coulee Dam to the U.S.- Canadian border. A fish bioenergetics model based on Stockwell and Johnson (1999) was developed. The water quality model output is used by the fish model to estimate fish growth potential within the system.

A summary of the calibration statistics for the year 2000 are presented in Table 9. A discussion of the calibration follows.

Table 9. Calibration statistics summary, 2000.

| Parameter/Constituent | Unit | Count | ME* | AME* | RMS* |
|--|----------------------------|-------|---------|---------|---------|
| Grand Coulee Dam Forebay | | | | | |
| Daily-average ELWS** | m | 366 | 0.01 | 0.01 | 0.01 |
| Hourly-average ELWS** | m | 8762 | -0.03 | 0.17 | 0.17 |
| LRFEP vertical profile stations | | | | | |
| Total Dissolved Solids | mg/L | 1308 | -0.18 | 0.39 | 0.44 |
| Dissolved Oxygen | mg/L | 1308 | -0.01 | 0.14 | 0.16 |
| pH | - | 1308 | -0.42 | 0.43 | 0.43 |
| Water Temperature | °C | 1308 | -0.14 | 0.49 | 0.50 |
| Below Grand Coulee Dam | | | | | |
| Orthophosphate | mg/L-P | 25 | 0.00072 | 0.00097 | 0.00097 |
| Ammonium | mg/L-N | 25 | -0.005 | 0.0067 | 0.0067 |
| Nitrate plus nitrite | mg/L-N | 25 | -0.017 | 0.036 | 0.036 |
| Alkalinity | mg/L- CaCO ₃ | 25 | 0.89 | 4.18 | 4.18 |

* ME=mean error, AME=absolute mean error, RMS=root mean square error, see Appendix G.

** ELWS = Elevation, Water Surface.

Hydrodynamics

The hydrodynamic calibration to the 2000 data is acceptable. The daily-average water surface elevation is within 1 cm, and the hourly-average elevation is within 17 cm with little bias. The water balance flows have a minimal bias and the magnitudes are within the range of river discharge measurement error. While the water surface elevation calibration for 2001 and 2002 appears good, the water balances for both years show a strong bias towards losing water from the river. A further investigation into the cause of this apparent bias is recommended prior to using the model over these time-frames.

Water Temperature

The most comprehensive and representative temperature data are from the LRFEP vertical profiles. The calibrated model shows a slight bias toward underpredicting temperature (by ~0.1 °C), which is acceptable. The RMS error, representing the typical difference in temperature

between the model and data, is 0.5 °C, which is considered a good calibration. The calibration of the station data in the Columbia River is good, but there is greater error in the calibration of the Spokane River data.

Alkalinity and pH

While the alkalinity calibration is acceptable, the pH calibration needs further work. There is an apparent data conflict between the USGS/Environment Canada data and the LRFEP data. Additionally, the high pH values seen in the LRFEP data in the fall need to be explained given the low carbonate and algal activity of the system. This conflict is largely the result of a faulty probe.

Ammonium and Nitrate plus Nitrite

The nitrogen balance in the system is largely conservative, so the calibration is straightforward. As with the bulk of the parameters, while the Columbia River calibration is good, the calibration at the Spokane River station shows more disagreement.

Orthophosphate

While the orthophosphate calibration is acceptable, it could be improved. A further investigation into the spring spike in phosphorous is warranted. Speculation suggests that attached algae may be a significant player in the spring phosphorous balance prior to the spring phytoplankton bloom. The phosphorous balance is relatively conservative otherwise. Phosphorous concentrations are also very low (near the detection limit), so sampling error may be significant. Within the W2 model, phosphorous calibration included including sorption onto sediments and adjusting the organic matter stoichiometry. An organic matter stoichiometry study may provide some insight into the nutrient loadings of the system.

Total Dissolved Solids

The total dissolved solids calibration is acceptable. TDS is largely conservative.

Algae

The algae calibration is acceptable in terms of total algae available for zooplankton consumption. The spring diatom bloom is generally captured. A general problem occurs in matching the dieback in June/July seen over most of the system, which is presumably due to either the hydrodynamic change in reservoir operation or in the spring zooplankton blooms. The algal populations within Lake Roosevelt experience significant washout and have low nutrient availability. While most stations have a good total algae calibration, the Hawk Creek station underpredicts total algae. W2 is largely a pelagic model, so a further consideration of the littoral primary production may be significant in terms of zooplankton, and hence fish, ecosystems.

Zooplankton

Zooplankton sampling presents many challenges, as does zooplankton modeling. In general, zooplankton concentrations range over 3 to 4 orders of magnitude over a year. The calibration concentration is typically within 1 order of magnitude, and much of the discrepancy occurs during the late winter and early spring. Fortunately, the fish growth potential model is not sensitive to zooplankton prey densities within 1 order of magnitude. Improvements to the zooplankton calibration would incorporate suggestions from biologists as to the important changes in driving factors during the seasonal shifts from winter to spring to summer and an improved understanding of the effects of zooplankton washout and retardation.

Fish Growth Potential

The fish growth potential suggests that growth is limited in the winter by prey availability and in the summer by warm temperatures relative to the available prey. The Spokane River arm demonstrates higher growth potential than the lower Columbia River reaches. This is consistent with anecdotal data and prevailing opinions amongst the fisheries biologists. Further improvements to the fish bioenergetics model would include identification of potential cold water refugia and investigation into the actual fish movement behavior such as diel migration/foraging, spatial distribution in terms of location and depths, and seasonal changes in that distribution. It is important to note that the fish growth potential does not consider entrainment, predation, nor spawning—all of which are not well understood within the Lake Roosevelt system for kokanee salmon.

Recommendations for further study

1. Spokane River arm. As previously recognized by the LRFEP, the Spokane River exhibits complicated hydrodynamics. Changes in Spokane River flow influence the river arm in terms of plunge point and stratification, and the mainstem Columbia River influences much of the lacustrine portion of the Spokane River arm. Given the biological importance of the Spokane River arm to kokanee, three programs would greatly benefit diagnostic understanding of the arm. The first would be to improve the boundary condition data, especially nutrients, flow, and temperature. The second would be to add a second monitoring station between the existing station 4.0, which is about at the riverine/lacustrine transition point, and the confluence with the Columbia River mainstem. An independent model of the Spokane River arm may provide insights into the relationships amongst the Columbia River, the boundary conditions, and the arm hydrodynamics. The third program would include a study of fish seasonal distribution within the entire system as well as shifts in diel vertical distributions. A comparison of kokanee behavior within the Spokane River arm with their behavior in the lower third of the Columbia River might suggest important behavioral factors.
2. Model Improvements.
 - a. The model could be calibrated to new data taken recently to extend the calibration period.

- b. The model could be assessed for decreasing the computational time to run a simulation. Currently, it takes a couple days of running the model to do full water quality, fish bioenergetics and hydrodynamics for a full-year for the entire system.
- 3. Enhancements. The following enhancements could be added to the model to improve its utility as a management tool.
 - a. The model could be tied into a chemical equilibrium model for metals to evaluate the effect of metals on fish uptake. Fish growth and metals accumulation could be tracked.
 - b. The fish energetics model could be coupled with a smart particle tracking algorithm such as the Numerical Fish Surrogate used by the Corps of Engineers to look at fish movement. The current model evaluated fish either at fixed locations (in a cage) or at a fixed segment but allowing vertical migration. There was no consideration of horizontal movement considering both transport by the fluid and volitional movement of the fish. With this capability, one could release fish, let's say from a hatchery, and have the model track their movement through the lake and their growth.

References

- Beauchamp, D. A.; Stewart, D. J.; and Thomas, G. L. (1989) "Corroboration of a bioenergetics model for sockeye salmon." *Transactions of the American Fisheries Society*, 118: 597–607.
- Bevelhimer, M. S.; and Adams, S. M. (1993) "A bioenergetic analysis of diel vertical migration by kokanee salmon, *Oncorhynchus nerka*." *Can. J. Fish. Aquat. Sci.* 50: 2336–2349.
- Brett, J. R. (1971) "Satiation time, appetite, and maximum food intake of sockeye salmon (*Oncorhynchus nerka*)." *J. Fish. Res. Board Can.* 28: 409-415.
- Brett, J. R.; and Groves, T. D. D. (1979) "Physiological energetics." Pages 279-353 in W. S. Hoar, D. J. Randall, and J. R. Brett, editors. *Fish physiology*, volume 8. Academic Press, New York.
- Butterwick, C.; Heaney, S.I.; and Talling, J.F. (2005) "Diversity in the influence of temperature on the growth rates of freshwater algae, and its ecological relevance." *Freshwater Biology*, 50, pp. 291-300.
- Cole, T. M. and Wells, S. A. (2006) "CE-QUAL-W2: A Two-Dimensional, Laterally Averaged, Hydrodynamic and Water Quality Model, v. 3.5." Instruction Report EL-06-1.
- Cole, T. M. and Wells, S. A. (2004) "CE-QUAL-W2: A Two-Dimensional, Laterally Averaged, Hydrodynamic and Water Quality Model, v. 3.2." Instruction Report EL-03-1.
- Downing, J. A.; and Rigler, F. H. (1984) "A Manual on methods for the assessment of secondary productivity in fresh waters." Blackwell Scientific, 2nd. ed.
- Eggers, D. M. (1977) "The nature of prey selection by planktivorous fish." *Ecology*, 58: 46-59.
- Hewett, S. W.; and Johnson, B. L. (1987) "A generalized bioenergetics model of fish growth for microcomputers." University of Wisconsin, Sea Grant Technical Report WIS-SG-87-245, Madison.
- Link, J.; and Edsall, T. A. (1996) "The effect of light on Lake Herring (*Coregonus artedii*) reactive volume." *Hydrobiologia* 332: 131-140.
- McKillip, M. L.; Annear, R. A.; and Wells, S. W. (2006) "Lake Roosevelt Model: Boundary Conditions and Set-up." Technical Report EWR-01-05, Department of Civil and Environmental Engineering, Portland State University. Portland, Oregon.
- Reynolds, C. S. (1984) *The ecology of freshwater phytoplankton*. Cambridge University Press. New York, New York.

Stockwell, J. D.; and Johnson, B. M. (1999) "Field evaluation of a biogenetics-based foraging model for kokanee (*Oncorhynchus nerka*)." Canadian Journal of Fisheries and Aquatic Science, 56 (suppl. 1), pp.140-151.

Stockwell, J. D.; and Johnson, B. M.. (1997) "Refinement and calibration of a bio-energetics-based foraging model for kokanee (*Oncorhynchus nerka*)." Canadian Journal of Fisheries and Aquatic Science, 54, pp.2659-2676.

Thornton, K. W.; and Lessem, A .S. (1978) "A temperature algorithm for modifying biological rates." Trans. Am. Fish. Soc. 107: 284–287.

Appendix A: Water quality boundary condition generation

Overview

The following water quality constituents are used as model inputs at the upstream boundary condition for each river.

- Alkalinity
- pH
- Dissolved oxygen
- Total dissolved solids (TDS)
- Inorganic suspended solids (ISS)
- Orthophosphate or soluble reactive phosphorous (ORP or PO₄)
- Ammonium (NH₃ or NH₄)
- Nitrate plus nitrite (NO_x)
- Algae
- Organic matter
 - Labile dissolved organic matter (LDOM)
 - Refractory dissolved organic matter (RDOM)
 - Labile particulate organic matter (LPOM)
 - Refractory particulate organic matter (RPOM)
- Total inorganic carbon (TIC)
- Generic constituents
 - Water age
 - Tracer
 - Coliform bacteria

Additional data are used to generate the model inputs:

- Total suspended solids (TSS)
- Conductivity

The water quality data are typically reported biweekly or monthly and each constituent is not always reported at the same time as any other constituent. The W2 model, however, updates all water quality boundary conditions at the same time. To resolve the differences in data reporting dates and times and reporting frequencies, the general approach is to build a daily-frequency time-series data set. Thus, each day of the model simulation will have one updated water quality boundary condition point. For each constituent, the method used to generate the interpolated or estimated values (some boundaries lack data) can differ among the boundaries. Several common methods are linear interpolation, flow weighting, constant value, and using another boundary's values. Each is discussed briefly.

Linear interpolation

The constituent data are converted into a decimal day format to allow for ease of calculation. Linear interpolation is used to generate values at 12:00 pm for each day.

Flow weighting

Several of the constituents at the upstream boundary of the mainstem Columbia River are only available prior to the confluence of the Columbia and Pend Oreille Rivers. To generate the daily time-series boundary condition values, daily-average flow values and daily water quality constituents are determined. Linear interpolation is used to determine the water quality values and to fill in any data gaps in the daily-average flow. The mainstem Columbia River upstream water quality constituent boundary condition the flow weighted average of the branch constituent values.

Constant value

For water quality constituent boundary condition values where the data are sparse or lacking and the model is expected to be insensitive to the values, a constant value may be assumed. The value is selected as being a representative annual average value. An examples is the coliform bacteria concentrations at the Banks Lake return flows which are small (<10% of the mainstem flow) and infrequent (8 days in 2000.)

Another boundary site's values

Some boundaries lack data for many constituents, the Sanpoil River, for example. An initial estimation of the constituent values may be taken from the other boundaries. Linear regressions often do not show a strong correlation between boundaries and constituents. The advantage of using a different boundary's values over a constant value or a estimated value from a poorly correlated relationship is that the seasonal pattern is likely to be preserved in terms of magnitude and variation. Calibration may show the need to further improve the boundary conditions using this approach.

Conductivity

Conductivity is not a model input. The conductivity data are used to generate the total dissolved solids boundary condition (see the section Total dissolved solids.) For each boundary, the data frequency, data source, period of record, and method used to generate daily values is shown in Table 10.

Table 10. Conductivity boundary condition generation summary.

| Boundary | Data frequency | Data source | Data period | Method |
|---|-----------------------|--------------------|--------------------|--|
| <u>Mainstem Columbia River</u> | * | * | * | Daily flow weighted average value from Columbia R. and Pend Oreille R. values. |
| <i>Columbia prior to confluence</i> | biweekly | BC08NE0001 | JAN99 to DEC04 | Linear interpolation between data points |
| <i>Pend Oreille prior to confluence</i> | biweekly | BC08NE0029 | JAN99 to DEC04 | Linear interpolation between data points |
| | | | | |
| <u>Kettle River</u> | biweekly | BC08NN0021 | JAN99 to SEP04 | Linear interpolation between data points |
| <u>Colville River</u> | -- | none | -- | Use mainstem Columbia River values |
| <u>Spokane River</u> | monthly | USGS 12433000 | JAN99 to SEP03 | Linear interpolation between data points |
| <u>Sanpoil River</u> | -- | none | -- | Use mainstem Columbia River values |
| <u>Banks Lake return flow</u> | monthly | USGS 12436500 | OCT01 to SEP04 | Average value of entire data set; constant value |

Coliform bacteria

Coliform bacteria are included in the model for completeness. For each boundary, the data frequency, data source, period of record, and method used to generate daily values is shown in Table 11.

Table 11. Coliform bacteria boundary condition generation summary.

| Boundary | Data frequency | Data source | Data period | Method |
|---|-----------------------|--------------------|-----------------------------------|--|
| <u>Mainstem Columbia River</u> | * | * | * | Daily flow weighted average value from Columbia R. and Pend Oreille R. values. |
| <i>Columbia prior to confluence</i> | biweekly | BC08NE0001 | JAN99 to DEC04 | Linear interpolation between data points |
| <i>Pend Oreille prior to confluence</i> | biweekly | BC08NE0029 | FEB00 to DEC04 | Linear interpolation between data points; repeat first value to prior to start of data set |
| | | | | |
| <u>Kettle River</u> | monthly | USGS 12404900 | JAN99 to SEP04 | Linear interpolation between data points |
| <u>Colville River</u> | monthly | USGS 12409000 | OCT94 to SEP95; OCT99 to SEP00 | Averaged values by month, repeated for all years; linear interpolation between calculated points |
| <u>Spokane River</u> | -- | none | -- | Arbitrary value of 1.0; constant value |
| <u>Sanpoil River</u> | -- | none | -- | Arbitrary value of 1.0; constant value |
| <u>Banks Lake return flow</u> | monthly | USGS 12436500 | JAN99 to SEP04 | Average value of entire data set; constant value |

Alkalinity

Alkalinity is an important boundary condition used to model the carbonate chemistry. For each boundary, the data frequency, data source, period of record, and method used to generate daily values is shown in Table 12.

Table 12. Alkalinity boundary condition generation summary.

| Boundary | Data frequency | Data source | Data period | Method |
|---|-----------------------|--------------------|--------------------|--|
| <u>Mainstem Columbia River</u> | * | * | * | Daily flow weighted average value from Columbia R. and Pend Oreille R. values. |
| <i>Columbia prior to confluence</i> | biweekly | BC08NE0001 | JAN99 to DEC04 | Linear interpolation between data points |
| <i>Pend Oreille prior to confluence</i> | biweekly | BC08NE0029 | JAN99 to DEC04 | Linear interpolation between data points |
| | | | | |
| <u>Kettle River</u> | -- | none | -- | Use mainstem Columbia River values |
| <u>Colville River</u> | -- | none | -- | Use mainstem Columbia River values |
| <u>Spokane River</u> | monthly | USGS 12433000 | MAY99 to APR00 | Average value of entire data set; constant value |
| <u>Sanpoil River</u> | -- | none | -- | Use mainstem Columbia River values |
| <u>Banks Lake return flow</u> | -- | none | -- | Use mainstem Columbia River values |

Dissolved oxygen

Dissolved oxygen is an important water quality constituent to abiotic and biological system. For each boundary, the data frequency, data source, period of record, and method used to generate daily values is shown in Table 13.

Table 13. Dissolved oxygen boundary condition generation summary.

| Boundary | Data frequency | Data source | Data period | Method |
|--------------------------------|-----------------------|-----------------------------------|-----------------------------------|---|
| <u>Mainstem Columbia River</u> | monthly | LRFEP sta 0.0 | JAN99 to JAN02 | Linear interpolation between data points |
| <u>Kettle River</u> | monthly | USGS 12404900 | JAN99 to SEP04 | Linear interpolation between data points |
| <u>Colville River</u> | monthly | USGS 12409000 | OCT94 to SEP95; OCT99 to SEP00 | Use DSBC; the annual variation at GCL is similar to the limited Colville River data variation |
| <u>Spokane River</u> | hourly | AvistaCorp. (Little Falls Dam) | APR99 to MAR02 | Daily average values |
| <u>Sanpoil River</u> | monthly | LRFEP sta 8.8 | JAN99 to JAN02 | Linear interpolation between data points |
| <u>Banks Lake return flow</u> | monthly | USGS 12436500 | JAN99 to SEP04 | Linear interpolation between data points |

Orthophosphate

Orthophosphate, taken to the equivalent to soluble reactive phosphorous, is an important nutrient. For each boundary, the data frequency, data source, period of record, and method used to generate daily values is shown in Table 14

Table 14. Orthophosphate boundary condition generation summary.

| Boundary | Data frequency | Data source | Data period | Method |
|--------------------------------|-----------------------|------------------------------------|-----------------------------------|--|
| <u>Mainstem Columbia River</u> | monthly | LRFEP sta 0.0 | JAN99 to JAN02 | Linear interpolation between data points |
| <u>Kettle River</u> | monthly | USGS 12404900 | JAN99 to SEP04 | Linear interpolation between data points |
| <u>Colville River</u> | -- | none | -- | Use Kettle River values |
| <u>Spokane River</u> | monthly | USGS 12433000; LRFEP sta 4.0 | JAN99 to APR00; JAN99 to JAN02 | Historical monthly average values; adjusted in 2000 to match LRFEP data. |
| <u>Sanpoil River</u> | -- | none | -- | Use mainstem Columbia River values |
| <u>Banks Lake return flow</u> | monthly | USGS 12436500 | JAN99 to SEP04 | Linear interpolation between data points |

Ammonium

Ammonium is an important nutrient. For each boundary, the data frequency, data source, period of record, and method used to generate daily values is shown in Table 15.

Table 15. Ammonium boundary condition generation summary.

| Boundary | Data frequency | Data source | Data period | Method |
|--------------------------------|-----------------------|--------------------|--------------------|--|
| <u>Mainstem Columbia River</u> | monthly | LRFEP sta 0.0 | JAN99 to JAN02 | Linear interpolation between data points |
| <u>Kettle River</u> | monthly | USGS 12404900 | JAN99 to SEP04 | Linear interpolation between data points |
| <u>Colville River</u> | -- | none | -- | Use Kettle River values |
| <u>Spokane River</u> | monthly | USGS 12433000 | JAN99 to SEP03 | Linear interpolation between data points |
| <u>Sanpoil River</u> | -- | none | -- | Use mainstem Columbia River values |
| <u>Banks Lake return flow</u> | monthly | USGS 12436500 | JAN99 to SEP04 | Linear interpolation between data points |

Nitrate plus nitrite

Nitrate plus nitrate, or NO_x, is an important nutrient. For each boundary, the data frequency, data source, period of record, and method used to generate daily values is shown in Table 16.

Table 16. Nitrate plus nitrite boundary condition generation summary.

| Boundary | Data frequency | Data source | Data period | Method |
|--------------------------------|-----------------------|--------------------|--------------------|--|
| <u>Mainstem Columbia River</u> | monthly | LRFEP sta 0.0 | JAN99 to JAN02 | Linear interpolation between data points |
| <u>Kettle River</u> | monthly | USGS 12404900 | JAN99 to SEP04 | Linear interpolation between data points |
| <u>Colville River</u> | -- | none | -- | Use Kettle River values |
| <u>Spokane River</u> | monthly | USGS 12433000 | JAN99 to SEP03 | Linear interpolation between data points |
| <u>Sanpoil River</u> | -- | none | -- | Use mainstem Columbia River values |
| <u>Banks Lake return flow</u> | monthly | USGS 12436500 | JAN99 to SEP04 | Linear interpolation between data points |

Chlorophyll-a (algae)

Algae data are available in-stream. Data include biovolume, organism density, speciation, and chlorophyll-a concentrations. The chlorophyll-a data are collected at three relative depths in the water column: lower, middle, and upper.

The values are numerically averaged to provide an single concentration for each sampling date and location. Linear interpolation is used to generate a chlorophyll-a concentration for each simulation day. The values are then converted from chlorophyll-a to algal dry weight using a stoichiometric ratio [ACHLA] of 100.

For each boundary, the data frequency, data source, period of record, and method used to generate daily values is shown in Table 17.

Table 17. Chlorophyll-a (algae) boundary condition generation summary.

| Boundary | Data frequency | Data source | Data period | Method |
|--------------------------------|-----------------------|--------------------|--------------------|--|
| <u>Mainstem Columbia River</u> | monthly | LRFEP sta 0.0 | JAN99 to JAN02 | Linear interpolation between data points |
| <u>Kettle River</u> | -- | none | -- | Use mainstem Columbia River values |
| <u>Colville River</u> | -- | none | -- | Not used |
| <u>Spokane River</u> | monthly | LRFEP sta 4.0 | JAN99 to JAN02 | Linear interpolation between data points |
| <u>Sanpoil River</u> | monthly | LRFEP sta 8.0 | JAN99 to JAN02 | Linear interpolation between data points |
| <u>Banks Lake return flow</u> | -- | none | -- | Not used |

The algal biomass is partitioned among the 3 algal functional groups (diatoms, mixed greens, and cyanobacteria) using the system wide ratio of the data at all sites. Fractions were generated for days of available data (roughly monthly) and linearly interpolated between to generate daily fractions. Figure 74 shows the interpolated values (lines) and the monthly values (icons). The daily fractions were applied at each boundary condition.

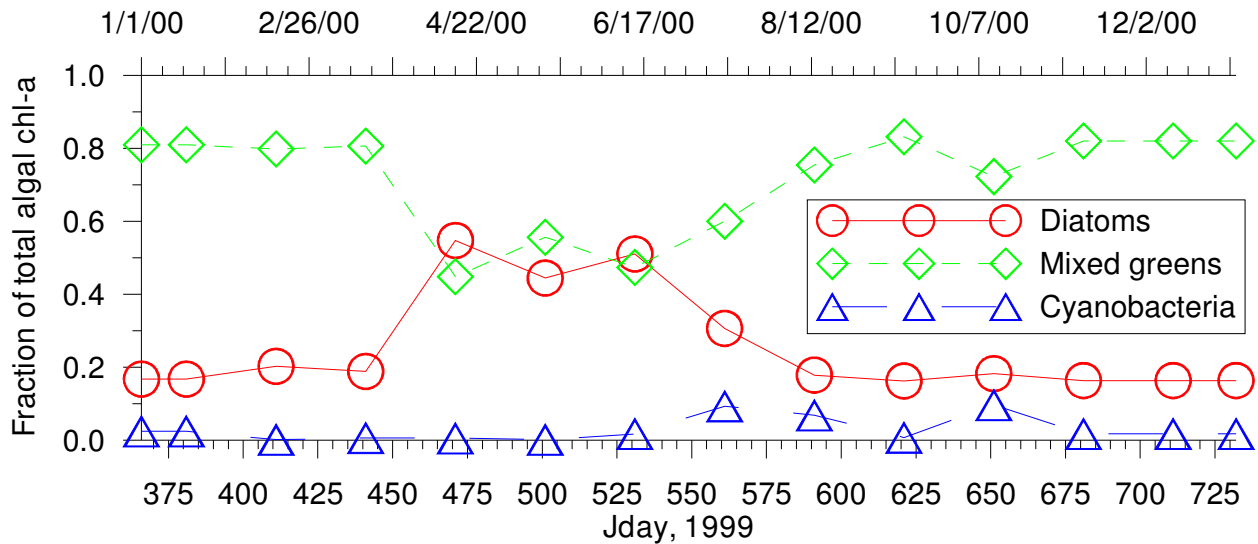


Figure 74. System wide daily fraction of total algal chl-a.

pH

pH is an important boundary condition used to model the carbonate chemistry. For each boundary, the data frequency, data source, period of record, and method used to generate daily values is shown in Table 18. pH is not directly input into the model, but is used to build the total inorganic carbon (TIC) input.

Table 18. pH boundary condition generation summary.

| Boundary | Data frequency | Data source | Data period | Method |
|--------------------------------|----------------|------------------|-----------------------------------|--|
| <u>Mainstem Columbia River</u> | monthly | USGS 12400520 | JAN99 to SEP03 | Linear interpolation between data points |
| <u>Kettle River</u> | monthly | USGS 12404900 | JAN99 to SEP02 | Linear interpolation between data points |
| <u>Colville River</u> | monthly | USGS 12409000 | OCT94 to SEP95; OCT99 to SEP00 | Not used |
| <u>Spokane River</u> | monthly | USGS 12433000 | JAN99 to SEP03 | Linear interpolation between data points |
| <u>Sanpoil River</u> | -- | none | -- | Not used |
| <u>Banks Lake return flow</u> | monthly | USGS 12436500 | JAN99 to SEP04 | Linear interpolation between data points |

Total dissolved solids

Total dissolved solids data are not available at the model boundaries; however, in stream data are collected by the LRFEP. An arithmetic average (as opposed to a volume weighted average) of the TDS and conductivity data at the most upstream station (sta 0.0) yielded the relationship,

$$\text{TDS (mg/L)} = 0.64 \cdot \text{Conductivity } (\mu\text{S/cm})$$

This ratio is consistent with other lacustrine systems and is used to estimate TDS from the conductivity data. Figure 75 is a plot of the linear regression.

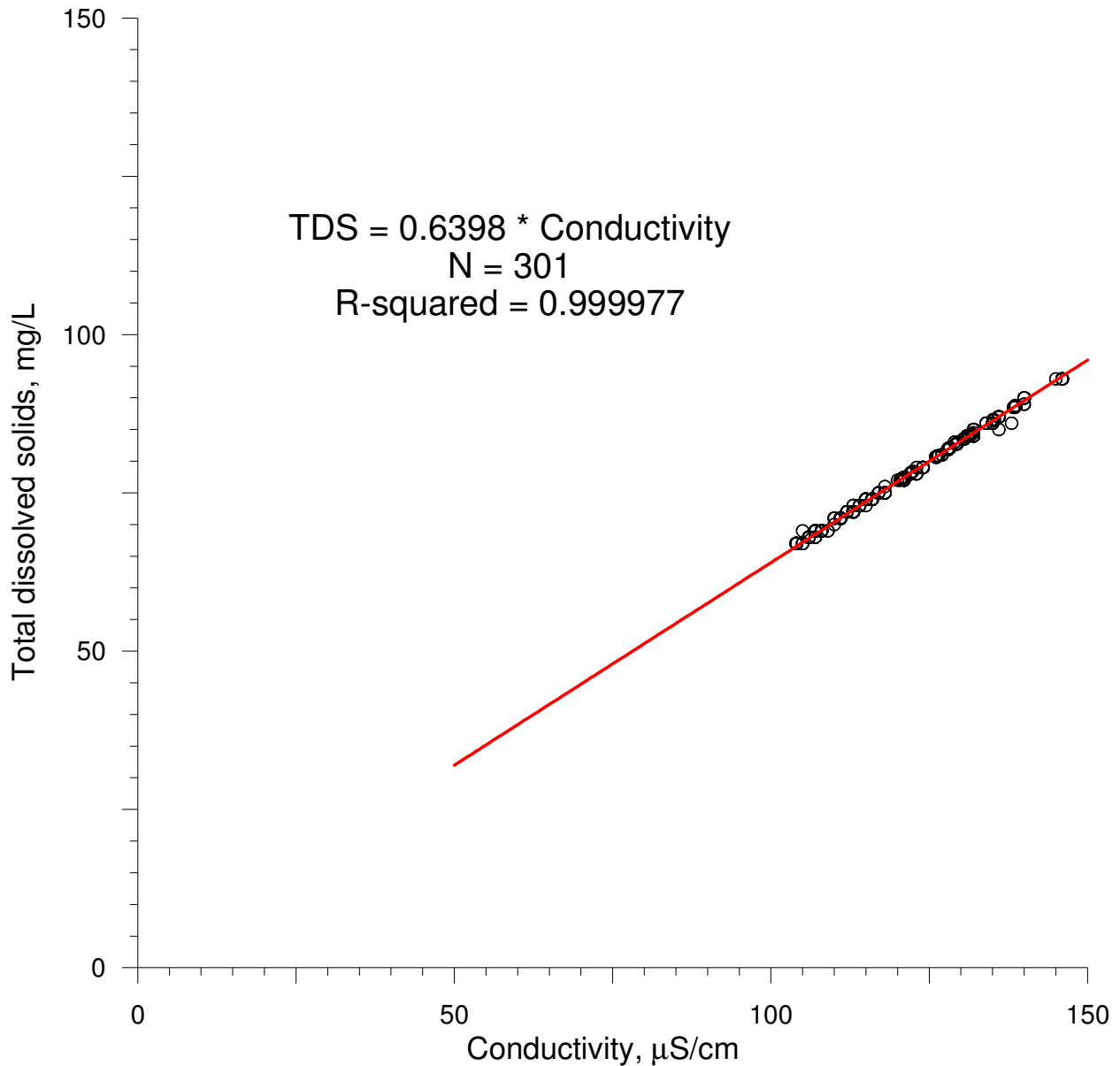


Figure 75. TDS:Conductivity ratio at LRFEP station 0.0, 1999 to 2002.

Total inorganic carbon and total organic matter

Total inorganic carbon

From carbonate chemistry, total inorganic carbon is a function of alkalinity, pH, and temperature.

Dissolved organic carbon

Dissolved organic carbon data are ultimately used to generate the constituents ISS, LDOM, RDOM, LPOM, RPOM. DOC data are only collected at the Environment Canada stations. For each boundary, the data frequency, data source, period of record, and method used to generate daily values is shown in Table 19.

Table 19. Dissolved organic carbon boundary condition generation summary.

| Boundary | Data frequency | Data source | Data period | Method |
|---|-----------------------|--------------------|--------------------|--|
| <u>Mainstem Columbia River</u> | * | * | * | Daily flow weighted average value from Columbia R. and Pend Oreille R. values. |
| <i>Columbia prior to confluence</i> | biweekly | BC08NE0001 | JAN99 to DEC04 | Linear interpolation between data points |
| <i>Pend Oreille prior to confluence</i> | biweekly | BC08NE0029 | JAN99 to DEC04 | Linear interpolation between data points |
| | | | | |
| <u>Kettle River</u> | biweekly | BC08NN0021 | JAN99 to DEC04 | Linear interpolation between data points |
| <u>Colville River</u> | -- | none | -- | Use mainstem Columbia River values |
| <u>Spokane River</u> | -- | none | -- | Use mainstem Columbia River values |
| <u>Sanpoil River</u> | -- | none | -- | Use mainstem Columbia River values |
| <u>Banks Lake return flow</u> | -- | none | -- | Use mainstem Columbia River values |

Total organic carbon

Total organic carbon is used to calculate TOM. However, only dissolved organic carbon data are available, so a DOC:POC ratio of 8:1 is assumed. TOC is then equal to $1.125 \cdot \text{DOC}$.

Total organic matter

Total organic matter is calculated from TOC using a carbon-biomass ratio, $\delta_c = 0.65$ and the relationship, $\text{TOM} = \text{TOC} / \delta_c - \text{algae}$.

Total suspended solids

Total suspended solids are not directly model input. TSS data are used to calculate the inorganic suspended solids. For each boundary, the data frequency, data source, period of record, and method used to generate daily values is shown in Table 20.

Table 20. Total suspended solids boundary condition generation summary.

| Boundary | Data frequency | Data source | Data period | Method |
|--------------------------------|----------------|------------------|----------------|--|
| <u>Mainstem Columbia River</u> | monthly | USGS 12400520 | JAN99 to SEP03 | Linear interpolation between data points |
| <u>Kettle River</u> | monthly | USGS 12404900 | JAN99 to SEP04 | Linear interpolation between data points |
| <u>Colville River</u> | -- | none | -- | Use mainstem Columbia River values |
| <u>Spokane River</u> | -- | none | -- | Use mainstem Columbia River values |
| <u>Sanpoil River</u> | -- | none | -- | Use mainstem Columbia River values |
| <u>Banks Lake return flow</u> | monthly | USGS 12436500 | JAN99 to SEP04 | Linear interpolation between data points |

Total inorganic suspended solids (ISS)

Inorganic suspended solids are calculated from the relationship,

$$\text{ISS (mg/L)} = \text{TSS (mg/L)} - \text{POM (mg/L)} - \text{algae (mg/L-dry weight)}$$

A minimum value of 0.1 mg/L was used.

Dissolved organic matter

Dissolved organic matter is calculated from DOC using the relationship, $\text{DOM} = \text{DOC} / \delta_c$. The DOM constituent is subdivided into a labile and refractory compartment using $f_{\text{L DOM}} = 0.5$ such that,

$$\text{LDOM} = f_{\text{LDOM}} * \text{DOM}$$

and

$$\text{RDOM} = (1 - f_{\text{LDOM}}) * \text{DOM}$$

Particulate organic matter

Particulate organic matter is calculated by,

$$\text{POM} = \text{TOM} - \text{DOM}$$

The POM constituent is subdivided into a labile and refractory compartment using $f_{\text{LPOM}} = 0.5$ such that,

$$\text{LPOM} = f_{\text{LPOM}} * \text{POM}$$

and

$$\text{RPOM} = (1 - f_{\text{LPOM}}) * \text{POM}$$

Zooplankton

Zooplankton data were collected at 17, 33, and 66 m tow depths. The tows were organized into the functional groups of cladocera and copepoda. The cladocera group was formed by adding the biomass on days with both Daphnia and “other cladocera” data available. The copepoda group was divided into herbivorous and omnivorous sub-groups. The omnivorous group was assumed to be 10% of the copepoda biomass.

For each boundary, the data frequency, data source, period of record, and method used to generate daily values is shown in Table 21.

Table 21. Zooplankton boundary condition generation summary.

| Boundary | Data frequency | Data source | Data period | Method |
|--------------------------------|-----------------------|--------------------|--------------------|--|
| <u>Mainstem Columbia River</u> | monthly | LRFEP sta 0.0 | JAN99 to JAN02 | Linear interpolation between data points |
| <u>Kettle River</u> | -- | none | -- | Use mainstem Columbia River values |
| <u>Colville River</u> | -- | none | -- | Not used |
| <u>Spokane River</u> | monthly | LRFEP sta 4.0 | JAN99 to JAN02 | Linear interpolation between data points |
| <u>Sanpoil River</u> | monthly | LRFEP sta 8.0 | JAN99 to JAN02 | Linear interpolation between data points |
| <u>Banks Lake return flow</u> | -- | none | -- | Not used |

Data and Daily Boundary Condition Model Input Comparison

Mainstem Columbia River

The mainstem Columbia River has 2 stations near the upstream boundary containing dissolved oxygen and nutrient data. The stations showed general agreement. While the USGS gage #12400520 is closer to the boundary, the LRFEP sta 0.0 data were used (echoed) to better match the downstream in-stream data.

For constituents with flow weighting of the Pend Oreille River and Columbia River data, the model inputs should fall between the two values.

The Columbia River water quality boundary condition inputs and the data used to build them (when applicable) are shown in Figure 76 through Figure 89.

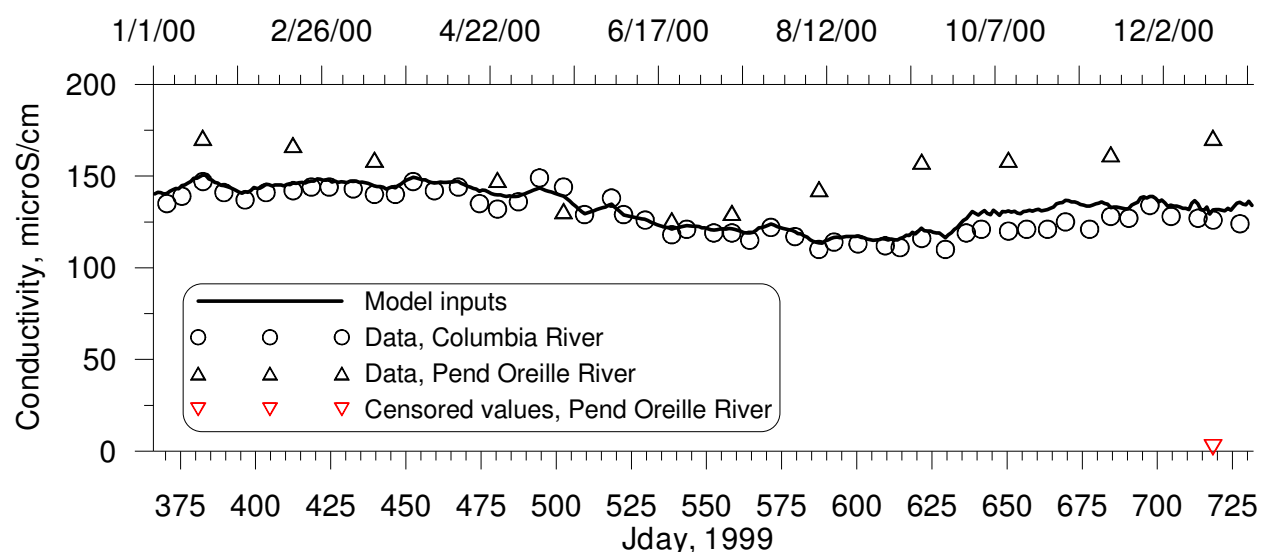


Figure 76. Mainstem Columbia River conductivity boundary condition, 2000.

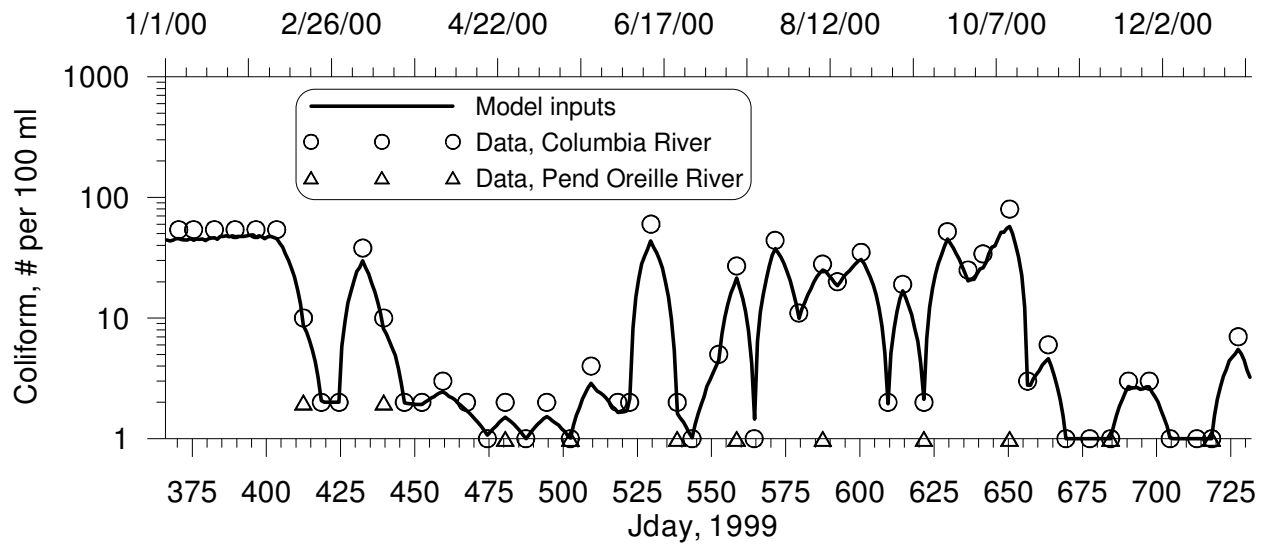


Figure 77. Mainstem Columbia River coliform boundary condition, 2000.

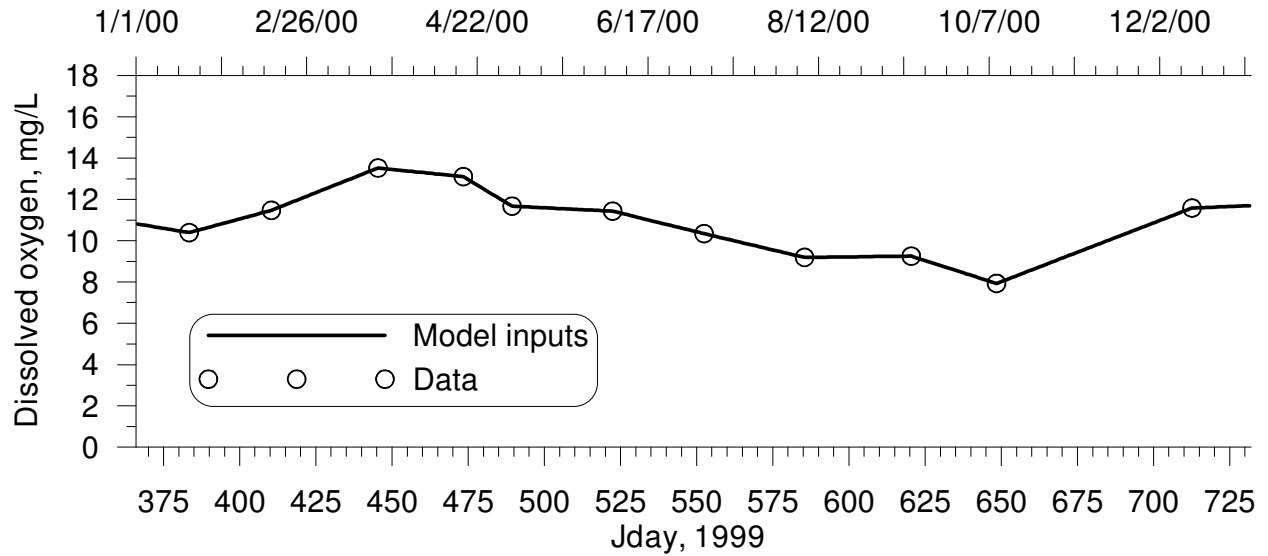


Figure 78. Mainstem Columbia River dissolved oxygen boundary condition, 2000.

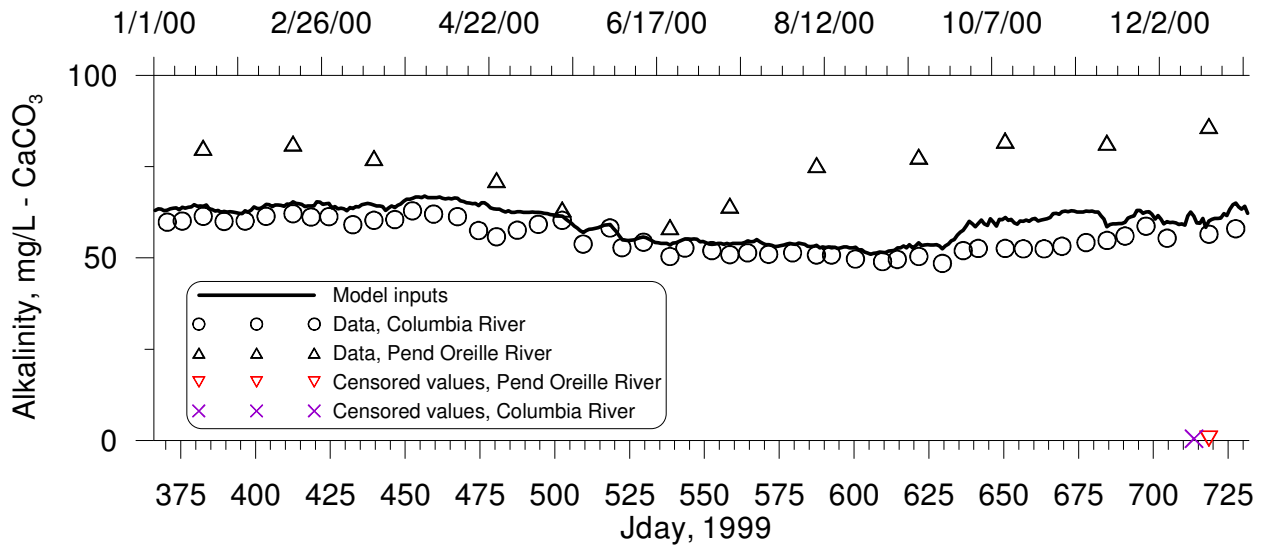


Figure 79. Mainstem Columbia River alkalinity boundary condition, 2000.

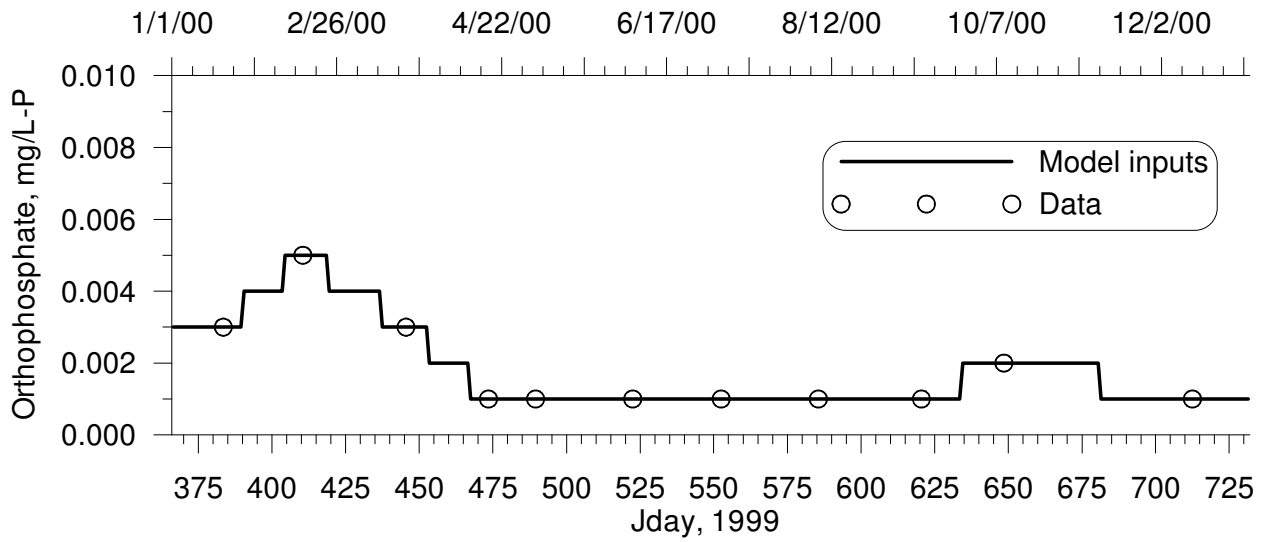


Figure 80. Mainstem Columbia River orthophosphate boundary condition, 2000.

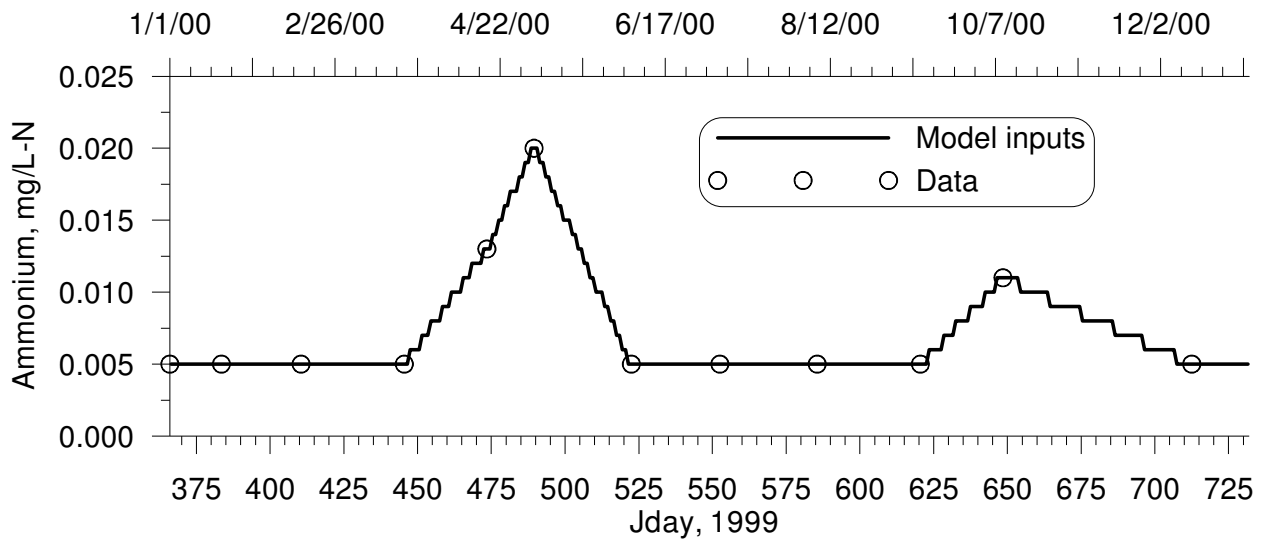


Figure 81. Mainstem Columbia River ammonium boundary condition, 2000.

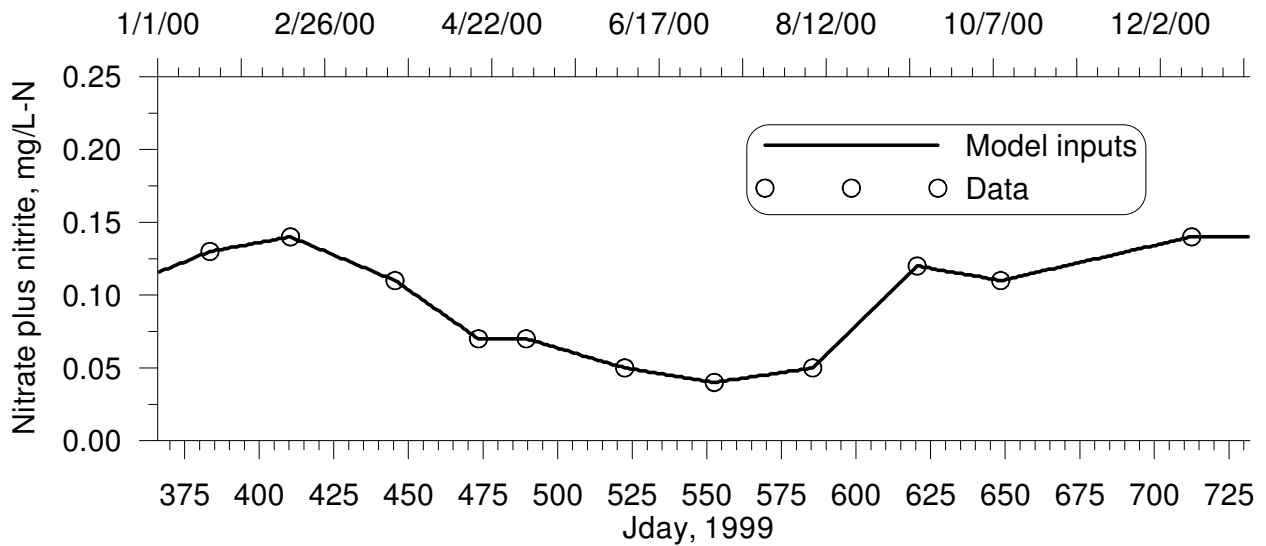


Figure 82. Mainstem Columbia River nitrate plus nitrite boundary condition, 2000.

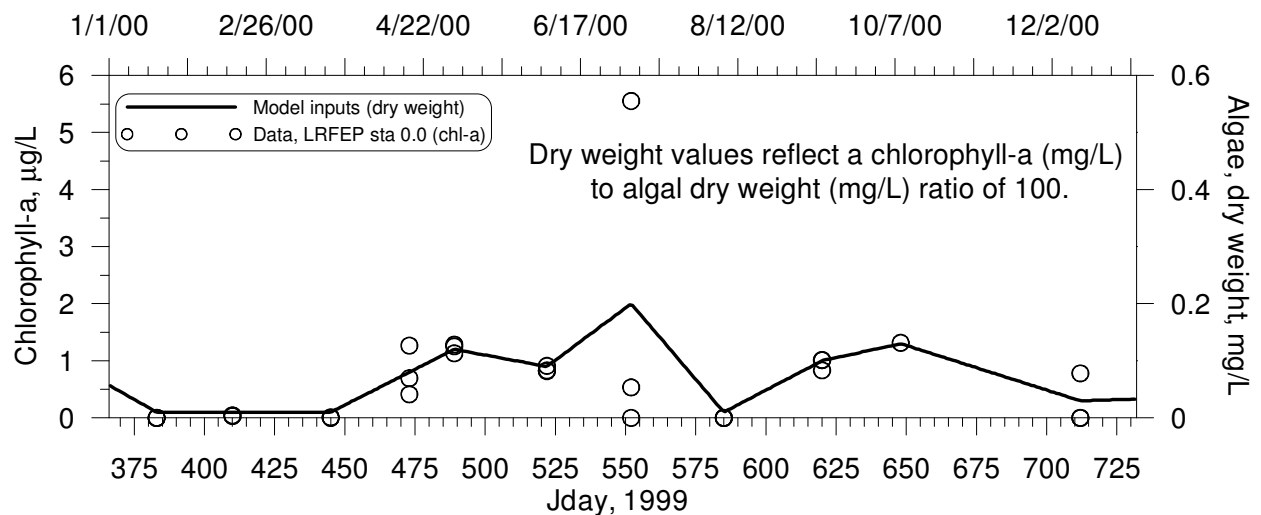


Figure 83. Mainstem Columbia River algal dry weight boundary condition, 2000.

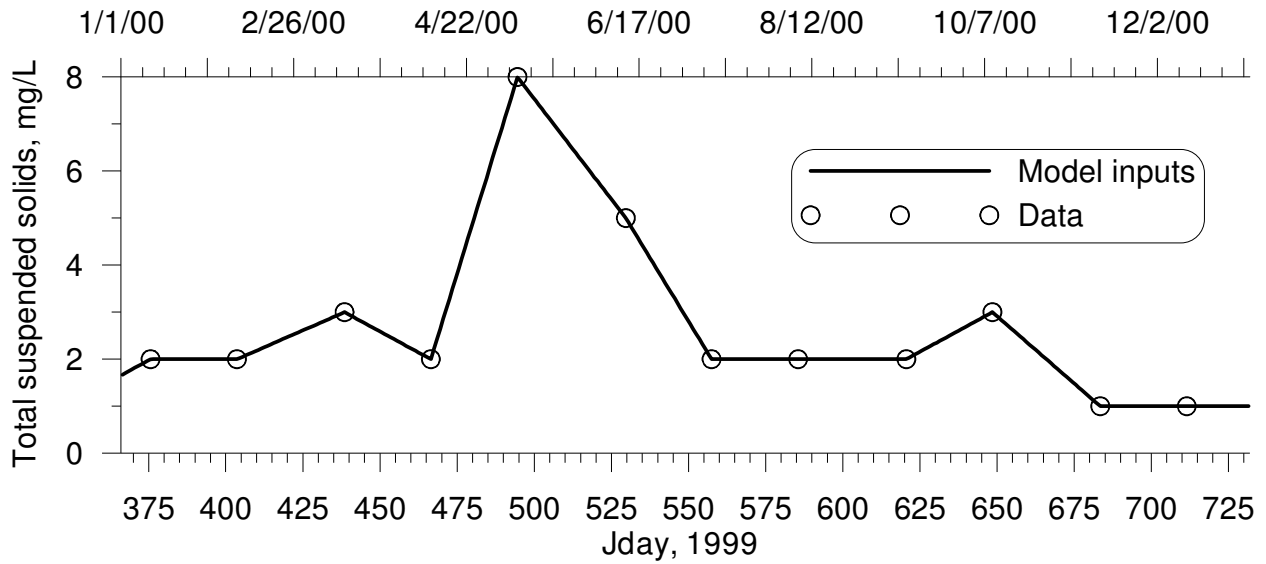


Figure 84. Mainstem Columbia River total suspended solids boundary condition, 2000.

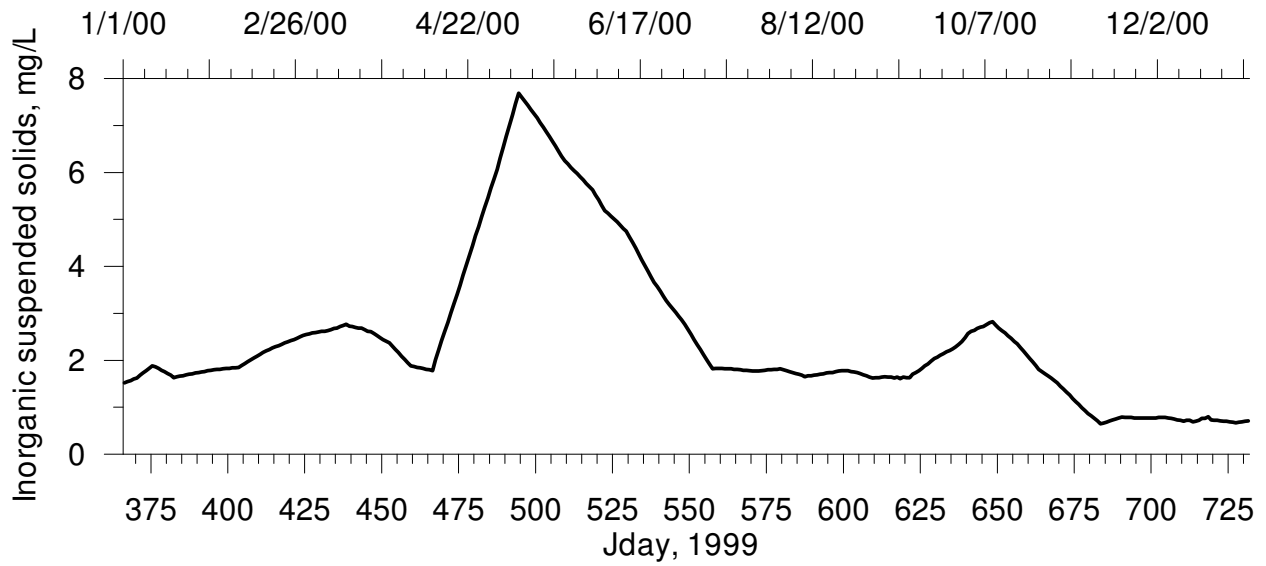


Figure 85. Mainstem Columbia River inorganic suspended solids boundary condition, 2000.

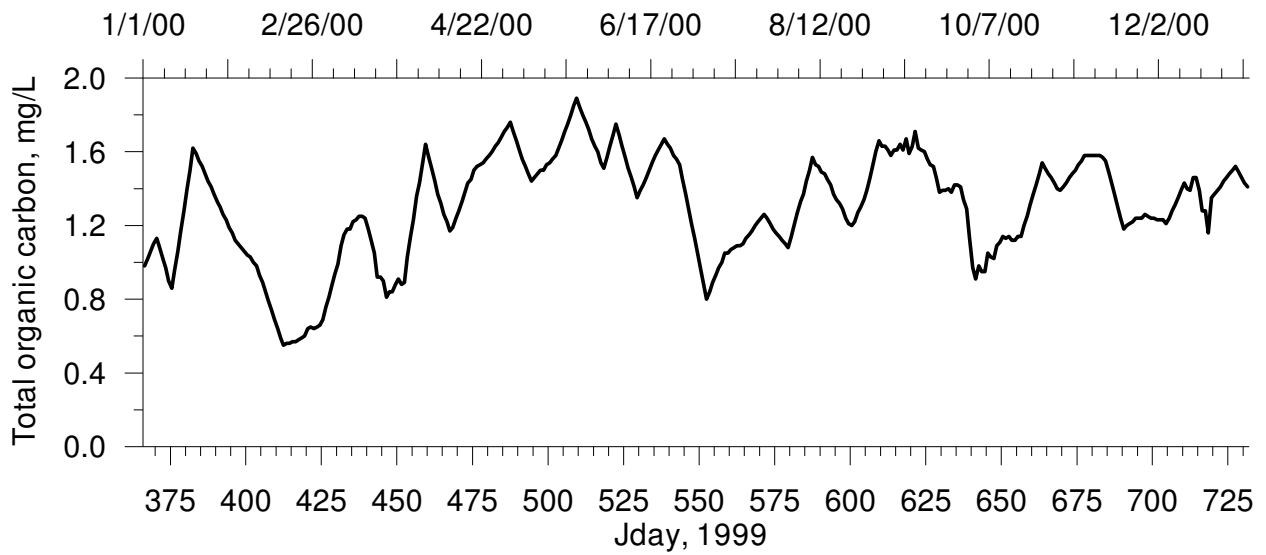


Figure 86. Mainstem Columbia River total organic carbon boundary condition, 2000.

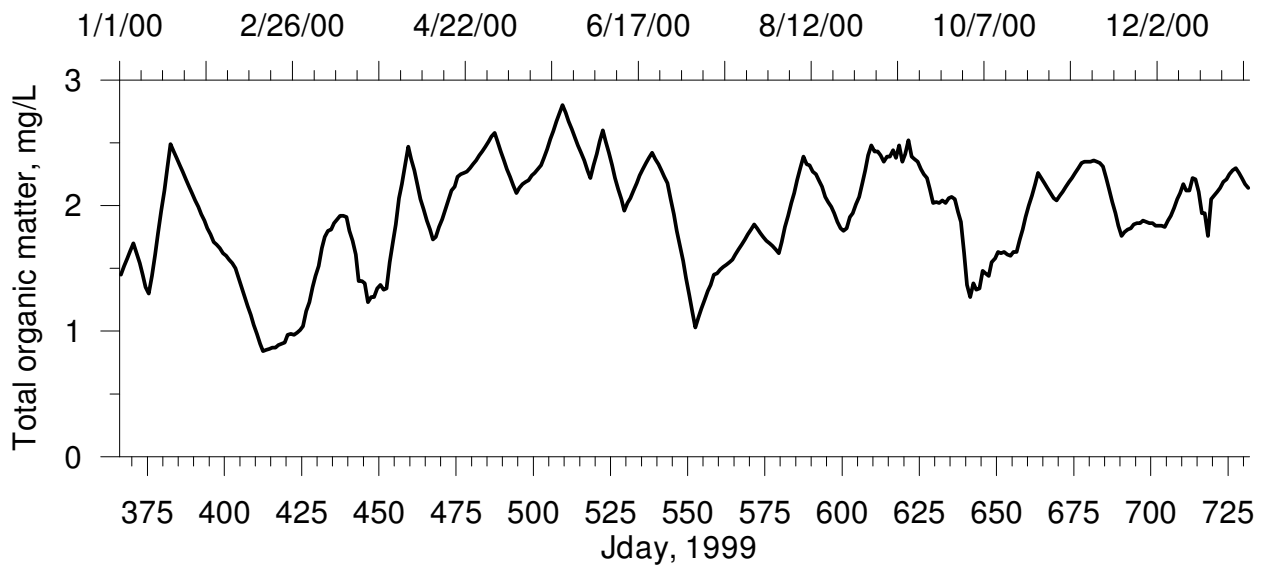


Figure 87. Mainstem Columbia River total organic matter boundary condition, 2000.

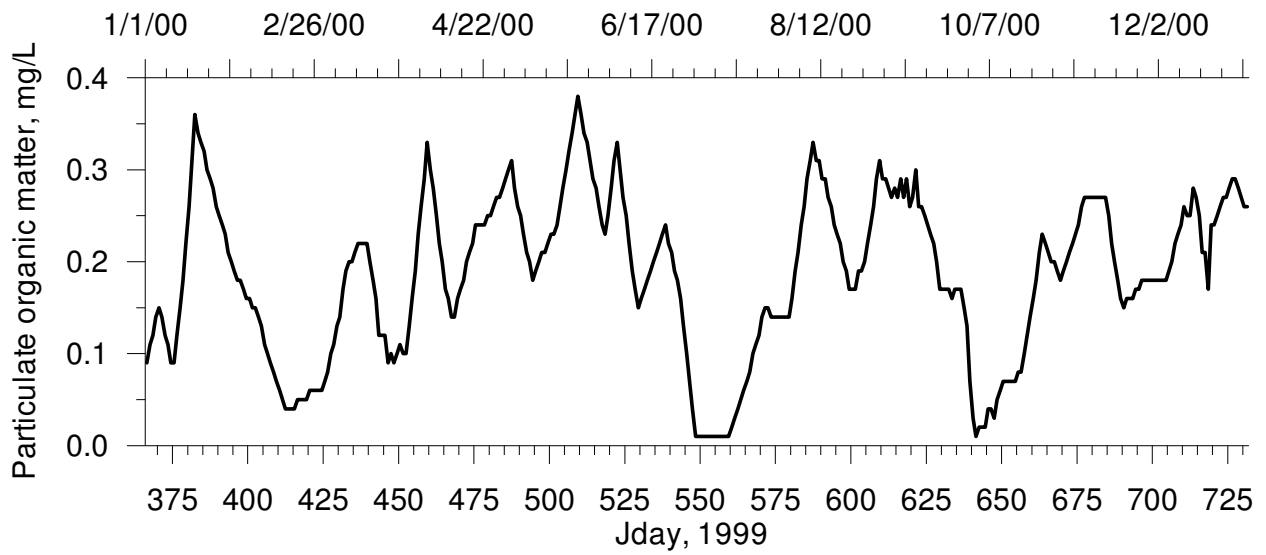


Figure 88. Mainstem Columbia River particulate organic matter boundary condition, 2000.

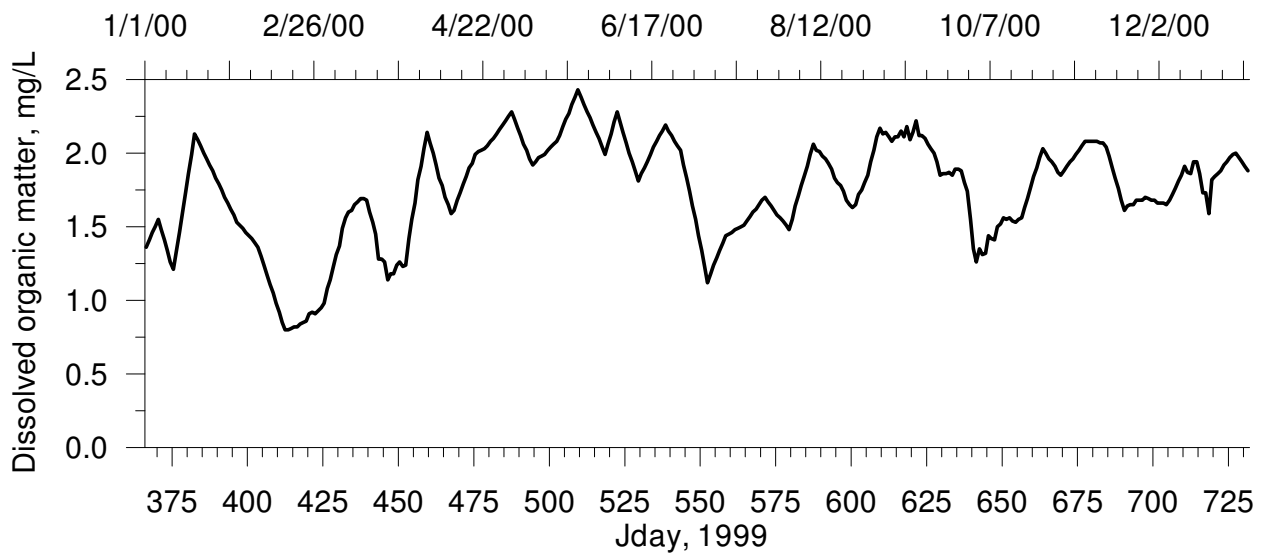


Figure 89. Mainstem Columbia River dissolved organic matter boundary condition, 2000.

Kettle River

The Kettle River water quality boundary condition inputs and the data used to build them (when applicable) are shown in Figure 90 through Figure 95. For constituents that used the Columbia River boundary condition, refer to the Columbia River section.

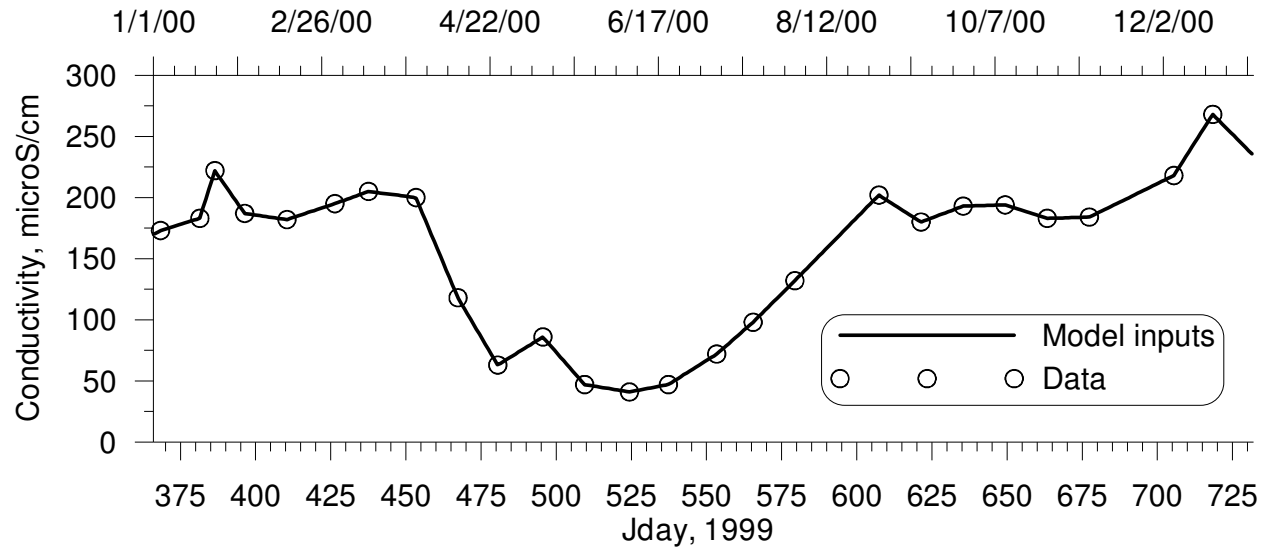


Figure 90. Kettle River conductivity boundary condition, 2000.

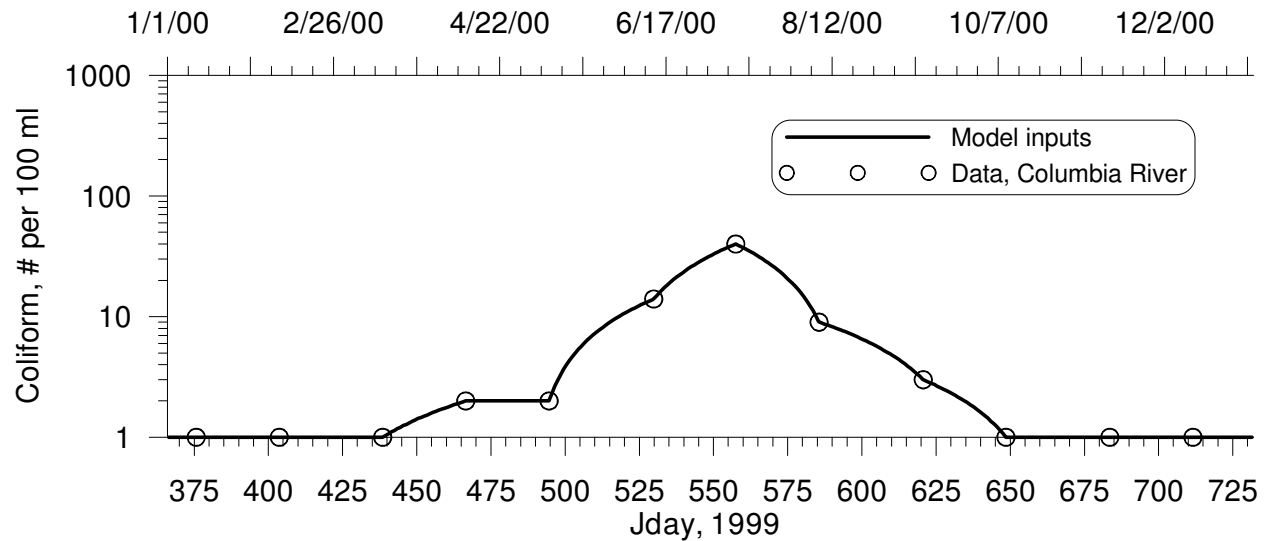


Figure 91. Kettle River coliform boundary condition, 2000.

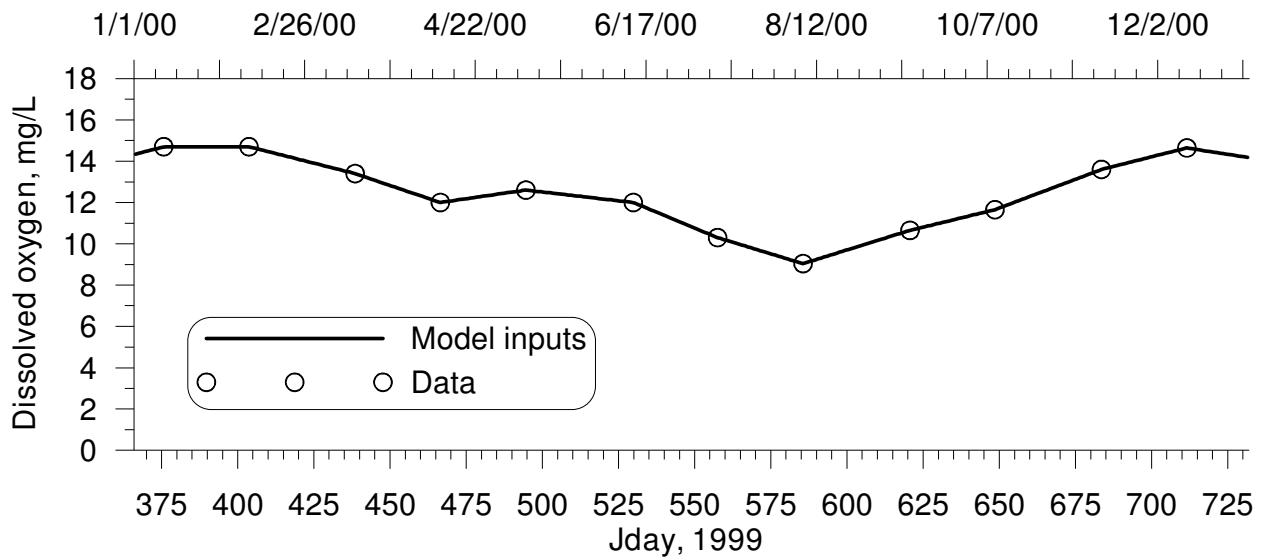


Figure 92. Kettle River dissolved oxygen boundary condition, 2000.

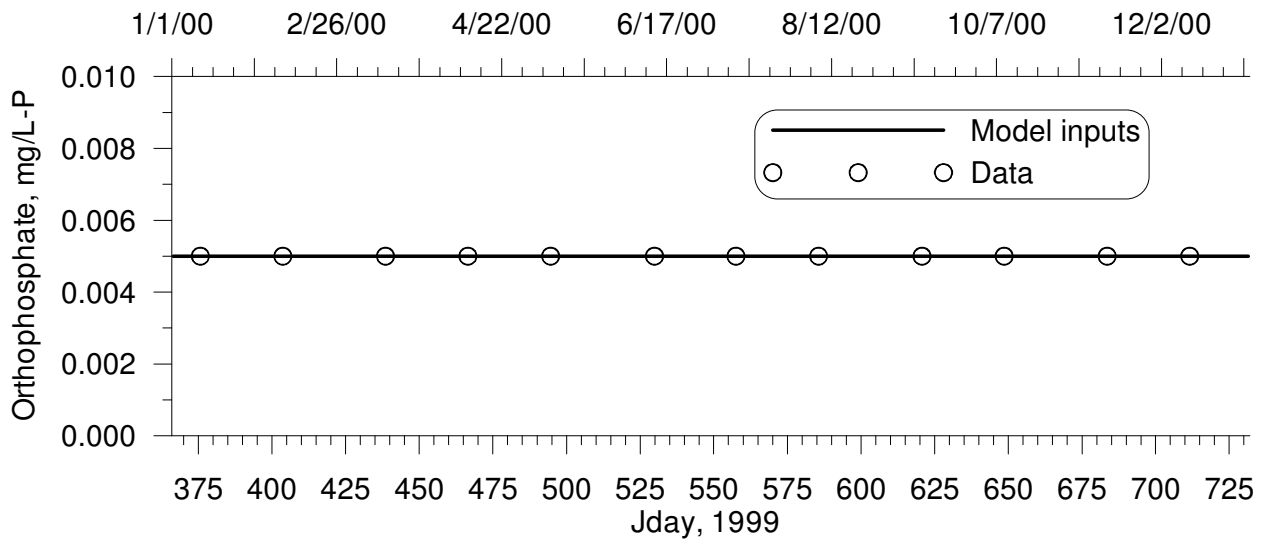


Figure 93. Kettle River orthophosphate boundary condition, 2000.

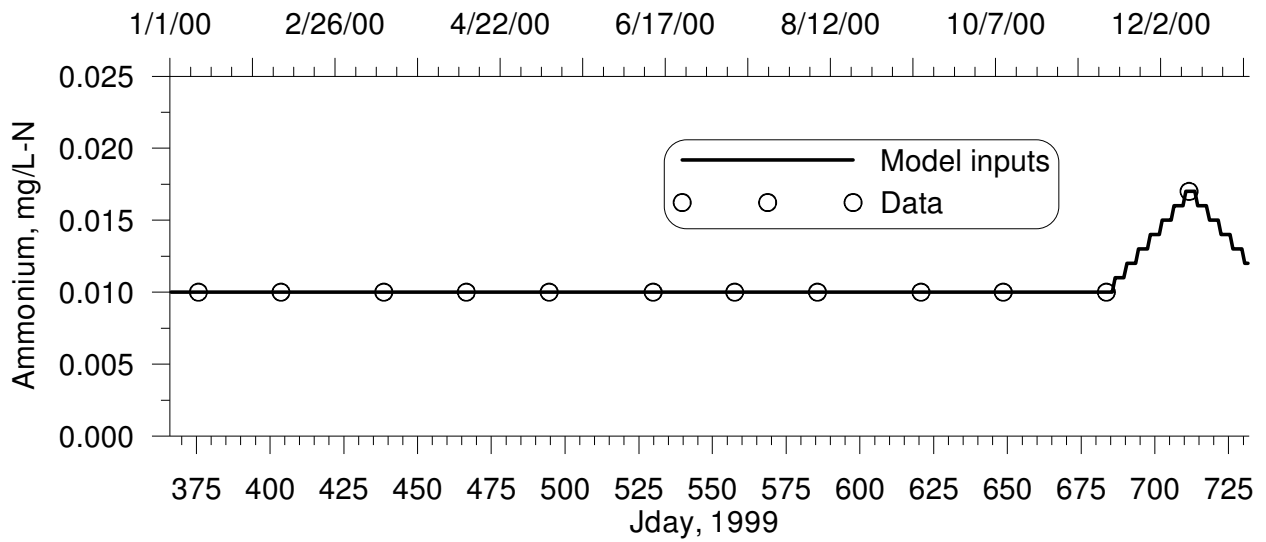


Figure 94. Kettle River ammonium boundary condition, 2000.

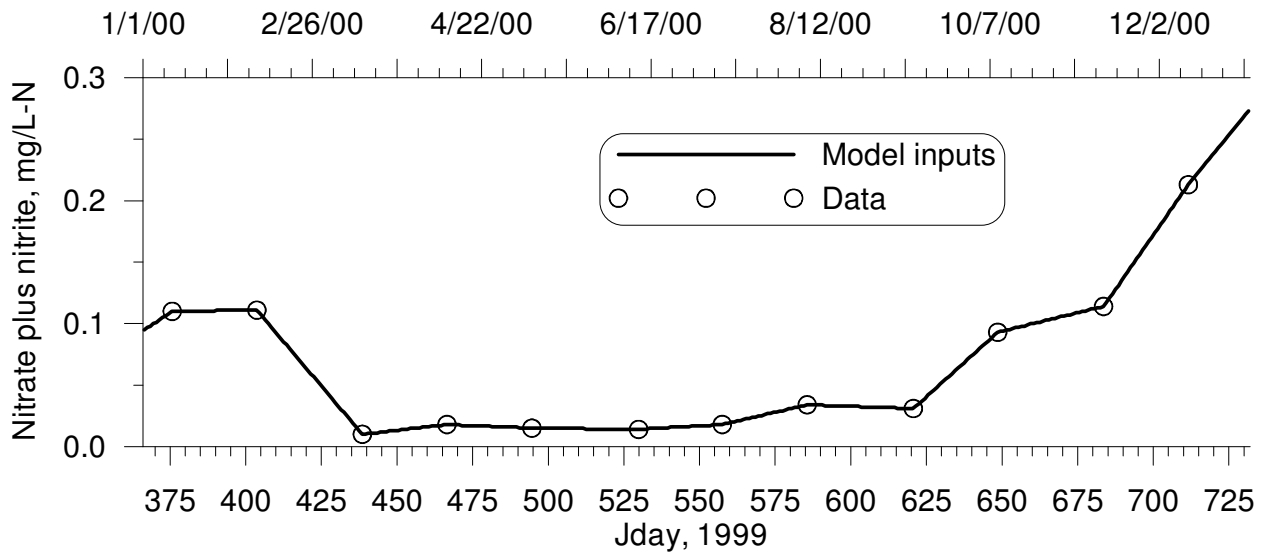


Figure 95. Kettle River nitrate plus nitrite boundary condition, 2000.

Colville River

The Colville River water quality boundary condition inputs and the data used to build them (when applicable) are shown in Figure 96 through Figure 99. For constituents that used the Columbia River boundary condition, refer to the Columbia River section.

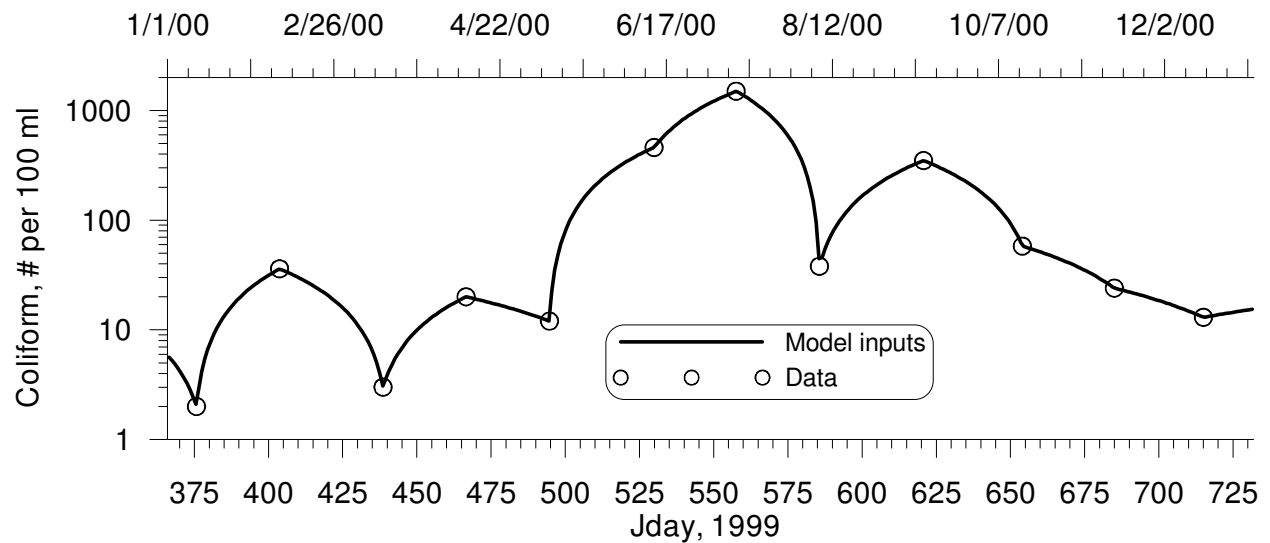


Figure 96. Colville River coliform boundary condition, 2000.

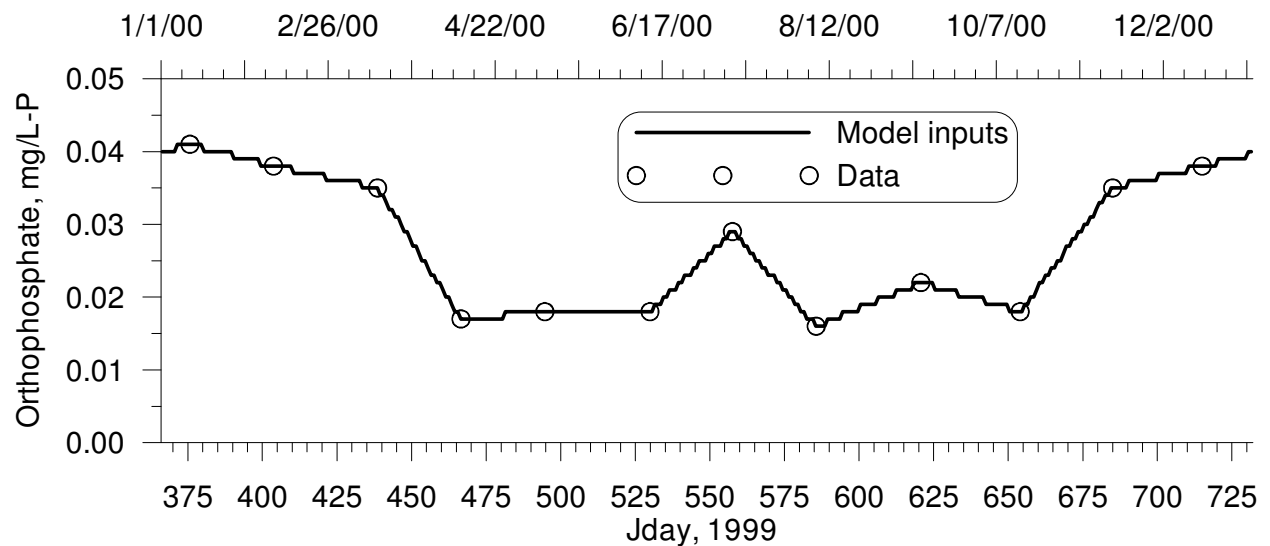


Figure 97. Colville River orthophosphate boundary condition, 2000.

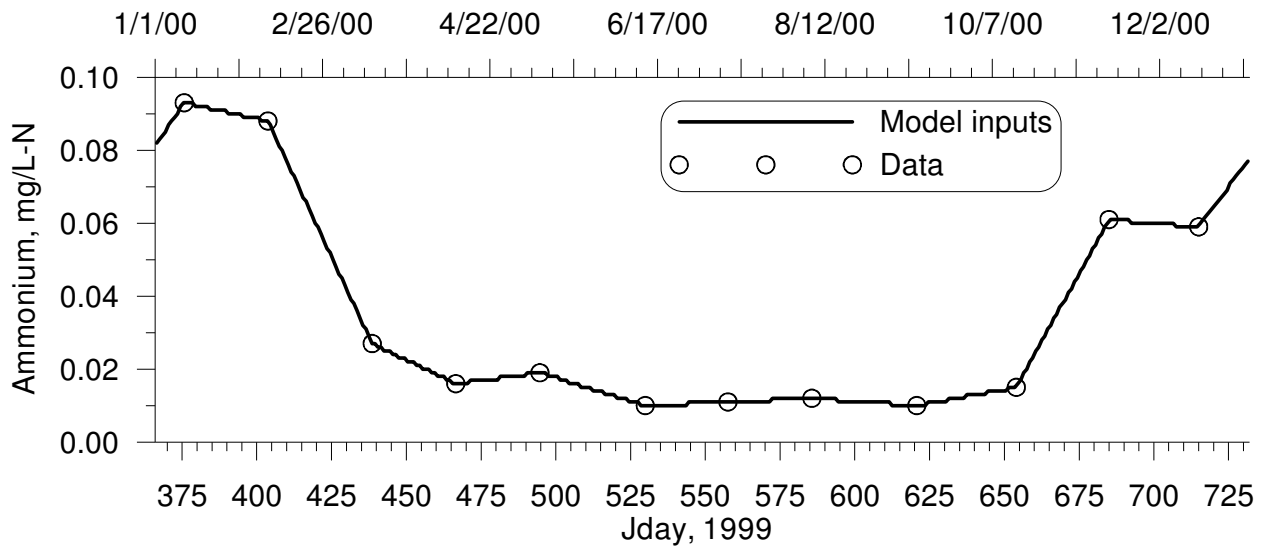


Figure 98. Colville River ammonium boundary condition, 2000.

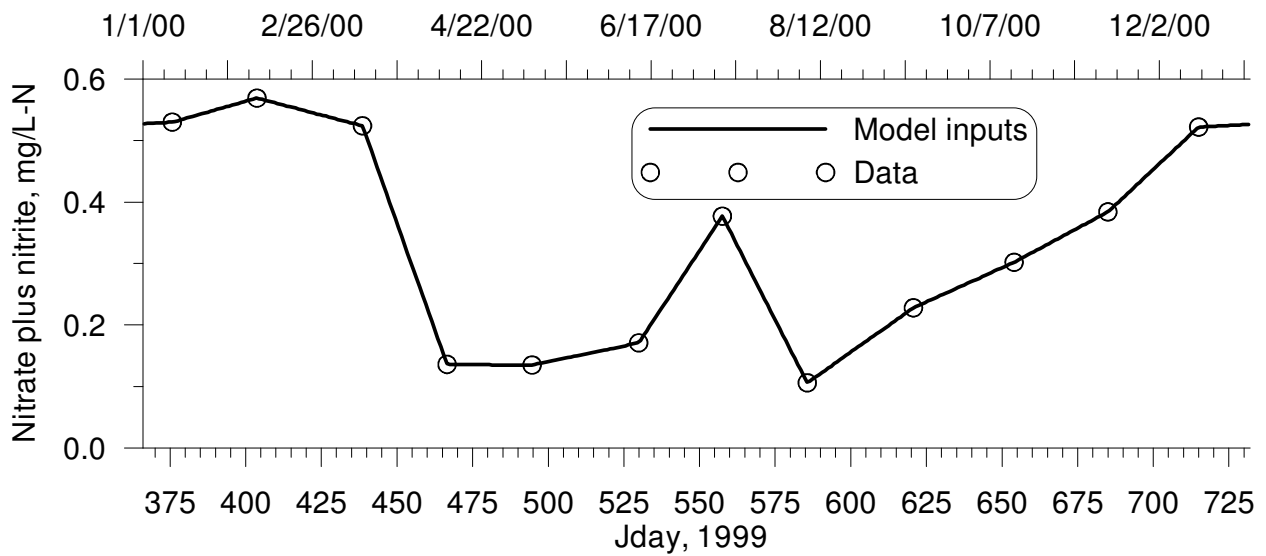


Figure 99. Colville River nitrate plus nitrite boundary condition, 2000.

Spokane River

The Spokane River water quality boundary condition inputs and the data used to build them (when applicable) are shown in Figure 100 through Figure 105. For constituents that used the Columbia River boundary condition, refer to the Columbia River section. The monthly averaged phosphorous inputs were adjusted in 2000 by adding actual data points from the LRFEP station 4.0 on J349, J383, and J411.

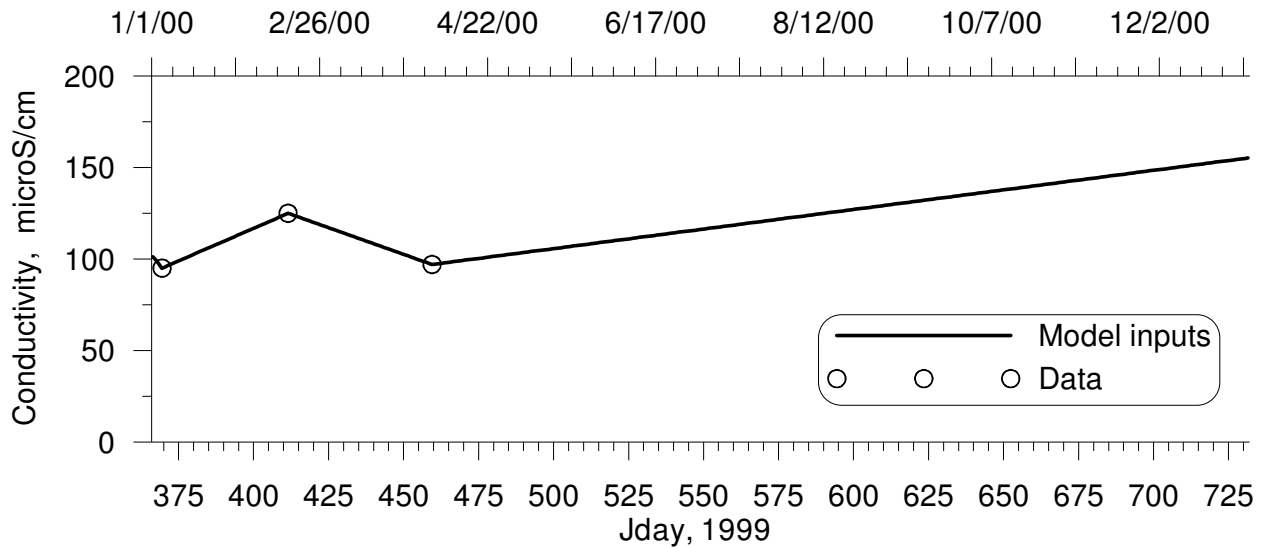


Figure 100. Spokane River conductivity boundary condition, 2000.

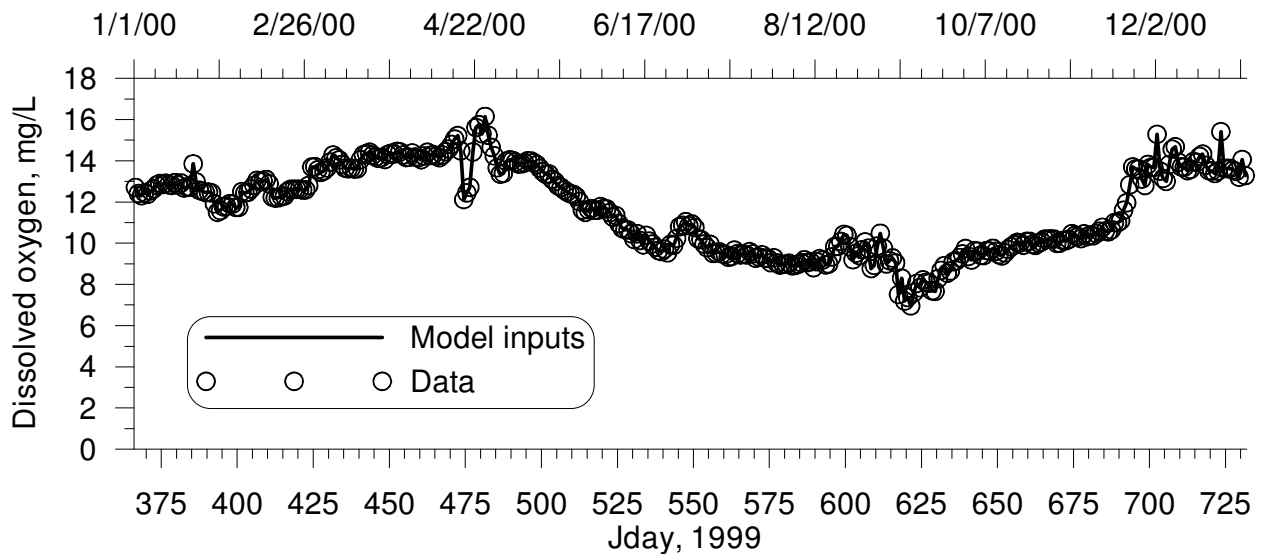


Figure 101. Spokane River dissolved oxygen boundary condition, 2000.

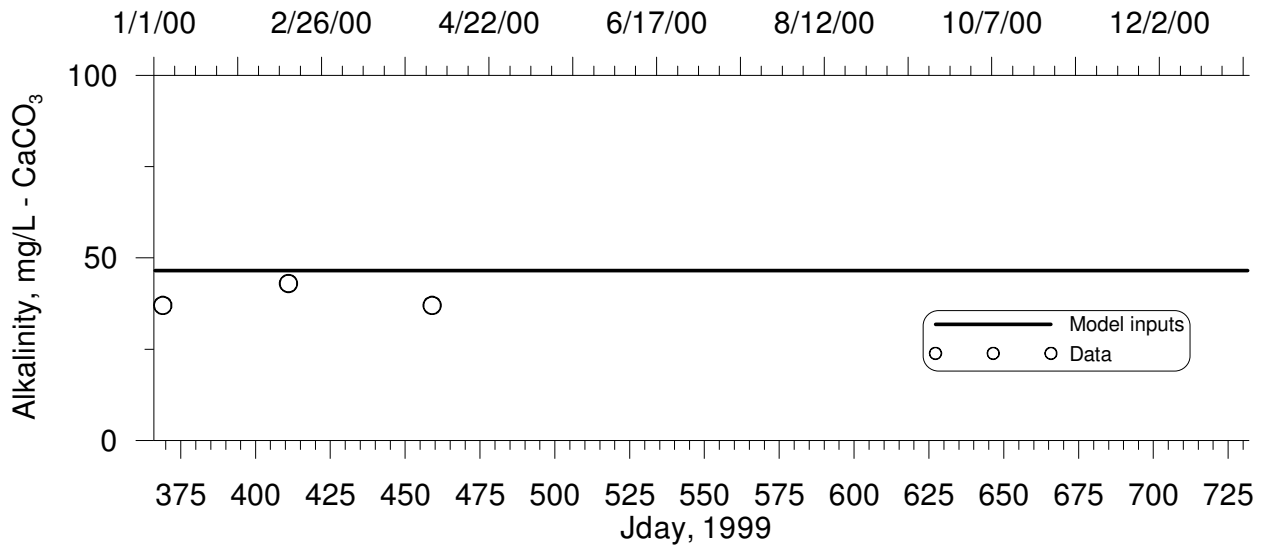


Figure 102. Spokane River alkalinity boundary condition, 2000.

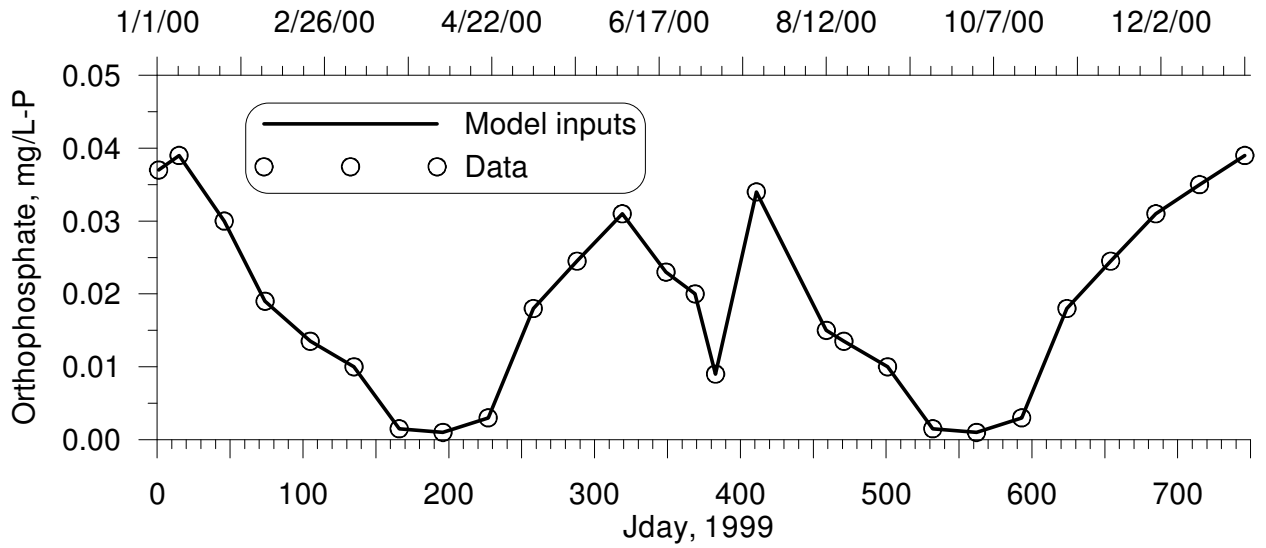


Figure 103. Spokane River orthophosphate boundary condition, 2000.

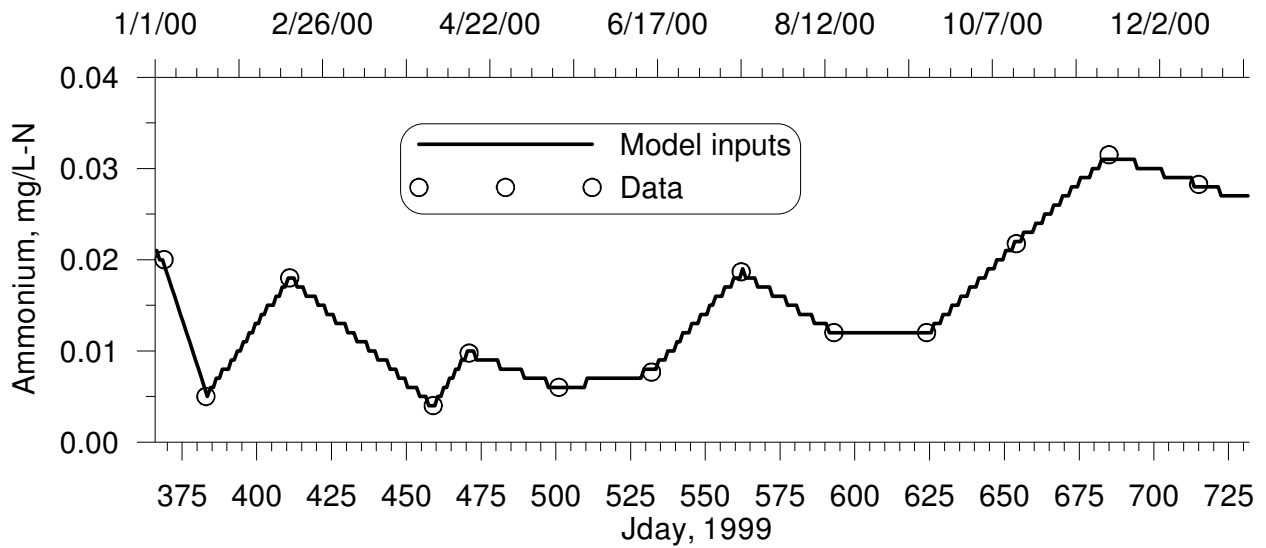


Figure 104. Spokane River ammonium boundary condition, 2000.

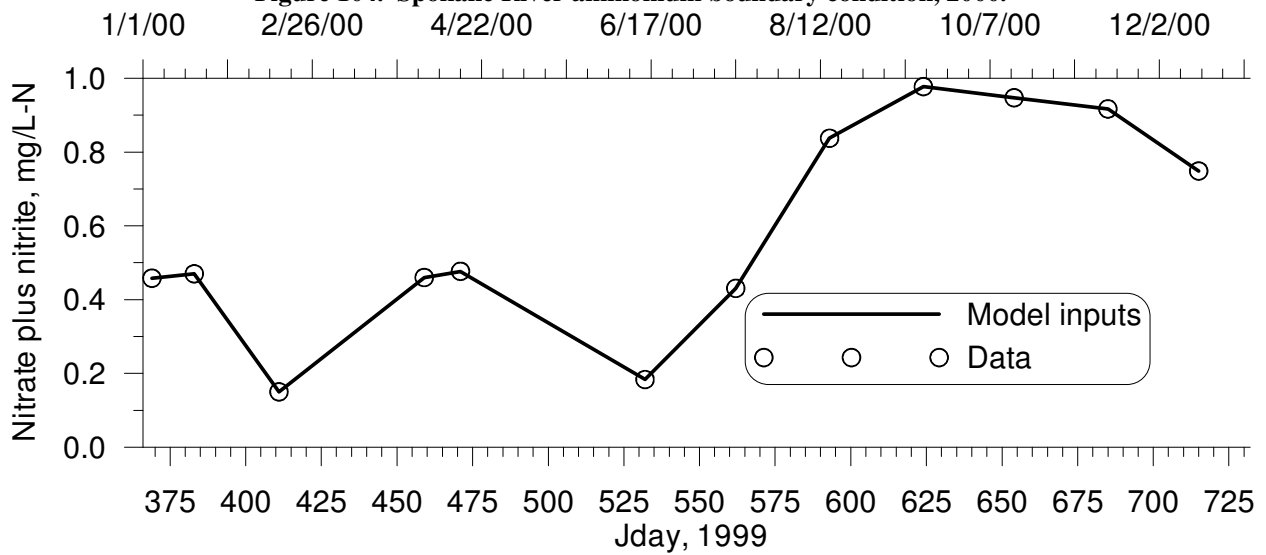


Figure 105. Spokane River nitrate plus nitrite boundary condition, 2000.

Sanpoil River

Dissolved oxygen is the only constituent available for the Sanpoil River. The water quality boundary condition input is shown as Figure 106. For constituents that used the Columbia River boundary condition, refer to the Columbia River section.

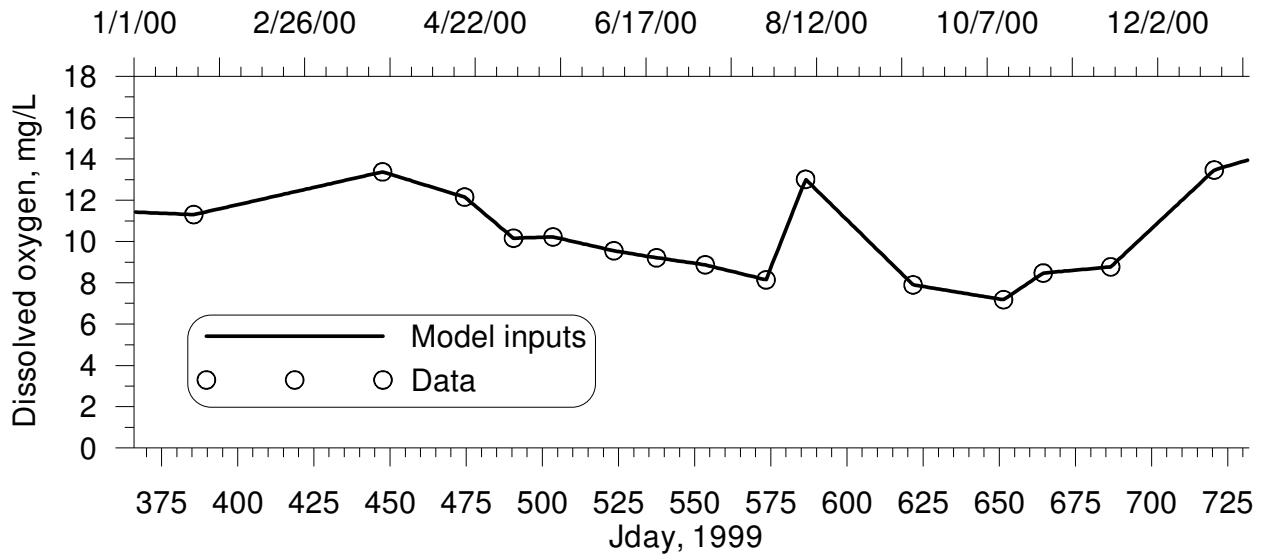


Figure 106. Sanpoil River dissolved oxygen boundary condition, 2000.

Banks Lake Return Flows

The Banks Lake return flow water quality boundary condition inputs and the data used to build them (when applicable) are shown in Figure 107 through Figure 112. For constituents that used the Columbia River boundary condition, refer to the Columbia River section. Note that the water quality constituent values only effect the model results during periods of inflow (marked with blue squares on the date axes). During periods of outflow (withdrawal), the boundary condition values are not used.

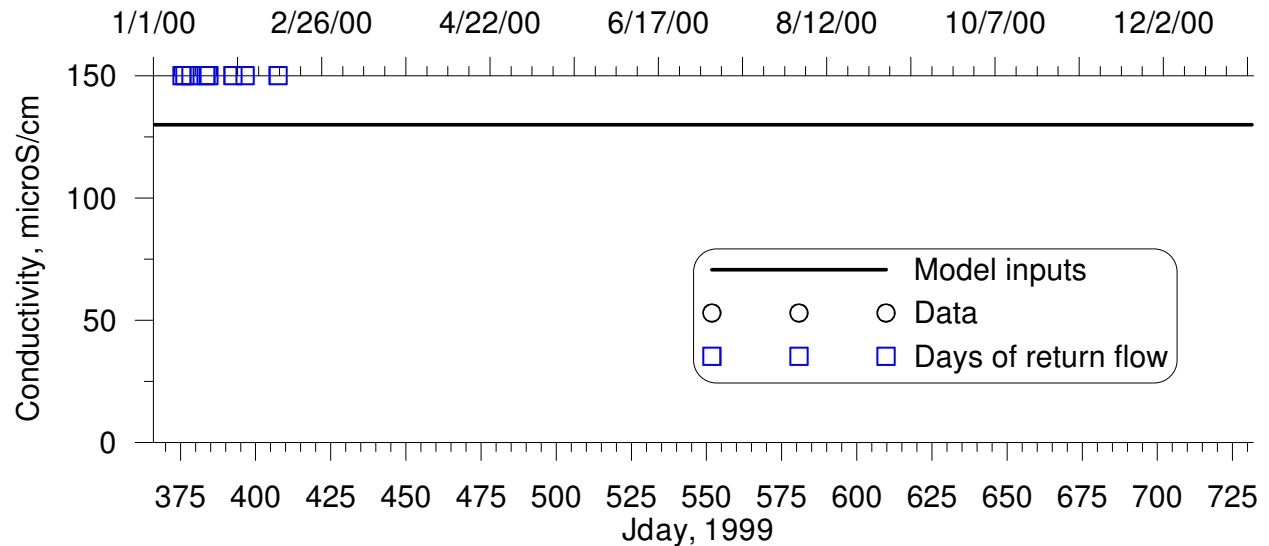


Figure 107. Banks Lake return flow conductivity boundary condition, 2000.

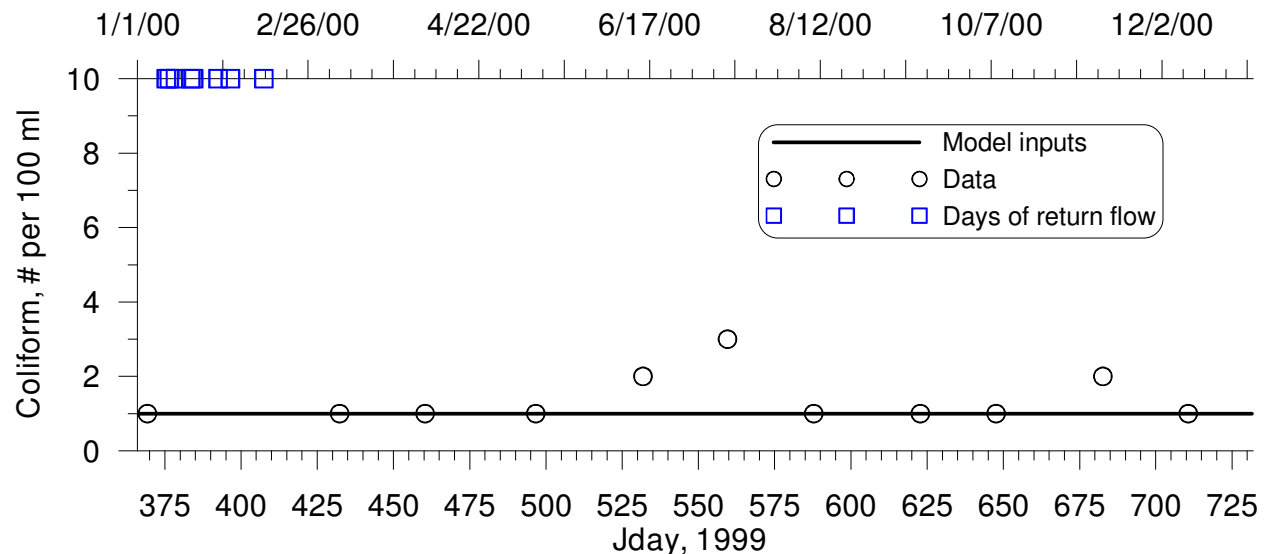


Figure 108. Banks Lake return flow coliform boundary condition, 2000.

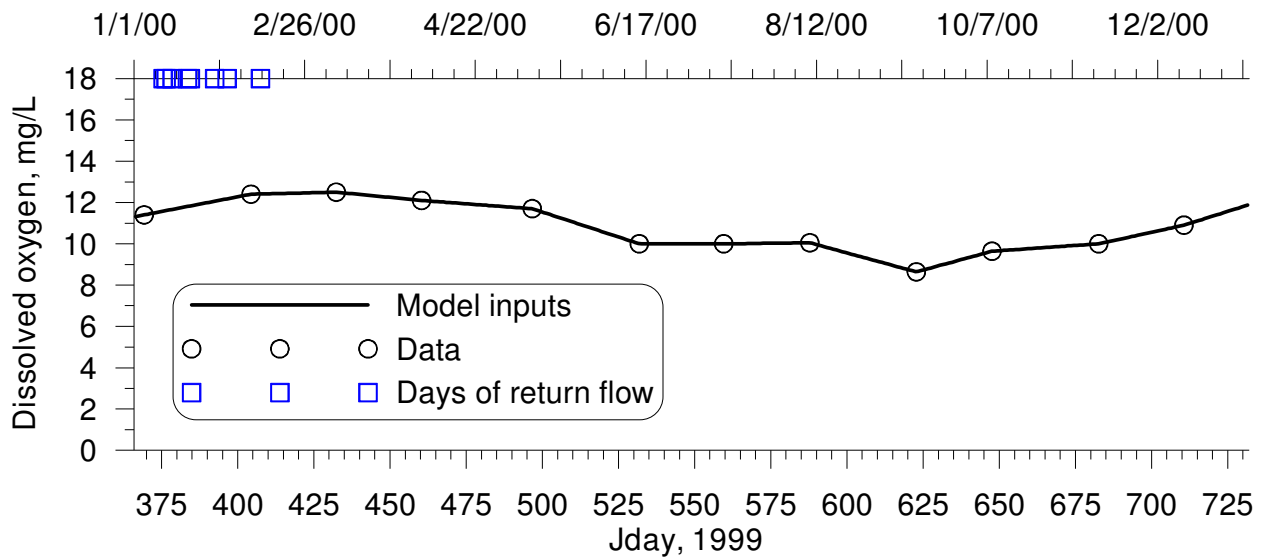


Figure 109. Banks Lake return flow dissolved oxygen boundary condition, 2000.

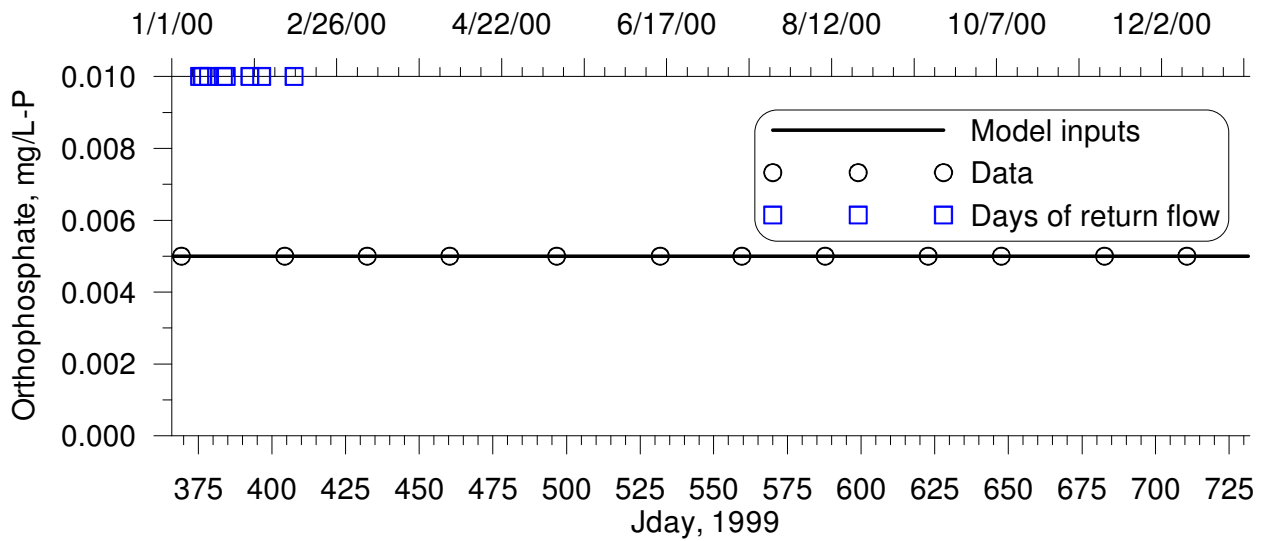


Figure 110. Banks Lake return flow orthophosphate boundary condition, 2000.

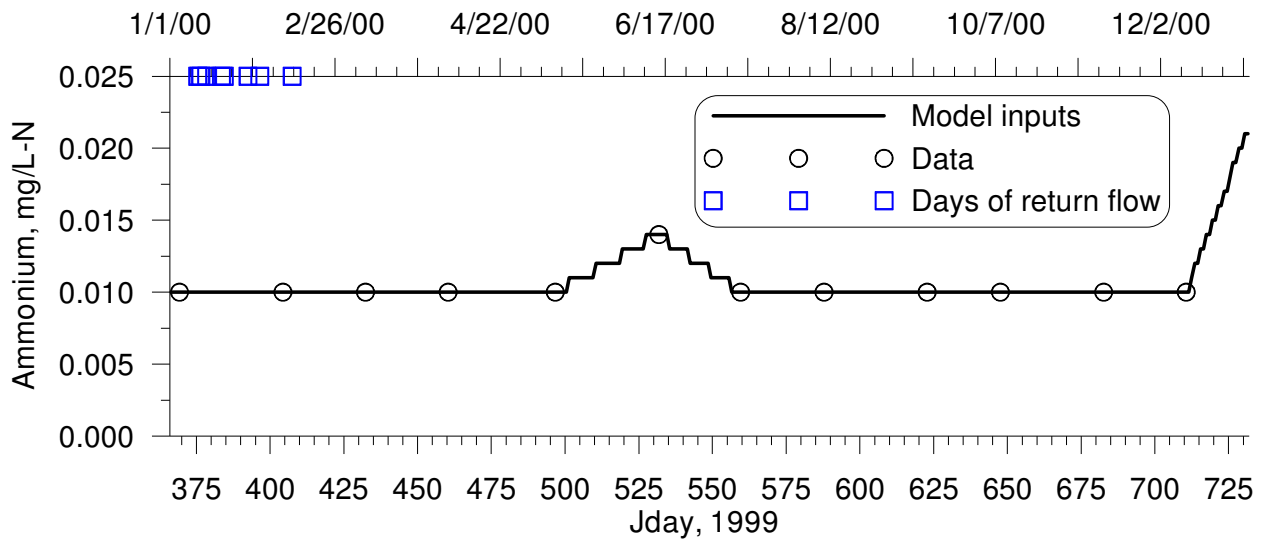


Figure 111. Banks Lake return flow ammonium boundary condition, 2000.

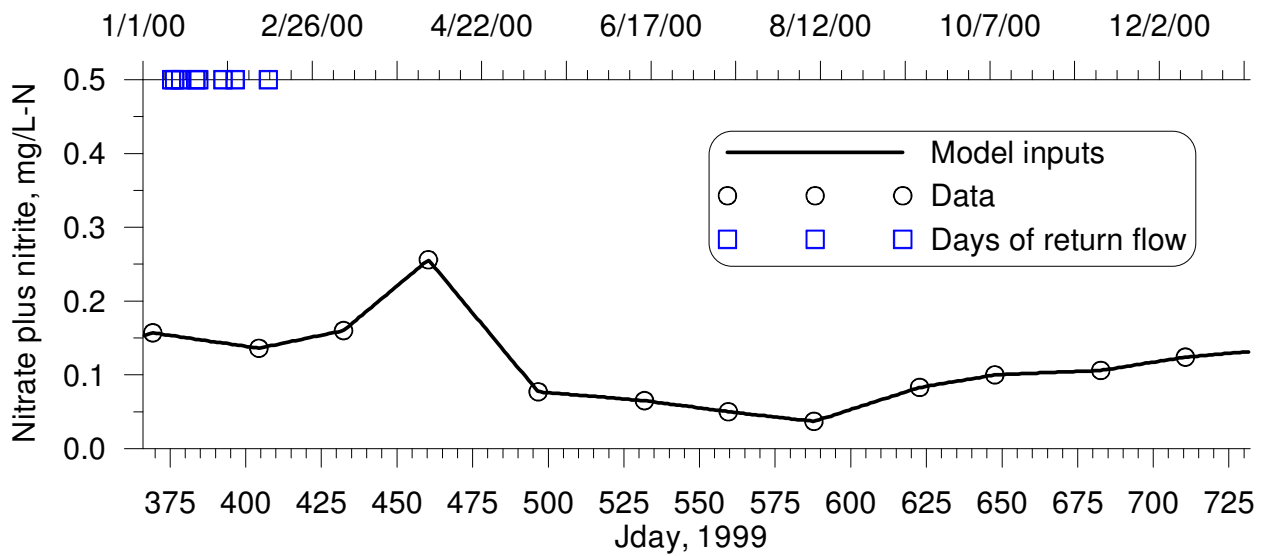


Figure 112. Banks Lake return flow nitrate plus nitrite boundary condition, 2000.

Appendix B: W2 control files

W2_con.npt

A sample W2_con.npt file is shown below. A summary of the water quality parameters is include as Appendix H.

```

W2 Model Version 3.5

TITLE C .....TITLE.....
Version 3.5 Lake Roosevelt Model
Riverine reaches of Columbia and Spokane Rivers broken into several WBS
for KT thickness control
Jday 1.0 is Jan 1, 1999

Run 123

Tom Cole, WES; Scott Wells, PSU; Rob Annear, PSU; Chris Berger, PSU

GRID          NWB      NBR      IMX      KMX
              25      25      583      76

IN/OUTFL      NTR      NST      NIW      NWD      NGT      NSP      NPI      NPU
              1       4       0       1       0       1       0       0

CONSTITU      NGC      NSS      NAL      NEP      NBOD     NMC      NZP
              3       1       3       1       1       0       3

MISCELL      NDAY
              100

TIME CON      TMSTRT   TMEND   YEAR
              366.000 731.000 1999

DLT CON      NDT      DLTMIN
              8      0.10000

DLT DATE      DLTD     DLTD     DLTD     DLTD     DLTD     DLTD     DLTD     DLTD     DLTD
              366.000 366.050 366.100 455.0    490.0    530.0    550.0    702.0

DLT MAX      DLTMAX   DLTMAX   DLTMAX   DLTMAX   DLTMAX   DLTMAX   DLTMAX   DLTMAX   DLTMAX
              5.00000 10.0000 45.00000 30.00    30.00    15.00    60.00    30.00

DLT FRN      DLTF     DLTF     DLTF     DLTF     DLTF     DLTF     DLTF     DLTF     DLTF
              0.90000 0.90000 0.90000 0.90000 0.90000 0.90000 0.90000 0.90000 0.90000 0.90000

DLT LIM1     VISC     CELC
WB 1         ON       ON
WB 2         ON       ON
WB 3         ON       ON
WB 4         ON       ON
WB 5         ON       ON
WB 6         ON       ON
WB 7         ON       ON
WB 8         ON       ON
WB 9         ON       ON
WB 10        ON       ON
WB 11        ON       ON
WB 12        ON       ON
WB 13        ON       ON
WB 14        ON       ON
WB 15        ON       ON
WB 16        ON       ON
WB 17        ON       ON
WB 18        ON       ON
WB 19        ON       ON
WB 20        ON       ON

```

| | | |
|-------|----|----|
| WB 21 | ON | ON |
| WB 22 | ON | ON |
| WB 23 | ON | ON |
| WB 24 | ON | ON |
| WB 25 | ON | ON |

| BRANCH G | US | DS | UHS | DHS | UQB | DQB | NLMIN | SLOPE | |
|----------|-----|-----|-----|-----|-----|-----|-------|---------|---------------|
| BR1 | 2 | 19 | 0 | 22 | 0 | 0 | 1 | 0.00055 | Mainstem |
| BR2 | 22 | 31 | 19 | 34 | 0 | 0 | 1 | 0.00000 | |
| BR3 | 34 | 43 | 31 | 46 | 0 | 0 | 1 | 0.00000 | |
| BR4 | 46 | 55 | 43 | 58 | 0 | 0 | 1 | 0.00000 | |
| BR5 | 58 | 67 | 55 | 70 | 0 | 0 | 1 | 0.00000 | |
| BR6 | 70 | 79 | 67 | 82 | 0 | 0 | 1 | 0.00000 | |
| BR7 | 82 | 91 | 79 | 94 | 0 | 0 | 1 | 0.00000 | |
| BR8 | 94 | 103 | 91 | 106 | 0 | 0 | 1 | 0.00000 | |
| BR9 | 106 | 115 | 103 | 118 | 0 | 0 | 1 | 0.00000 | |
| BR10 | 118 | 127 | 115 | 130 | 0 | 0 | 1 | 0.00000 | |
| BR11 | 130 | 139 | 127 | 142 | 0 | 0 | 1 | 0.00000 | |
| BR12 | 142 | 192 | 139 | 195 | 0 | 0 | 1 | 0.00000 | KFLW mainstem |
| BR13 | 195 | 273 | 192 | 276 | 0 | 0 | 1 | 0.00000 | SBMW mainstem |
| BR14 | 276 | 317 | 273 | 0 | 0 | 0 | 1 | 0.00000 | GCGW mainstem |
| BR15 | 320 | 334 | 0 | 144 | 0 | 0 | 1 | 0.00000 | Kettle R |
| BR16 | 337 | 340 | 0 | 147 | 0 | 0 | 1 | 0.00000 | Marcus Flats |
| BR17 | 343 | 346 | 0 | 154 | 0 | 0 | 1 | 0.00000 | Colville R |
| BR18 | 349 | 360 | 0 | 363 | 0 | 0 | 1 | 0.00067 | Spokane R |
| BR19 | 363 | 374 | 360 | 377 | 0 | 0 | 1 | 0.00067 | |
| BR20 | 377 | 388 | 374 | 391 | 0 | 0 | 1 | 0.00067 | |
| BR21 | 391 | 402 | 388 | 405 | 0 | 0 | 1 | 0.00067 | |
| BR22 | 405 | 416 | 402 | 419 | 0 | 0 | 1 | 0.00067 | |
| BR23 | 419 | 544 | 416 | 248 | 0 | 0 | 1 | 0.00000 | |
| BR24 | 547 | 554 | 0 | 255 | 0 | 0 | 1 | 0.00000 | Hawk Cr |
| BR25 | 557 | 582 | 0 | 289 | 0 | 0 | 1 | 0.00000 | Sanpoil R |

| LOCATION | LAT | LONG | EBOT | BS | BE | JBDN |
|----------|---------|---------|---------|----|----|------|
| WB 1 | 48.9667 | 117.659 | 261.000 | 1 | 1 | 1 |
| WB 2 | 48.9367 | 117.735 | 261.000 | 2 | 2 | 2 |
| WB 3 | 48.9163 | 117.786 | 261.000 | 3 | 3 | 3 |
| WB 4 | 48.8854 | 117.833 | 261.000 | 4 | 4 | 4 |
| WB 5 | 48.8609 | 117.885 | 261.000 | 5 | 5 | 5 |
| WB 6 | 48.8292 | 117.926 | 261.000 | 6 | 6 | 6 |
| WB 7 | 48.8132 | 117.976 | 261.000 | 7 | 7 | 7 |
| WB 8 | 48.7869 | 118.010 | 261.000 | 8 | 8 | 8 |
| WB 9 | 48.7541 | 118.047 | 261.000 | 9 | 9 | 9 |
| WB 10 | 48.7145 | 118.039 | 261.000 | 10 | 10 | 10 |
| WB 11 | 48.6772 | 118.037 | 261.000 | 11 | 11 | 11 |
| WB 12 | 48.5078 | 118.168 | 261.000 | 12 | 12 | 12 |
| WB 13 | 48.0380 | 118.372 | 261.000 | 13 | 13 | 13 |
| WB 14 | 47.9281 | 118.821 | 261.000 | 14 | 14 | 14 |
| WB 15 | 48.7071 | 118.118 | 261.000 | 15 | 15 | 15 |
| WB 16 | 48.6229 | 118.084 | 261.000 | 16 | 16 | 16 |
| WB 17 | 48.5710 | 118.098 | 261.000 | 17 | 17 | 17 |
| WB 18 | 47.8240 | 117.934 | 261.000 | 18 | 18 | 18 |
| WB 19 | 47.8305 | 117.971 | 261.000 | 19 | 19 | 19 |
| WB 20 | 47.8213 | 118.002 | 261.000 | 20 | 20 | 20 |
| WB 21 | 47.8012 | 118.023 | 261.000 | 21 | 21 | 21 |
| WB 22 | 47.7957 | 118.060 | 261.000 | 22 | 22 | 22 |
| WB 23 | 47.9002 | 118.170 | 261.000 | 23 | 23 | 23 |
| WB 24 | 47.8143 | 118.349 | 261.000 | 24 | 24 | 24 |
| WB 25 | 48.0142 | 118.670 | 261.000 | 25 | 25 | 25 |

| INIT CND | T2I | ICEI | WTYPEC |
|----------|---------|---------|--------|
| WB 1 | -2.0000 | 0.00000 | FRESH |
| WB 2 | -2.0000 | 0.00000 | FRESH |
| WB 3 | -2.0000 | 0.00000 | FRESH |
| WB 4 | -2.0000 | 0.00000 | FRESH |
| WB 5 | -2.0000 | 0.00000 | FRESH |
| WB 6 | -2.0000 | 0.00000 | FRESH |
| WB 7 | -2.0000 | 0.00000 | FRESH |
| WB 8 | -2.0000 | 0.00000 | FRESH |
| WB 9 | -2.0000 | 0.00000 | FRESH |
| WB 10 | -2.0000 | 0.00000 | FRESH |
| WB 11 | -2.0000 | 0.00000 | FRESH |
| WB 12 | -2.0000 | 0.00000 | FRESH |
| WB 13 | -2.0000 | 0.00000 | FRESH |
| WB 14 | -2.0000 | 0.00000 | FRESH |
| WB 15 | -2.0000 | 0.00000 | FRESH |
| WB 16 | -2.0000 | 0.00000 | FRESH |
| WB 17 | -2.0000 | 0.00000 | FRESH |
| WB 18 | -2.0000 | 0.00000 | FRESH |
| WB 19 | -2.0000 | 0.00000 | FRESH |
| WB 20 | -2.0000 | 0.00000 | FRESH |

| | | | |
|-------|---------|---------|-------|
| WB 21 | -2.0000 | 0.00000 | FRESH |
| WB 22 | -2.0000 | 0.00000 | FRESH |
| WB 23 | -2.0000 | 0.00000 | FRESH |
| WB 24 | -2.0000 | 0.00000 | FRESH |
| WB 25 | -2.0000 | 0.00000 | FRESH |

| CALCULAT | VBC | EBC | MBC | PQC | EVC | PRC |
|----------|-----|-----|-----|-----|-----|-----|
| WB 1 | ON | ON | ON | OFF | ON | OFF |
| WB 2 | ON | ON | ON | OFF | ON | OFF |
| WB 3 | ON | ON | ON | OFF | ON | OFF |
| WB 4 | ON | ON | ON | OFF | ON | OFF |
| WB 5 | ON | ON | ON | OFF | ON | OFF |
| WB 6 | ON | ON | ON | OFF | ON | OFF |
| WB 7 | ON | ON | ON | OFF | ON | OFF |
| WB 8 | ON | ON | ON | OFF | ON | OFF |
| WB 9 | ON | ON | ON | OFF | ON | OFF |
| WB 10 | ON | ON | ON | OFF | ON | OFF |
| WB 11 | ON | ON | ON | OFF | ON | OFF |
| WB 12 | ON | ON | ON | OFF | ON | OFF |
| WB 13 | ON | ON | ON | OFF | ON | OFF |
| WB 14 | ON | ON | ON | OFF | ON | OFF |
| WB 15 | ON | ON | ON | OFF | ON | OFF |
| WB 16 | ON | ON | ON | OFF | ON | OFF |
| WB 17 | ON | ON | ON | OFF | ON | OFF |
| WB 18 | ON | ON | ON | OFF | ON | OFF |
| WB 19 | ON | ON | ON | OFF | ON | OFF |
| WB 20 | ON | ON | ON | OFF | ON | OFF |
| WB 21 | ON | ON | ON | OFF | ON | OFF |
| WB 22 | ON | ON | ON | OFF | ON | OFF |
| WB 23 | ON | ON | ON | OFF | ON | OFF |
| WB 24 | ON | ON | ON | OFF | ON | OFF |
| WB 25 | ON | ON | ON | OFF | ON | OFF |

| DEAD SEA | WINDC | QINC | QOUTC | HEATC |
|----------|-------|------|-------|-------|
| WB 1 | ON | ON | ON | ON |
| WB 2 | ON | ON | ON | ON |
| WB 3 | ON | ON | ON | ON |
| WB 4 | ON | ON | ON | ON |
| WB 5 | ON | ON | ON | ON |
| WB 6 | ON | ON | ON | ON |
| WB 7 | ON | ON | ON | ON |
| WB 8 | ON | ON | ON | ON |
| WB 9 | ON | ON | ON | ON |
| WB 10 | ON | ON | ON | ON |
| WB 11 | ON | ON | ON | ON |
| WB 12 | ON | ON | ON | ON |
| WB 13 | ON | ON | ON | ON |
| WB 14 | ON | ON | ON | ON |
| WB 15 | ON | ON | ON | ON |
| WB 16 | ON | ON | ON | ON |
| WB 17 | ON | ON | ON | ON |
| WB 18 | ON | ON | ON | ON |
| WB 19 | ON | ON | ON | ON |
| WB 20 | ON | ON | ON | ON |
| WB 21 | ON | ON | ON | ON |
| WB 22 | ON | ON | ON | ON |
| WB 23 | ON | ON | ON | ON |
| WB 24 | ON | ON | ON | ON |
| WB 25 | ON | ON | ON | ON |

| INTERPOL | QINIC | DTRIC | HDIC |
|----------|-------|-------|------|
| BR1 | ON | ON | ON |
| BR2 | ON | ON | ON |
| BR3 | ON | ON | ON |
| BR4 | ON | ON | ON |
| BR5 | ON | ON | ON |
| BR6 | ON | ON | ON |
| BR7 | ON | ON | ON |
| BR8 | ON | ON | ON |
| BR9 | ON | ON | ON |
| BR10 | ON | ON | ON |
| BR11 | ON | ON | ON |
| BR12 | ON | ON | ON |
| BR13 | ON | ON | ON |
| BR14 | ON | ON | ON |
| BR15 | ON | ON | ON |
| BR16 | ON | ON | ON |
| BR17 | ON | ON | ON |
| BR18 | ON | ON | ON |
| BR19 | ON | ON | ON |
| BR20 | ON | ON | ON |

| | | | |
|------|----|----|----|
| BR21 | ON | ON | ON |
| BR22 | ON | ON | ON |
| BR23 | ON | ON | ON |
| BR24 | ON | ON | ON |
| BR25 | ON | ON | ON |

| HEAT EXCH | SLHTC | SROC | RHEVAP | METIC | FETCHC | AFW | BFW | CFW | WINDH | IWIND |
|-----------|-------|------|--------|-------|--------|---------|---------|---------|---------|-------|
| WB 1 | TERM | ON | OFF | ON | OFF | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 2 | TERM | ON | OFF | ON | OFF | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 3 | TERM | ON | OFF | ON | OFF | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 4 | TERM | ON | OFF | ON | OFF | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 5 | TERM | ON | OFF | ON | OFF | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 6 | TERM | ON | OFF | ON | OFF | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 7 | TERM | ON | OFF | ON | OFF | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 8 | TERM | ON | OFF | ON | OFF | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 9 | TERM | ON | OFF | ON | OFF | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 10 | TERM | ON | OFF | ON | OFF | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 11 | TERM | ON | OFF | ON | OFF | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 12 | TERM | ON | OFF | ON | ON | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 13 | TERM | ON | OFF | ON | ON | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 14 | TERM | ON | OFF | ON | ON | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 15 | TERM | ON | OFF | ON | OFF | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 16 | TERM | ON | OFF | ON | OFF | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 17 | TERM | ON | OFF | ON | OFF | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 18 | TERM | ON | OFF | ON | OFF | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 19 | TERM | ON | OFF | ON | OFF | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 20 | TERM | ON | OFF | ON | OFF | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 21 | TERM | ON | OFF | ON | OFF | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 22 | TERM | ON | OFF | ON | OFF | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 23 | TERM | ON | OFF | ON | ON | 9.20000 | 0.46000 | 2.00000 | 2.00000 | -1 |
| WB 24 | TERM | ON | OFF | ON | OFF | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |
| WB 25 | TERM | ON | OFF | ON | ON | 9.20000 | 0.46000 | 2.00000 | 2.00000 | 0 |

| ICE COVE | ICEC | SLICEC | ALBEDO | HWICE | BICE | GICE | ICEMIN | ICET2 |
|----------|------|--------|---------|---------|---------|---------|---------|---------|
| WB 1 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 2 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 3 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 4 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 5 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 6 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 7 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 8 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 9 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 10 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 11 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 12 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 13 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 14 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 15 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 16 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 17 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 18 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 19 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 20 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 21 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 22 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 23 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 24 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |
| WB 25 | OFF | DETAIL | 0.25000 | 10.0000 | 0.60000 | 0.07000 | 0.05000 | 3.00000 |

| TRANSPOR | SLTRC | THETA |
|----------|----------|---------|
| WB 1 | ULTIMATE | 0.55000 |
| WB 2 | ULTIMATE | 0.55000 |
| WB 3 | ULTIMATE | 0.55000 |
| WB 4 | ULTIMATE | 0.55000 |
| WB 5 | ULTIMATE | 0.55000 |
| WB 6 | ULTIMATE | 0.55000 |
| WB 7 | ULTIMATE | 0.55000 |
| WB 8 | ULTIMATE | 0.55000 |
| WB 9 | ULTIMATE | 0.55000 |
| WB 10 | ULTIMATE | 0.55000 |
| WB 11 | ULTIMATE | 0.55000 |
| WB 12 | ULTIMATE | 0.55000 |
| WB 13 | ULTIMATE | 0.55000 |
| WB 14 | ULTIMATE | 0.55000 |
| WB 15 | ULTIMATE | 0.55000 |
| WB 16 | ULTIMATE | 0.55000 |
| WB 17 | ULTIMATE | 0.55000 |
| WB 18 | ULTIMATE | 0.55000 |
| WB 19 | ULTIMATE | 0.55000 |
| WB 20 | ULTIMATE | 0.55000 |

WB 21 ULTIMATE 0.55000
 WB 22 ULTIMATE 0.55000
 WB 23 ULTIMATE 0.55000
 WB 24 ULTIMATE 0.55000
 WB 25 ULTIMATE 0.55000

| HYD COEF | AX | DX | CBHE | TSED | FI | TSEDF | FRICC |
|----------|---------|---------|---------|---------|---------|---------|-------|
| WB 1 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 2 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 3 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 4 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 5 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 6 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 7 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 8 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 9 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 10 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 11 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 12 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 13 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 14 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 15 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 16 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 17 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 18 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 19 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 20 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 21 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 22 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 23 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 24 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |
| WB 25 | 1.00000 | 1.00000 | 0.30000 | 11.5000 | 0.01000 | 1.00000 | MANN |

| EDDY VISC | AZC | AZSLC | AZMAX | PHISET |
|-----------|-----|-------|---------|--------|
| WB 1 | W2N | IMP | 1.00000 | 0.0 |
| WB 2 | W2N | IMP | 1.00000 | 0.0 |
| WB 3 | W2N | IMP | 1.00000 | 0.0 |
| WB 4 | W2N | IMP | 1.00000 | 0.0 |
| WB 5 | W2N | IMP | 1.00000 | 0.0 |
| WB 6 | W2N | IMP | 1.00000 | 0.0 |
| WB 7 | W2N | IMP | 1.00000 | 0.0 |
| WB 8 | W2N | IMP | 1.00000 | 0.0 |
| WB 9 | W2N | IMP | 1.00000 | 0.0 |
| WB 10 | W2N | IMP | 1.00000 | 0.0 |
| WB 11 | W2N | IMP | 1.00000 | 0.0 |
| WB 12 | W2 | IMP | 1.00000 | 0.0 |
| WB 13 | W2 | IMP | 1.00000 | 0.0 |
| WB 14 | W2 | IMP | 1.00000 | 0.0 |
| WB 15 | W2 | IMP | 1.00000 | 0.0 |
| WB 16 | W2 | IMP | 1.00000 | 0.0 |
| WB 17 | W2 | IMP | 1.00000 | 0.0 |
| WB 18 | W2N | IMP | 0.01000 | 0.0 |
| WB 19 | W2N | IMP | 0.01000 | 0.0 |
| WB 20 | W2N | IMP | 0.01000 | 0.0 |
| WB 21 | W2N | IMP | 0.01000 | 0.0 |
| WB 22 | W2N | IMP | 0.01000 | 0.0 |
| WB 23 | W2 | IMP | 0.01000 | 0.0 |
| WB 24 | W2 | IMP | 1.00000 | 0.0 |
| WB 25 | W2 | IMP | 1.00000 | 0.0 |

| N STRUC | NSTR |
|---------|------|
| BR1 | 0 |
| BR2 | 0 |
| BR3 | 0 |
| BR4 | 0 |
| BR5 | 0 |
| BR6 | 0 |
| BR7 | 0 |
| BR8 | 0 |
| BR9 | 0 |
| BR10 | 0 |
| BR11 | 0 |
| BR12 | 0 |
| BR13 | 0 |
| BR14 | 4 |
| BR15 | 0 |
| BR16 | 0 |
| BR17 | 0 |
| BR18 | 0 |
| BR19 | 0 |
| BR20 | 0 |

BR21 0
 BR22 0
 BR23 0
 BR24 0
 BR25 0

| STR INT | STRIC | STRIC | STRIC | STRIC | STRIC | STRIC | STRIC | STRIC | STRIC | STRIC |
|---------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| BR 1 | | | | | | | | | | |
| BR 2 | | | | | | | | | | |
| BR 3 | | | | | | | | | | |
| BR 4 | | | | | | | | | | |
| BR 5 | | | | | | | | | | |
| BR 6 | | | | | | | | | | |
| BR 7 | | | | | | | | | | |
| BR 8 | | | | | | | | | | |
| BR 9 | | | | | | | | | | |
| BR 10 | | | | | | | | | | |
| BR 11 | | | | | | | | | | |
| BR 12 | | | | | | | | | | |
| BR 13 | | | | | | | | | | |
| BR 14 | ON | ON | ON | ON | | | | | | |
| BR 15 | | | | | | | | | | |
| BR 16 | | | | | | | | | | |
| BR 17 | | | | | | | | | | |
| BR 18 | | | | | | | | | | |
| BR 19 | | | | | | | | | | |
| BR 20 | | | | | | | | | | |
| BR 21 | | | | | | | | | | |
| BR 22 | | | | | | | | | | |
| BR 23 | | | | | | | | | | |
| BR 24 | | | | | | | | | | |
| BR 25 | | | | | | | | | | |

| STR TOP | KTSTR | KTSTR | KTSTR | KTSTR | KTSTR | KTSTR | KTSTR | KTSTR | KTSTR | KTSTR |
|---------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| BR1 | | | | | | | | | | |
| BR2 | | | | | | | | | | |
| BR3 | | | | | | | | | | |
| BR4 | | | | | | | | | | |
| BR5 | | | | | | | | | | |
| BR6 | | | | | | | | | | |
| BR7 | | | | | | | | | | |
| BR8 | | | | | | | | | | |
| BR9 | | | | | | | | | | |
| BR10 | | | | | | | | | | |
| BR11 | | | | | | | | | | |
| BR12 | | | | | | | | | | |
| BR13 | | | | | | | | | | |
| BR14 | 2 | 2 | 2 | 2 | | | | | | |
| BR15 | | | | | | | | | | |
| BR16 | | | | | | | | | | |
| BR17 | | | | | | | | | | |
| BR18 | | | | | | | | | | |
| BR19 | | | | | | | | | | |
| BR20 | | | | | | | | | | |
| BR21 | | | | | | | | | | |
| BR22 | | | | | | | | | | |
| BR23 | | | | | | | | | | |
| BR24 | | | | | | | | | | |
| BR25 | | | | | | | | | | |

| STR BOT | KBSTR | KBSTR | KBSTR | KBSTR | KBSTR | KBSTR | KBSTR | KBSTR | KBSTR | KBSTR |
|---------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| BR1 | | | | | | | | | | |
| BR2 | | | | | | | | | | |
| BR3 | | | | | | | | | | |
| BR4 | | | | | | | | | | |
| BR5 | | | | | | | | | | |
| BR6 | | | | | | | | | | |
| BR7 | | | | | | | | | | |
| BR8 | | | | | | | | | | |
| BR9 | | | | | | | | | | |
| BR10 | | | | | | | | | | |
| BR11 | | | | | | | | | | |
| BR12 | | | | | | | | | | |
| BR13 | | | | | | | | | | |
| BR14 | 70 | 36 | 70 | 70 | | | | | | |
| BR15 | | | | | | | | | | |
| BR16 | | | | | | | | | | |
| BR17 | | | | | | | | | | |
| BR18 | | | | | | | | | | |
| BR19 | | | | | | | | | | |
| BR20 | | | | | | | | | | |

BR21
 BR22
 BR23
 BR24
 BR25

| STR | SINK | SINKC | SINKC | SINKC | SINKC | SINKC | SINKC | SINKC | SINKC | SINKC |
|------|------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| BR1 | | | | | | | | | | |
| BR2 | | | | | | | | | | |
| BR3 | | | | | | | | | | |
| BR4 | | | | | | | | | | |
| BR5 | | | | | | | | | | |
| BR6 | | | | | | | | | | |
| BR7 | | | | | | | | | | |
| BR8 | | | | | | | | | | |
| BR9 | | | | | | | | | | |
| BR10 | | | | | | | | | | |
| BR11 | | | | | | | | | | |
| BR12 | | | | | | | | | | |
| BR13 | | | | | | | | | | |
| BR14 | | LINE | LINE | LINE | LINE | | | | | |
| BR15 | | | | | | | | | | |
| BR16 | | | | | | | | | | |
| BR17 | | | | | | | | | | |
| BR18 | | | | | | | | | | |
| BR19 | | | | | | | | | | |
| BR20 | | | | | | | | | | |
| BR21 | | | | | | | | | | |
| BR22 | | | | | | | | | | |
| BR23 | | | | | | | | | | |
| BR24 | | | | | | | | | | |
| BR25 | | | | | | | | | | |

| STR | ELEV | ESTR | ESTR | ESTR | ESTR | ESTR | ESTR | ESTR | ESTR | ESTR |
|------|------|---------|---------|---------|---------|------|------|------|------|------|
| BR1 | | | | | | | | | | |
| BR2 | | | | | | | | | | |
| BR3 | | | | | | | | | | |
| BR4 | | | | | | | | | | |
| BR5 | | | | | | | | | | |
| BR6 | | | | | | | | | | |
| BR7 | | | | | | | | | | |
| BR8 | | | | | | | | | | |
| BR9 | | | | | | | | | | |
| BR10 | | | | | | | | | | |
| BR11 | | | | | | | | | | |
| BR12 | | | | | | | | | | |
| BR13 | | | | | | | | | | |
| BR14 | | 317.300 | 364.400 | 350.500 | 320.000 | | | | | |
| BR15 | | | | | | | | | | |
| BR16 | | | | | | | | | | |
| BR17 | | | | | | | | | | |
| BR18 | | | | | | | | | | |
| BR19 | | | | | | | | | | |
| BR20 | | | | | | | | | | |
| BR21 | | | | | | | | | | |
| BR22 | | | | | | | | | | |
| BR23 | | | | | | | | | | |
| BR24 | | | | | | | | | | |
| BR25 | | | | | | | | | | |

| STR | WIDT | WSTR | WSTR | WSTR | WSTR | WSTR | WSTR | WSTR | WSTR | WSTR |
|------|------|---------|--------|---------|---------|------|------|------|------|------|
| BR1 | | | | | | | | | | |
| BR2 | | | | | | | | | | |
| BR3 | | | | | | | | | | |
| BR4 | | | | | | | | | | |
| BR5 | | | | | | | | | | |
| BR6 | | | | | | | | | | |
| BR7 | | | | | | | | | | |
| BR8 | | | | | | | | | | |
| BR9 | | | | | | | | | | |
| BR10 | | | | | | | | | | |
| BR11 | | | | | | | | | | |
| BR12 | | | | | | | | | | |
| BR13 | | | | | | | | | | |
| BR14 | | 500.000 | 60.000 | 100.000 | 100.000 | | | | | |
| BR15 | | | | | | | | | | |
| BR16 | | | | | | | | | | |
| BR17 | | | | | | | | | | |
| BR18 | | | | | | | | | | |
| BR19 | | | | | | | | | | |
| BR20 | | | | | | | | | | |

BR21
BR22
BR23
BR24
BR25

| | | | | | | | | | | |
|---------------------|----------------|------------------|-------------------|-------------------|-------------------|-----------------|-----------------|--------|---------------|-------|
| PIPES | IUPI | IDPI | EUPI | EDPI | WPI | DLXPI | FPI | FMINPI | WTHLC | |
| PIPE UP | PUPIC | ETUPI | EBUPI | KTUPI | KBUPI | | | | | |
| PIPE DOWN | PDPIC | ETDPI | EBDPI | KTDPI | KBDPI | | | | | |
| SPILLWAY SP 2 | IUSP 317 | IDSP 0 | ESP 395.200 | A1SP 2000.00 | B1SP 1.50000 | A2SP 900.000 | B2SP 1.50000 | | WTHLC DOWN | |
| SPILL UP SP 2 | PUSPC DISTR | ETUSP 0.00000 | EBUSP 0.00000 | KTUSP 2 | KBUSP 70 | | | | | |
| SPILL DOWN SP 2 | PDSPC DISTR | ETUSP 0.00000 | EBUSP 0.00000 | KTDSP 2 | KBDSP 70 | | | | | |
| SPILL GAS SP 2 | GASSPC OFF | EQSP 0 | AGASSP 0.00000 | BGASSP 0.00000 | CGASSP 0.00000 | | | | | |
| GATES | IUGT | IDGT | EGT | A1GT | B1GT | G1GT | A2GT | B2GT | G2GT | WTHLC |
| GATE WEIR | GTA1 | GTB1 | GTA2 | GTB2 | DYNVAR | | | | | |
| GATE UP | PUGTC | ETUGT | EBUGT | KTUGT | KBUGT | | | | | |
| GATE DOWN | PDGTC | ETDGT | EBDGT | KTDGT | KBDGT | | | | | |
| GATE GAS | GASGTC | EQGT | AGASGT | BGASGT | CGASGT | | | | | |
| PUMPS 1 | IUPU | IDPU | EPU | STRTPU | ENDPU | EONPU | EOFFPU | QPU | WTHLC | |
| PUMPS 2 | PPUC | ETPU | EBPU | KTPU | KBPU | | | | | |
| WEIR SEG | IWR | IWR | IWR | IWR | IWR | IWR | IWR | IWR | IWR | |
| WEIR TOP | KTWR | KTWR | KTWR | KTWR | KTWR | KTWR | KTWR | KTWR | KTWR | |
| WEIR BOT | KBWR | KBWR | KBWR | KBWR | KBWR | KBWR | KBWR | KBWR | KBWR | |
| WD INT | WDIC ON | WDIC | WDIC | WDIC | WDIC | WDIC | WDIC | WDIC | WDIC | |
| WD SEG | IWD 317 | IWD | IWD | IWD | IWD | IWD | IWD | IWD | IWD | |
| WD ELEV | EWD 363.65 | EWD | EWD | EWD | EWD | EWD | EWD | EWD | EWD | |
| WD TOP | KTWD 2 | KTWD | KTWD | KTWD | KTWD | KTWD | KTWD | KTWD | KTWD | |
| WD BOT | KBWD 70 | KBWD | KBWD | KBWD | KBWD | KBWD | KBWD | KBWD | KBWD | |
| TRIB PLA SPECIFY | PTRC DISTR | PTRC DISTR | PTRC SPECIFY | PTRC | PTRC | PTRC | PTRC | PTRC | PTRC | |
| TRIB INT | TRIC ON | TRIC ON | TRIC | TRIC | TRIC | TRIC | TRIC | TRIC | TRIC | |
| TRIB SEG | ITR 316 | ITR 144 | ITR 316 | ITR | ITR | ITR | ITR | ITR | ITR | |

| | | | | | | | | | | |
|-----------|------------------|------------------|--------|--------|--------|--------|--------|--------|--------|--------|
| TRIB TOP | ELTRT 368.000 | ELTRT 368.000 | ELTRT | ELTRT | ELTRT | ELTRT | ELTRT | ELTRT | ELTRT | ELTRT |
| TRIB BOT | ELTRB 359.300 | ELTRB 359.300 | ELTRB | ELTRB | ELTRB | ELTRB | ELTRB | ELTRB | ELTRB | ELTRB |
| DST TRIB | DTRC | DTRC | DTRC | DTRC | DTRC | DTRC | DTRC | DTRC | DTRC | DTRC |
| BR 1 | ON | | | | | | | | | |
| BR 2 | ON | | | | | | | | | |
| BR 3 | ON | | | | | | | | | |
| BR 4 | ON | | | | | | | | | |
| BR 5 | ON | | | | | | | | | |
| BR 6 | ON | | | | | | | | | |
| BR 7 | ON | | | | | | | | | |
| BR 8 | ON | | | | | | | | | |
| BR 9 | ON | | | | | | | | | |
| BR 10 | ON | | | | | | | | | |
| BR 11 | ON | | | | | | | | | |
| BR 12 | ON | | | | | | | | | |
| BR 13 | ON | | | | | | | | | |
| BR 14 | ON | | | | | | | | | |
| BR 15 | OFF | | | | | | | | | |
| BR 16 | OFF | | | | | | | | | |
| BR 17 | OFF | | | | | | | | | |
| BR 18 | OFF | | | | | | | | | |
| BR 19 | OFF | | | | | | | | | |
| BR 20 | OFF | | | | | | | | | |
| BR 21 | OFF | | | | | | | | | |
| BR 22 | OFF | | | | | | | | | |
| BR 23 | OFF | | | | | | | | | |
| BR 24 | OFF | | | | | | | | | |
| BR 25 | OFF | | | | | | | | | |
| HYD PRIN | HPRWBC | HPRWBC | HPRWBC | HPRWBC | HPRWBC | HPRWBC | HPRWBC | HPRWBC | HPRWBC | HPRWBC |
| NVIOL | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| U | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| W | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| T | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| RHO | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| AZ | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| SHEAR | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| ST | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| SB | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| ADMX | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| DM | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| HDG | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| ADMZ | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| HPG | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| GRAV | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| SNP PRINT | SNPC | NSNP | NISNP | | | | | | | |

| | | | |
|-------|----|---|----|
| WB 1 | ON | 4 | 18 |
| WB 2 | ON | 4 | 10 |
| WB 3 | ON | 4 | 10 |
| WB 4 | ON | 4 | 10 |
| WB 5 | ON | 4 | 10 |
| WB 6 | ON | 2 | 3 |
| WB 7 | ON | 2 | 3 |
| WB 8 | ON | 2 | 3 |
| WB 9 | ON | 2 | 3 |
| WB 10 | ON | 2 | 3 |
| WB 11 | ON | 2 | 3 |
| WB 12 | ON | 2 | 13 |
| WB 13 | ON | 4 | 27 |
| WB 14 | ON | 2 | 11 |
| WB 15 | ON | 3 | 15 |
| WB 16 | ON | 2 | 2 |
| WB 17 | ON | 2 | 2 |
| WB 18 | ON | 2 | 2 |
| WB 19 | ON | 2 | 2 |
| WB 20 | ON | 2 | 2 |
| WB 21 | ON | 2 | 2 |
| WB 22 | ON | 2 | 2 |
| WB 23 | ON | 2 | 14 |
| WB 24 | ON | 2 | 2 |
| WB 25 | ON | 2 | 9 |

| SNP DATE | SNPD | SNPD | SNPD | SNPD | SNPD | SNPD | SNPD | SNPD | SNPD |
|----------|---------|---------|--------|-------|------|------|------|------|------|
| WB 1 | 1.00000 | 366.000 | 590.0 | 640.0 | | | | | |
| WB 2 | 1.00000 | 366.000 | 590.0 | 640.0 | | | | | |
| WB 3 | 1.00000 | 366.000 | 590.0 | 640.0 | | | | | |
| WB 4 | 1.00000 | 366.000 | 590.0 | 640.0 | | | | | |
| WB 5 | 1.00000 | 366.000 | 590.0 | 640.0 | | | | | |
| WB 6 | 1.00000 | 366.000 | | | | | | | |
| WB 7 | 1.00000 | 366.000 | | | | | | | |
| WB 8 | 1.00000 | 366.000 | | | | | | | |
| WB 9 | 1.00000 | 366.000 | | | | | | | |
| WB 10 | 1.00000 | 366.000 | | | | | | | |
| WB 11 | 1.00000 | 366.000 | | | | | | | |
| WB 12 | 1.00000 | 366.000 | | | | | | | |
| WB 13 | 1.00000 | 366.000 | 415.00 | 460.0 | | | | | |
| WB 14 | 1.00000 | 366.000 | | | | | | | |
| WB 15 | 1.00000 | 366.000 | 370.00 | | | | | | |
| WB 16 | 1.00000 | 366.000 | | | | | | | |
| WB 17 | 1.00000 | 366.000 | | | | | | | |
| WB 18 | 1.00000 | 366.000 | | | | | | | |
| WB 19 | 1.00000 | 366.000 | | | | | | | |
| WB 20 | 1.00000 | 366.000 | | | | | | | |
| WB 21 | 1.00000 | 366.000 | | | | | | | |
| WB 22 | 1.00000 | 366.000 | | | | | | | |
| WB 23 | 1.00000 | 366.000 | | | | | | | |
| WB 24 | 1.00000 | 366.000 | | | | | | | |
| WB 25 | 1.00000 | 366.000 | | | | | | | |

| SNP FREQ | SNPF | SNPF | SNPF | SNPF | SNPF | SNPF | SNPF | SNPF | SNPF |
|----------|---------|---------|---------|------|------|------|------|------|------|
| WB 1 | 2.00000 | 2.50000 | 0.250 | 2.00 | | | | | |
| WB 2 | 2.00000 | 2.50000 | 0.250 | 2.00 | | | | | |
| WB 3 | 2.00000 | 2.50000 | 0.250 | 2.00 | | | | | |
| WB 4 | 2.00000 | 2.50000 | 0.250 | 2.00 | | | | | |
| WB 5 | 2.00000 | 2.50000 | 0.250 | 2.00 | | | | | |
| WB 6 | 2.00000 | 2.50000 | | | | | | | |
| WB 7 | 2.00000 | 2.50000 | | | | | | | |
| WB 8 | 2.00000 | 2.50000 | | | | | | | |
| WB 9 | 2.00000 | 2.50000 | | | | | | | |
| WB 10 | 2.00000 | 2.50000 | | | | | | | |
| WB 11 | 2.00000 | 2.50000 | | | | | | | |
| WB 12 | 2.00000 | 2.50000 | | | | | | | |
| WB 13 | 2.00000 | 2.50000 | 0.500 | 2.50 | | | | | |
| WB 14 | 2.00000 | 2.50000 | | | | | | | |
| WB 15 | 2.00000 | 0.10000 | 2.50000 | | | | | | |
| WB 16 | 2.00000 | 2.50000 | | | | | | | |
| WB 17 | 2.00000 | 2.50000 | | | | | | | |
| WB 18 | 2.00000 | 2.50000 | | | | | | | |
| WB 19 | 2.00000 | 2.50000 | | | | | | | |
| WB 20 | 2.00000 | 2.50000 | | | | | | | |
| WB 21 | 2.00000 | 2.50000 | | | | | | | |
| WB 22 | 2.00000 | 2.50000 | | | | | | | |
| WB 23 | 2.00000 | 2.50000 | | | | | | | |
| WB 24 | 2.00000 | 2.50000 | | | | | | | |
| WB 25 | 2.00000 | 2.50000 | | | | | | | |

| SNP SEG | ISNP | ISNP | ISNP | ISNP | ISNP | ISNP | ISNP | ISNP | ISNP |
|---------|------|------|------|------|------|------|------|------|------|
|---------|------|------|------|------|------|------|------|------|------|

| | | | | | | | | | |
|-------|-----|-----|-----|-----|-----|-----|-----|-----|-----|
| WB 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 |
| WB 2 | 22 | 23 | 24 | 25 | 26 | 27 | 28 | 29 | 30 |
| | 31 | | | | | | | | |
| WB 3 | 34 | 35 | 36 | 37 | 38 | 39 | 40 | 41 | 42 |
| | 43 | | | | | | | | |
| WB 4 | 46 | 47 | 48 | 49 | 50 | 51 | 52 | 53 | 54 |
| | 55 | | | | | | | | |
| WB 5 | 58 | 59 | 60 | 61 | 62 | 63 | 64 | 65 | 66 |
| | 67 | | | | | | | | |
| WB 6 | 70 | 75 | 79 | | | | | | |
| WB 7 | 82 | 85 | 91 | | | | | | |
| WB 8 | 94 | 98 | 103 | | | | | | |
| WB 9 | 106 | 110 | 115 | | | | | | |
| WB 10 | 118 | 122 | 127 | | | | | | |
| WB 11 | 130 | 135 | 139 | | | | | | |
| WB 12 | 142 | 143 | 144 | 145 | 146 | 147 | 148 | 149 | 150 |
| | 160 | 170 | 180 | 192 | | | | | |
| WB 13 | 195 | 210 | 220 | 221 | 222 | 223 | 224 | 225 | 226 |
| | 227 | 228 | 229 | 230 | 231 | 232 | 233 | 234 | 235 |
| | 236 | 237 | 238 | 239 | 240 | 250 | 260 | 270 | 273 |
| WB 14 | 276 | 290 | 300 | 310 | 311 | 312 | 313 | 314 | 315 |
| | 316 | 317 | | | | | | | |
| WB 15 | 320 | 321 | 322 | 323 | 324 | 325 | 326 | 327 | 328 |
| | 329 | 330 | 331 | 332 | 333 | 334 | | | |
| WB 16 | 337 | 340 | | | | | | | |
| WB 17 | 343 | 346 | | | | | | | |
| WB 18 | 349 | 360 | | | | | | | |
| WB 19 | 363 | 374 | | | | | | | |
| WB 20 | 377 | 388 | | | | | | | |
| WB 21 | 391 | 402 | | | | | | | |
| WB 22 | 405 | 416 | | | | | | | |
| WB 23 | 419 | 430 | 440 | 450 | 460 | 470 | 480 | 490 | 500 |
| | 510 | 520 | 530 | 540 | 544 | | | | |
| WB 24 | 547 | 554 | | | | | | | |
| WB 25 | 557 | 560 | 565 | 570 | 575 | 576 | 577 | 580 | 582 |

| SCR PRINT | SCRC | NSCR | | | | | | | |
|-----------|------|------|--|--|--|--|--|--|--|
| WB 1 | OFF | 1 | | | | | | | |
| WB 2 | OFF | 1 | | | | | | | |
| WB 3 | OFF | 1 | | | | | | | |
| WB 4 | OFF | 1 | | | | | | | |
| WB 5 | OFF | 1 | | | | | | | |
| WB 6 | OFF | 1 | | | | | | | |
| WB 7 | OFF | 1 | | | | | | | |
| WB 8 | OFF | 1 | | | | | | | |
| WB 9 | OFF | 1 | | | | | | | |
| WB 10 | OFF | 1 | | | | | | | |
| WB 11 | OFF | 1 | | | | | | | |
| WB 12 | OFF | 1 | | | | | | | |
| WB 13 | OFF | 1 | | | | | | | |
| WB 14 | ON | 2 | | | | | | | |
| WB 15 | OFF | 1 | | | | | | | |
| WB 16 | OFF | 1 | | | | | | | |
| WB 17 | OFF | 1 | | | | | | | |
| WB 18 | OFF | 1 | | | | | | | |
| WB 19 | OFF | 1 | | | | | | | |
| WB 20 | OFF | 1 | | | | | | | |
| WB 21 | OFF | 1 | | | | | | | |
| WB 22 | OFF | 1 | | | | | | | |
| WB 23 | OFF | 1 | | | | | | | |
| WB 24 | OFF | 1 | | | | | | | |
| WB 25 | OFF | 1 | | | | | | | |

| SCR DATE | SCRD | SCRD | SCRD | SCRD | SCRD | SCRD | SCRD | SCRD | SCRD |
|----------|---------|---------|------|------|------|------|------|------|------|
| WB 1 | 108.000 | | | | | | | | |
| WB 2 | 108.000 | | | | | | | | |
| WB 3 | 108.000 | | | | | | | | |
| WB 4 | 108.000 | | | | | | | | |
| WB 5 | 108.000 | | | | | | | | |
| WB 6 | 108.000 | | | | | | | | |
| WB 7 | 108.000 | | | | | | | | |
| WB 8 | 108.000 | | | | | | | | |
| WB 9 | 108.000 | | | | | | | | |
| WB 10 | 108.000 | | | | | | | | |
| WB 11 | 108.000 | | | | | | | | |
| WB 12 | 108.000 | | | | | | | | |
| WB 13 | 108.000 | | | | | | | | |
| WB 14 | 1.00000 | 366.000 | | | | | | | |
| WB 15 | 108.000 | | | | | | | | |
| WB 16 | 108.000 | | | | | | | | |

| | |
|-------|---------|
| WB 17 | 108.000 |
| WB 18 | 108.000 |
| WB 19 | 108.000 |
| WB 20 | 108.000 |
| WB 21 | 108.000 |
| WB 22 | 108.000 |
| WB 23 | 108.000 |
| WB 24 | 108.000 |
| WB 25 | 108.000 |

| SCR | FREQ | SCR | SCR | SCR | SCR | SCR | SCR | SCR | SCR | SCR |
|-------|---------|---------|-----|-----|-----|-----|-----|-----|-----|-----|
| WB 1 | 0.00010 | | | | | | | | | |
| WB 2 | 0.00010 | | | | | | | | | |
| WB 3 | 0.00010 | | | | | | | | | |
| WB 4 | 0.00010 | | | | | | | | | |
| WB 5 | 0.00010 | | | | | | | | | |
| WB 6 | 0.00010 | | | | | | | | | |
| WB 7 | 0.00010 | | | | | | | | | |
| WB 8 | 0.10000 | | | | | | | | | |
| WB 9 | 0.00010 | | | | | | | | | |
| WB 10 | 0.00010 | | | | | | | | | |
| WB 11 | 0.00010 | | | | | | | | | |
| WB 12 | 0.00010 | | | | | | | | | |
| WB 13 | 0.00010 | | | | | | | | | |
| WB 14 | 0.10000 | 0.10000 | | | | | | | | |
| WB 15 | 0.00010 | | | | | | | | | |
| WB 16 | 0.00010 | | | | | | | | | |
| WB 17 | 0.00010 | | | | | | | | | |
| WB 18 | 0.00010 | | | | | | | | | |
| WB 19 | 0.00010 | | | | | | | | | |
| WB 20 | 0.00010 | | | | | | | | | |
| WB 21 | 0.00010 | | | | | | | | | |
| WB 22 | 0.00010 | | | | | | | | | |
| WB 23 | 0.00010 | | | | | | | | | |
| WB 24 | 0.00010 | | | | | | | | | |
| WB 25 | 0.00010 | | | | | | | | | |

| PRF | PLOT | PRFC | NPRF | NIPRF |
|-------|------|------|------|-------|
| WB 1 | OFF | | 2 | 1 |
| WB 2 | OFF | | 2 | 1 |
| WB 3 | OFF | | 2 | 1 |
| WB 4 | OFF | | 2 | 1 |
| WB 5 | OFF | | 2 | 1 |
| WB 6 | OFF | | 2 | 1 |
| WB 7 | OFF | | 2 | 1 |
| WB 8 | OFF | | 2 | 1 |
| WB 9 | OFF | | 2 | 1 |
| WB 10 | OFF | | 2 | 1 |
| WB 11 | ON | | 2 | 1 |
| WB 12 | ON | | 2 | 2 |
| WB 13 | ON | | 2 | 3 |
| WB 14 | ON | | 2 | 3 |
| WB 15 | OFF | | 2 | 1 |
| WB 16 | OFF | | 2 | 1 |
| WB 17 | OFF | | 2 | 1 |
| WB 18 | OFF | | 2 | 1 |
| WB 19 | OFF | | 2 | 1 |
| WB 20 | OFF | | 2 | 1 |
| WB 21 | OFF | | 2 | 1 |
| WB 22 | OFF | | 2 | 1 |
| WB 23 | ON | | 2 | 1 |
| WB 24 | ON | | 2 | 1 |
| WB 25 | ON | | 2 | 2 |

| PRF | DATE | PRFD | PRFD | PRFD | PRFD | PRFD | PRFD | PRFD | PRFD |
|-------|---------|---------|------|------|------|------|------|------|------|
| WB 1 | 1.50000 | 366.500 | | | | | | | |
| WB 2 | 1.50000 | 366.500 | | | | | | | |
| WB 3 | 1.50000 | 366.500 | | | | | | | |
| WB 4 | 1.50000 | 366.500 | | | | | | | |
| WB 5 | 1.50000 | 366.500 | | | | | | | |
| WB 6 | 1.50000 | 366.500 | | | | | | | |
| WB 7 | 1.50000 | 366.500 | | | | | | | |
| WB 8 | 1.50000 | 366.500 | | | | | | | |
| WB 9 | 1.50000 | 366.500 | | | | | | | |
| WB 10 | 1.50000 | 366.500 | | | | | | | |
| WB 11 | 1.50000 | 366.500 | | | | | | | |
| WB 12 | 1.50000 | 366.500 | | | | | | | |
| WB 13 | 1.50000 | 366.500 | | | | | | | |
| WB 14 | 1.50000 | 366.500 | | | | | | | |
| WB 15 | 1.50000 | 366.500 | | | | | | | |
| WB 16 | 1.50000 | 366.500 | | | | | | | |

| | | |
|-------|---------|---------|
| WB 17 | 1.50000 | 366.500 |
| WB 18 | 1.50000 | 366.500 |
| WB 19 | 1.50000 | 366.500 |
| WB 20 | 1.50000 | 366.500 |
| WB 21 | 1.50000 | 366.500 |
| WB 22 | 1.50000 | 366.500 |
| WB 23 | 1.50000 | 366.500 |
| WB 24 | 1.50000 | 366.500 |
| WB 25 | 1.50000 | 366.500 |

| PRF | FREQ | PRFF | PRFF | PRFF | PRFF | PRFF | PRFF | PRFF | PRFF | PRFF |
|-------|------|---------|---------|------|------|------|------|------|------|------|
| WB 1 | | 1.00000 | 1.00000 | | | | | | | |
| WB 2 | | 1.00000 | 1.00000 | | | | | | | |
| WB 3 | | 1.00000 | 1.00000 | | | | | | | |
| WB 4 | | 1.00000 | 1.00000 | | | | | | | |
| WB 5 | | 1.00000 | 1.00000 | | | | | | | |
| WB 6 | | 1.00000 | 1.00000 | | | | | | | |
| WB 7 | | 1.00000 | 1.00000 | | | | | | | |
| WB 8 | | 1.00000 | 1.00000 | | | | | | | |
| WB 9 | | 1.00000 | 1.00000 | | | | | | | |
| WB 10 | | 1.00000 | 1.00000 | | | | | | | |
| WB 11 | | 1.00000 | 1.00000 | | | | | | | |
| WB 12 | | 1.00000 | 1.00000 | | | | | | | |
| WB 13 | | 1.00000 | 1.00000 | | | | | | | |
| WB 14 | | 1.00000 | 1.00000 | | | | | | | |
| WB 15 | | 1.00000 | 1.00000 | | | | | | | |
| WB 16 | | 1.00000 | 1.00000 | | | | | | | |
| WB 17 | | 1.00000 | 1.00000 | | | | | | | |
| WB 18 | | 1.00000 | 1.00000 | | | | | | | |
| WB 19 | | 1.00000 | 1.00000 | | | | | | | |
| WB 20 | | 1.00000 | 1.00000 | | | | | | | |
| WB 21 | | 1.00000 | 1.00000 | | | | | | | |
| WB 22 | | 1.00000 | 1.00000 | | | | | | | |
| WB 23 | | 1.00000 | 1.00000 | | | | | | | |
| WB 24 | | 1.00000 | 1.00000 | | | | | | | |
| WB 25 | | 1.00000 | 1.00000 | | | | | | | |

| PRF | SEG | IPRF | IPRF | IPRF | IPRF | IPRF | IPRF | IPRF | IPRF | IPRF |
|-------|-----|------|------|------|------|------|------|------|------|------|
| WB 1 | | 0 | | | | | | | | |
| WB 2 | | 0 | | | | | | | | |
| WB 3 | | 0 | | | | | | | | |
| WB 4 | | 0 | | | | | | | | |
| WB 5 | | 0 | | | | | | | | |
| WB 6 | | 0 | | | | | | | | |
| WB 7 | | 0 | | | | | | | | |
| WB 8 | | 0 | | | | | | | | |
| WB 9 | | 0 | | | | | | | | |
| WB 10 | | 0 | | | | | | | | |
| WB 11 | | 133 | | | | | | | | |
| WB 12 | | 151 | 188 | | | | | | | |
| WB 13 | | 211 | 247 | 252 | | | | | | |
| WB 14 | | 287 | 290 | 313 | | | | | | |
| WB 15 | | 0 | | | | | | | | |
| WB 16 | | 0 | | | | | | | | |
| WB 17 | | 344 | | | | | | | | |
| WB 18 | | 0 | | | | | | | | |
| WB 19 | | 0 | | | | | | | | |
| WB 20 | | 0 | | | | | | | | |
| WB 21 | | 0 | | | | | | | | |
| WB 22 | | 0 | | | | | | | | |
| WB 23 | | 475 | | | | | | | | |
| WB 24 | | 553 | | | | | | | | |
| WB 25 | | 558 | 576 | | | | | | | |

| SPR | PLOT | SPRC | NSPR | NISPR |
|-------|------|------|------|-------|
| WB 1 | | OFF | 0 | 0 |
| WB 2 | | OFF | 0 | 0 |
| WB 3 | | OFF | 0 | 0 |
| WB 4 | | OFF | 0 | 0 |
| WB 5 | | OFF | 0 | 0 |
| WB 6 | | OFF | 0 | 0 |
| WB 7 | | OFF | 0 | 0 |
| WB 8 | | OFF | 0 | 0 |
| WB 9 | | OFF | 0 | 0 |
| WB 10 | | OFF | 0 | 0 |
| WB 11 | | OFF | 0 | 0 |
| WB 12 | | OFF | 0 | 0 |
| WB 13 | | OFF | 14 | 3 |
| WB 14 | | OFF | 0 | 0 |
| WB 15 | | ON | 1 | 14 |
| WB 16 | | OFF | 0 | 0 |

| | | | |
|-------|-----|---|---|
| WB 17 | OFF | 0 | 0 |
| WB 18 | OFF | 0 | 0 |
| WB 19 | OFF | 0 | 0 |
| WB 20 | OFF | 0 | 0 |
| WB 21 | OFF | 0 | 0 |
| WB 22 | OFF | 0 | 0 |
| WB 23 | OFF | 0 | 0 |
| WB 24 | OFF | 0 | 0 |
| WB 25 | OFF | 0 | 0 |

| SPR DATE | SPRD | SPRD | SPRD | SPRD | SPRD | SPRD | SPRD | SPRD | SPRD |
|----------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| WB 1 | | | | | | | | | |
| WB 2 | | | | | | | | | |
| WB 3 | | | | | | | | | |
| WB 4 | | | | | | | | | |
| WB 5 | | | | | | | | | |
| WB 6 | | | | | | | | | |
| WB 7 | | | | | | | | | |
| WB 8 | | | | | | | | | |
| WB 9 | | | | | | | | | |
| WB 10 | | | | | | | | | |
| WB 11 | | | | | | | | | |
| WB 12 | | | | | | | | | |
| WB 13 | 448.500 | 475.500 | 491.500 | 502.500 | 524.500 | 536.500 | 554.500 | 572.500 | 587.500 |
| WB 14 | 599.500 | 622.500 | 634.500 | 650.500 | 663.500 | | | | |
| WB 15 | 368.000 | | | | | | | | |
| WB 16 | | | | | | | | | |
| WB 17 | | | | | | | | | |
| WB 18 | | | | | | | | | |
| WB 19 | | | | | | | | | |
| WB 20 | | | | | | | | | |
| WB 21 | | | | | | | | | |
| WB 22 | | | | | | | | | |
| WB 23 | | | | | | | | | |
| WB 24 | | | | | | | | | |
| WB 25 | | | | | | | | | |

| SPR FREQ | SPRF | SPRF | SPRF | SPRF | SPRF | SPRF | SPRF | SPRF | SPRF |
|----------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| WB 1 | | | | | | | | | |
| WB 2 | | | | | | | | | |
| WB 3 | | | | | | | | | |
| WB 4 | | | | | | | | | |
| WB 5 | | | | | | | | | |
| WB 6 | | | | | | | | | |
| WB 7 | | | | | | | | | |
| WB 8 | | | | | | | | | |
| WB 9 | | | | | | | | | |
| WB 10 | | | | | | | | | |
| WB 11 | | | | | | | | | |
| WB 12 | | | | | | | | | |
| WB 13 | 500.000 | 500.000 | 500.000 | 500.000 | 500.000 | 500.000 | 500.000 | 500.000 | 500.000 |
| WB 14 | 500.000 | 500.000 | 500.000 | 500.000 | 500.000 | | | | |
| WB 15 | 2.00 | | | | | | | | |
| WB 16 | | | | | | | | | |
| WB 17 | | | | | | | | | |
| WB 18 | | | | | | | | | |
| WB 19 | | | | | | | | | |
| WB 20 | | | | | | | | | |
| WB 21 | | | | | | | | | |
| WB 22 | | | | | | | | | |
| WB 23 | | | | | | | | | |
| WB 24 | | | | | | | | | |
| WB 25 | | | | | | | | | |

| SPR SEG | ISPR | ISPR | ISPR | ISPR | ISPR | ISPR | ISPR | ISPR | ISPR |
|---------|------|------|------|------|------|------|------|------|------|
| WB 1 | | | | | | | | | |
| WB 2 | | | | | | | | | |
| WB 3 | | | | | | | | | |
| WB 4 | | | | | | | | | |
| WB 5 | | | | | | | | | |
| WB 6 | | | | | | | | | |
| WB 7 | | | | | | | | | |
| WB 8 | | | | | | | | | |
| WB 9 | | | | | | | | | |
| WB 10 | | | | | | | | | |
| WB 11 | | | | | | | | | |
| WB 12 | | | | | | | | | |
| WB 13 | 99 | 101 | 103 | | | | | | |
| WB 14 | | | | | | | | | |

| | | | | | | | | | |
|-------|-----|-----|-----|-----|-----|-----|-----|-----|-----|
| WB 15 | 320 | 321 | 322 | 323 | 324 | 325 | 326 | 327 | 328 |
| | 329 | 330 | 331 | 332 | 333 | | | | |
| WB 16 | | | | | | | | | |
| WB 17 | | | | | | | | | |
| WB 18 | | | | | | | | | |
| WB 19 | | | | | | | | | |
| WB 20 | | | | | | | | | |
| WB 21 | | | | | | | | | |
| WB 22 | | | | | | | | | |
| WB 23 | | | | | | | | | |
| WB 24 | | | | | | | | | |
| WB 25 | | | | | | | | | |

| VPL PLOT | VPLC | NVPL |
|----------|------|------|
| WB 1 | OFF | 1 |
| WB 2 | OFF | 1 |
| WB 3 | OFF | 1 |
| WB 4 | OFF | 1 |
| WB 5 | OFF | 1 |
| WB 6 | OFF | 1 |
| WB 7 | OFF | 1 |
| WB 8 | OFF | 1 |
| WB 9 | OFF | 1 |
| WB 10 | OFF | 1 |
| WB 11 | OFF | 1 |
| WB 12 | OFF | 1 |
| WB 13 | OFF | 1 |
| WB 14 | OFF | 1 |
| WB 15 | OFF | 1 |
| WB 16 | OFF | 1 |
| WB 17 | OFF | 1 |
| WB 18 | OFF | 1 |
| WB 19 | OFF | 1 |
| WB 20 | OFF | 1 |
| WB 21 | OFF | 1 |
| WB 22 | OFF | 1 |
| WB 23 | OFF | 1 |
| WB 24 | OFF | 1 |
| WB 25 | OFF | 1 |

| VPL DATE | VPLD | VPLD | VPLD | VPLD | VPLD | VPLD | VPLD | VPLD | VPLD |
|----------|---------|------|------|------|------|------|------|------|------|
| WB 1 | 63.5000 | | | | | | | | |
| WB 2 | 63.5000 | | | | | | | | |
| WB 3 | 63.5000 | | | | | | | | |
| WB 4 | 63.5000 | | | | | | | | |
| WB 5 | 63.5000 | | | | | | | | |
| WB 6 | 63.5000 | | | | | | | | |
| WB 7 | 63.5000 | | | | | | | | |
| WB 8 | 63.5000 | | | | | | | | |
| WB 9 | 63.5000 | | | | | | | | |
| WB 10 | 63.5000 | | | | | | | | |
| WB 11 | 63.5000 | | | | | | | | |
| WB 12 | 63.5000 | | | | | | | | |
| WB 13 | 63.5000 | | | | | | | | |
| WB 14 | 63.5000 | | | | | | | | |
| WB 15 | 63.5000 | | | | | | | | |
| WB 16 | 63.5000 | | | | | | | | |
| WB 17 | 63.5000 | | | | | | | | |
| WB 18 | 63.5000 | | | | | | | | |
| WB 19 | 63.5000 | | | | | | | | |
| WB 20 | 63.5000 | | | | | | | | |
| WB 21 | 63.5000 | | | | | | | | |
| WB 22 | 63.5000 | | | | | | | | |
| WB 23 | 63.5000 | | | | | | | | |
| WB 24 | 63.5000 | | | | | | | | |
| WB 25 | 63.5000 | | | | | | | | |

| VPL FREQ | VPLF | VPLF | VPLF | VPLF | VPLF | VPLF | VPLF | VPLF | VPLF |
|----------|---------|------|------|------|------|------|------|------|------|
| WB 1 | 1.00000 | | | | | | | | |
| WB 2 | 1.00000 | | | | | | | | |
| WB 3 | 1.00000 | | | | | | | | |
| WB 4 | 1.00000 | | | | | | | | |
| WB 5 | 1.00000 | | | | | | | | |
| WB 6 | 1.00000 | | | | | | | | |
| WB 7 | 1.00000 | | | | | | | | |
| WB 8 | 1.00000 | | | | | | | | |
| WB 9 | 1.00000 | | | | | | | | |
| WB 10 | 1.00000 | | | | | | | | |
| WB 11 | 1.00000 | | | | | | | | |
| WB 12 | 1.00000 | | | | | | | | |
| WB 13 | 1.00000 | | | | | | | | |

| | |
|-------|---------|
| WB 14 | 1.00000 |
| WB 15 | 1.00000 |
| WB 16 | 1.00000 |
| WB 17 | 1.00000 |
| WB 18 | 1.00000 |
| WB 19 | 1.00000 |
| WB 20 | 1.00000 |
| WB 21 | 1.00000 |
| WB 22 | 1.00000 |
| WB 23 | 1.00000 |
| WB 24 | 1.00000 |
| WB 25 | 1.00000 |

| CPL PLOT | CPLC | NCPL |
|----------|------|------|
| WB 1 | ON | 2 |
| WB 2 | ON | 2 |
| WB 3 | ON | 2 |
| WB 4 | ON | 2 |
| WB 5 | ON | 2 |
| WB 6 | ON | 2 |
| WB 7 | ON | 2 |
| WB 8 | ON | 2 |
| WB 9 | ON | 2 |
| WB 10 | ON | 2 |
| WB 11 | ON | 2 |
| WB 12 | ON | 2 |
| WB 13 | ON | 2 |
| WB 14 | ON | 2 |
| WB 15 | ON | 2 |
| WB 16 | ON | 2 |
| WB 17 | ON | 2 |
| WB 18 | ON | 2 |
| WB 19 | ON | 2 |
| WB 20 | ON | 2 |
| WB 21 | ON | 2 |
| WB 22 | ON | 2 |
| WB 23 | ON | 2 |
| WB 24 | ON | 2 |
| WB 25 | ON | 2 |

| CPL DATE | CPLD | CPLD | CPLD | CPLD | CPLD | CPLD | CPLD | CPLD | CPLD |
|----------|---------|---------|--------|--------|------|------|------|------|------|
| WB 1 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 2 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 3 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 4 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 5 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 6 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 7 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 8 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 9 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 10 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 11 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 12 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 13 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 14 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 15 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 16 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 17 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 18 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 19 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 20 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 21 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 22 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 23 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 24 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |
| WB 25 | 1.00000 | 365.000 | 366.00 | 600.00 | | | | | |

| CPL FREQ | CPLF | CPLF | CPLF | CPLF | CPLF | CPLF | CPLF | CPLF | CPLF |
|----------|---------|---------|-------|------|------|------|------|------|------|
| WB 1 | 1.00000 | 2.00000 | 0.100 | 2.00 | | | | | |
| WB 2 | 1.00000 | 2.00000 | 0.100 | 2.00 | | | | | |
| WB 3 | 1.00000 | 2.00000 | 0.100 | 2.00 | | | | | |
| WB 4 | 1.00000 | 2.00000 | 0.100 | 2.00 | | | | | |
| WB 5 | 1.00000 | 2.00000 | 0.100 | 2.00 | | | | | |
| WB 6 | 1.00000 | 2.00000 | 0.100 | 2.00 | | | | | |
| WB 7 | 1.00000 | 2.00000 | 0.100 | 2.00 | | | | | |
| WB 8 | 1.00000 | 2.00000 | 0.100 | 2.00 | | | | | |
| WB 9 | 1.00000 | 2.00000 | 0.100 | 2.00 | | | | | |
| WB 10 | 1.00000 | 2.00000 | 0.100 | 2.00 | | | | | |
| WB 11 | 1.00000 | 2.00000 | 0.100 | 2.00 | | | | | |
| WB 12 | 1.00000 | 2.00000 | 0.100 | 2.00 | | | | | |
| WB 13 | 1.00000 | 2.00000 | 0.100 | 2.00 | | | | | |

| | | | | |
|-------|---------|---------|-------|------|
| WB 14 | 1.00000 | 2.00000 | 0.100 | 2.00 |
| WB 15 | 1.00000 | 2.00000 | | |
| WB 16 | 1.00000 | 2.00000 | | |
| WB 17 | 1.00000 | 2.00000 | | |
| WB 18 | 1.00000 | 2.00000 | | |
| WB 19 | 1.00000 | 2.00000 | | |
| WB 20 | 1.00000 | 2.00000 | | |
| WB 21 | 1.00000 | 2.00000 | | |
| WB 22 | 1.00000 | 2.00000 | | |
| WB 23 | 1.00000 | 2.00000 | | |
| WB 24 | 1.00000 | 2.00000 | | |
| WB 25 | 1.00000 | 2.00000 | | |

| FLUXES | FLXC | NFLX |
|--------|------|------|
| WB 1 | OFF | 0 |
| WB 2 | OFF | 0 |
| WB 3 | OFF | 0 |
| WB 4 | OFF | 0 |
| WB 5 | OFF | 0 |
| WB 6 | OFF | 0 |
| WB 7 | OFF | 0 |
| WB 8 | OFF | 0 |
| WB 9 | OFF | 0 |
| WB 10 | OFF | 0 |
| WB 11 | OFF | 0 |
| WB 12 | OFF | 0 |
| WB 13 | OFF | 0 |
| WB 14 | OFF | 0 |
| WB 15 | OFF | 0 |
| WB 16 | OFF | 0 |
| WB 17 | OFF | 0 |
| WB 18 | OFF | 0 |
| WB 19 | OFF | 0 |
| WB 20 | OFF | 0 |
| WB 21 | OFF | 0 |
| WB 22 | OFF | 0 |
| WB 23 | OFF | 0 |
| WB 24 | OFF | 0 |
| WB 25 | OFF | 0 |

| FLX DATE | FLXD | FLXD | FLXD | FLXD | FLXD | FLXD | FLXD | FLXD | FLXD |
|----------|-------|------|------|------|------|------|------|------|------|
| WB 1 | 366.0 | | | | | | | | |
| WB 2 | 366.0 | | | | | | | | |
| WB 3 | 366.0 | | | | | | | | |
| WB 4 | 366.0 | | | | | | | | |
| WB 5 | 366.0 | | | | | | | | |
| WB 6 | 366.0 | | | | | | | | |
| WB 7 | 366.0 | | | | | | | | |
| WB 8 | 366.0 | | | | | | | | |
| WB 9 | 366.0 | | | | | | | | |
| WB 10 | 366.0 | | | | | | | | |
| WB 11 | 366.0 | | | | | | | | |
| WB 12 | 366.0 | | | | | | | | |
| WB 13 | 366.0 | | | | | | | | |
| WB 14 | 366.0 | | | | | | | | |
| WB 15 | 366.0 | | | | | | | | |
| WB 16 | 366.0 | | | | | | | | |
| WB 17 | 366.0 | | | | | | | | |
| WB 18 | 366.0 | | | | | | | | |
| WB 19 | 366.0 | | | | | | | | |
| WB 20 | 366.0 | | | | | | | | |
| WB 21 | 366.0 | | | | | | | | |
| WB 22 | 366.0 | | | | | | | | |
| WB 23 | 366.0 | | | | | | | | |
| WB 24 | 366.0 | | | | | | | | |
| WB 25 | 366.0 | | | | | | | | |

| FLX FREQ | FLXF | FLXF | FLXF | FLXF | FLXF | FLXF | FLXF | FLXF | FLXF |
|----------|------|------|------|------|------|------|------|------|------|
| WB 1 | 2.0 | | | | | | | | |
| WB 2 | 2.0 | | | | | | | | |
| WB 3 | 2.0 | | | | | | | | |
| WB 4 | 2.0 | | | | | | | | |
| WB 5 | 2.0 | | | | | | | | |
| WB 6 | 2.0 | | | | | | | | |
| WB 7 | 2.0 | | | | | | | | |
| WB 8 | 2.0 | | | | | | | | |
| WB 9 | 2.0 | | | | | | | | |
| WB 10 | 2.0 | | | | | | | | |
| WB 11 | 2.0 | | | | | | | | |
| WB 12 | 2.0 | | | | | | | | |
| WB 13 | 2.0 | | | | | | | | |

WB 14 2.0
 WB 15 2.0
 WB 16 2.0
 WB 17 2.0
 WB 18 2.0
 WB 19 2.0
 WB 20 2.0
 WB 21 2.0
 WB 22 2.0
 WB 23 2.0
 WB 24 2.0
 WB 25 2.0

TSR PLOT TSRC NISR NITSR
 ON 2 80

TSR DATE TSRD TSRD TSRD TSRD TSRD TSRD TSRD TSRD TSRD
 1.00000 366.000

TSR FREQ TSRF TSRF TSRF TSRF TSRF TSRF TSRF TSRF TSRF
 0.05000 0.05000

TSR SEG ITSR ITSR ITSR ITSR ITSR ITSR ITSR ITSR ITSR
 2 5 10 15 19 22 31 34 43
 46 55 58 67 70 79 82 91 94
 103 106 115 118 127 130 139 142 150
 160 170 180 192 195 210 220 223 230
 240 250 260 273 276 290 300 317 320
 334 337 340 343 346 349 360 363 374
 377 388 391 402 405 416 419 544 547
 554 557 582 317 475 313 133 151 188
 211 247 252 287 344 553 558 576

TSR LAYE ETSR ETSR ETSR ETSR ETSR ETSR ETSR ETSR ETSR
 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000
 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000
 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000
 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000
 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000
 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000
 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000
 0.00000 0.00000 0.00000 -18.000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000
 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000 0.00000

WITH OUT WDOC NWDO NIWDO
 ON 1 4

WITH DAT WDOD WDOD WDOD WDOD WDOD WDOD WDOD WDOD WDOD
 366.000 0.00000

WITH FRE WDOF WDOF WDOF WDOF WDOF WDOF WDOF WDOF WDOF
 0.10000 0.10000

WITH SEG IWDO IWDO IWDO IWDO IWDO IWDO IWDO IWDO IWDO
 192 273 317 544

RESTART RSOC NRSO RSIC
 ON 1 OFF

RSO DATE RSOD RSOD RSOD RSOD RSOD RSOD RSOD RSOD RSOD
 450.000 575.000

RSO FREQ RSOF RSOF RSOF RSOF RSOF RSOF RSOF RSOF RSOF
 100.000 500.000

CST COMP CCC LIMC CUF
 ON ON 12

CST ACTIVE CAC
 TDS ON
 Age ON
 Tracer ON
 Coliform ON
 ISS1 ON
 PO4 ON
 NH4 ON
 NO3 ON
 DSI OFF
 PSI OFF
 FE OFF
 LDOM ON

| | |
|--------|-----|
| RDOM | ON |
| LPOM | ON |
| RPOM | ON |
| BOD1 | ON |
| ALG1 | ON |
| ALG2 | ON |
| ALG3 | ON |
| DO | ON |
| TIC | ON |
| ALK | ON |
| ZOO1 | ON |
| ZOO2 | ON |
| ZOO3 | ON |
| LDOM_P | OFF |
| RDOM_P | OFF |
| LPOM_P | OFF |
| RPOM_P | OFF |
| LDOM_N | OFF |
| RDOM_N | OFF |
| LPOM_N | OFF |
| RPOM_N | OFF |

| | | | | | | | | | |
|----------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| CST DERI | CDWBC | CDWBC | CDWBC | CDWBC | CDWBC | CDWBC | CDWBC | CDWBC | CDWBC |
| DOC | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| POC | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| TOC | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| DON | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| PON | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| TON | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| TKN | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| TN | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| DOP | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| POP | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| TOP | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| TP | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| APR | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| CHLA | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| ATOT | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| %DO | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| TSS | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| TISS | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| CBOD | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| pH | ON | ON | ON | ON | ON | ON | ON | ON | ON |

| | | | | | | | | | | |
|----------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|
| | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| ALG2 | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| ALG3 | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| DO | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| TIC | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| ALK | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| | ON | ON | ON | ON | ON | ON | ON | ON | ON | ON |
| ZOO1 | ON | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | ON |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | ON |
| ZOO2 | ON | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | ON |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | ON |
| ZOO3 | ON | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | ON |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | ON |
| LDOMP | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| RDOMP | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| LPOMP | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| RPOMP | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| LDOMN | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| RDOMN | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| LPOMN | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| RPOMN | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| CTR CON | CTRTRC | CTRTRC | CTRTRC | CTRTRC | CTRTRC | CTRTRC | CTRTRC | CTRTRC | CTRTRC | CTRTRC |
| IDS | ON | ON | | | | | | | | |
| Age | ON | ON | | | | | | | | |
| Tracer | ON | ON | | | | | | | | |
| Coliform | ON | ON | | | | | | | | |
| ISS1 | ON | ON | | | | | | | | |
| PO4 | ON | ON | | | | | | | | |
| NH4 | ON | ON | | | | | | | | |
| NO3 | ON | ON | | | | | | | | |
| DSI | OFF | OFF | | | | | | | | |
| PSI | OFF | OFF | | | | | | | | |
| FE | OFF | OFF | | | | | | | | |
| LDOM | ON | ON | | | | | | | | |
| RDOM | ON | ON | | | | | | | | |
| LPOM | ON | ON | | | | | | | | |
| RPOM | ON | ON | | | | | | | | |
| BOD1 | ON | ON | | | | | | | | |
| ALG1 | ON | ON | | | | | | | | |
| ALG2 | ON | ON | | | | | | | | |
| ALG3 | ON | ON | | | | | | | | |
| DO | ON | ON | | | | | | | | |
| TIC | ON | ON | | | | | | | | |
| ALK | ON | ON | | | | | | | | |
| ZOO1 | OFF | OFF | | | | | | | | |
| ZOO2 | OFF | OFF | | | | | | | | |
| ZOO3 | OFF | OFF | | | | | | | | |
| LDOM_P | OFF | OFF | | | | | | | | |
| RDOM_P | OFF | OFF | | | | | | | | |
| LPOM_P | OFF | OFF | | | | | | | | |
| RPOM_P | OFF | OFF | | | | | | | | |
| LDOM_N | OFF | OFF | | | | | | | | |

| | | | | | | | | | |
|----------|---------|---------|---------|---------|-----|------|-----|-----|-----|
| DO | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| TIC | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| ALK | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| ZOO1 | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| ZOO2 | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| ZOO3 | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| LDOMP | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| RDOMP | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| LPOMP | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| RPOMP | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| LDOMN | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| RDOMN | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| LPOMN | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| RPOMN | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| EX COEF | EXH2O | EXSS | EXOM | BETA | EXC | EXIC | | | |
| WB 1 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 2 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 3 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 4 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 5 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 6 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 7 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 8 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 9 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 10 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 11 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 12 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 13 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 14 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 15 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 16 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 17 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 18 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 19 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 20 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 21 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 22 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 23 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 24 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| WB 25 | 0.25000 | 0.10000 | 0.10000 | 0.45000 | OFF | OFF | | | |
| ALG EX | EXA | EXA | EXA | EXA | EXA | EXA | EXA | EXA | EXA |
| | 0.20 | 0.20 | 0.20 | | | | | | |
| ZOO EX | EXZ | EXZ | EXZ | EXZ | EXZ | EXZ | EXZ | EXZ | EXZ |
| | 0.2 | 0.2 | 0.2 | | | | | | |
| MACRO EX | EXM | EXM | EXM | EXM | EXM | EXM | | | |
| | 0.0100 | | | | | | | | |
| GENERIC | CGQ10 | CG0DK | CG1DK | CGS | | | | | |

| | | | | | | | | | | |
|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------------|
| Age | 0.00 | -1.0 | 0.0 | 0.0 | | | | | | |
| Tracer | 0.00 | 0.0 | 0.0 | 0.0 | | | | | | |
| Colform | 1.04 | 0.0 | 0.5 | 0.0 | | | | | | |
| S SOLIDS | SSS | SEDRC | TAUCR | | | | | | | |
| SS# 1 | 1.000 | OFF | 1.0E-05 | | | | | | | |
| ALGAL RATE | AG | AR | AE | AM | AS | AHSP | AHSN | AHSSI | ASAT | |
| ALG1 | 2.50000 | 0.0400 | 0.04000 | 0.08000 | 0.50000 | 0.00200 | 0.01400 | 0.00000 | 85.0000 | diatoms |
| ALG2 | 2.50000 | 0.0400 | 0.04000 | 0.10000 | 0.05000 | 0.00300 | 0.01400 | 0.00000 | 36.0000 | greens |
| ALG3 | 1.00000 | 0.1000 | 0.04000 | 0.08000 | 0.10000 | 0.00300 | 0.01400 | 0.00000 | 75.0000 | cyanobacteria |
| ALGAL TEMP | AT1 | AT2 | AT3 | AT4 | AK1 | AK2 | AK3 | AK4 | | |
| ALG1 | 1.00000 | 10.0000 | 20.0000 | 28.0000 | 0.20000 | 0.99000 | 0.99000 | 0.10000 | | |
| ALG2 | 1.00000 | 10.0000 | 20.0000 | 25.0000 | 0.20000 | 0.99000 | 0.99000 | 0.10000 | | |
| ALG3 | 10.0000 | 33.0000 | 35.0000 | 40.0000 | 0.05000 | 0.99000 | 0.99000 | 0.10000 | | |
| ALG STOI | ALGP | ALGN | ALGC | ALGSI | ACHLA | ALPOM | ANEQN | ANPR | | |
| ALG1 | 0.00300 | 0.03400 | 0.45000 | 0.32000 | 150.0 | 0.50000 | 2 | 0.00100 | | |
| ALG2 | 0.00200 | 0.04000 | 0.53500 | 0.00000 | 150.0 | 0.50000 | 2 | 0.00100 | | |
| ALG3 | 0.00800 | 0.08000 | 0.47500 | 0.00000 | 150.0 | 0.50000 | 2 | 0.00100 | | |
| EPIPHYTE | EPIC | EPIC | EPIC | EPIC | EPIC | EPIC | EPIC | EPIC | EPIC | |
| EPI1 | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | |
| EPI PRIN | EPRC | EPRC | EPRC | EPRC | EPRC | EPRC | EPRC | EPRC | EPRC | |
| EPI1 | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | |
| EPI INIT | EPICI | EPICI | EPICI | EPICI | EPICI | EPICI | EPICI | EPICI | EPICI | |
| EPI1 | 0.00300 | 0.00300 | 0.00300 | 0.00300 | 0.00300 | 0.00300 | 0.00300 | 0.00300 | 0.00300 | |
| | 0.00300 | 0.00300 | 0.00300 | 0.00500 | 0.01000 | 0.00300 | 0.00300 | 0.00300 | 0.01000 | |
| | 0.01000 | 0.01000 | 0.01000 | 0.01000 | 0.01000 | 0.00500 | 0.00500 | 0.00500 | | |
| EPI RATE | EG | ER | EE | EM | EB | EHSP | EHSN | EHSSI | | |
| EPI1 | 2.00000 | 0.0200 | 0.04000 | 0.05000 | 0.00100 | 0.00300 | 0.01400 | 0.00000 | | |
| EPI HALF | ESAT | EHS | ENEQN | ENPR | | | | | | |
| EPI1 | 85.0000 | 15.0000 | 2 | 0.00100 | | | | | | |
| EPI TEMP | ET1 | ET2 | ET3 | ET4 | EK1 | EK2 | EK3 | EK4 | | |
| EPI1 | 8.00000 | 10.0000 | 20.0000 | 30.0000 | 0.10000 | 0.99000 | 0.99000 | 0.10000 | | |
| EPI STOI | EP | EN | EC | ESI | ECHLA | EPOM | | | | |
| EPI1 | 0.01000 | 0.04000 | 0.50000 | 0.00000 | 100.000 | 0.50000 | | | | |
| ZOOP RATE | ZMAX | ZRESP | ZMORT | ZEFFIC | PREFP | ZOOMIN | ZS2P | | | |
| | 2.50 | 0.100 | 0.10 | 0.60 | 0.10 | 0.0010 | 0.30 | | | |
| | 2.50 | 0.100 | 0.10 | 0.60 | 0.10 | 0.0010 | 0.50 | | | |
| | 2.50 | 0.100 | 0.10 | 0.60 | 0.10 | 0.0010 | 0.30 | | | |
| ZOOP ALGP | PREFA | PREFA | PREFA | PREFA | PREFA | PREFA | PREFA | PREFA | PREFA | |
| Zoo1 | 0.35 | 0.55 | 0.00 | | | | | | | |
| Zoo2 | 0.30 | 0.40 | 0.00 | | | | | | | |
| Zoo3 | 0.35 | 0.55 | 0.00 | | | | | | | |
| ZOOP ZOOP | PREFZ | PREFZ | PREFZ | PREFZ | PREFZ | PREFZ | PREFZ | PREFZ | PREFZ | |
| Zoo1 | 0.00 | 0.00 | 0.00 | | | | | | | |
| Zoo2 | 0.10 | 0.00 | 0.10 | | | | | | | |
| Zoo3 | 0.00 | 0.00 | 0.00 | | | | | | | |
| ZOOP TEMP | ZOOT1 | ZOOT2 | ZOOT3 | ZOOT4 | ZOOK1 | ZOOK2 | ZOOK3 | ZOOK4 | | |
| | 1.0 | 10.0 | 25.0 | 36.0 | 0.3 | 0.9 | 0.98 | 0.200 | | |
| | 1.0 | 10.0 | 25.0 | 36.0 | 0.3 | 0.9 | 0.98 | 0.200 | | |
| | 1.0 | 10.0 | 25.0 | 36.0 | 0.3 | 0.9 | 0.98 | 0.200 | | |
| ZOOP STO | ZP | ZN | ZC | | | | | | | |
| | 0.00700 | 0.08000 | 0.45000 | | | | | | | |
| | 0.00700 | 0.08000 | 0.45000 | | | | | | | |
| | 0.00700 | 0.08000 | 0.45000 | | | | | | | |
| MACROPHYT | MACWBC | MACWBC | MACWBC | MACWBC | MACWBC | MACWBC | MACWBC | MACWBC | MACWBC | |
| Mac1 | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | |
| MAC PRINT | MPRWBC | MPRWBC | MPRWBC | MPRWBC | MPRWBC | MPRWBC | MPRWBC | MPRWBC | MPRWBC | |
| Mac1 | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | |

| | | | | | | | | | |
|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF | OFF |
| MAC INI | MACWBCI | MACWBCI | MACWBCI | MACWBCI | MACWBCI | MACWBCI | MACWBCI | MACWBCI | MACWBCI |
| Mac1 | 0.00001 | 0.00001 | 0.00001 | 0.00001 | 0.00001 | 0.00001 | 0.00001 | 0.00001 | 0.00001 |
| | 0.00001 | 0.00001 | 0.00001 | 0.00001 | 0.00001 | 0.00001 | 0.00001 | 0.00001 | 0.00001 |
| | 0.00001 | 0.00001 | 0.00001 | 0.00001 | 0.00001 | 0.00001 | 0.00001 | 0.00001 | 0.00001 |
| MAC RATE | MG | MR | MM | MSAT | MHSP | MHSN | MHSC | MPOM | LRPMAC |
| Mac 1 | 0.30 | 0.05 | 0.05 | 30.0 | 0.0 | 0.0 | 0.0 | 0.9 | 0.2 |
| MAC SED | PSED | NSED | | | | | | | |
| MAC 1 | 0.5 | 0.5 | | | | | | | |
| MAC DIST | MBMP | MMAX | | | | | | | |
| Mac 1 | 40.0 | 500.0 | | | | | | | |
| MAC DRAG | CDSTEM | DWV | DMSA | ANORM | | | | | |
| Mac 1 | 2.0 | 7e4 | 8.00 | 0.80 | | | | | |
| MAC TEMP | MT1 | MT2 | MT3 | MT4 | MK1 | MK2 | MK3 | MK4 | |
| Mac 1 | 7.0 | 15.0 | 24.0 | 34.0 | 0.1 | 0.99 | 0.99 | 0.01 | |
| MAC STOICH | MP | MN | MC | | | | | | |
| Mac 1 | 0.005 | 0.08 | 0.45 | | | | | | |
| DOM | LDOMDK | RDOMDK | LRDDK | | | | | | |
| WB 1 | 0.01000 | 0.00100 | 0.01000 | | | | | | |
| WB 2 | 0.01000 | 0.00100 | 0.01000 | | | | | | |
| WB 3 | 0.01000 | 0.00100 | 0.01000 | | | | | | |
| WB 4 | 0.01000 | 0.00100 | 0.01000 | | | | | | |
| WB 5 | 0.01000 | 0.00100 | 0.01000 | | | | | | |
| WB 6 | 0.01000 | 0.00100 | 0.01000 | | | | | | |
| WB 7 | 0.01000 | 0.00100 | 0.01000 | | | | | | |
| WB 8 | 0.01000 | 0.00100 | 0.01000 | | | | | | |
| WB 9 | 0.01000 | 0.00100 | 0.01000 | | | | | | |
| WB 10 | 0.01000 | 0.00100 | 0.01000 | | | | | | |
| WB 11 | 0.01000 | 0.00100 | 0.01000 | | | | | | |
| WB 12 | 0.01000 | 0.00100 | 0.01000 | | | | | | |
| WB 13 | 0.01000 | 0.00100 | 0.01000 | | | | | | |
| WB 14 | 0.01000 | 0.00100 | 0.01000 | | | | | | |
| WB 15 | 0.01000 | 0.00100 | 0.01000 | | | | | | |
| WB 16 | 0.01000 | 0.00100 | 0.01000 | | | | | | |
| WB 17 | 0.01000 | 0.00100 | 0.01000 | | | | | | |
| WB 18 | 0.10000 | 0.00100 | 0.01000 | | | | | | |
| WB 19 | 0.10000 | 0.00100 | 0.01000 | | | | | | |
| WB 20 | 0.10000 | 0.00100 | 0.01000 | | | | | | |
| WB 21 | 0.10000 | 0.00100 | 0.01000 | | | | | | |
| WB 22 | 0.10000 | 0.00100 | 0.01000 | | | | | | |
| WB 23 | 0.10000 | 0.00100 | 0.01000 | | | | | | |
| WB 24 | 0.01000 | 0.00100 | 0.01000 | | | | | | |
| WB 25 | 0.01000 | 0.00100 | 0.01000 | | | | | | |
| POM | LPOMDK | RPOMDK | LRPDK | POMS | | | | | |
| WB 1 | 0.05000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 2 | 0.05000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 3 | 0.05000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 4 | 0.05000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 5 | 0.05000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 6 | 0.05000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 7 | 0.05000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 8 | 0.05000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 9 | 0.05000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 10 | 0.05000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 11 | 0.05000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 12 | 0.05000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 13 | 0.05000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 14 | 0.05000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 15 | 0.05000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 16 | 0.05000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 17 | 0.05000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 18 | 0.25000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 19 | 0.25000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 20 | 0.25000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 21 | 0.25000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 22 | 0.25000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 23 | 0.25000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 24 | 0.05000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| WB 25 | 0.05000 | 0.00500 | 0.01000 | 0.10000 | | | | | |
| OM STOIC | ORGP | ORGN | ORGC | ORGSI | | | | | |

| | | | | |
|-------|---------|---------|---------|---------|
| WB 1 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 2 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 3 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 4 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 5 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 6 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 7 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 8 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 9 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 10 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 11 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 12 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 13 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 14 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 15 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 16 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 17 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 18 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 19 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 20 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 21 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 22 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 23 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 24 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |
| WB 25 | 0.00500 | 0.05000 | 0.45000 | 0.18000 |

| OM RATE | OMT1 | OMT2 | OMK1 | OMK2 |
|---------|---------|---------|---------|---------|
| WB 1 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 2 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 3 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 4 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 5 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 6 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 7 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 8 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 9 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 10 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 11 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 12 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 13 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 14 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 15 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 16 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 17 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 18 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 19 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 20 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 21 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 22 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 23 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 24 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |
| WB 25 | 4.00000 | 30.0000 | 0.10000 | 0.99000 |

| CBOD | KBOD | TBOD | RBOD | CBODS |
|-------|---------|---------|---------|-------|
| BOD 1 | 0.25080 | 1.01470 | 1.85000 | 0.0 |

| CBOD STOIC | BODP | BODN | BODC |
|------------|---------|---------|---------|
| BOD 1 | 0.00400 | 0.06000 | 0.32000 |

| PHOSPHOR | PO4R | PARTP |
|----------|---------|---------|
| WB 1 | 0.00100 | 0.20000 |
| WB 2 | 0.00100 | 0.20000 |
| WB 3 | 0.00100 | 0.20000 |
| WB 4 | 0.00100 | 0.20000 |
| WB 5 | 0.00100 | 0.20000 |
| WB 6 | 0.00100 | 0.20000 |
| WB 7 | 0.00100 | 0.20000 |
| WB 8 | 0.00100 | 0.20000 |
| WB 9 | 0.00100 | 0.20000 |
| WB 10 | 0.00100 | 0.20000 |
| WB 11 | 0.00100 | 0.20000 |
| WB 12 | 0.00100 | 0.20000 |
| WB 13 | 0.00100 | 0.20000 |
| WB 14 | 0.00100 | 0.20000 |
| WB 15 | 0.00100 | 0.20000 |
| WB 16 | 0.00100 | 0.20000 |
| WB 17 | 0.00100 | 0.20000 |
| WB 18 | 0.00100 | 0.20000 |
| WB 19 | 0.00100 | 0.20000 |
| WB 20 | 0.00100 | 0.20000 |
| WB 21 | 0.00100 | 0.20000 |

| | | |
|-------|---------|---------|
| WB 22 | 0.00100 | 0.20000 |
| WB 23 | 0.00100 | 0.20000 |
| WB 24 | 0.00100 | 0.20000 |
| WB 25 | 0.00100 | 0.20000 |

| AMMONIUM | NH4R | NH4DK |
|----------|---------|---------|
| WB 1 | 0.00100 | 0.12000 |
| WB 2 | 0.00100 | 0.12000 |
| WB 3 | 0.00100 | 0.12000 |
| WB 4 | 0.00100 | 0.12000 |
| WB 5 | 0.00100 | 0.12000 |
| WB 6 | 0.00100 | 0.12000 |
| WB 7 | 0.00100 | 0.12000 |
| WB 8 | 0.00100 | 0.12000 |
| WB 9 | 0.00100 | 0.12000 |
| WB 10 | 0.00100 | 0.12000 |
| WB 11 | 0.00100 | 0.12000 |
| WB 12 | 0.00100 | 0.12000 |
| WB 13 | 0.00100 | 0.12000 |
| WB 14 | 0.00100 | 0.12000 |
| WB 15 | 0.00100 | 0.12000 |
| WB 16 | 0.00100 | 0.12000 |
| WB 17 | 0.00100 | 0.12000 |
| WB 18 | 0.00100 | 0.12000 |
| WB 19 | 0.00100 | 0.12000 |
| WB 20 | 0.00100 | 0.12000 |
| WB 21 | 0.00100 | 0.12000 |
| WB 22 | 0.00100 | 0.12000 |
| WB 23 | 0.00100 | 0.12000 |
| WB 24 | 0.00100 | 0.12000 |
| WB 25 | 0.00100 | 0.12000 |

| NH4 RATE | NH4T1 | NH4T2 | NH4K1 | NH4K2 |
|----------|---------|---------|---------|---------|
| WB 1 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 2 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 3 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 4 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 5 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 6 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 7 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 8 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 9 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 10 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 11 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 12 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 13 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 14 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 15 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 16 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 17 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 18 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 19 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 20 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 21 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 22 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 23 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 24 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 25 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |

| NITRATE | NO3DK | NO3S |
|---------|---------|---------|
| WB 1 | 0.03000 | 1.00000 |
| WB 2 | 0.03000 | 1.00000 |
| WB 3 | 0.03000 | 1.00000 |
| WB 4 | 0.03000 | 1.00000 |
| WB 5 | 0.03000 | 1.00000 |
| WB 6 | 0.03000 | 1.00000 |
| WB 7 | 0.03000 | 1.00000 |
| WB 8 | 0.03000 | 1.00000 |
| WB 9 | 0.03000 | 1.00000 |
| WB 10 | 0.03000 | 1.00000 |
| WB 11 | 0.03000 | 1.00000 |
| WB 12 | 0.03000 | 1.00000 |
| WB 13 | 0.03000 | 1.00000 |
| WB 14 | 0.03000 | 1.00000 |
| WB 15 | 0.03000 | 1.00000 |
| WB 16 | 0.03000 | 1.00000 |
| WB 17 | 0.03000 | 1.00000 |
| WB 18 | 0.03000 | 1.00000 |
| WB 19 | 0.03000 | 1.00000 |
| WB 20 | 0.03000 | 1.00000 |
| WB 21 | 0.03000 | 1.00000 |

| | | |
|-------|---------|---------|
| WB 22 | 0.03000 | 1.00000 |
| WB 23 | 0.03000 | 1.00000 |
| WB 24 | 0.03000 | 1.00000 |
| WB 25 | 0.03000 | 1.00000 |

| NO3 RATE | NO3T1 | NO3T2 | NO3K1 | NO3K2 |
|----------|---------|---------|---------|---------|
| WB 1 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 2 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 3 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 4 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 5 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 6 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 7 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 8 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 9 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 10 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 11 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 12 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 13 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 14 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 15 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 16 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 17 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 18 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 19 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 20 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 21 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 22 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 23 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 24 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |
| WB 25 | 5.00000 | 25.0000 | 0.10000 | 0.99000 |

| SILICA | DSIR | PSIS | PSIDK | PARTSI |
|--------|---------|---------|---------|---------|
| WB 1 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 2 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 3 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 4 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 5 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 6 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 7 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 8 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 9 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 10 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 11 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 12 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 13 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 14 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 15 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 16 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 17 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 18 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 19 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 20 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 21 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 22 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 23 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 24 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |
| WB 25 | 0.10000 | 0.00000 | 0.30000 | 0.20000 |

| IRON | FER | FES |
|-------|---------|---------|
| WB 1 | 0.05000 | 2.00000 |
| WB 2 | 0.05000 | 2.00000 |
| WB 3 | 0.05000 | 2.00000 |
| WB 4 | 0.05000 | 2.00000 |
| WB 5 | 0.05000 | 2.00000 |
| WB 6 | 0.05000 | 2.00000 |
| WB 7 | 0.05000 | 2.00000 |
| WB 8 | 0.05000 | 2.00000 |
| WB 9 | 0.05000 | 2.00000 |
| WB 10 | 0.05000 | 2.00000 |
| WB 11 | 0.05000 | 2.00000 |
| WB 12 | 0.05000 | 2.00000 |
| WB 13 | 0.05000 | 2.00000 |
| WB 14 | 0.05000 | 2.00000 |
| WB 15 | 0.05000 | 2.00000 |
| WB 16 | 0.05000 | 2.00000 |
| WB 17 | 0.05000 | 2.00000 |
| WB 18 | 0.05000 | 2.00000 |
| WB 19 | 0.05000 | 2.00000 |
| WB 20 | 0.05000 | 2.00000 |
| WB 21 | 0.05000 | 2.00000 |

| | | |
|-------|---------|---------|
| WB 22 | 0.05000 | 2.00000 |
| WB 23 | 0.05000 | 2.00000 |
| WB 24 | 0.05000 | 2.00000 |
| WB 25 | 0.05000 | 2.00000 |

| | |
|---------|---------|
| SED CO2 | CO2R |
| WB 1 | 0.10000 |
| WB 2 | 0.10000 |
| WB 3 | 0.10000 |
| WB 4 | 0.10000 |
| WB 5 | 0.10000 |
| WB 6 | 0.10000 |
| WB 7 | 0.10000 |
| WB 8 | 0.10000 |
| WB 9 | 0.10000 |
| WB 10 | 0.10000 |
| WB 11 | 0.10000 |
| WB 12 | 0.10000 |
| WB 13 | 0.10000 |
| WB 14 | 0.10000 |
| WB 15 | 0.10000 |
| WB 16 | 0.10000 |
| WB 17 | 0.10000 |
| WB 18 | 0.10000 |
| WB 19 | 0.10000 |
| WB 20 | 0.10000 |
| WB 21 | 0.10000 |
| WB 22 | 0.10000 |
| WB 23 | 0.10000 |
| WB 24 | 0.10000 |
| WB 25 | 0.10000 |

| | | |
|----------|---------|---------|
| STOICH 1 | O2NH4 | O2OM |
| WB 1 | 4.57000 | 1.40000 |
| WB 2 | 4.57000 | 1.40000 |
| WB 3 | 4.57000 | 1.40000 |
| WB 4 | 4.57000 | 1.40000 |
| WB 5 | 4.57000 | 1.40000 |
| WB 6 | 4.57000 | 1.40000 |
| WB 7 | 4.57000 | 1.40000 |
| WB 8 | 4.57000 | 1.40000 |
| WB 9 | 4.57000 | 1.40000 |
| WB 10 | 4.57000 | 1.40000 |
| WB 11 | 4.57000 | 1.40000 |
| WB 12 | 4.57000 | 1.40000 |
| WB 13 | 4.57000 | 1.40000 |
| WB 14 | 4.57000 | 1.40000 |
| WB 15 | 4.57000 | 1.40000 |
| WB 16 | 4.57000 | 1.40000 |
| WB 17 | 4.57000 | 1.40000 |
| WB 18 | 4.57000 | 1.40000 |
| WB 19 | 4.57000 | 1.40000 |
| WB 20 | 4.57000 | 1.40000 |
| WB 21 | 4.57000 | 1.40000 |
| WB 22 | 4.57000 | 1.40000 |
| WB 23 | 4.57000 | 1.40000 |
| WB 24 | 4.57000 | 1.40000 |
| WB 25 | 4.57000 | 1.40000 |

| | | |
|----------|---------|---------|
| STOICH 2 | O2AR | O2AG |
| ALG1 | 1.10000 | 1.40000 |
| ALG2 | 1.10000 | 1.40000 |
| ALG3 | 1.10000 | 1.40000 |

| | | |
|----------|---------|---------|
| STOICH 3 | O2ER | O2EG |
| EPI1 | 1.10000 | 1.40000 |

| | |
|----------|------|
| STOICH 4 | O2ZR |
| ZOO1 | 1.11 |
| ZOO2 | 1.12 |
| ZOO3 | 1.13 |

| | | |
|----------|------|------|
| STOICH 5 | O2MR | O2MG |
| MAC1 | 1.1 | 1.4 |

| | |
|----------|---------|
| O2 LIMIT | KDO |
| | 0.10000 |

| | | | | | | | | |
|----------|------|-------|---------|---------|---------|---------|---------|---------|
| SEDIMENT | SEDC | SEDCI | SEDCI | SEDCI | SEDCI | SEDCI | SEDCI | SEDCI |
| WB 1 | OFF | OFF | 0.00000 | 0.10000 | 0.10000 | 1.00000 | 1.00000 | 1.00000 |
| WB 2 | OFF | OFF | 0.00000 | 0.10000 | 0.10000 | 1.00000 | 1.00000 | 1.00000 |

WB 2 bth2_2m_r56b.npt
WB 3 bth3_2m_e2.npt
WB 4 bth4_2m_r56b.npt
WB 5 bth5_2m_r56b.npt
WB 6 bth6_2m_r56b.npt
WB 7 bth7_2m_r56b.npt
WB 8 bth8_2m_e.npt
WB 9 bth9_2m_e.npt
WB 10 bth10_2m_e.npt
WB 11 bth11_2m_e_mann1.npt
WB 12 bth12_2m_e_mann1.npt
WB 13 bth13_2m_e_mann1.npt
WB 14 bth14_2m_e_mann1.npt
WB 15 bth15_kettle_2m_r53wq.npt
WB 16 bth16_marcus_2m_e.npt
WB 17 bth17_colville_2m_r56b.npt
WB 18 bth18_spokane_2m_e9.npt
WB 19 bth19_spokane_2m_e9.npt
WB 20 bth20_spokane_2m_e9.npt
WB 21 bth21_spokane_2m_e9.npt
WB 22 bth22_spokane_2m_e9.npt
WB 23 bth23_spokane_2m_e.npt
WB 24 bth24_hawk_2m_e.npt
WB 25 bth25_sanpoil_2m_e.npt

MET FILE.....METFN.....

WB 1 nmet_kflw_est_wb1.npt
WB 2 nmet_kflw_est_wb2.npt
WB 3 nmet_kflw_est_wb3.npt
WB 4 nmet_kflw_est_wb4.npt
WB 5 nmet_kflw_est_wb5.npt
WB 6 nmet_kflw_est_wb6.npt
WB 7 nmet_kflw_est_wb7.npt
WB 8 nmet_kflw_est_wb8.npt
WB 9 nmet_kflw_est_wb9.npt
WB 10 nmet_kflw_est_wb10.npt
WB 11 nmet_kflw_est_wb11.npt
WB 12 nmet_kflw_est_wb12.npt
WB 13 nmet_sbmw_est_wb13.npt
WB 14 nmet_gcgw_est_wb14.npt
WB 15 nmet_kflw_est_wb15.npt
WB 16 nmet_kflw_est_wb16.npt
WB 17 nmet_kflw_est_wb17.npt
WB 18 nmet_sbmw_est_wb18.npt
WB 19 nmet_sbmw_est_wb19.npt
WB 20 nmet_sbmw_est_wb20.npt
WB 21 nmet_sbmw_est_wb21.npt
WB 22 nmet_sbmw_est_wb22.npt
WB 23 nmet_sbmw_est_wb23.npt
WB 24 nmet_gcgw_est_wb24.npt
WB 25 nmet_gcgw_est_wb25.npt

EXT FILE.....EXTFN.....

WB 1 ext_photic_wb1.npt
WB 2 ext_photic_wb2.npt
WB 3 ext_photic_wb3.npt
WB 4 ext_photic_wb4.npt
WB 5 ext_photic_wb5.npt
WB 6 ext_photic_wb6.npt
WB 7 ext_photic_wb7.npt
WB 8 ext_photic_wb8.npt
WB 9 ext_photic_wb9.npt
WB 10 ext_photic_wb10.npt
WB 11 ext_photic_wb11.npt
WB 12 ext_photic_wb12.npt
WB 13 ext_photic_wb13.npt
WB 14 ext_photic_wb14.npt
WB 15 ext_photic_wb15.npt
WB 16 ext_photic_wb16.npt
WB 17 ext_photic_wb17.npt
WB 18 ext_spokane18.npt
WB 19 ext_spokane19.npt
WB 20 ext_spokane20.npt
WB 21 ext_spokane21.npt
WB 22 ext_spokane22.npt
WB 23 ext_spokane23.npt
WB 24 ext_photic_wb24.npt
WB 25 ext_photic_wb25.npt

VPR FILE.....VPRFN.....

WB 1 vpr00wb1.npt

WB 2 vpr00wb2.npt
WB 3 vpr00wb3.npt
WB 4 vpr00wb4.npt
WB 5 vpr00wb5.npt
WB 6 vpr00wb6.npt
WB 7 vpr00wb7.npt
WB 8 vpr00wb8.npt
WB 9 vpr00wb9.npt
WB 10 vpr00wb10.npt
WB 11 vpr00wb11.npt
WB 12 vpr00wb12.npt
WB 13 vpr00wb13.npt
WB 14 vpr00wb14.npt
WB 15 vpr00wb15.npt
WB 16 vpr00wb16.npt
WB 17 vpr00wb17.npt
WB 18 vpr00wb18.npt
WB 19 vpr00wb19.npt
WB 20 vpr00wb20.npt
WB 21 vpr00wb21.npt
WB 22 vpr00wb22.npt
WB 23 vpr00wb23.npt
WB 24 vpr00wb24.npt
WB 25 vpr00wb25.npt

LPR FILE.....LPRFN.....

WB 1 lpr_wb01.npt
WB 2 lpr_wb02.npt
WB 3 lpr_wb03.npt
WB 4 lpr_wb04.npt
WB 5 lpr_wb05.npt
WB 6 lpr_wb06.npt
WB 7 lpr_wb07.npt
WB 8 lpr_wb08.npt
WB 9 lpr_wb09.npt
WB 10 lpr_wb10.npt
WB 11 lpr_wb11.npt
WB 12 lpr_wb12.npt
WB 13 lpr_wb13.npt
WB 14 lpr_wb14.npt
WB 15 lpr_wb15.npt
WB 16 lpr_wb16.npt
WB 17 lpr_wb17.npt
WB 18 lpr_wb18.npt
WB 19 lpr_wb19.npt
WB 20 lpr_wb20.npt
WB 21 lpr_wb21.npt
WB 22 lpr_wb22.npt
WB 23 lpr_wb23.npt
WB 24 lpr_wb24.npt
WB 25 lpr_wb25.npt

QIN FILE.....QINFN.....

BR1 qin_usbc99_adj1.npt
BR2 qin_br2.npt
BR3 qin_br3.npt
BR4 qin_br4.npt
BR5 qin_br5.npt
BR6 qin_br6.npt
BR7 qin_br7.npt
BR8 qin_br8.npt
BR9 qin_br9.npt
BR10 qin_br10.npt
BR11 qin_br11.npt
BR12 qin_br12.npt
BR13 qin_br13.npt
BR14 qin_br14.npt
BR15 qin_kettle99.npt
BR16 qin_marcus99.npt
BR17 qin_colville_censored.npt
BR18 qin_spokane_avista.npt
BR19 qin_br16.npt
BR20 qin_br17.npt
BR21 qin_br18.npt
BR22 qin_br19.npt
BR23 qin_br20.npt
BR24 qin_hawk99.npt
BR25 qin_sanpoil99.npt

TIN FILE.....TINFN.....

BR1 tin_usbc_censored_2000.npt

BR2 tin_br2.npt
 BR3 tin_br3.npt
 BR4 tin_br4.npt
 BR5 tin_br5.npt
 BR6 tin_br6.npt
 BR7 tin_br7.npt
 BR8 tin_br8.npt
 BR9 tin_br9.npt
 BR10 tin_br10.npt
 BR11 tin_br11.npt
 BR12 tin_br12.npt
 BR13 tin_br13.npt
 BR14 tin_br14.npt
 BR15 tin_kettle99.npt
 BR16 tin_marcus99.npt
 BR17 tin_colville_censored99.npt
 BR18 tin_spokane_minT.npt
 BR19 tin_br19.npt
 BR20 tin_br20.npt
 BR21 tin_br21.npt
 BR22 tin_br22.npt
 BR23 tin_br23.npt
 BR24 tin_hawk99.npt
 BR25 tin_sanpoil99.npt

CIN FILE.....CINFN.....

BR1 cin_columbia_r121.npt
 BR2 cin_temp2.npt
 BR3 cin_temp3.npt
 BR4 cin_temp4.npt
 BR5 cin_temp5.npt
 BR6 cin_temp6.npt
 BR7 cin_temp7.npt
 BR8 cin_temp8.npt
 BR9 cin_temp9.npt
 BR10 cin_temp10.npt
 BR11 cin_temp11.npt
 BR12 cin_temp12.npt
 BR13 cin_temp13.npt
 BR14 cin_temp14.npt
 BR15 cin_kettle1.npt
 BR16 cin_temp16.npt
 BR17 cin_colville1.npt
 BR18 cin_spokane_r121.npt
 BR19 cin_temp19.npt
 BR20 cin_temp20.npt
 BR21 cin_temp21.npt
 BR22 cin_temp22.npt
 BR23 cin_temp23.npt
 BR24 cin_temp24.npt
 BR25 cin_sanpoil1.npt

QOT FILE.....QOTFN.....

BR1 qot_br1.npt
 BR2 qot_br2.npt
 BR3 qot_br3.npt
 BR4 qot_br4.npt
 BR5 qot_br5.npt
 BR6 qot_br6.npt
 BR7 qot_br7.npt
 BR8 qot_br8.npt
 BR9 qot_br9.npt
 BR10 qot_br10.npt
 BR11 qot_br11.npt
 BR12 qot_br12.npt
 BR13 qot_br13.npt
 BR14 qot_gcl_4structures.npt
 BR15 qot_br15.npt
 BR16 qot_br16.npt
 BR17 qot_br17.npt
 BR18 qot_br18.npt
 BR19 qot_br19.npt
 BR20 qot_br20.npt
 BR21 qot_br21.npt
 BR22 qot_br22.npt
 BR23 qot_br23.npt
 BR24 qot_br24.npt
 BR25 qot_br25.npt

QTR FILE.....QTRFN.....

TR1 qtr_returnq.npt

TTR FILE.....TTRFN.....
TR1 ttr_banks99.npt

CTR FILE.....CTRFN.....
TR1 ctr_usbc_banks.npt

QDT FILE.....QDTFN.....
BR1 qdt_br1.npt
BR2 qdt_br2.npt
BR3 qdt_br3.npt
BR4 qdt_br4.npt
BR5 qdt_br5.npt
BR6 qdt_br6.npt
BR7 qdt_br7.npt
BR8 qdt_br8.npt
BR9 qdt_br9.npt
BR10 qdt_br10.npt
BR11 qdt_br11.npt
BR12 qdt_br12.npt
BR13 qdt_br13.npt
BR14 qdt_br14.npt
BR15 qdt_br15.npt
BR16 qdt_br16.npt
BR17 qdt_br17.npt
BR18 qdt_br18.npt
BR19 qdt_br19.npt
BR20 qdt_br20.npt
BR21 qdt_br21.npt
BR22 qdt_br22.npt
BR23 qdt_br23.npt
BR24 qdt_br24.npt
BR25 qdt_br25.npt

TDT FILE.....TDTFN.....
BR1 tdt_br1.npt
BR2 tdt_br2.npt
BR3 tdt_br3.npt
BR4 tdt_br4.npt
BR5 tdt_br5.npt
BR6 tdt_br6.npt
BR7 tdt_br7.npt
BR8 tdt_br8.npt
BR9 tdt_br9.npt
BR10 tdt_br10.npt
BR11 tdt_br11.npt
BR12 tdt_br12.npt
BR13 tdt_br13.npt
BR14 tdt_br14.npt
BR15 tdt_br15.npt
BR16 tdt_br16.npt
BR17 tdt_br17.npt
BR18 tdt_br18.npt
BR19 tdt_br19.npt
BR20 tdt_br20.npt
BR21 tdt_br21.npt
BR22 tdt_br22.npt
BR23 tdt_br23.npt
BR24 tdt_br24.npt
BR25 tdt_br25.npt

CDT FILE.....CDFN.....
BR1 cdt_usbc1.npt
BR2 cdt_usbc2.npt
BR3 cdt_usbc3.npt
BR4 cdt_usbc4.npt
BR5 cdt_usbc5.npt
BR6 cdt_usbc6.npt
BR7 cdt_usbc7.npt
BR8 cdt_usbc8.npt
BR9 cdt_usbc9.npt
BR10 cdt_usbc10.npt
BR11 cdt_usbc11.npt
BR12 cdt_usbc12.npt
BR13 cdt_usbc13.npt
BR14 cdt_usbc14.npt
BR15 cdt_br15.npt
BR16 cdt_br16.npt
BR17 cdt_br17.npt
BR18 cdt_br18.npt
BR19 cdt_br19.npt

BR20 cdt_br20.npt
BR21 cdt_br21.npt
BR22 cdt_br22.npt
BR23 cdt_br23.npt
BR24 cdt_br24.npt
BR25 cdt_br25.npt

PRE FILE.....PREFN.....
BR1 pre_br1.npt - not used
BR2 pre_br2.npt - not used
BR3 pre_br3.npt - not used
BR4 pre_br4.npt - not used
BR5 pre_br5.npt - not used
BR6 pre_br6.npt - not used
BR7 pre_br7.npt - not used
BR8 pre_br8.npt - not used
BR9 pre_br9.npt - not used
BR10 pre_br10.npt - not used
BR11 pre_br11.npt - not used
BR12 pre_br12.npt - not used
BR13 pre_br13.npt - not used
BR14 pre_br14.npt - not used
BR15 pre_br14.npt - not used
BR16 pre_br16.npt - not used
BR17 pre_br17.npt - not used
BR18 pre_br18.npt - not used
BR19 pre_br19.npt - not used
BR20 pre_br20.npt - not used
BR21 pre_br21.npt - not used
BR22 pre_br22.npt - not used
BR23 pre_br23.npt - not used
BR24 pre_br24.npt - not used
BR25 pre_br25.npt - not used

TPR FILE.....TPRFN.....
BR1 tpr_br1.npt - not used
BR2 tpr_br2.npt - not used
BR3 tpr_br3.npt - not used
BR4 tpr_br4.npt - not used
BR5 tpr_br5.npt - not used
BR6 tpr_br6.npt - not used
BR7 tpr_br7.npt - not used
BR8 tpr_br8.npt - not used
BR9 tpr_br9.npt - not used
BR10 tpr_br10.npt - not used
BR11 tpr_br11.npt - not used
BR12 tpr_br12.npt - not used
BR13 tpr_br13.npt - not used
BR14 tpr_br14.npt - not used
BR15 tpr_br14.npt - not used
BR16 tpr_br16.npt - not used
BR17 tpr_br17.npt - not used
BR18 tpr_br18.npt - not used
BR19 tpr_br19.npt - not used
BR20 tpr_br20.npt - not used
BR21 tpr_br21.npt - not used
BR22 tpr_br22.npt - not used
BR23 tpr_br23.npt - not used
BR24 tpr_br24.npt - not used
BR25 tpr_br25.npt - not used

CPR FILE.....CPRFN.....
BR1 cpr_br1.npt - not used
BR2 cpr_br2.npt - not used
BR3 cpr_br3.npt - not used
BR4 cpr_br4.npt - not used
BR5 cpr_br5.npt - not used
BR6 cpr_br6.npt - not used
BR7 cpr_br7.npt - not used
BR8 cpr_br8.npt - not used
BR9 cpr_br9.npt - not used
BR10 cpr_br10.npt - not used
BR11 cpr_br11.npt - not used
BR12 cpr_br12.npt - not used
BR13 cpr_br13.npt - not used
BR14 cpr_br14.npt - not used
BR15 cpr_br14.npt - not used
BR16 cpr_br16.npt - not used
BR17 cpr_br17.npt - not used
BR18 cpr_br18.npt - not used
BR19 cpr_br19.npt - not used

BR20 cpr_br20.npt - not used
BR21 cpr_br21.npt - not used
BR22 cpr_br22.npt - not used
BR23 cpr_br23.npt - not used
BR24 cpr_br24.npt - not used
BR25 cpr_br25.npt - not used

EUH FILE.....EUHFN.....
BR1 euh_br1.npt
BR2 euh_br2.npt
BR3 euh_br3.npt
BR4 euh_br4.npt
BR5 euh_br5.npt
BR6 euh_br6.npt
BR7 euh_br7.npt
BR8 euh_br8.npt
BR9 euh_br9.npt
BR10 euh_br10.npt
BR11 euh_br11.npt
BR12 euh_br12.npt
BR13 euh_br13.npt
BR14 euh_br14.npt
BR15 euh_br15.npt
BR16 euh_br16.npt
BR17 euh_br17.npt
BR18 euh_br18.npt
BR19 euh_br19.npt
BR20 euh_br20.npt
BR21 euh_br21.npt
BR22 euh_br22.npt
BR23 euh_br23.npt
BR24 euh_br24.npt
BR25 euh_br25.npt

TUH FILE.....TUHFN.....
BR1 tuh_br1.npt
BR2 tuh_br2.npt
BR3 tuh_br3.npt
BR4 tuh_br4.npt
BR5 tuh_br5.npt
BR6 tuh_br6.npt
BR7 tuh_br7.npt
BR8 tuh_br8.npt
BR9 tuh_br9.npt
BR10 tuh_br10.npt
BR11 tuh_br11.npt
BR12 tuh_br12.npt
BR13 tuh_br13.npt
BR14 tuh_br14.npt
BR15 tuh_br15.npt
BR16 tuh_br16.npt
BR17 tuh_br17.npt
BR18 tuh_br18.npt
BR19 tuh_br19.npt
BR20 tuh_br20.npt
BR21 tuh_br21.npt
BR22 tuh_br22.npt
BR23 tuh_br23.npt
BR24 tuh_br24.npt
BR25 tuh_br25.npt

CUH FILE.....CUHFN.....
BR1 cuh_br1.npt
BR2 cuh_br2.npt
BR3 cuh_br3.npt
BR4 cuh_br4.npt
BR5 cuh_br5.npt
BR6 cuh_br6.npt
BR7 cuh_br7.npt
BR8 cuh_br8.npt
BR9 cuh_br9.npt
BR10 cuh_br10.npt
BR11 cuh_br11.npt
BR12 cuh_br12.npt
BR13 cuh_br13.npt
BR14 cuh_br14.npt
BR15 cuh_br15.npt
BR16 cuh_br16.npt
BR17 cuh_br17.npt
BR18 cuh_br18.npt
BR19 cuh_br19.npt

BR20 cuh_br20.npt
BR21 cuh_br21.npt
BR22 cuh_br22.npt
BR23 cuh_br23.npt
BR24 cuh_br24.npt
BR25 cuh_br25.npt

EDH FILE.....EDHFN.....

BR1 edh_br1.npt
BR2 edh_br2.npt
BR3 edh_br3.npt
BR4 edh_br4.npt
BR5 edh_br5.npt
BR6 edh_br6.npt
BR7 edh_br7.npt
BR8 edh_br8.npt
BR9 edh_br9.npt
BR10 edh_br10.npt
BR11 edh_br11.npt
BR12 edh_br12.npt
BR13 edh_br13.npt
BR14 edh_br14.npt
BR15 edh_br15.npt
BR16 edh_br16.npt
BR17 edh_br17.npt
BR18 edh_br18.npt
BR19 edh_br19.npt
BR20 edh_br20.npt
BR21 edh_br21.npt
BR22 edh_br22.npt
BR23 edh_br23.npt
BR24 edh_br24.npt
BR25 edh_br25.npt

TDH FILE.....TDHFN.....

BR1 tdh_br1.npt
BR2 tdh_br2.npt
BR3 tdh_br3.npt
BR4 tdh_br4.npt
BR5 tdh_br5.npt
BR6 tdh_br6.npt
BR7 tdh_br7.npt
BR8 tdh_br8.npt
BR9 tdh_br9.npt
BR10 tdh_br10.npt
BR11 tdh_br11.npt
BR12 tdh_br12.npt
BR13 tdh_br13.npt
BR14 tdh_br14.npt
BR15 tdh_br15.npt
BR16 tdh_br16.npt
BR17 tdh_br17.npt
BR18 tdh_br18.npt
BR19 tdh_br19.npt
BR20 tdh_br20.npt
BR21 tdh_br21.npt
BR22 tdh_br22.npt
BR23 tdh_br23.npt
BR24 tdh_br24.npt
BR25 tdh_br25.npt

CDH FILE.....CDHFN.....

BR1 cdh_br1.npt
BR2 cdh_br2.npt
BR3 cdh_br3.npt
BR4 cdh_br4.npt
BR5 cdh_br5.npt
BR6 cdh_br6.npt
BR7 cdh_br7.npt
BR8 cdh_br8.npt
BR9 cdh_br9.npt
BR10 cdh_br10.npt
BR11 cdh_br11.npt
BR12 cdh_br12.npt
BR13 cdh_br13.npt
BR14 cdh_br14.npt
BR15 cdh_br15.npt
BR16 cdh_br16.npt
BR17 cdh_br17.npt
BR18 cdh_br18.npt
BR19 cdh_br19.npt

BR20 cdh_br20.npt
BR21 cdh_br21.npt
BR22 cdh_br22.npt
BR23 cdh_br23.npt
BR24 cdh_br24.npt
BR25 cdh_br25.npt

SNP FILE.....SNPFN.....

WB 1 snp1.opt
WB 2 snp2.opt
WB 3 snp3.opt
WB 4 snp4.opt
WB 5 snp5.opt
WB 6 snp6.opt
WB 7 snp7.opt
WB 8 snp8.opt
WB 9 snp9.opt
WB 10 snp10.opt
WB 11 snp11.opt
WB 12 snp12.opt
WB 13 snp13.opt
WB 14 snp14.opt
WB 15 snp15.opt
WB 16 snp16.opt
WB 17 snp17.opt
WB 18 snp18.opt
WB 19 snp19.opt
WB 20 snp20.opt
WB 21 snp21.opt
WB 22 snp22.opt
WB 23 snp23.opt
WB 24 snp24.opt
WB 25 snp25.opt

PRF FILE.....PRFFN.....

WB 1 prf1.opt
WB 2 prf2.opt
WB 3 prf3.opt
WB 4 prf4.opt
WB 5 prf5.opt
WB 6 prf6.opt
WB 7 prf7.opt
WB 8 prf8.opt
WB 9 prf9.opt
WB 10 prf10.opt
WB 11 prf11.opt
WB 12 prf12.opt
WB 13 prf13.opt
WB 14 prf14.opt
WB 15 prf15.opt
WB 16 prf16.opt
WB 17 prf17.opt
WB 18 prf18.opt
WB 19 prf19.opt
WB 20 prf20.opt
WB 21 prf21.opt
WB 22 prf22.opt
WB 23 prf23.opt
WB 24 prf24.opt
WB 25 prf25.opt

VPL FILE.....VPLFN.....

WB 1 vpl1.opt
WB 2 vpl2.opt
WB 3 vpl3.opt
WB 4 vpl4.opt
WB 5 vpl5.opt
WB 6 vpl6.opt
WB 7 vpl7.opt
WB 8 vpl8.opt
WB 9 vpl9.opt
WB 10 vpl10.opt
WB 11 vpl11.opt
WB 12 vpl12.opt
WB 13 vpl13.opt
WB 14 vpl14.opt
WB 15 vpl15.opt
WB 16 vpl16.opt
WB 17 vpl17.opt
WB 18 vpl18.opt
WB 19 vpl19.opt

WB 20 vpl20.opt
WB 21 vpl21.opt
WB 22 vpl22.opt
WB 23 vpl23.opt
WB 24 vpl24.opt
WB 25 vpl25.opt

CPL FILE.....CPLFN.....

WB 1 cpl1.opt
WB 2 cpl2.opt
WB 3 cpl3.opt
WB 4 cpl4.opt
WB 5 cpl5.opt
WB 6 cpl6.opt
WB 7 cpl7.opt
WB 8 cpl8.opt
WB 9 cpl9.opt
WB 10 cpl10.opt
WB 11 cpl11.opt
WB 12 cpl12.opt
WB 13 cpl13.opt
WB 14 cpl14.opt
WB 15 cpl15.opt
WB 16 cpl16.opt
WB 17 cpl17.opt
WB 18 cpl18.opt
WB 19 cpl19.opt
WB 20 cpl20.opt
WB 21 cpl21.opt
WB 22 cpl22.opt
WB 23 cpl23.opt
WB 24 cpl24.opt
WB 25 cpl25.opt

SPR FILE.....SPRFN.....

WB 1 spr1.opt
WB 2 spr2.opt
WB 3 spr3.opt
WB 4 spr4.opt
WB 5 spr5.opt
WB 6 spr6.opt
WB 7 spr7.opt
WB 8 spr8.opt
WB 9 spr9.opt
WB 10 spr10.opt
WB 11 spr11.opt
WB 12 spr12.opt
WB 13 spr13.opt
WB 14 spr14.opt
WB 15 spr15.opt
WB 16 spr16.opt
WB 17 spr17.opt
WB 18 spr18.opt
WB 19 spr19.opt
WB 20 spr20.opt
WB 21 spr21.opt
WB 22 spr22.opt
WB 23 spr23.opt
WB 24 spr24.opt
WB 25 spr25.opt

FLX FILE.....FLXFN.....

WB 1 flx1.opt
WB 2 flx2.opt
WB 3 flx3.opt
WB 4 flx4.opt
WB 5 flx5.opt
WB 6 flx6.opt
WB 7 flx7.opt
WB 8 flx8.opt
WB 9 flx9.opt
WB 10 flx10.opt
WB 11 flx11.opt
WB 12 flx12.opt
WB 13 flx13.opt
WB 14 flx14.opt
WB 15 flx15.opt
WB 16 flx16.opt
WB 17 flx17.opt
WB 18 flx18.opt
WB 19 flx19.opt

WB 20 flx20.opt
WB 21 flx21.opt
WB 22 flx22.opt
WB 23 flx23.opt
WB 24 flx24.opt
WB 25 flx25.opt

TSR FILE.....TSRFN.....
tsr.opt

WDO FILE.....WDOFN.....
wdo.opt

W2_bio_con.npt

A sample W2_bio_con.npt file is shown below. A further description is provided in Appendix D.

Lake Roosevelt Model
Fish Bioenergetics Input File

```

FISHCOMP  FCALC      FUF  FDIAGC  BESTC  DIELC  CMAXC
           ON        30    ON        ON    ON    OFF

BIO TIME  JSTART    JEND
           366.0    730.0

F KINETI  T1      T2      T3      T4      K1      K2      K3      K4
           3.0     20.0    20.0    24.0    0.58    0.98    0.98    0.50

F STOICH  FOXYCAL
           13569.4

F1 IC     MASS     LENGTH  STO_CON
           76.0     193.0   0.001

Z LENGTH  LZOO     LZOO     LZOO     LZOO     LZOO     LZOO     LZOO     LZOO
           1.0     1.0     1.3

Z MASS    MZOO     MZOO     MZOO     MZOO     MZOO     MZOO     MZOO     MZOO
           7.7e-3   4.9e-3   8.82e-5

Z STOICH  EZOO     EZOO     EZOO     EZOO     EZOO     EZOO     EZOO     EZOO
           2420.0   2420.0   2420.0

FORAGE    HANDLE    FVELA    FVELB    FVELC  THRESHC THRESHV THRESHZ  DIELLUX
           0.33    0.099    0.13    0.04505  ON      6.032    3        0.01

FGP PLOT  FGPC
           ON

FGP DATE  FGPD
           367.0

FGP FREQ  FGPF
           1.0

MASSTYPE  GEN      BEST     DIEL
           FXN    USER    SEG

MASS FXN  GIM      BIM      DIM      GALP    BALP    DALP    GTI      BTI      DTI
           76.0    76.0    76.0    0.0049  0.0049  0.0049  366.0    366.0    366.0

DIAG OPT  FISHC    BIOPARC  CONSC    DIGC    RESPC    SURFC    DEPTHC
           ON      ON      ON      ON      ON      ON      ON

SINGLEPT  SINGLEC  SINGFN  SINIBIO  TLC
           ON      143    312     OFF

ZAV FILE.....ZAVFN.....
           Zav.npt

FGP FILE.....FGPFN.....
WB1      FGP_wb1.opt

FXN FILE.....FXNFN.....
GEN      Gen_userfxn.npt
BEST     Best_userfxn.npt
DIEL     Diel_userfxn.npt

```


Appendix C: Fish Bioenergetic Parameter Formulation

While the bioenergetics and foraging model of Stockwell and Johnson (1997) forms the basis of the approach, the formulation of the bioenergetics parameters were resolved into consistent units of joules over each time step. The consumption parameter was left in units of prey per minute, and growth remained in units of mass (g). Table 22 lists the parameters and their units.

Table 22. Bioenergetics parameters summary.

| Parameter | Symbol | Units |
|-------------|--------|------------------|
| Consumption | C | #/min (prey) |
| Digestion | D | J (per timestep) |
| Excretion | U | J (per timestep) |
| Egestion | F | J (per timestep) |
| Respiration | R | J (per timestep) |
| SDA | S | J (per timestep) |
| Growth | G | g (per day) |

Growth

Growth is computed daily from the sum of the parameters shown in (1) and converted from energy to mass using the fish energy density, E_{fish} .

$$G = \frac{1}{E_{\text{fish}}} \sum_{\text{day}} (D - U - F - R - S) \quad (1)$$

$$G = \frac{[\text{J}]}{[\text{J} \cdot \text{g}^{-1}]}$$

$$G = \text{g (per day)}$$

Fish energy density

The allometric function reported by Hewitt & Johnson (1992) is used for the Kokanee energy density (Equation (2)).

$$E_{\text{fish}} \left[\frac{\text{J}}{\text{g}} \right] = \begin{cases} 4.1868 \cdot (1.8510 \cdot M + 1250) & \text{for } M \leq 196 \text{ g} \\ 4.1868 \cdot (1.1254 \cdot M + 1588) & \text{for } M > 196 \text{ g} \end{cases} \quad (2)$$

Digestion

Digestion is formulated as the energy derived from digested prey over the timestep. The energy from consumption is included in the digestion parameter. Consumption does not explicitly appear in the growth equation (Equation (1)). The approach (Equation (3)) is based on Bevelhimer and Adams (1993); physically, the terms represent the initial stomach content to be digested, the prey consumed during the time step, and the undigested stomach contents at the end of the time step. The last term allows for a time-lag between consumption and digestion. The time to digest the bulk of the a full stomach can range from a couple hours if near optimal temperature to roughly half a day if at very cold temperatures. The variables and their units are shown in Table 23.

The digestion parameter is a function of several other functions: the Thornton-Lessem function (see ancillary functions section), the digestion coefficient, the consumption parameter, and the stomach content at the start of the time step.

$$D = \left(\left(\underbrace{M_o}_{\text{initial content}} + \underbrace{(C \cdot m_z \cdot t) \cdot TL}_{\text{consumption}} \right) - \left(\underbrace{M_o e^{-r \cdot t / 60} + \frac{C \cdot m_z \cdot 60 \cdot TL}{r} (1 - e^{-r \cdot t / 60})}_{\text{undigested contents}} \right) \right) \cdot E_{\text{prey}} \quad (3)$$

$$D = \left(\left(g - \text{wet} + \left(\frac{\#}{\text{min}} \cdot g - \text{wet} \cdot \text{min} \right) \right) - \left(g - \text{wet} + \frac{\#}{\text{min}} \cdot g - \text{wet} \cdot \text{min} \right) \right) \left(\frac{J}{g - \text{wet}} \right)$$

D = [J] (per timestep)

Table 23. Digestion parameter variables and units.

| Variable | Units | Definition |
|-------------------|-----------------------|-----------------------------------|
| M_o | g-wet | Initial stomach content |
| C | #/min (e.g., Daphnia) | Consumption |
| m_z | g-wet | mass of a single prey zooplankton |
| t | (30) minutes | Timestep |
| TL | dimensionless | Thornton-Lessem function |
| r | dimensionless | digestion coefficient |
| E_{prey} | J/g-wet | Prey energy content (density) |
| 60 | minutes | unit conversion factor |

Digestion coefficient (function)

A kokanee digestion coefficient, r , is reported by Stockwell & Johnson (1999) and Bevelhimer and Adams (1993) and reproduced in Table 24. The Stockwell & Johnson formulation is used to avoid non-positive values which may occur at low temperatures. Additionally, the temperature input to the bioenergetics algorithms is constrained to a minimum value of 0.01 °C.

Table 24. Digestion coefficient formulations

| Equation | Source | # |
|------------------------------|-----------------------------|-------|
| $r = 0.014 \cdot T - 0.0154$ | Bevelhimer and Adams (1993) | (4) |
| $r = 0.014 \cdot T + 0.1135$ | Stockwell & Johnson (1999) | (5) |

Stomach content and capacity

Stomach content is updated after each time step. The remaining stomach content is composed of the undigested initial stomach content and the consumed prey undigested over the time step, as shown in Equation (6).

$$M_0^{new} = \underbrace{M_0 e^{-r \cdot t / 60}}_{\text{undigested initial content}} + \underbrace{\frac{C \cdot m_z \cdot 60 \cdot TL}{r} (1 - e^{-r \cdot t / 60})}_{\text{undigested consumption}} \quad (6)$$

Stomach capacity is formulated based on Brett (1971). If, at the start of any time step, the stomach content exceeds the stomach capacity, then consumption is zero for that time step.

$$\text{Capacity [g]} = \begin{cases} M \cdot (14.1 - 4.95 \cdot \log(M)) / 100 & \text{for } M \leq 253.5 \text{ g} \\ 0.022 \cdot M & \text{for } M > 253.5 \text{ g} \end{cases} \quad (7)$$

Consumption, Search Volume, and Reaction Distance

Consumption is based on the Stockwell & Johnson (1997) model (Equation (8)), and is a function of other functions: search volume (rate), and the Thornton-Lessem function (see ancillary functions section).

$$C = \frac{E \cdot z}{1 + E \cdot z \cdot h} \cdot TL \cdot 60 \quad (8)$$

$$C = \frac{\frac{\text{m}^3}{\text{s}} \cdot \frac{\#}{\text{m}^3}}{\frac{\text{m}^3}{\text{s}} \cdot \frac{\#}{\text{m}^3} \cdot \frac{\text{s}}{\text{min}}} \cdot 60\text{s}$$

$$C = \left[\frac{\#}{\text{min}} \right] \text{ (“prey” per timestep)}$$

The search volume (rate) is the simple cylinder suggested by Eggers (1977) (Equation (9)). The reaction distance is taken to be a sphere described by Link and Edsall (1996) (Equation (10)). The consumption, search volume, and reaction distance parameter variables and units are shown in Table 25

$$E \left[\frac{\text{m}^3}{\text{s}} \right] = \pi \cdot R_d^2 \cdot v \quad (9)$$

$$R_d [\text{m}] = 0.01 \cdot 4.9424 \cdot \text{lux}^{0.086} \quad (10)$$

Table 25. Consumption parameter variables and units.

| Variable | Units | Definition |
|----------|-------------------------|----------------------------|
| R_d | m | Predator reaction distance |
| lux | lux | Ambient light intensity |
| E | m^3 / s | Search rate |
| v | m / s | fish swimming speed |
| z | $\# / \text{m}^3$ | prey density |
| h | $\# / \text{s}$ | handling time |
| TL | dimensionless | Thornton-Lessem function |
| 60 | s / min | unit conversion factor |

Egestion & Excretion

Egestion (fecal wastes) and excretion (urinary wastes) are formulated using the approach of Hewett & Johnson (1987). Parameter variables and units are shown for egestion in Table 26 and for excretion in Table 27.

$$F = 0.455 \cdot T^{-0.222} \cdot D \quad (11)$$

$$F = \left(\frac{1}{^{\circ}\text{C}} \right)^{\circ}\text{C} \cdot \text{J} = \text{J}$$

$$F = [\text{J}] \text{ (per timestep)}$$

Table 26. Egestion parameter variables and units.

| Variable | Units | Definition |
|----------|------------------|---------------------|
| T | deg. C | Temperature |
| D | J (per timestep) | Digestion parameter |

$$U = 0.0233 \cdot T^{0.58} \cdot (D - F) \quad (12)$$

$$U = \left(\frac{1}{^{\circ}\text{C}} \right)^{\circ}\text{C} \cdot \text{J} = \text{J}$$

$$U = [\text{J}] \text{ (per timestep)}$$

Table 27. Excretion parameter variables and units.

| Variable | Units | Definition |
|----------|------------------|---------------------|
| T | deg. C | Temperature |
| D | J (per timestep) | Digestion parameter |
| F | J (per timestep) | Egestion parameter |

Specific Dynamic Action (SDA)

Specific dynamic action (the physiological cost of digesting a meal) is formulated using the approach of Brett & Groves (1979). The coefficient (0.172) in Equation (13) for kokanee is taken from Beauchamp, et al. (1989). The variables and units of SDA are shown in Table 28.

$$\begin{aligned} \text{SDA} &= 0.172 \cdot (D - F) && (13) \\ \text{SDA} &= [\text{J}] \text{ (per timestep)} \end{aligned}$$

Table 28. Specific dynamic action parameter variables and units.

| Variable | Units | Definition |
|-----------------|------------------|---------------------|
| D | J (per timestep) | Digestion parameter |
| F | J (per timestep) | Egestion parameter |

Respiration

The respiration parameter (Equation (14)) is formulated based on the approach of Beauchamp, et al. (1989) as are the activity, ACT, and cruising speed (velocity) functions, Equations (15) and (16), respectively. The respiration, activity, and cruising speed variables and units are shown in Table 29. The value for the oxycaloric conversion factor is that used by Stockwell & Johnson (1997) of 3241 cal/ g-O₂ (13569.4 J/ g-O₂).

The respiration formulation has included fish mass, and was converted from cal to J. The activity formulation was converted from velocity inputs in cm/s to units of m/s. The cruising speed was converted from units of cm/s to units of m/s.

$$R = 0.00143 \cdot M^{-0.209} \cdot e^{0.086 \cdot T} \cdot \text{ACT} \cdot \text{OXYCAL} \cdot \frac{t}{t_{\text{day}}} \cdot M \quad (14)$$

$$R = \frac{\text{g-O}_2}{(\sim \text{g-fish}) \cdot (\sim \text{g-fish})^{-0.209}} (\sim \text{g-fish})^{-0.209} \frac{\text{J}}{\text{g-O}_2} \cdot \frac{\text{min}}{\text{min}} \cdot \text{g-fish}$$

$$R = \left[\frac{\text{J}}{\text{g-fish}} \cdot \text{g-fish} \right] = [\text{J}] \text{ (per timestep)}$$

$$\text{ACT} = \exp(0.0234 \cdot v \cdot 100) \quad (15)$$

$$v = 0.01 \cdot 9.9 \cdot \exp(0.0405 \cdot T) \cdot M^{0.13} \quad (16)$$

Table 29. Respiration parameter variables and units.

| Variable | Units | Definition |
|------------------|----------------------|------------------------------|
| M | g-wet | Fish mass |
| T | deg. C | Temperature |
| ACT | dimensionless | Activity |
| OXYCAL | J / g-O ₂ | Oxycaloric conversion factor |
| t | (30) minutes | Timestep |
| t _{day} | (1440) minutes | Duration of a day |

Ancillary function values

The bioenergetics parameters are a function of other, ancillary functions. Table 30 reports the values of those functions for a 76 and 125 g kokanee at 4 and 20 °C to give a sense of the range and sensitivity of the ancillary functions.

Table 30. Representative ancillary function values for 76 and 125 g kokanee at 4 and 20 °C.

| Temperature: | 4 °C | | 20 °C | |
|--|------------------------|------------------------|------------------------|------------------------|
| Fish mass: | 76 g | 125 g | 76 g | 125 g |
| Ancillary function | | | | |
| E_{fish} , fish energy density (J/g) | 5822.5 | 6202.2 | 5822.5 | 6202.2 |
| TL function | 0.58 | 0.58 | 0.98 | 0.98 |
| r, digestion coefficient | 0.17 | 0.17 | 0.39 | 0.39 |
| Stomach capacity (g) | 3.64 | 4.65 | 3.64 | 4.65 |
| v, velocity (m/s) | 0.204 | 0.218 | 0.391 | 0.417 |
| ACT | 1.61 | 1.66 | 2.50 | 2.65 |
| at 50,000 lux (summer, noon, near the surface) | | | | |
| R_d , reaction distance (m) | 0.125 | 0.125 | 0.125 | 0.125 |
| E, search volume (m^3/s) | 10.00×10^{-3} | 10.70×10^{-3} | 19.19×10^{-3} | 20.47×10^{-3} |
| at 10 lux (crepuscular or at depth) | | | | |
| R_d , reaction distance (m) | 0.06 | 0.06 | 0.06 | 0.06 |
| E, search volume (m^3/s) | 2.31×10^{-3} | 2.46×10^{-3} | 4.42×10^{-3} | 4.72×10^{-3} |

The formulation of the Thornton-Lessem function is identical to that of Beachamp, et al. (1989), which was later used by Stockwell & Johnson (1997). A plot of the function is shown as Figure 113.

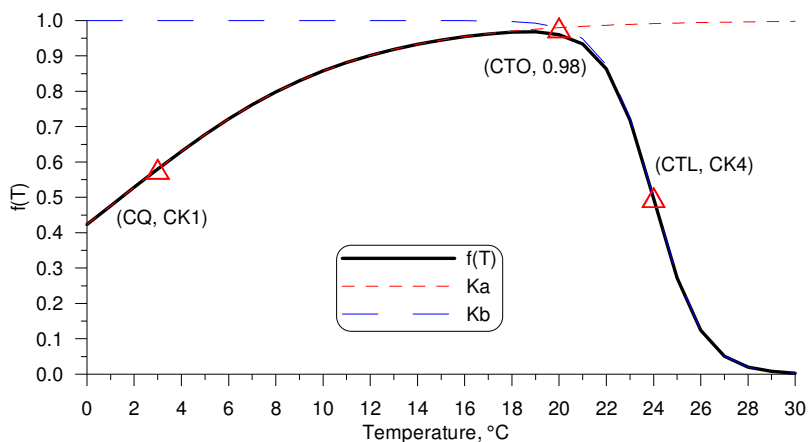


Figure 113. Plot of the Thornton-Lessem function.

Appendix D: Bioenergetics control file explanation

A separate control file, “W2_bio_con.npt,” was generated for the fish bioenergetics algorithm. A sample file is shown in Appendix B. This section describes the control file inputs to the fish bioenergetics routine. Some features are placeholders for future development and integration with the CE-QUAL-W2 code.

The input “type” can be of several formats: The ON/OFF format is used to turn on logical controls. The INTEGER format cannot contain decimals, and the REAL format must contain one decimal. The REAL format can also use scientific notation, but there must be a decimal in the value (e.g., 3.1e-5). The ON/OFF, INTEGER, and REAL formats are right justified, meaning that the rightmost character must be located at the right edge of the column. Each column has 8 spaces, and the first column is not read by the code, but can contain notes to the user. The CHARACTER format is used for words in the *fish mass* type cards and filenames in the *filenames* cards. The fish mass type card is also right justified and case-sensitive. The filenames are left justified, and are not case sensitive. Refer to the examples below.

The first 8 lines of the card are ignored. They can be used to make notes for the user.

Fish Computation

The FCALC value turns the routine on. FCALC and FUF are placeholders for integration within W2. FDIAGC allows for the output of diagnostic values. FDIAGC must be ON for BESTC and DIELC outputs. BESTC turns on the calculation and output of the 'best growth' algorithm. DIELC turns on the calculation and output of the 'vertical foraging' routine. CMAXC sets the model consumption to 100% of the practical maximum consumption, as outlined by Beauchamp, et al., 1989. This is a useful feature for comparing results to literature values and for evaluating maximum consumption scenarios.

| Input | Type | Description |
|--------|---------|---|
| FCALC | ON/OFF | Turns on fish calculations |
| FUF | INTEGER | Model timestep, in minutes |
| FDIAGC | ON/OFF | Turns on diagnostic calculations and output |
| BESTC | ON/OFF | Turns on "best growth" cell calculations and output. |
| DIELC | ON/OFF | Turns on "vertical foraging" calculations and output. |
| CMAXC | ON/OFF | Sets model consumption (prey/min) to 100% of the internally defined maximum practical consumption |

Example:

```
FISHCOMP  FCALC  FUF  FDIAGC  BESTC  DIELC  CMAXC
           ON    30   ON      ON    ON    ON
```

Bioenergetic time control

This card specifies the starting and ending dates for the fish model.

| Input | Type | Description |
|--------|------|-------------------------------|
| JSTART | REAL | Sets the starting julian date |
| JEND | REAL | Sets the ending julian date |

Example:

```
BIO TIME  JSTART  JEND
          366.0  730.0
```

Fish temperature kinetics

This card specifies the four pairs of temperature and function values in the Thornton-Lessem function.

| Input | Type | Description |
|-------|------|--|
| T1 | REAL | Thornton-Lessem T1 or CQ temperature |
| T2 | REAL | Thornton-Lessem T2 or CTO temperature |
| T3 | REAL | Thornton-Lessem T3 or CTM temperature |
| T4 | REAL | Thornton-Lessem T4 or CTL temperature |
| K1 | REAL | Thornton-Lessem K1 or CK1 kinetic term |
| K2 | REAL | Thornton-Lessem K2 or CK2 kinetic term |
| K3 | REAL | Thornton-Lessem K3 or CK3 kinetic term |
| K4 | REAL | Thornton-Lessem K4 or CK4 kinetic term |

Example:

```
F KINETI      T1      T2      T3      T4      K1      K2      K3      K4
              3.0     20.0    20.0    24.0    0.58    0.98    0.98    0.50
```

Fish oxycaloric stoichiometric constant

This card specifies the oxycaloric stoichiometric constant, OXYCAL in the literature.

| Input | Type | Description |
|---------|------|---|
| FOXYCAL | REAL | Fish oxycaloric stoichiometric constant |

Example:

```
F STOICH FOXYCAL
          13569.4
```

Fish physical properties initial conditions

This card specifies the initial physical fish properties for all cells. Fish length is currently not used for any calculations and is a placeholder for future use.

| Input | Type | Description |
|---------|------|--|
| MASS | REAL | Initial fish mass, in grams, at the starting date |
| LENGTH | REAL | Initial fish length, in mm, at the starting date |
| STO_CON | REAL | Initial fish stomach content, in grams, at the starting date |

Example:

```
F1 IC      MASS  LENGTH  STO_CON
          10.0   193.0   0.001
```

Zooplankton property cards

This card specifies the zooplankton population physical properties. These properties are steady and uniform. Each zooplankton group can have different properties. The LZOO values is not used for any calculation and is a placeholder for future development.

| Input | Type | Description |
|-------|------|---|
| LZOO | REAL | Average length of the zooplankton group population, in mm. |
| MZOO | REAL | Average mass of the zooplankton group population, in g/m ³ . |
| EZOO | REAL | Average energy content of the zooplankton group population, in J/g. |

Example:

| | | | | | | | | |
|----------|--------|--------|---------|------|------|------|------|------|
| Z LENGTH | LZOO | LZOO | LZOO | LZOO | LZOO | LZOO | LZOO | LZOO |
| | 1.0 | 1.0 | 1.3 | | | | | |
| Z MASS | MZOO | MZOO | MZOO | MZOO | MZOO | MZOO | MZOO | MZOO |
| | 7.7e-3 | 4.9e-3 | 5.73e-5 | | | | | |
| Z STOICH | EZOO | EZOO | EZOO | EZOO | EZOO | EZOO | EZOO | EZOO |
| | 2420.0 | 2420.0 | 2420.0 | | | | | |

Foraging constants

This card specifies user-defined foraging constants. HANDLE is the average prey handling time. Values for kokanee near 1.0 second per prey item are common in the literature; a minimum of 0.33 s/prey (Stockwell and Johnson, 1999) is recommended for kokanee. The Allometric fish velocity function can be defined by the user using FVELA, FVELB, and FVELC in the form of the velocity (m/s) shown below (fish mass, M, in grams; water temperature, T, in Celsius.)

$$v = A \cdot \exp(C \cdot T) \cdot M^B \quad (17)$$

At a minimum prey density threshold, selective (one prey group) feeding may be prescribed by THRESHC. The zooplankton group number, THRESHZ, and the critical density, THRESHV, are user defined.

When using the 'vertical foraging' strategy, activated by DIELC, the user may define daylight conditions using the DIELLUX value, in lux. This is the irradiance value in the surface layer of the model. As the sun rises and sets quickly, the model is not sensitive to small (one order of magnitude) changes in the value of DIELLUX.

| Input | Type | Description |
|---------|---------|--|
| HANDLE | REAL | Average prey handling time, in seconds |
| FVELA | REAL | Allometric fish velocity parameter, A |
| FVELB | REAL | Allometric fish velocity parameter, B |
| FVELC | REAL | Allometric fish velocity parameter, C |
| THRESHC | ON/OFF | Turns on preying upon a single zooplankton group at a threshold prey density |
| THRESHV | REAL | Critical prey density for threshold feeding |
| THRESHZ | INTEGER | The zooplankton group number for threshold feeding |
| DIELLUX | REAL | User defined light intensity for daylight conditions, in lux |

Example:

| FORAGE | HANDLE | FVELA | FVELB | FVELC | THRESHC | THRESHV | THRESHZ | DIELLUX |
|--------|--------|-------|-------|---------|---------|---------|---------|---------|
| | 0.50 | 0.099 | 0.13 | 0.04505 | ON | 6.032 | 3 | 0.01 |

Fish growth potential animation cards

This card allows for the fish growth potential output in a format ready for use by the TECPLOT software for the generation of graphic animations. Only one date, FGPD, is supported.

| Input | Type | Description |
|-------|--------|---|
| FGPC | ON/OFF | Turns on the fish growth potential output file. |
| FGPD | REAL | Specifies the starting Julian date of the fish growth potential output file |
| FGPF | REAL | Specifies the frequency of output to the fish growth potential output file, in days |

Example:

```
FGP PLOT   FGPC
           ON

FGP DATE   FGPD
           367.0

FGP FREQ   FGPF
           1.0
```

Fish mass type

This card specifies the method used to change the fish mass with time. The general algorithm uses CELL, FIXED, FXN, USER. The ‘best’ and ‘vertical foraging’ algorithms use SEG, FIXED, FXN, and USER.

FIXED allows the fish mass to remain constant at the initial value. FXN prescribes fish growth as an exponential function (see *fish mass function parameters*). This is useful for reproducing growth from field data. USER allows a user defined mass time-series, which need not be a “nice” mathematical function. CELL allows the daily predicted fish growth potential for each model cell to be applied to that cell. Similarly, SEG allows the predicted growth to be applied to *all* cells in the model segment.

| Input | Type | Description |
|-------|-----------|---|
| GEN | CHARACTER | Determines the method used to describe the fish mass time-series for the general function |
| BEST | CHARACTER | Determines the method used to describe the fish mass time-series for the best growth function |
| DIEL | CHARACTER | Determines the method used to describe the fish mass time-series for the vertical foraging function |

Example:

```
MASSTYPE      GEN      BEST      DIEL
              FIXED    SEG      FXN
```

Fish mass function parameters

This card specifies the parameters of the prescribed exponential fish mass time-series function. The form of the equation is shown below. M and IM are in grams; TI is in days.

$$M = IM \cdot \exp(ALP(JDAY - TI)) \quad (18)$$

| Input | Type | Description |
|-------|------|---|
| GIM | REAL | General function (by cell) initial mass |
| BIM | REAL | “Best growth” (by segment) initial mass |
| DIM | REAL | “Vertical foraging” (by segment) initial mass |
| GALP | REAL | General function alpha parameter |
| BALP | REAL | “Best growth” function alpha parameter |
| DALP | REAL | “Vertical foraging” function alpha parameter |
| GTI | REAL | General function initial time parameter |
| BTI | REAL | “Best growth” function initial time parameter |
| DTI | REAL | “Vertical foraging” function initial time parameter |

Example:

| | | | | | | | | | |
|----------|------|------|------|--------|--------|--------|-------|-------|-------|
| MASS FXN | GIM | BIM | DIM | GALP | BALP | DALP | GTI | BTI | DTI |
| | 76.0 | 76.0 | 76.0 | 0.0049 | 0.0049 | 0.0049 | 366.0 | 366.0 | 366.0 |

Diagnostic output controls

This card activates the diagnostic output. FISHC must be ON for any of the other outputs. These features are best used with the single location feature as output from all active segments will be reported in a single file. Output is reported daily at midnight.

| Input | Type | Description |
|---------|--------|---|
| FISHC | ON/OFF | Turns on all diagnostic output |
| BIOPARC | ON/OFF | Turns on the bioenergetics parameter output |
| CONSC | ON/OFF | Turns on the consumption diagnostic output |
| DIGC | ON/OFF | Turns on the digestion diagnostic output |
| RESPC | ON/OFF | Turns on the respiration diagnostic output |
| SURFC | ON/OFF | Turns on selected output for the surface cell |
| DEPTHC | ON/OFF | Turns on the reporting of the foraging layers for the best growth and vertical foraging routines. |

Example:

| | | | | | | | |
|----------|-------|---------|-------|------|-------|-------|--------|
| DIAG OPT | FISHC | BIOPARC | CONSC | DIGC | RESPC | SURFC | DEPTHC |
| | ON | ON | ON | ON | ON | ON | ON |

Single location controls

This card allows for computation of the bioenergetics routine at a single segment. This allows for faster computations. The TLC feature allows reporting of the calculated Thornton-Lessem function values. TLC should be set to OFF unless needed. To limit file size, the values are reported at midnight for the surface layer. Output is also only available if SINGLEC is ON.

| Input | Type | Description |
|---------|---------|--|
| SINGLEC | ON/OFF | Turns on fish calculations at a single location |
| SINGFN | INTEGER | Specifies the input file number for single location calculations |
| SINIBIO | INTEGER | Specifies the CE-QUAL-W2 model segment number for single location calculations |
| TLC | ON/OFF | Turns on the Thornton-Lessem function value output file. |

Example:

| | | | | |
|----------|---------|--------|---------|-----|
| SINGLEPT | SINGLEC | SINGFN | SINIBIO | TLC |
| | ON | 143 | 312 | OFF |

User defined input and output filenames

This card allows the user to specify some of the output filenames. ZAV is the zooplankton biological availability fraction (0 to 1) filename. The TECPLOT format fish growth potential file and the user defined fish mass time-series files can also be named. The filenames, including the file extension, cannot exceed 72 characters.

| Input | Type | Description |
|------------------|-----------|---|
| ZAV FILE | CHARACTER | The zooplankton availability input filename |
| FGP FILE | CHARACTER | The fish growth output filename |
| GEN FXN FILE | CHARACTER | The general (by cell) user defined fish mass input file |
| BEST FXN FILE | CHARACTER | The "best growth" (by segment) user defined fish mass input file |
| DIEL FXN FILE | CHARACTER | The "vertical foraging" (by segment) user defined fish mass input file |

Example:

```
ZAV FILE.....ZAVFN.....  
      Zav.npt  
  
FGP FILE.....FGPFN.....  
WB1      FGP_wb1.opt  
  
FXN FILE.....FXNFN.....  
GEN      Gen_userfxn.npt  
BEST     Best_userfxn.npt  
DIEL     Diel_userfxn.npt
```

Appendix E: Plots of weighted model results and model-data comparisons

Weighted algae and zooplankton are shown in the Biotic Modeling Approach and Calibration section.

Water quality constituent calibration is discussed in the section, Constituent Calibration Discussion. The discrepancy in model-data comparisons of pH are discussed in the section Alternate Boundary Conditions.

The sampling locations were shown in Figure 2 in the Monitoring Sites section.

The water quality data are often collected at varied depths. When a notable vertical gradient is exhibited in the data or the model results, the exact depth can be important. To better compare the model and data given the potential uncertainty in vertical location in the water column, model results are averaged over the upper 5 model layers (approximately 10 m). Values at the surface layer (from the *tsr.opt* output files) are also shown to give a sense of the vertical gradient exhibited by the model.

Model-data comparison plots of orthophosphate, ammonium, nitrate plus nitrite (NO_x), dissolved oxygen, alkalinity, pH, and total dissolved solids are shown where data are available.

Station 0.0

The water quality constituent concentrations at station 0.0 predicted by the model are largely an echo of the upstream boundary conditions. With the exception of pH, the boundary conditions are generated from the data at station 0.0. The travel time from the U.S.-Canadian border to station 0.0 ranges from roughly 0.5 days during spring drawdown to 2 days during the fall. As such, there is little time to influence the constituent concentrations.

The stair stepping of the model concentrations seen in the orthophosphate (Figure 114) and ammonium (Figure 115) is a result of the model input file having concentrations in 3 decimal places (0.001 increments).

The model exhibits a vertically well-mixed water column at station 0.0.

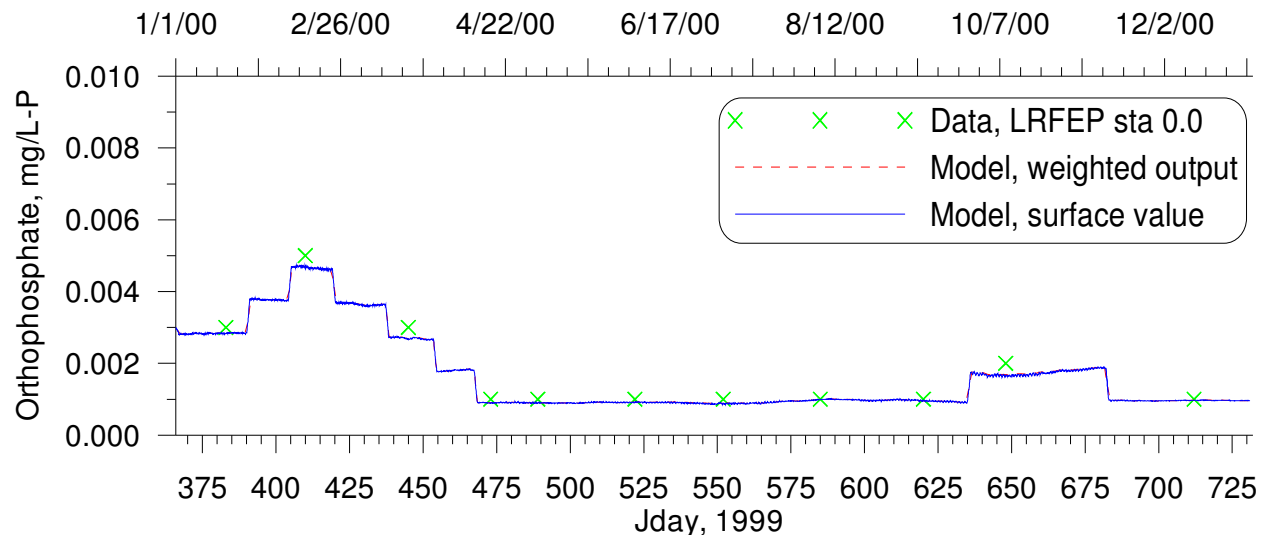


Figure 114. Model-data comparison of orthophosphate at LRFEP station 0.0.

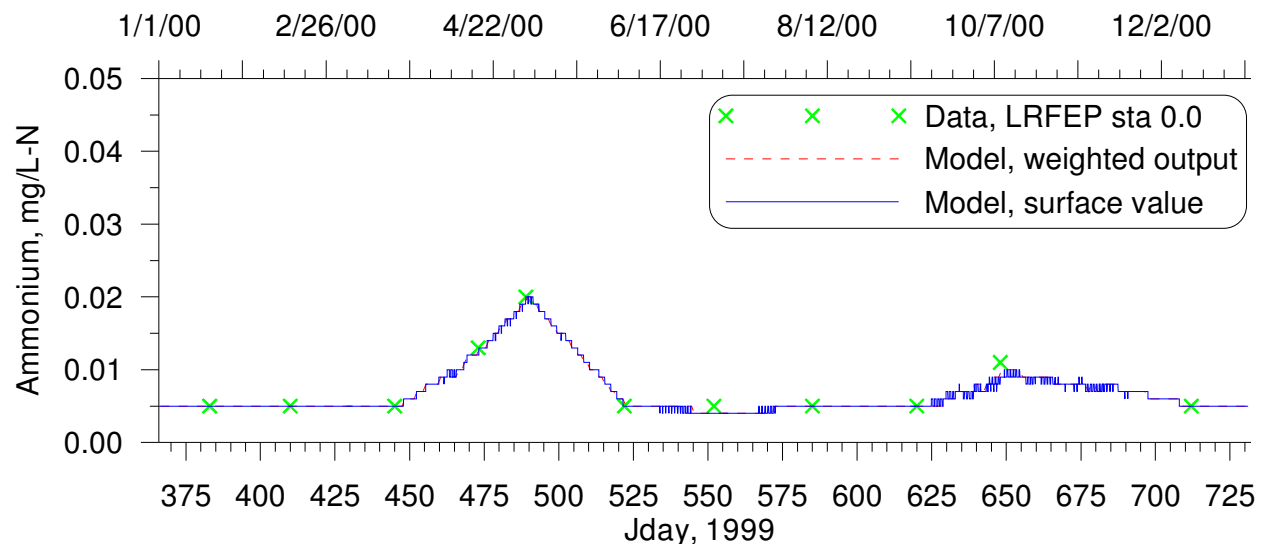


Figure 115. Model-data comparison of ammonium at LRFEP station 0.0.

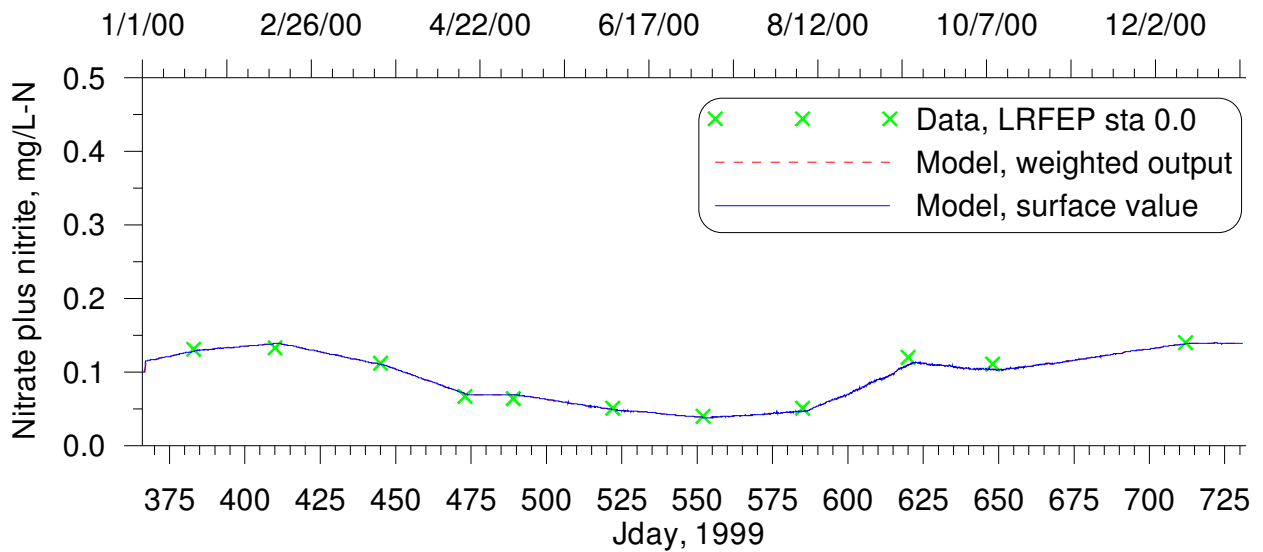


Figure 116. Model-data comparison of nitrate plus nitrite at LRFEP station 0.0.

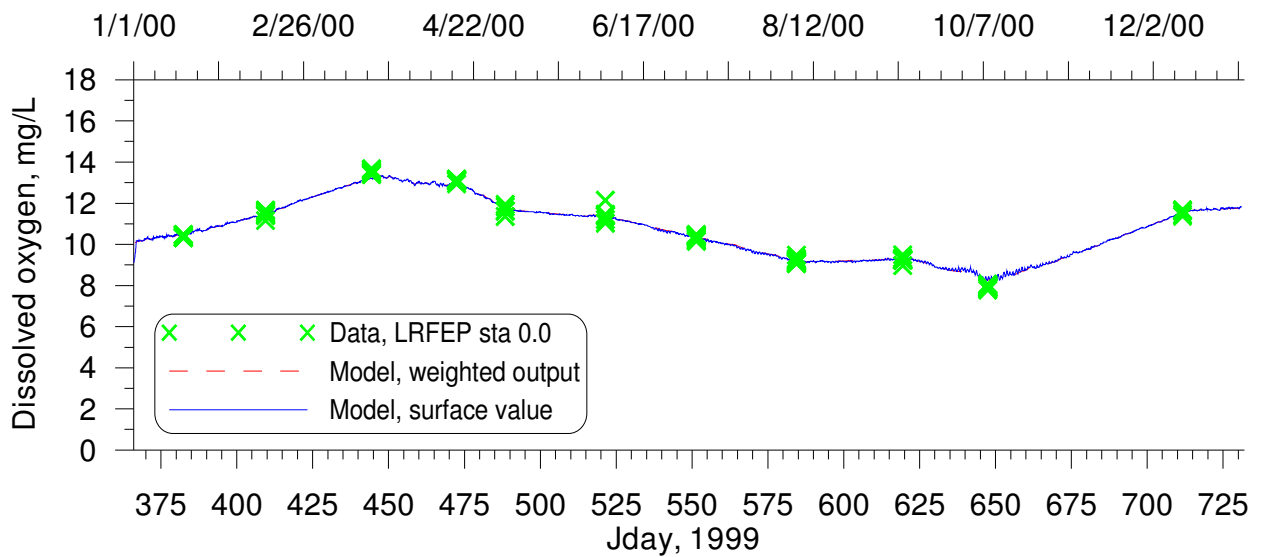


Figure 117. Model-data comparison of dissolved oxygen at LRFEP station 0.0.

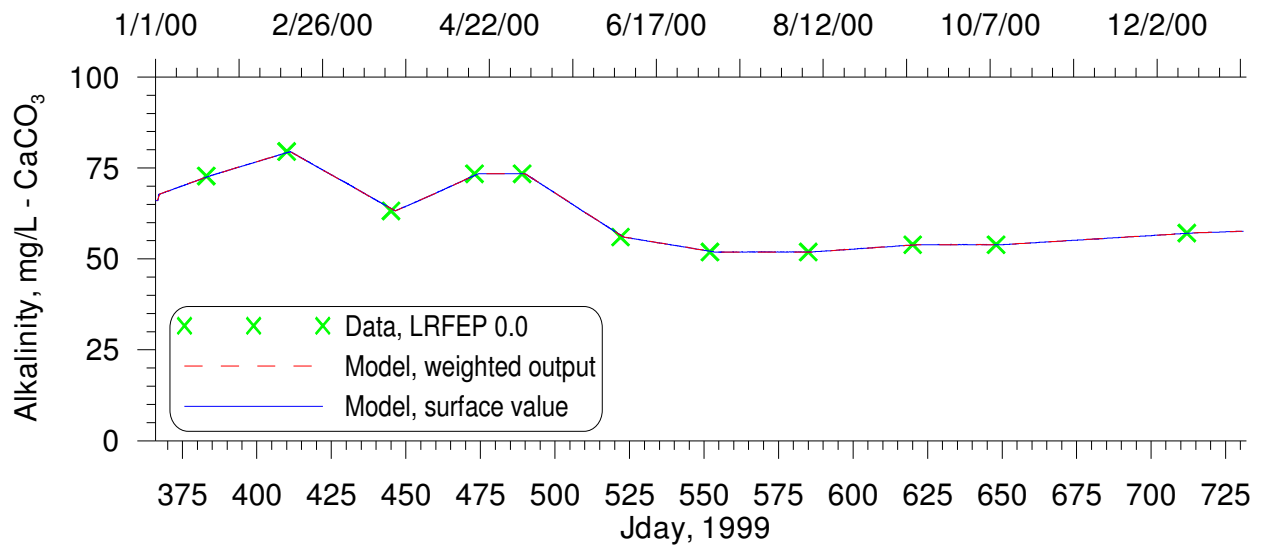


Figure 118. Model-data comparison of alkalinity at LRFEP station 0.0.

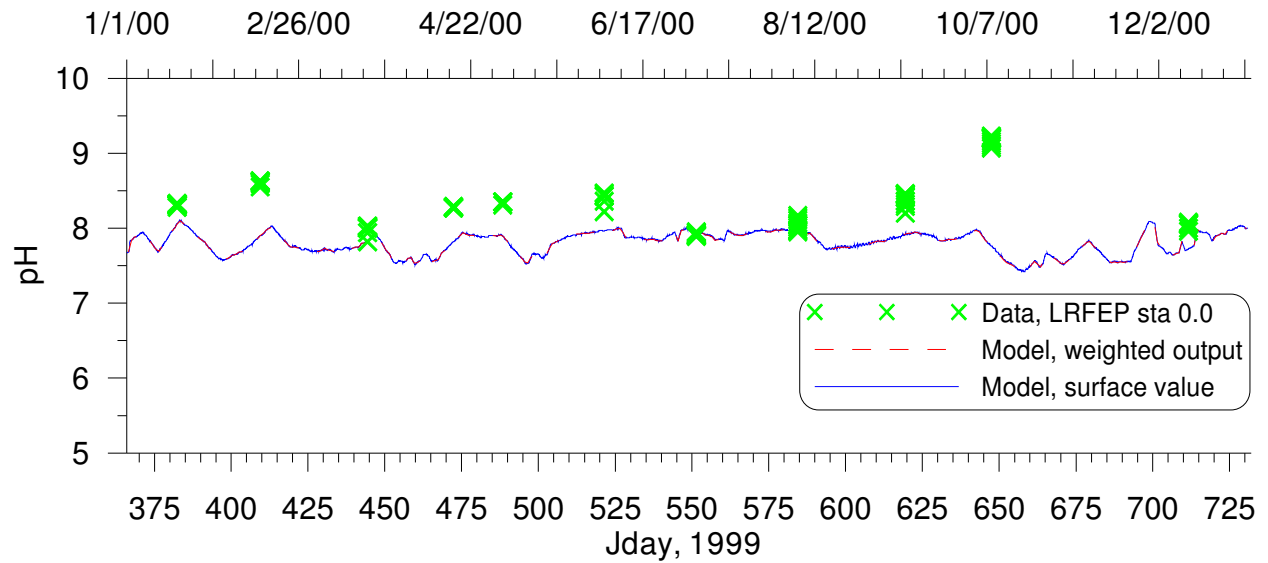


Figure 119. Model-data comparison of pH at LRFEP station 0.0.

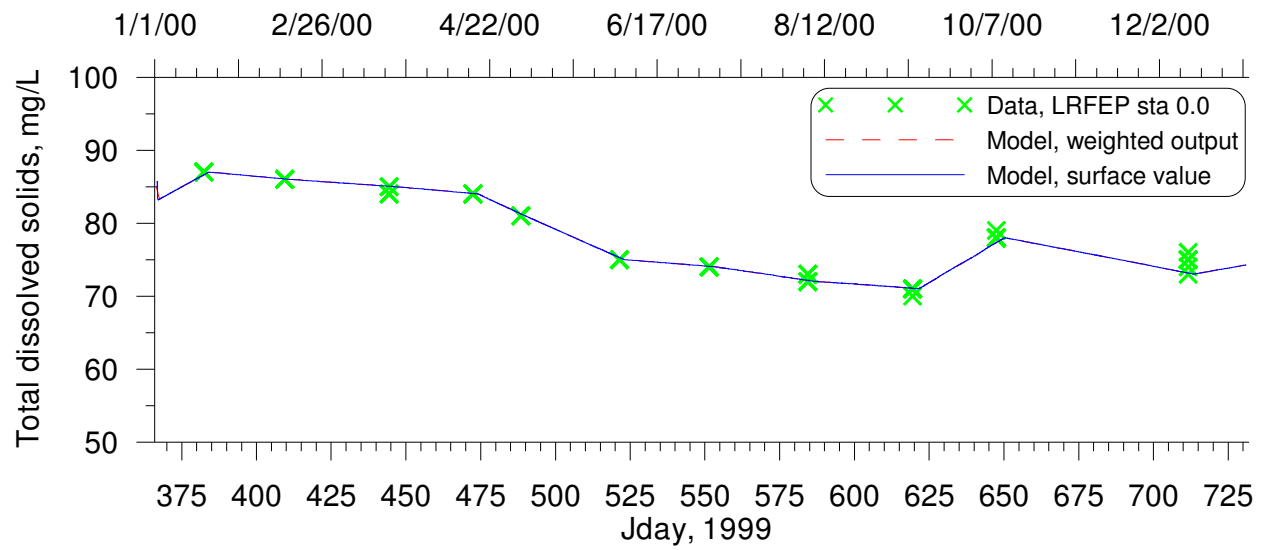


Figure 120. Model-data comparison of total dissolved solids at LRFEP station 0.0.

Station 1.0

The travel time to station 1.0 ranges roughly from 0.6 to 4 days from the U.S.-Canadian border, which is only slightly larger than the travel time to the nearby station 0.0. The water quality constituent concentrations are still largely an echo of the upstream boundary conditions. The model exhibits a vertically well-mixed water column.

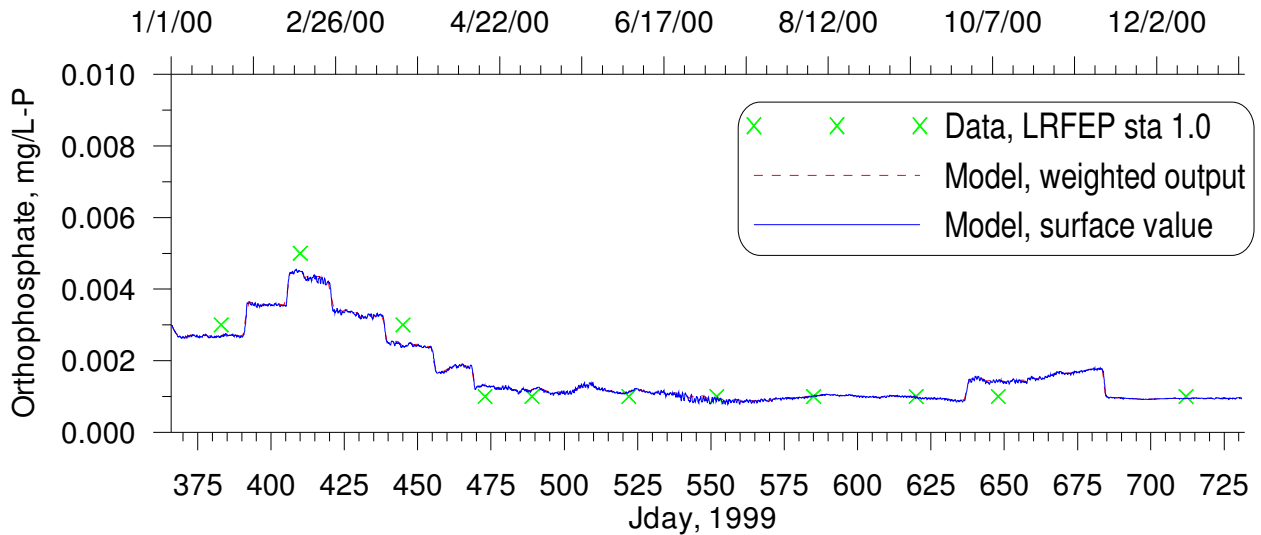


Figure 121. Model-data comparison of orthophosphate at LRFEP station 1.0.

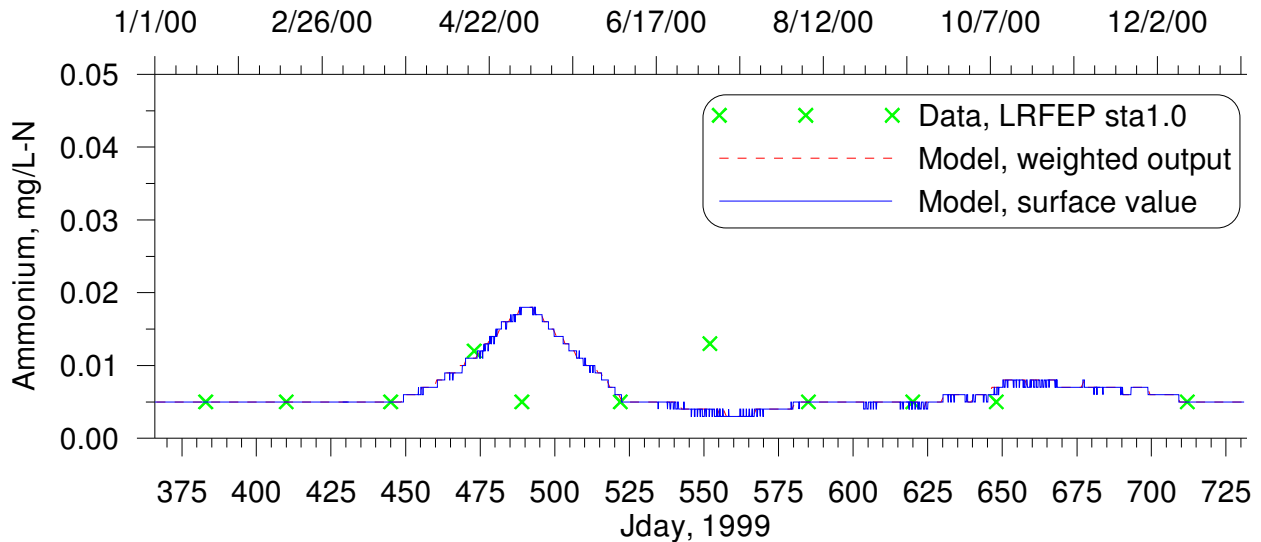


Figure 122. Model-data comparison of ammonium at LRFEP station 1.0.

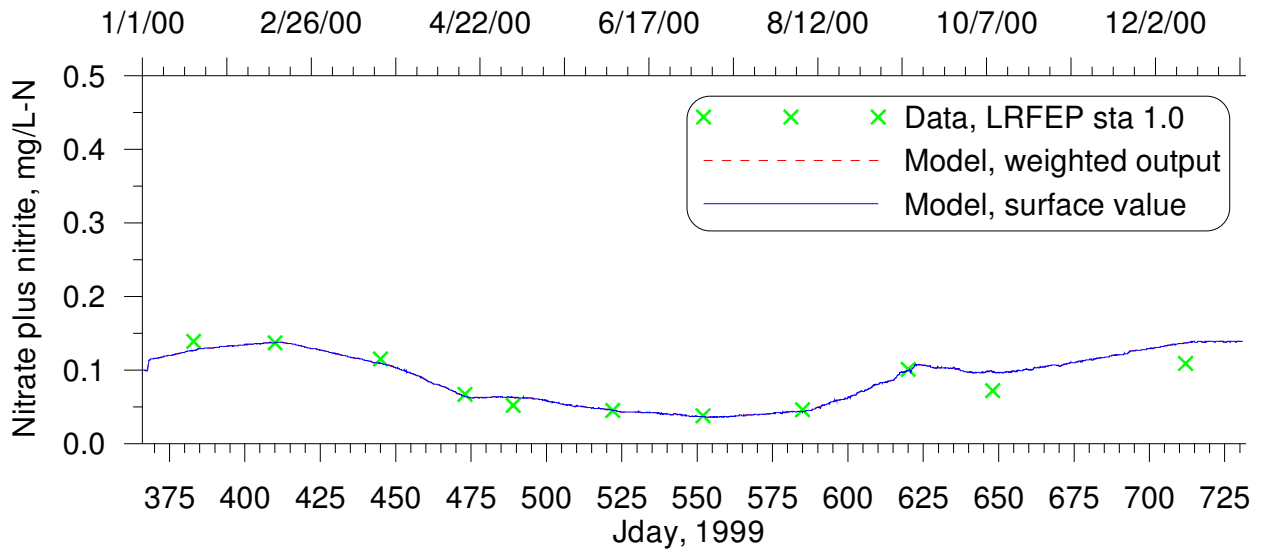


Figure 123. Model-data comparison of nitrate plus nitrite at LRFEP station 1.0.

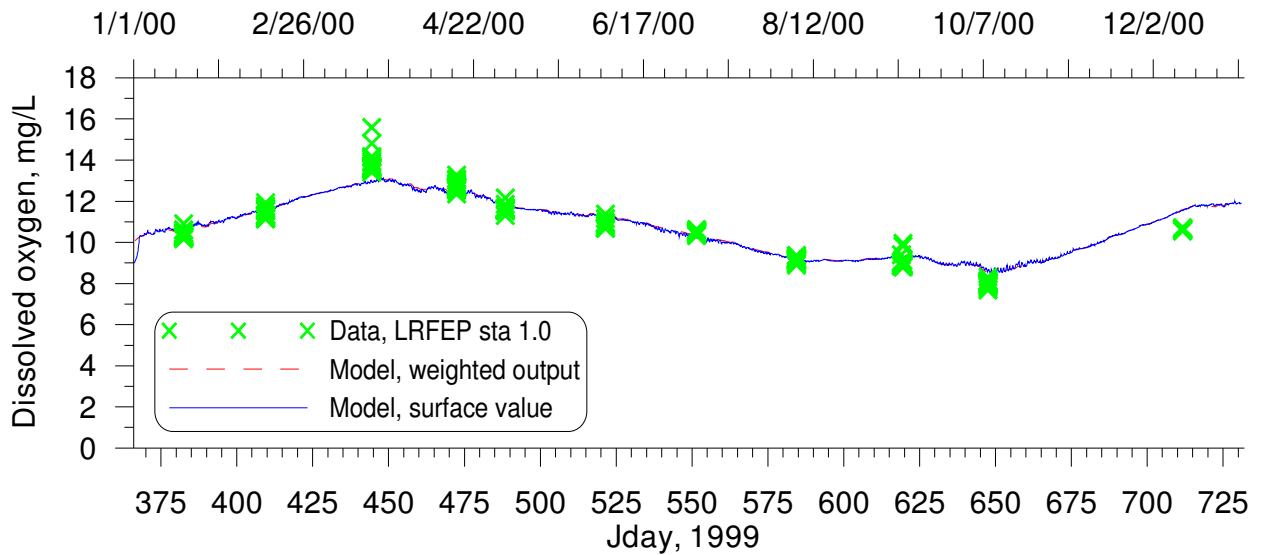


Figure 124. Model-data comparison of dissolved oxygen at LRFEP station 1.0.

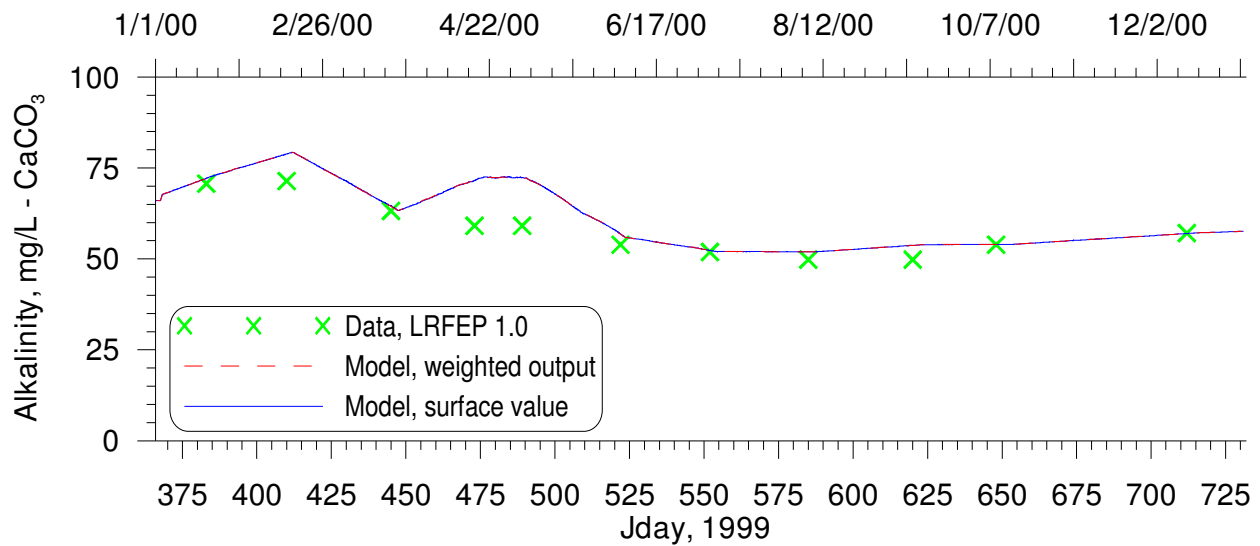


Figure 125. Model-data comparison of alkalinity at LRFEP station 1.0.

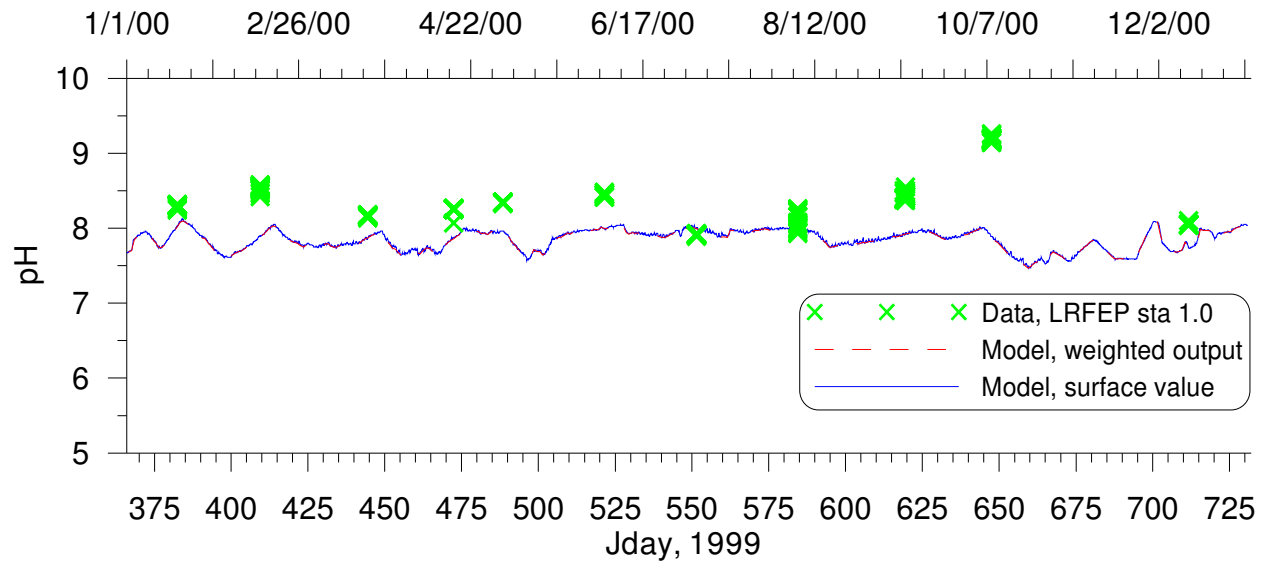


Figure 126. Model-data comparison of pH at LRFEP station 1.0.

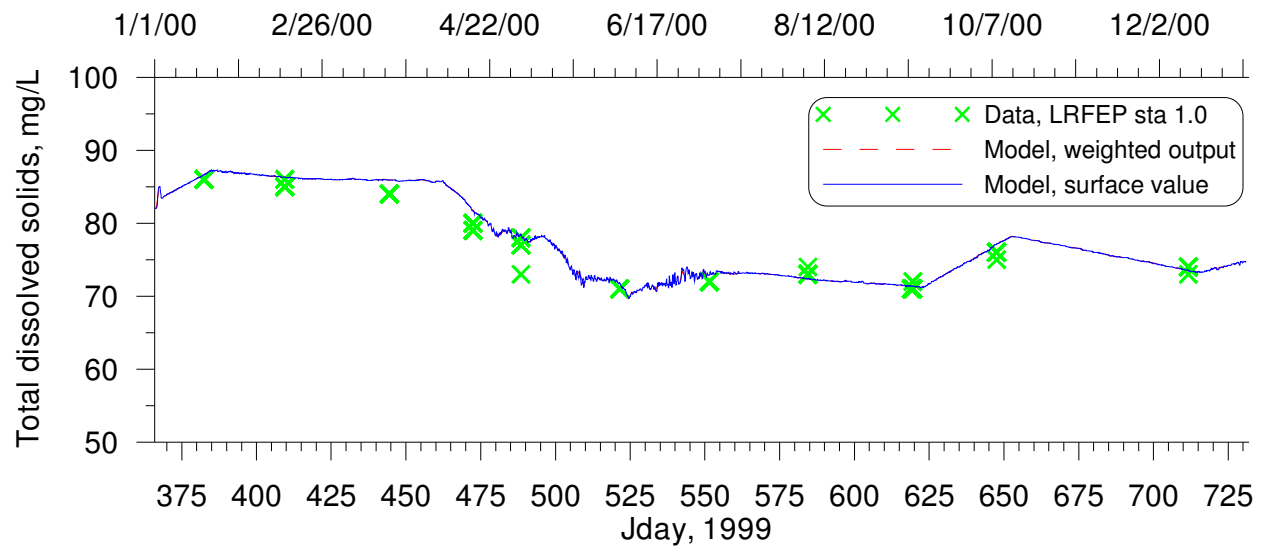


Figure 127. Model-data comparison of total dissolved solids at LRFEP station 1.0.

Station 2.0

The travel time to station 2.0 ranges roughly from 5 to 20 days from the U.S.-Canadian border. The model exhibits some small and isolated vertical gradients; the data show larger vertical gradients, however, especially for the dissolved oxygen (Figure 131) and total dissolved solids (TDS) (Figure 134) constituents. The TDS data at station 2.0 have much higher values than the upstream and downstream stations during spring drawdown. This could be the result of the nearby upstream tributary streams.

The stair stepping seen in the ammonium concentrations (Figure 129) is a result of the model output having 3 decimal places.

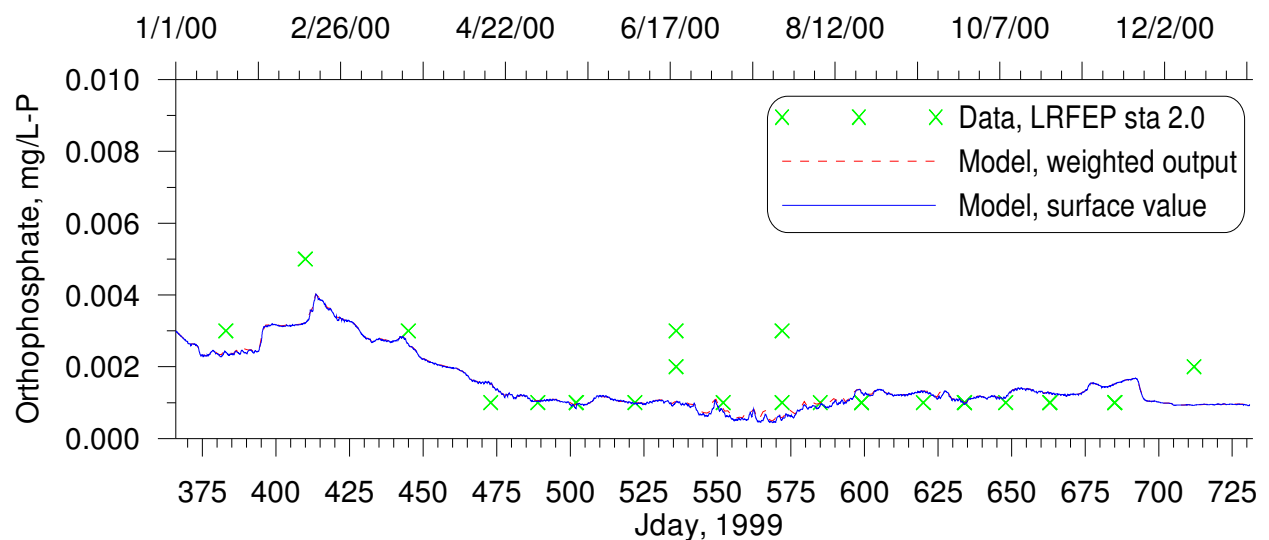


Figure 128. Model-data comparison of orthophosphate at LRFEP station 2.0.

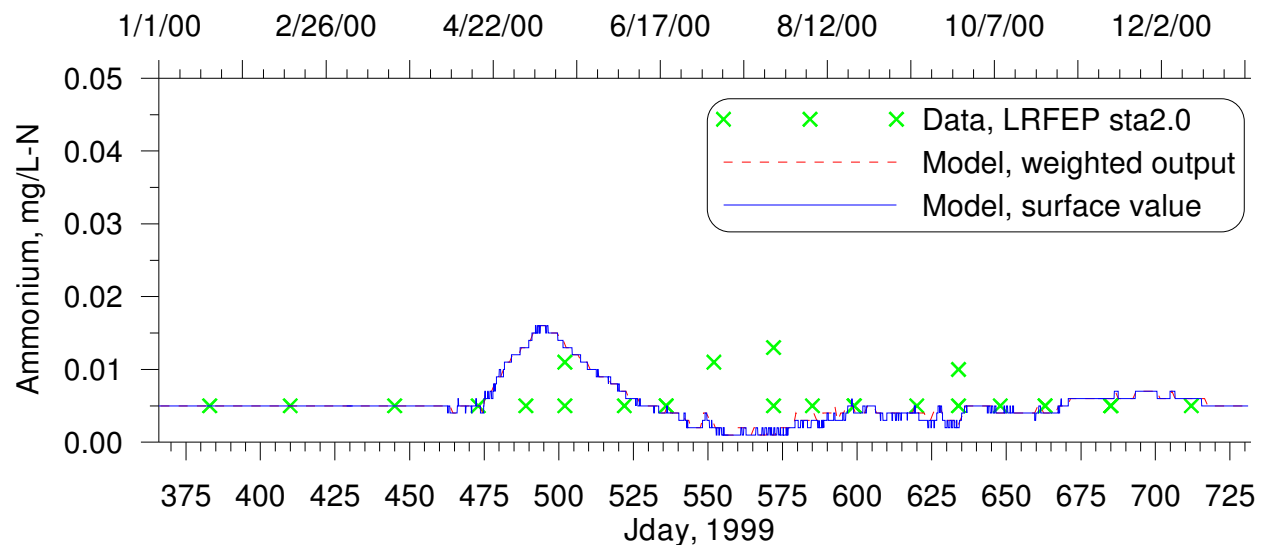


Figure 129. Model-data comparison of ammonium at LRFEP station 2.0.

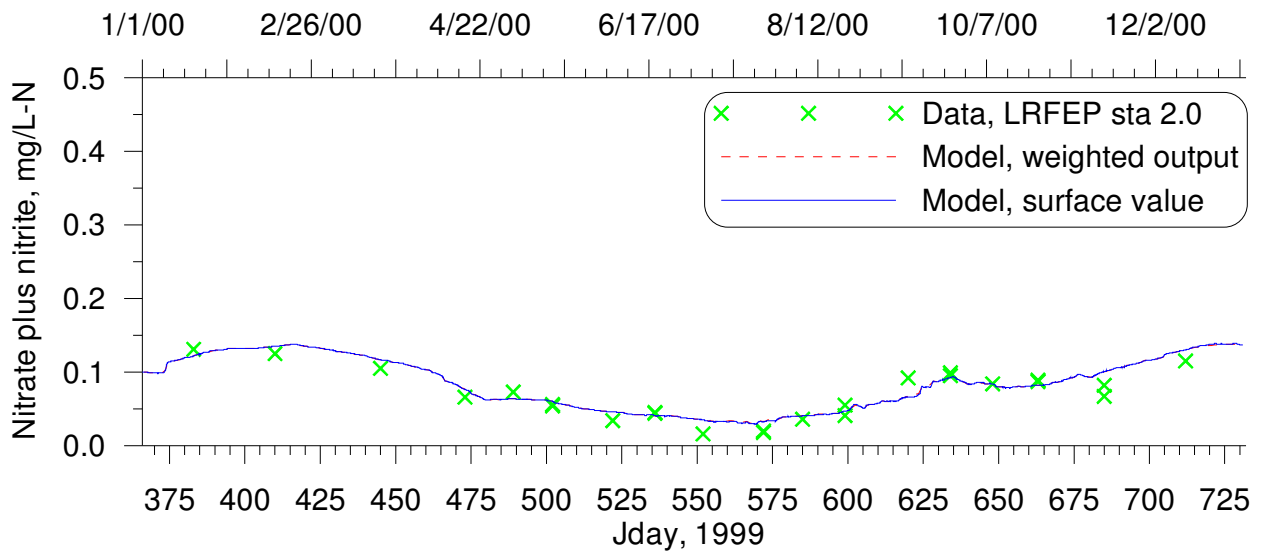


Figure 130. Model-data comparison of nitrate plus nitrite at LRFEP station 2.0.

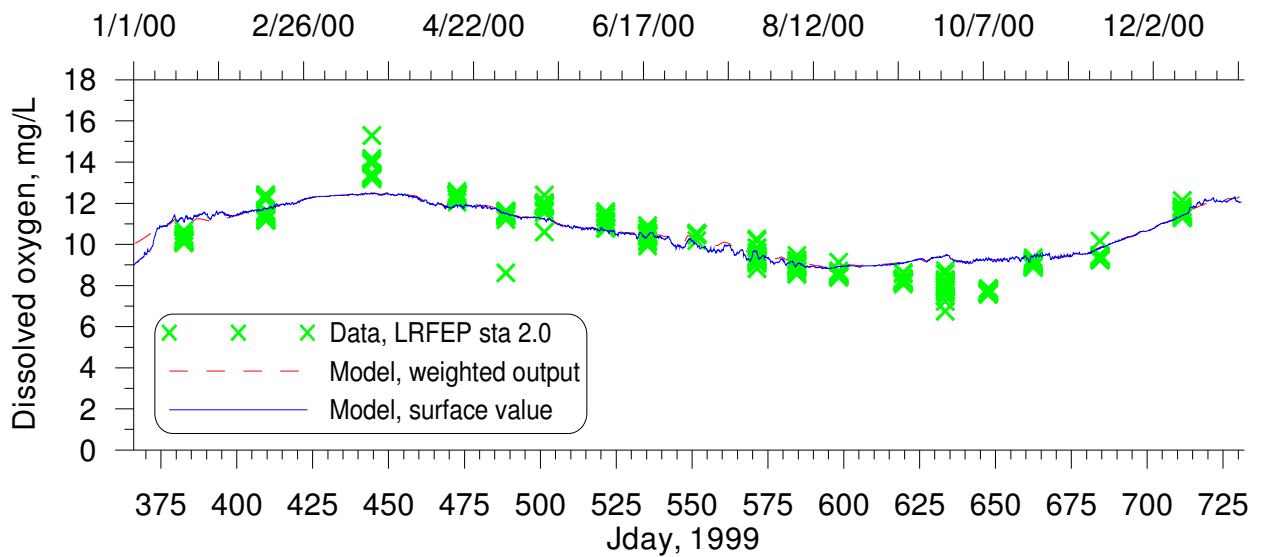


Figure 131. Model-data comparison of dissolved oxygen at LRFEP station 2.0.

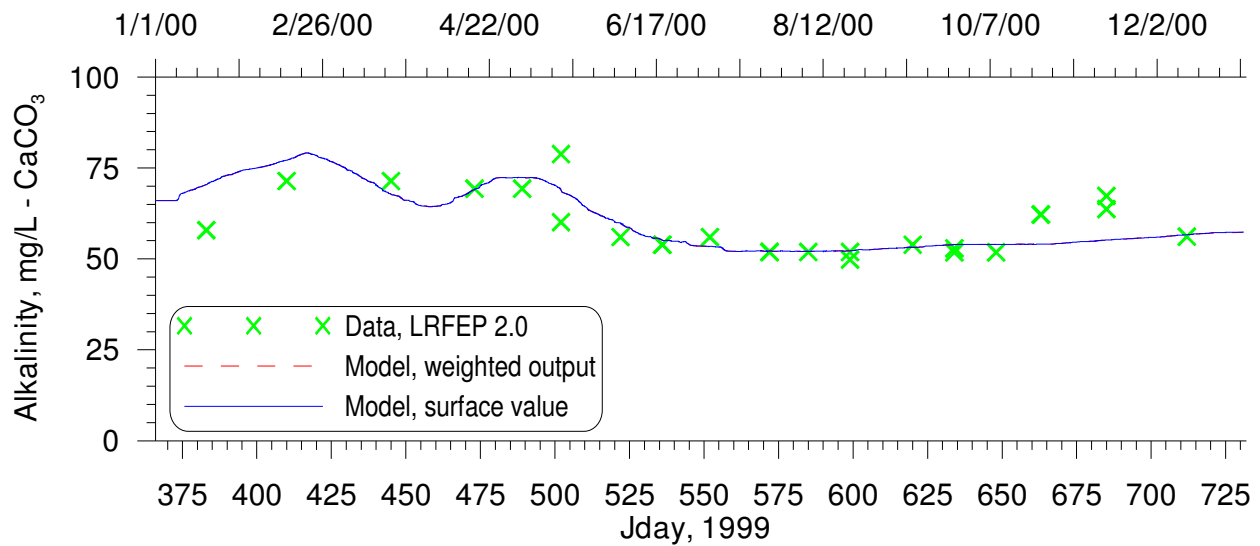


Figure 132. Model-data comparison of alkalinity at LRFEP station 2.0.

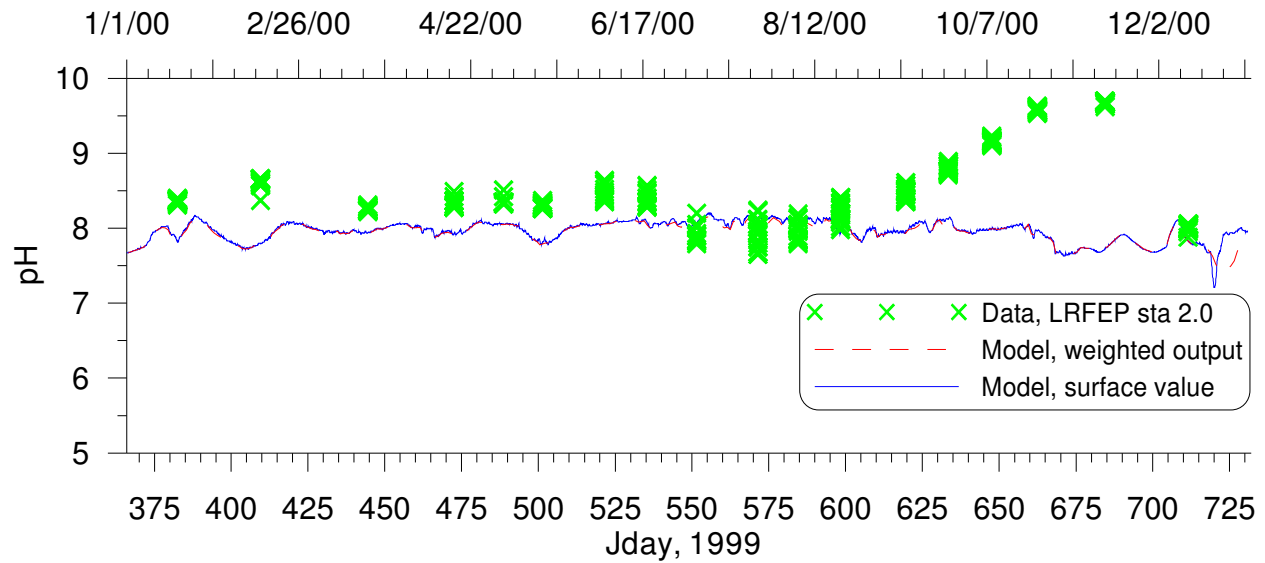


Figure 133. Model-data comparison of pH at LRFEP station 2.0.

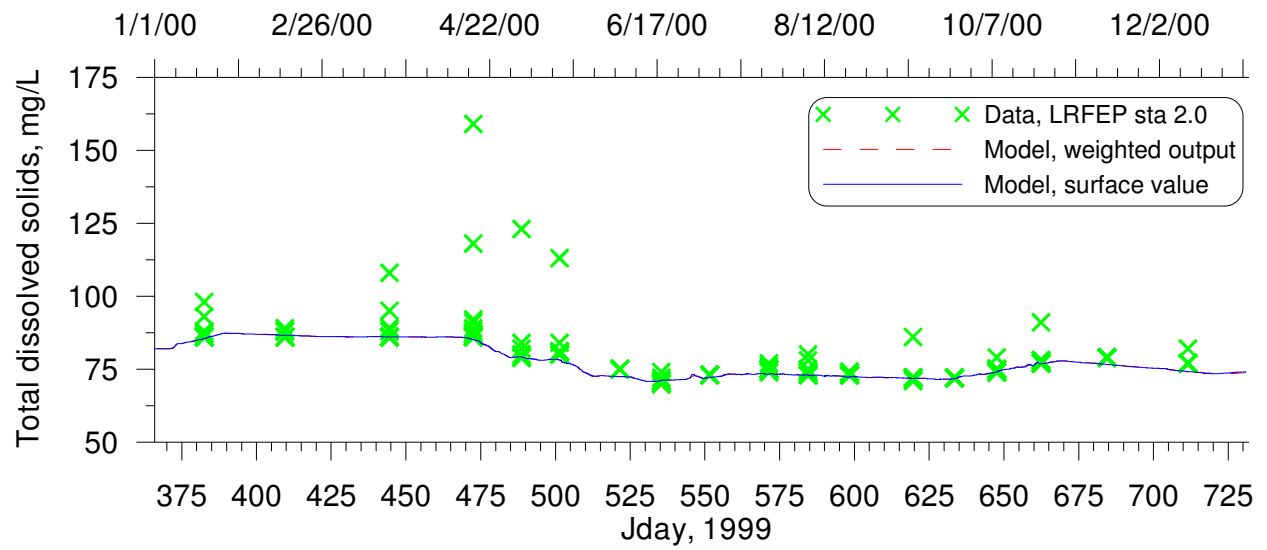


Figure 134. Model-data comparison of total dissolved solids at LRFEP station 2.0.

Station 3.0

The travel time to station 3.0 ranges roughly from 6 to 22 days from the U.S.-Canadian border. The model exhibits some small and isolated vertical gradients; the data show larger vertical gradients, however.

The stair stepping seen in the ammonium concentrations (Figure 136) is a result of the model output having 3 decimal places.

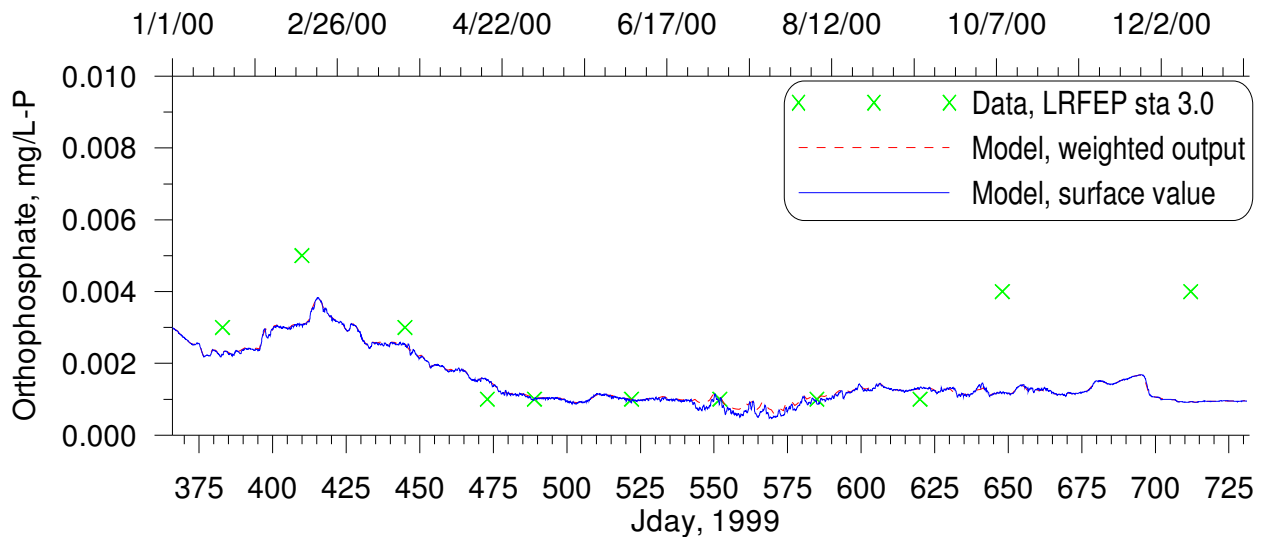


Figure 135. Model-data comparison of orthophosphate at LRFEP station 3.0.

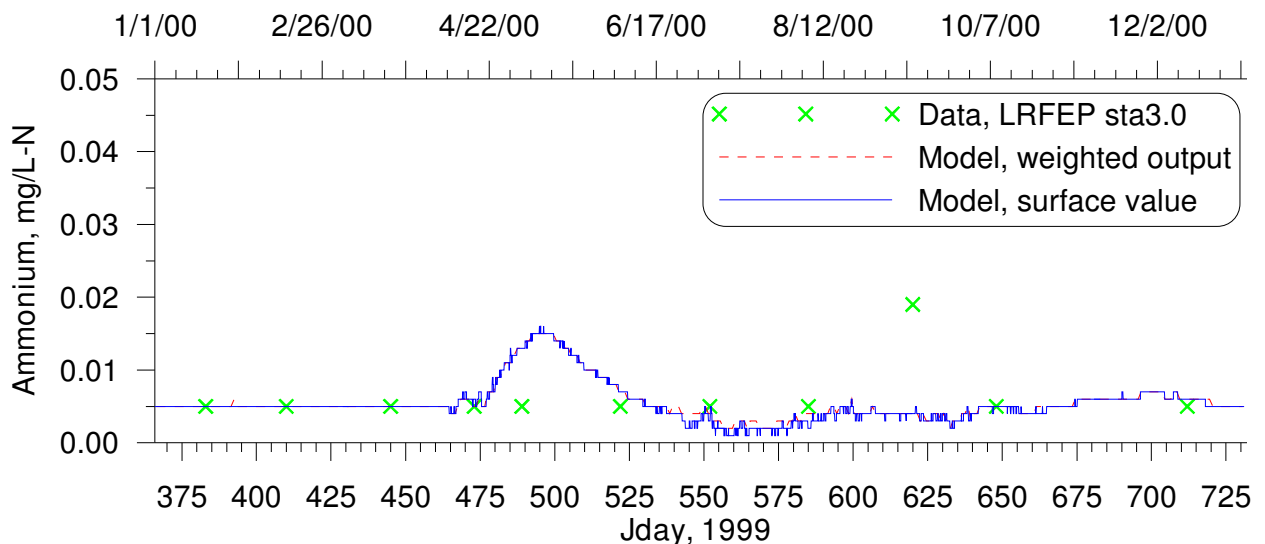


Figure 136. Model-data comparison of ammonium at LRFEP station 3.0.

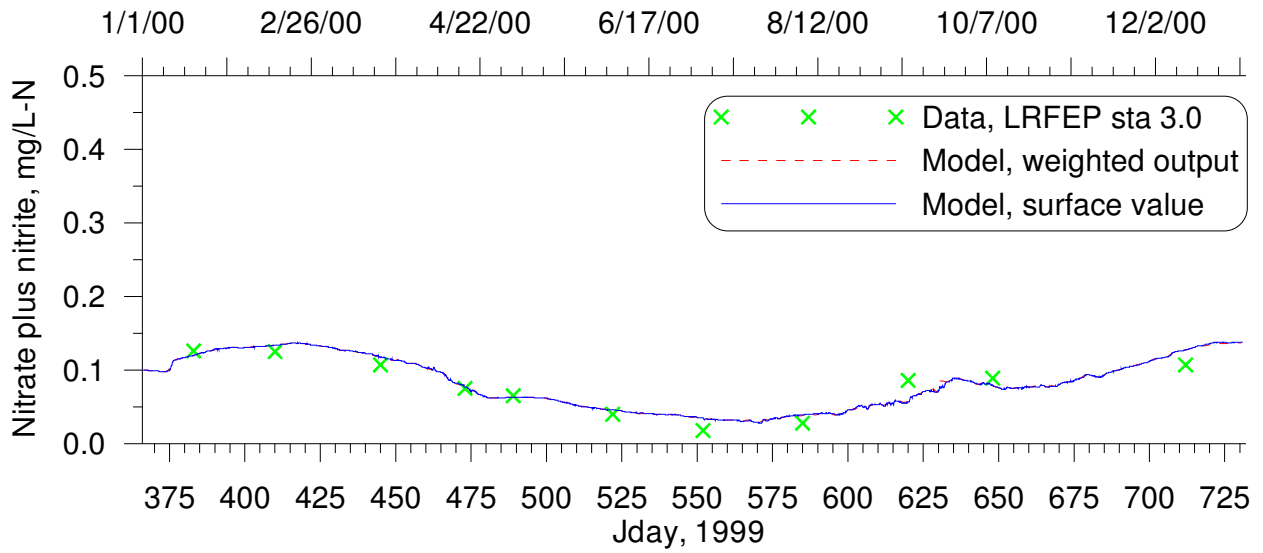


Figure 137. Model-data comparison of nitrate plus nitrite at LRFEP station 3.0.

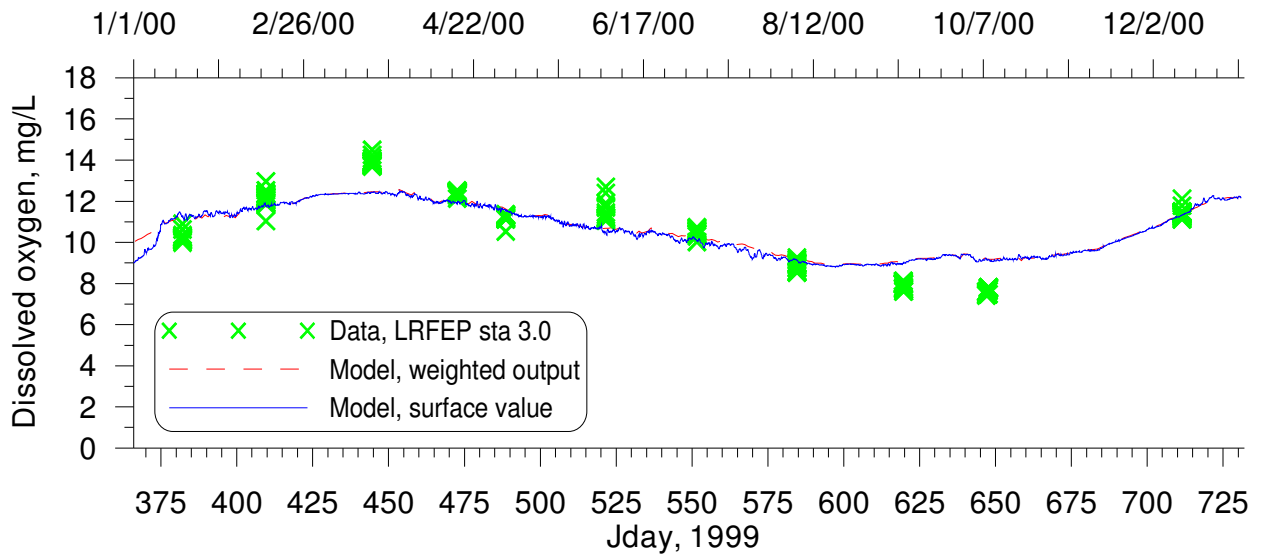


Figure 138. Model-data comparison of dissolved oxygen at LRFEP station 3.0.

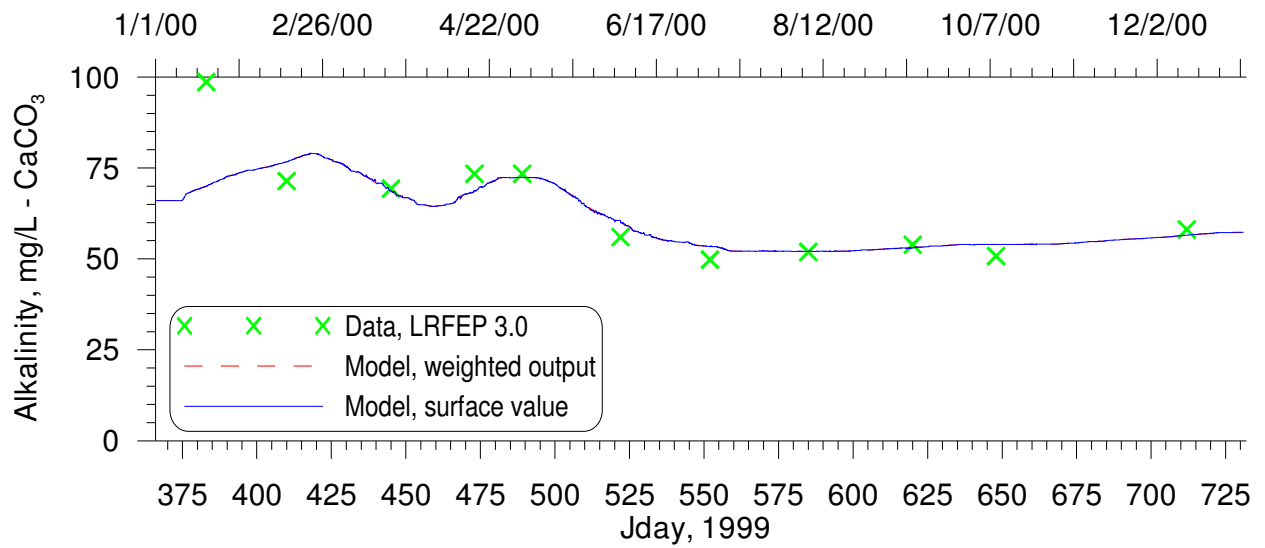


Figure 139. Model-data comparison of alkalinity at LRFEP station 3.0.

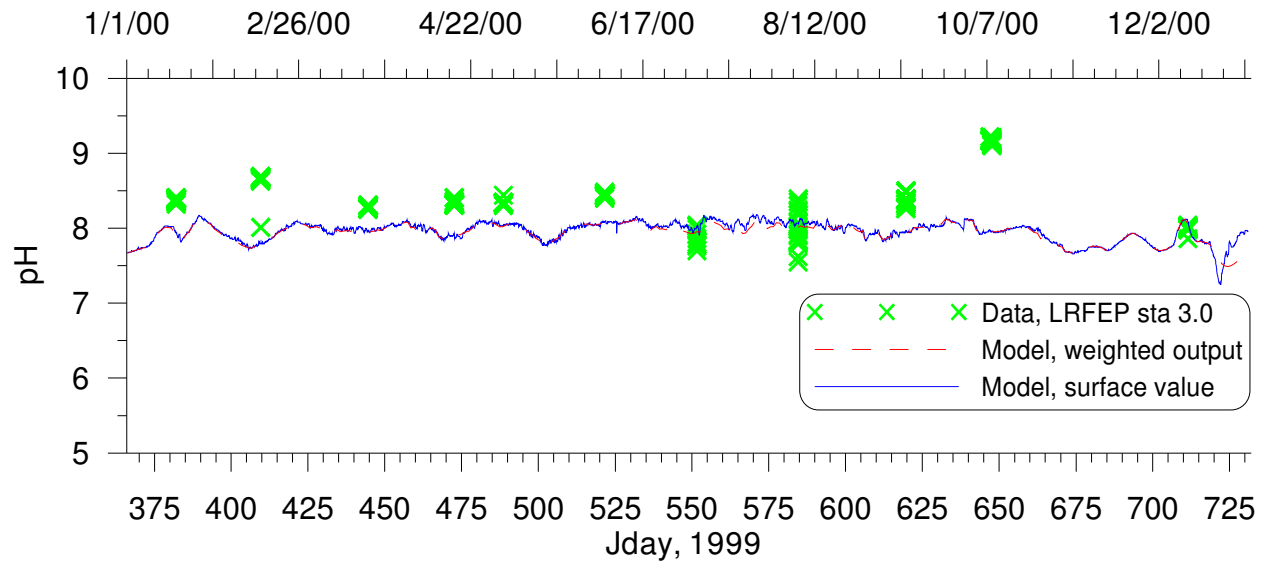


Figure 140. Model-data comparison of pH at LRFEP station 3.0.

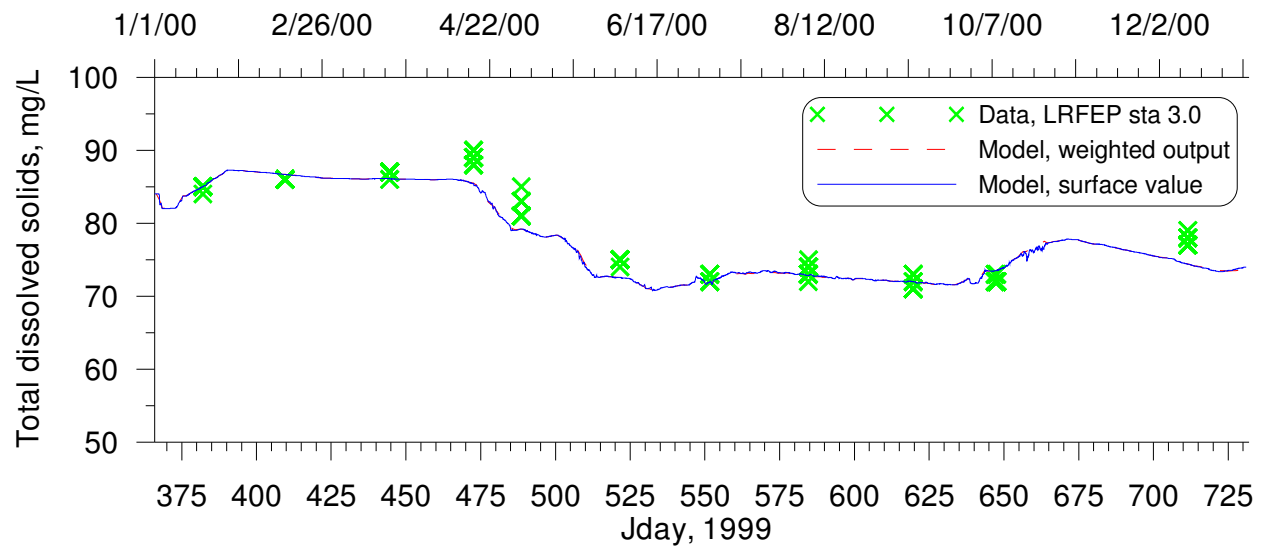


Figure 141. Model-data comparison of total dissolved solids at LRFEP station 3.0.

Station 4.0

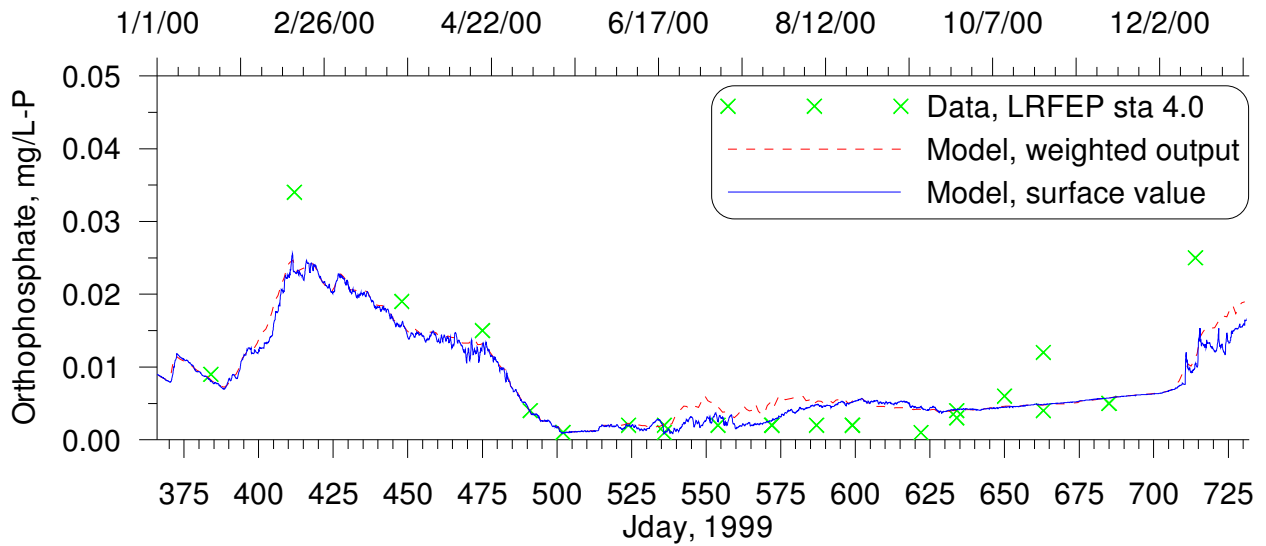


Figure 142. Model-data comparison of orthophosphate at LRFEP station 4.0.

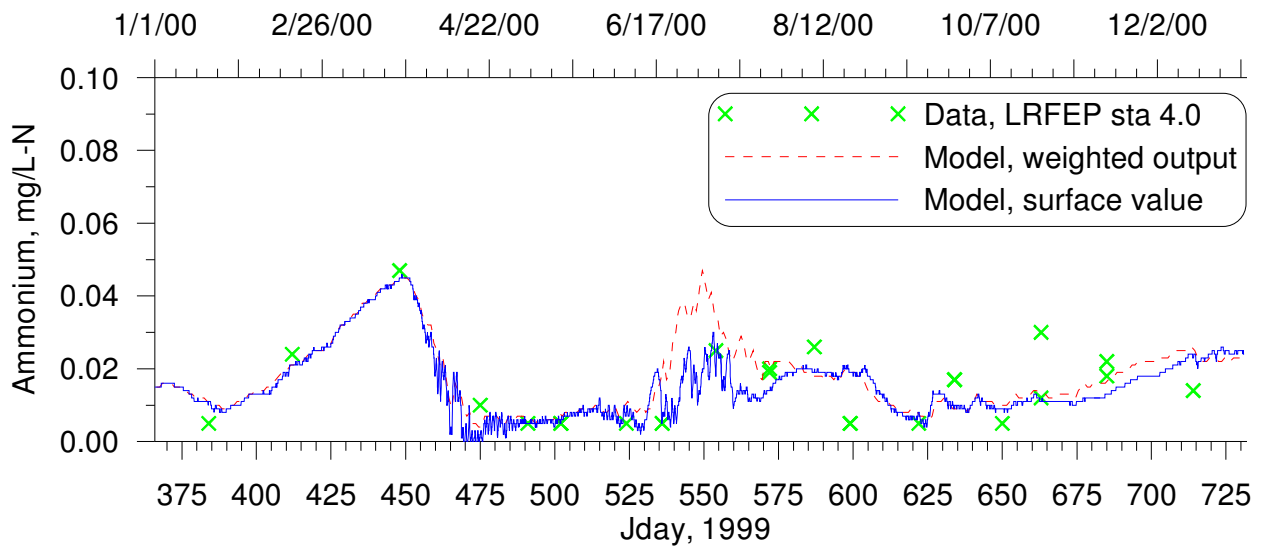


Figure 143. Model-data comparison of ammonium at LRFEP station 4.0.

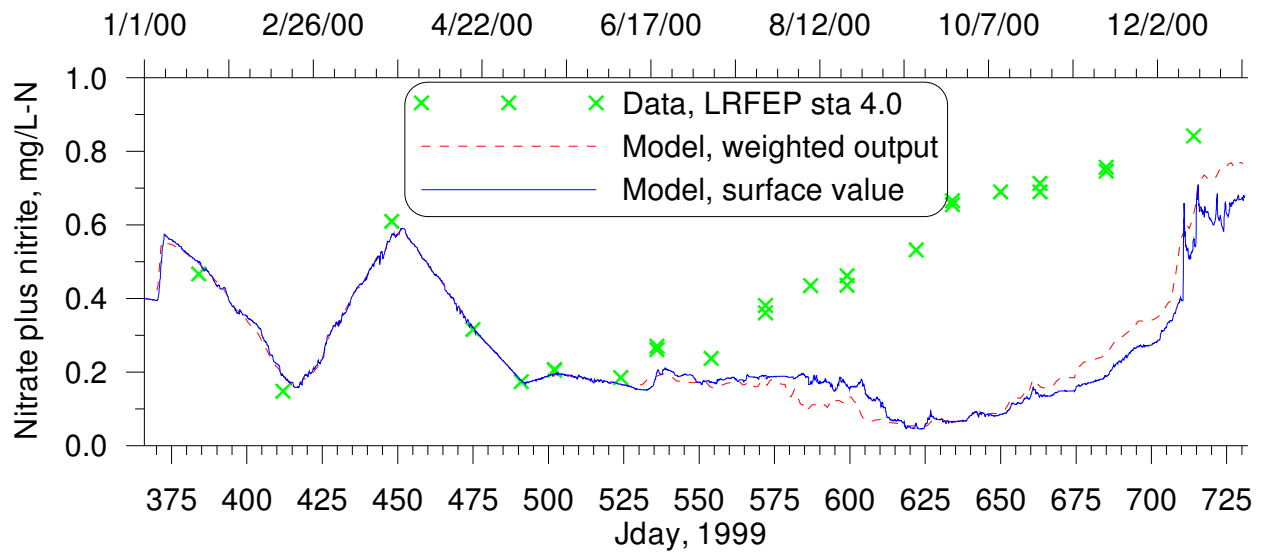


Figure 144. Model-data comparison of nitrate plus nitrite at LRFEP station 4.0.

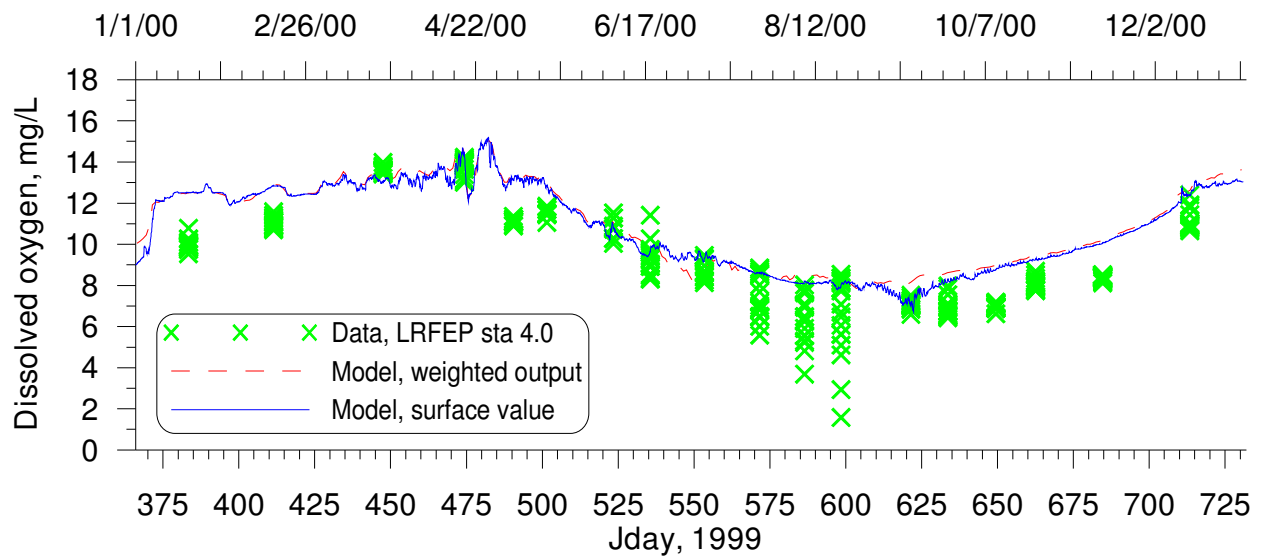


Figure 145. Model-data comparison of dissolved oxygen at LRFEP station 4.0.

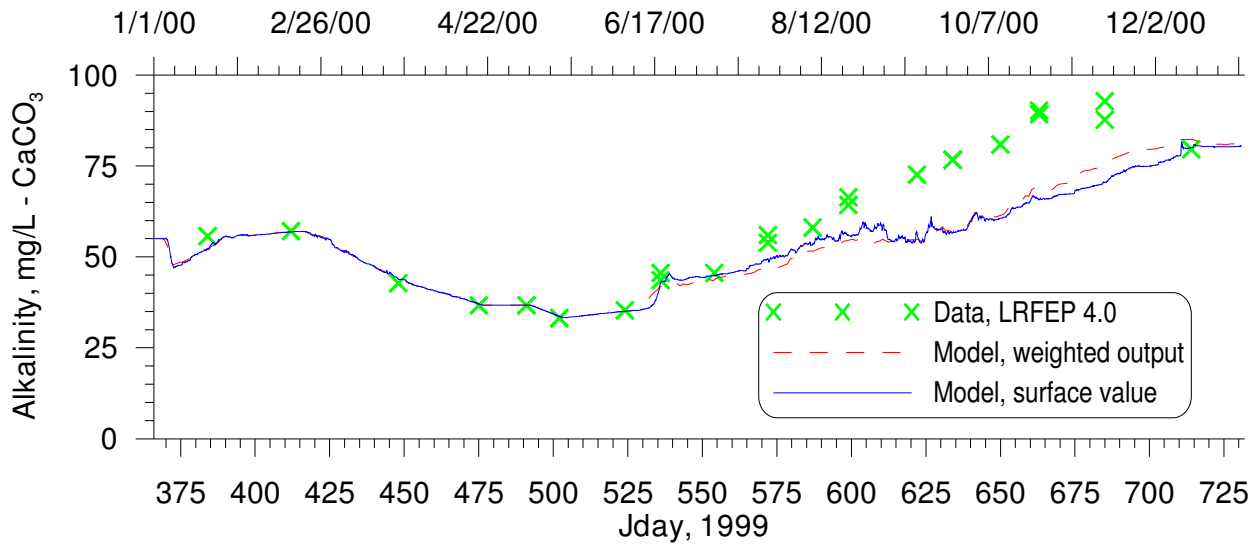


Figure 146. Model-data comparison of alkalinity at LRFEP station 4.0.

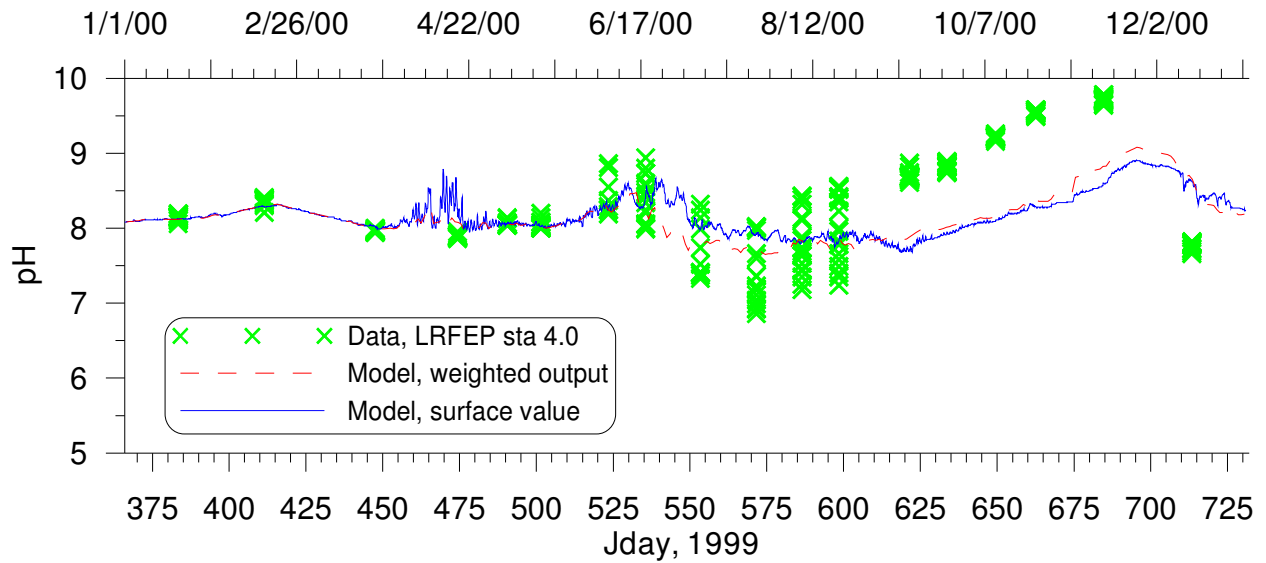


Figure 147. Model-data comparison of pH at LRFEP station 4.0.

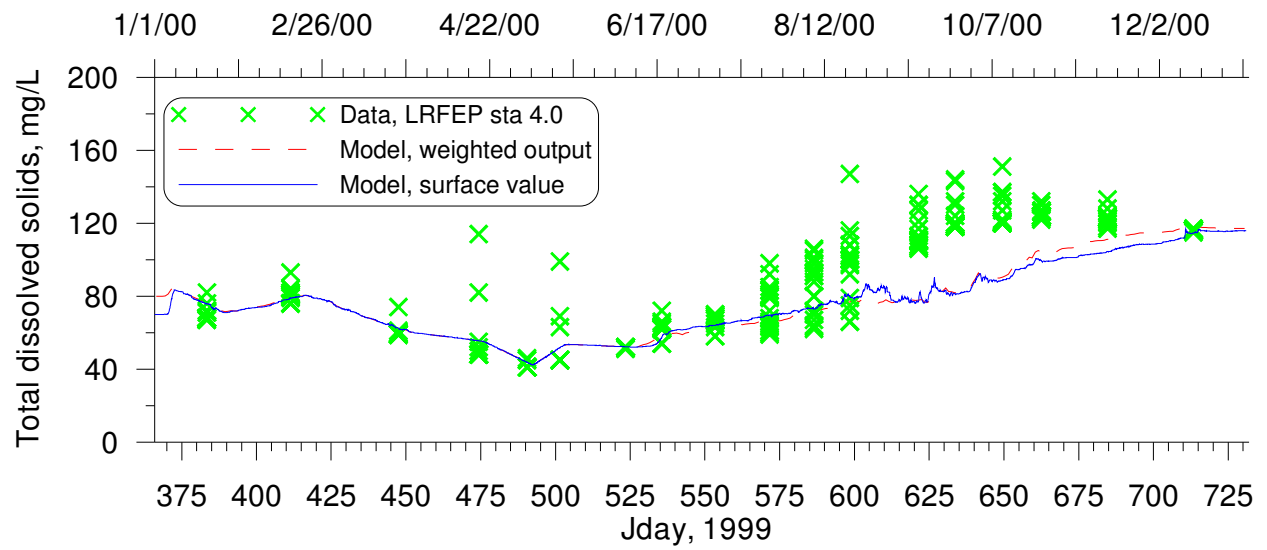


Figure 148 Model-data comparison of total dissolved solids at LRFEP station 4.0.

Station 5.5

The travel time to station 5.5 ranges roughly from 9 to 56 days from the U.S.-Canadian border. The model exhibits some vertical gradients at times for most of the constituents

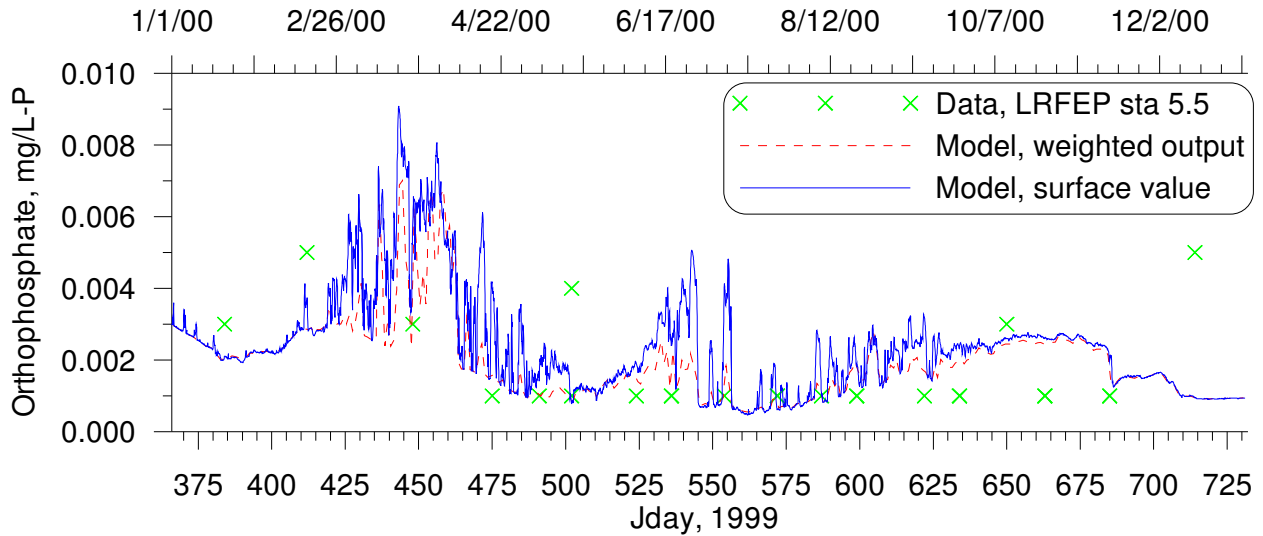


Figure 149. Model-data comparison of orthophosphate at LRFEP station 5.5.

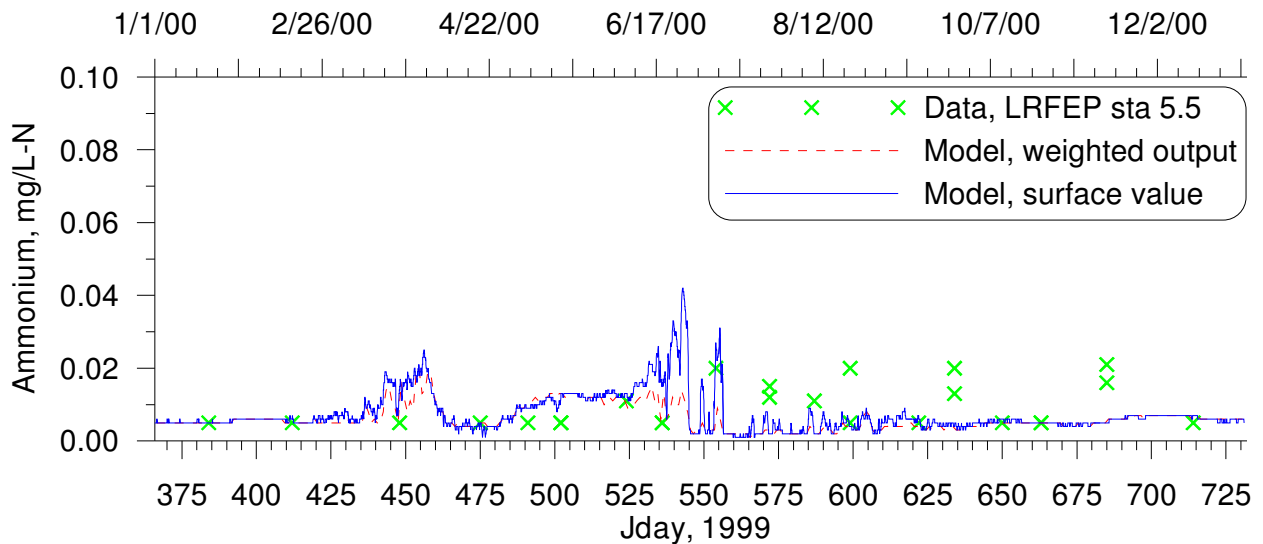


Figure 150. Model-data comparison of ammonium at LRFEP station 5.5.

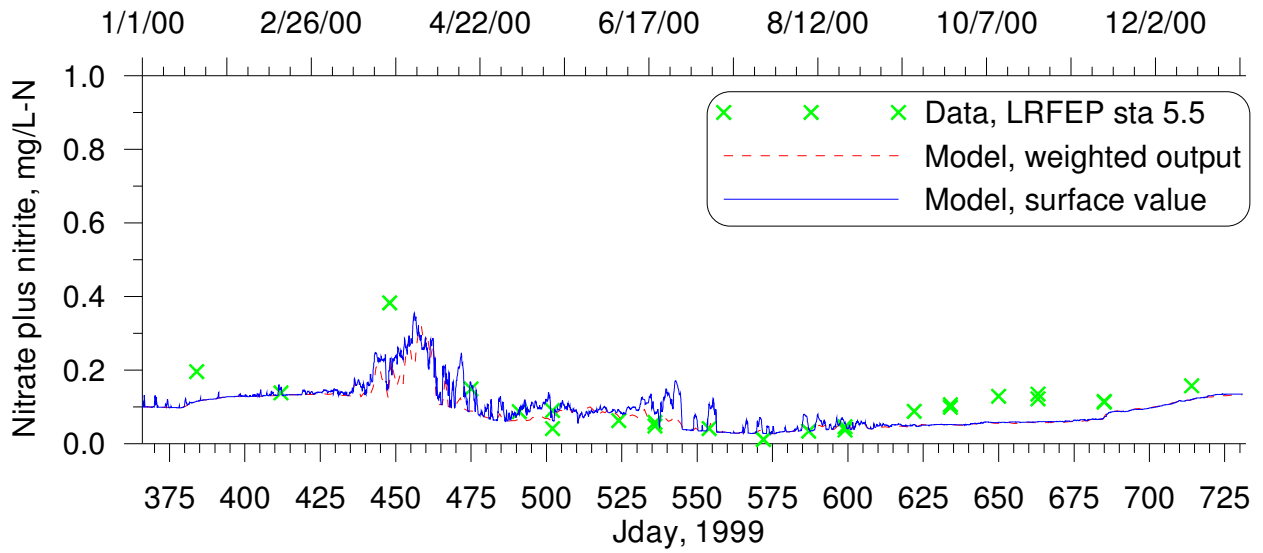


Figure 151. Model-data comparison of nitrate plus nitrite at LRFEP station 5.5.

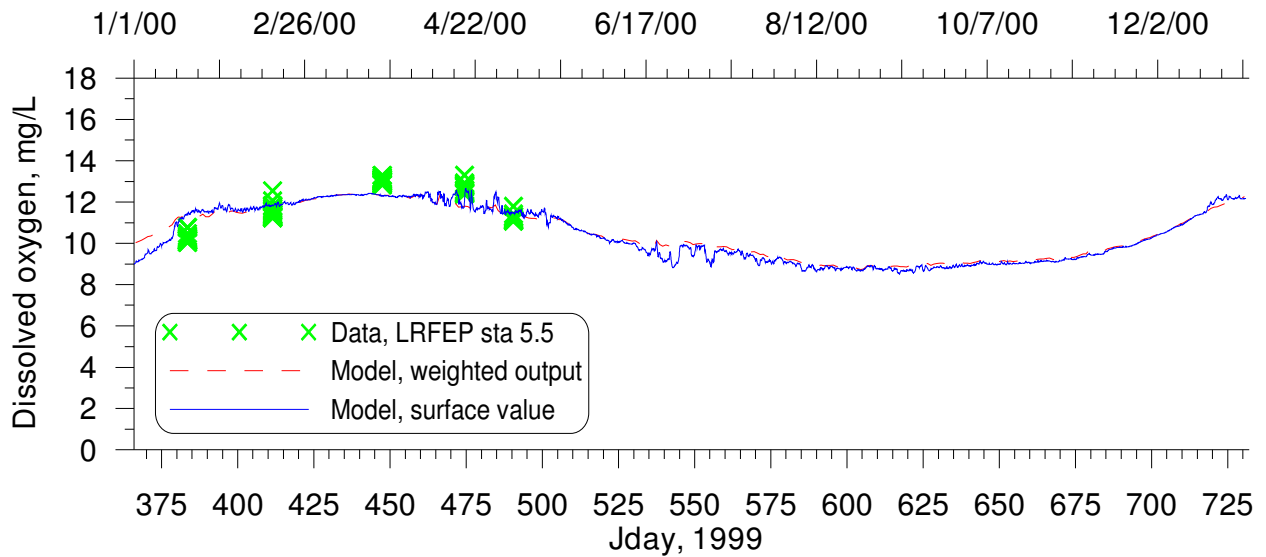


Figure 152. Model-data comparison of dissolved oxygen at LRFEP station 5.5.

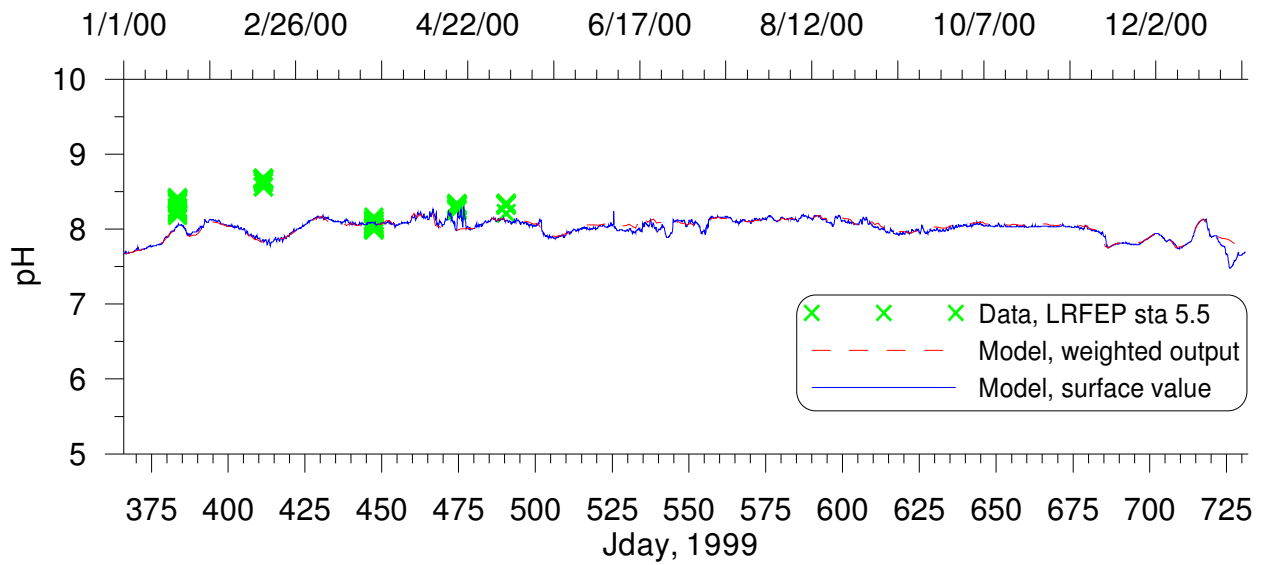


Figure 153. Model-data comparison of pH at LRFEP station 5.5.

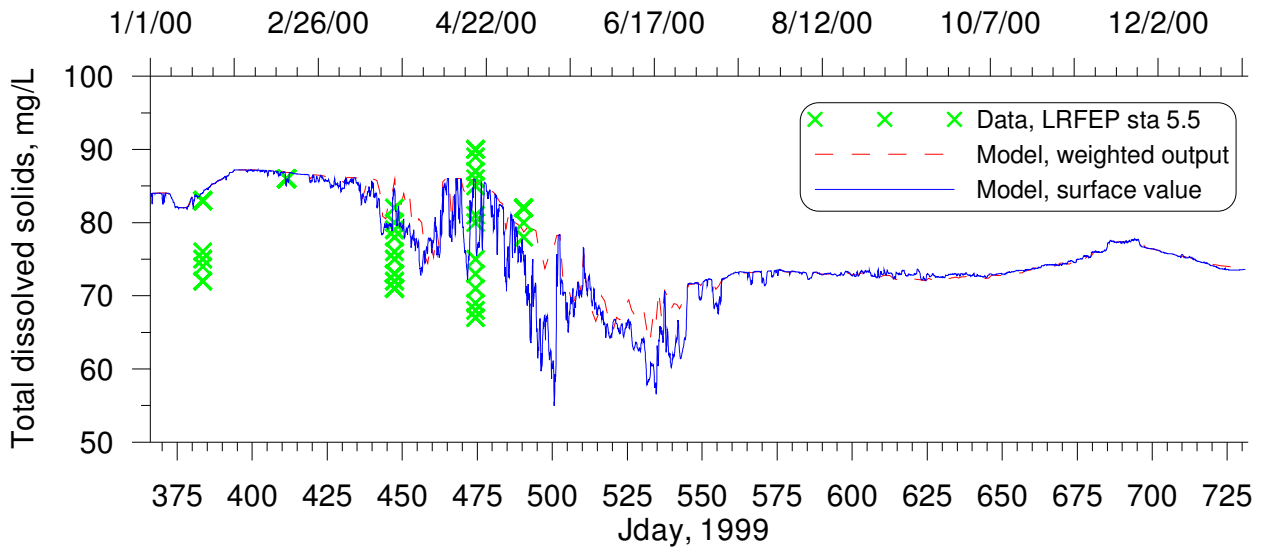


Figure 154. Model-data comparison of total dissolved solids at LRFEP station 5.5.

Station 6.0

Station 6.0 is located downstream of the confluence of the Spokane and Columbia Rivers. The noise seen in the orthophosphate (Figure 155), ammonium (Figure 156), and nitrate plus nitrite (Figure 157) is a result of algal activity.

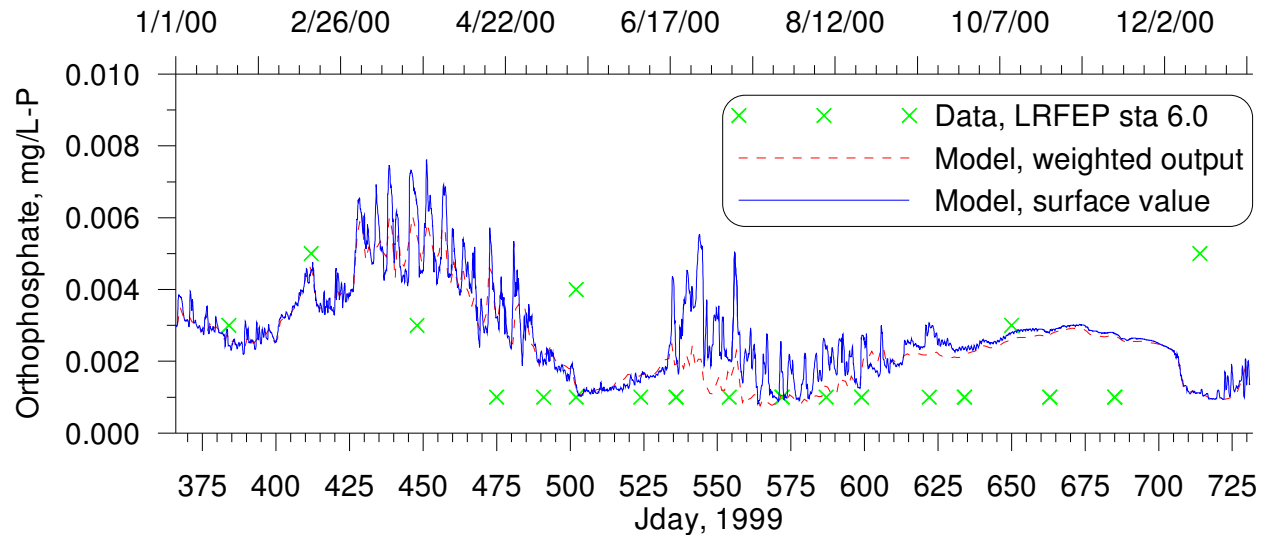


Figure 155. Model-data comparison of orthophosphate at LRFEP station 6.0.

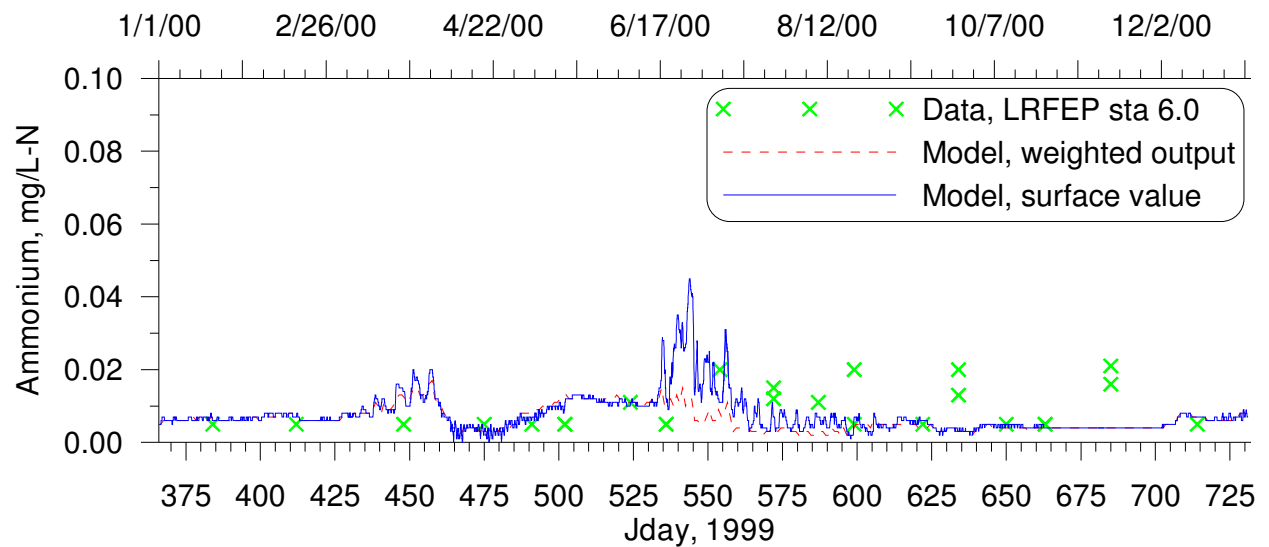


Figure 156. Model-data comparison of ammonium at LRFEP station 6.0.

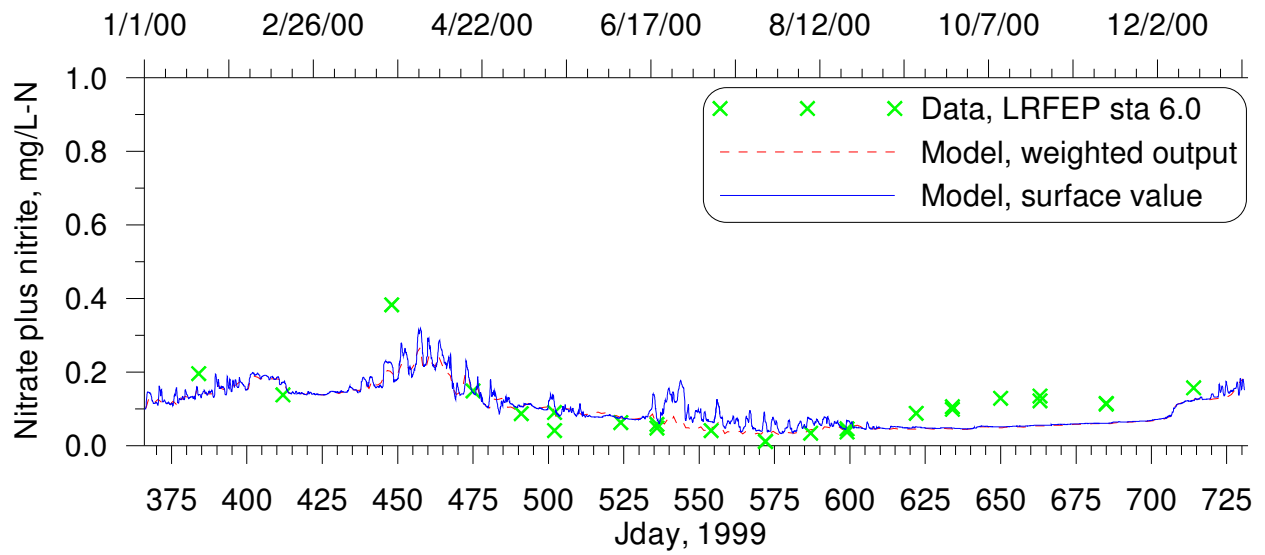


Figure 157. Model-data comparison of nitrate plus nitrite at LRFEP station 6.0.

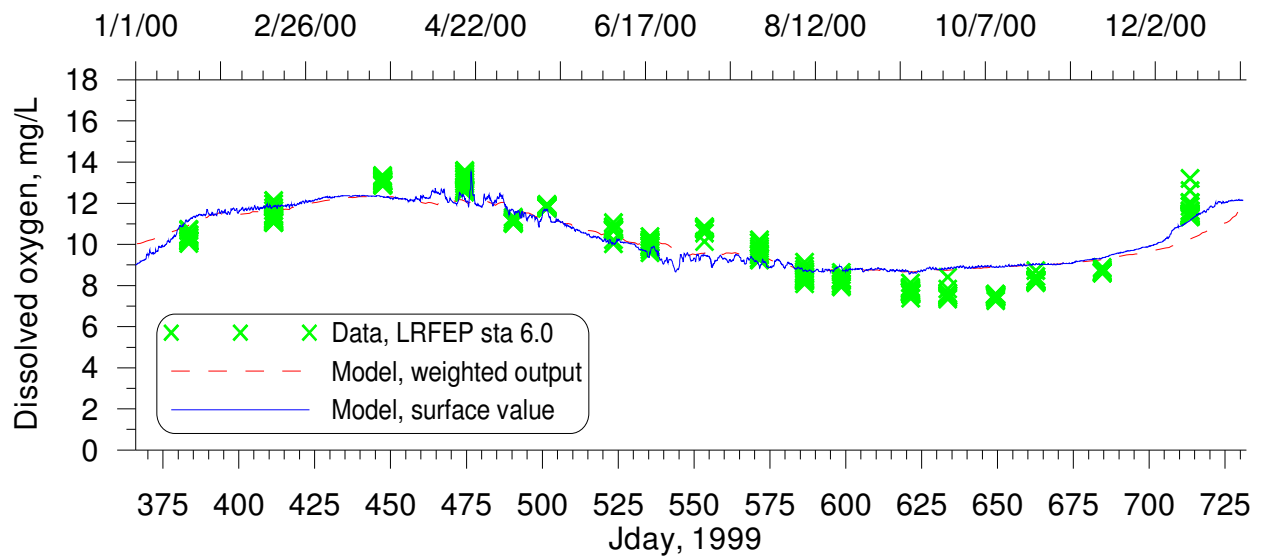


Figure 158. Model-data comparison of dissolved oxygen at LRFEP station 6.0.

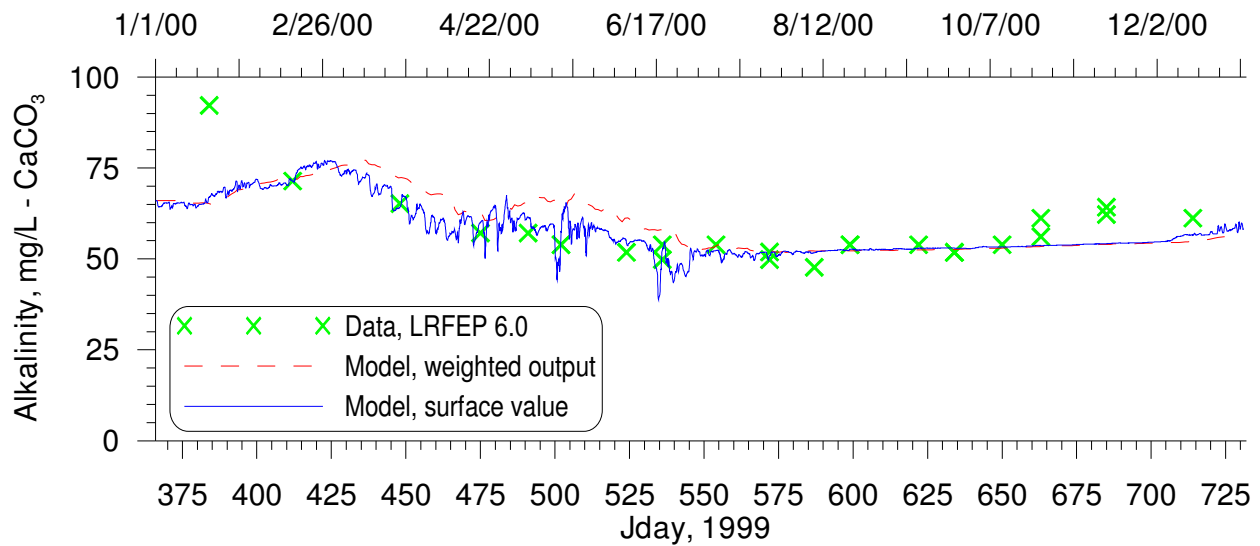


Figure 159. Model-data comparison of alkalinity at LRFEP station 6.0.

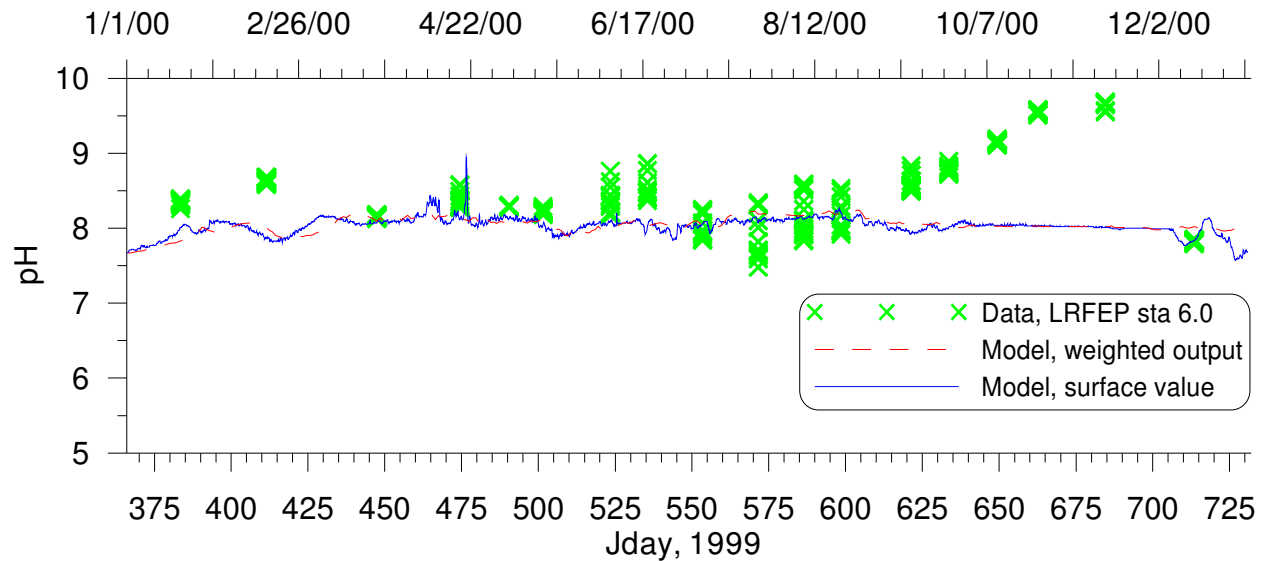


Figure 160. Model-data comparison of pH at LRFEP station 6.0.

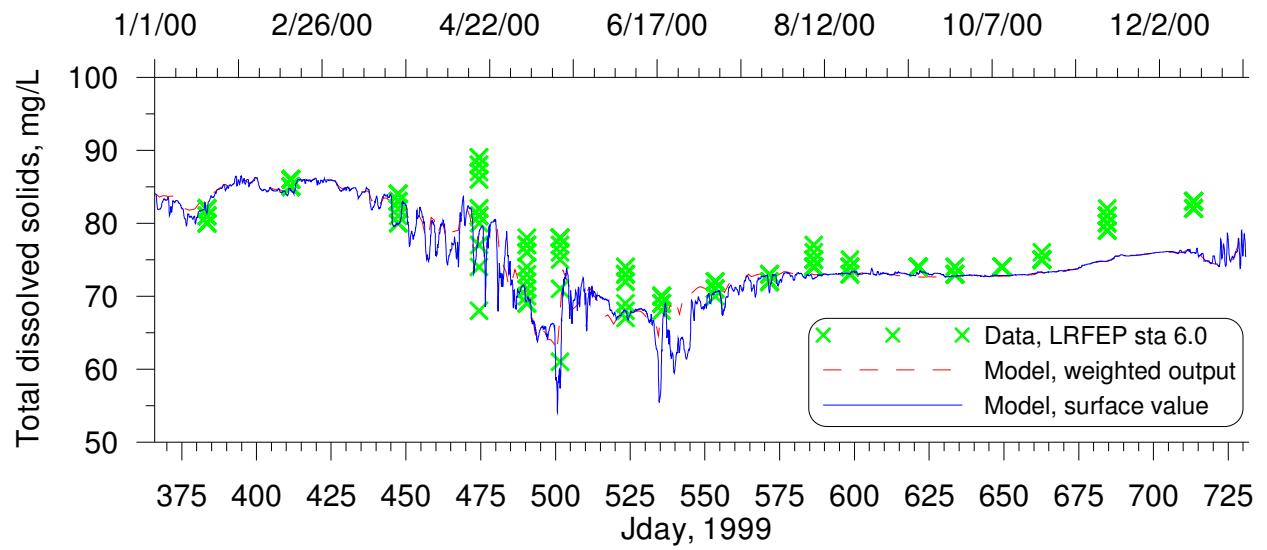


Figure 161. Model-data comparison of total dissolved solids at LRFEP station 6.0.

Station 6.5

Only dissolved oxygen (Figure 162), pH (Figure 163), and total dissolved solids (Figure 164) data are available at station 6.5. The model exhibits less vertical gradient than the data. The model also predicts that the constituent concentrations at station 6.5 are more influenced by the Columbia River mainstem than the data suggest. There are two likely explanations. One, the sampling location is more sheltered than the model's laterally averaged results would suggest. Two, the model does not include any inflow from Hawk Creek since no data were available, but the inflow of nutrients may be biologically significant. Refer also to the algae calibration section.

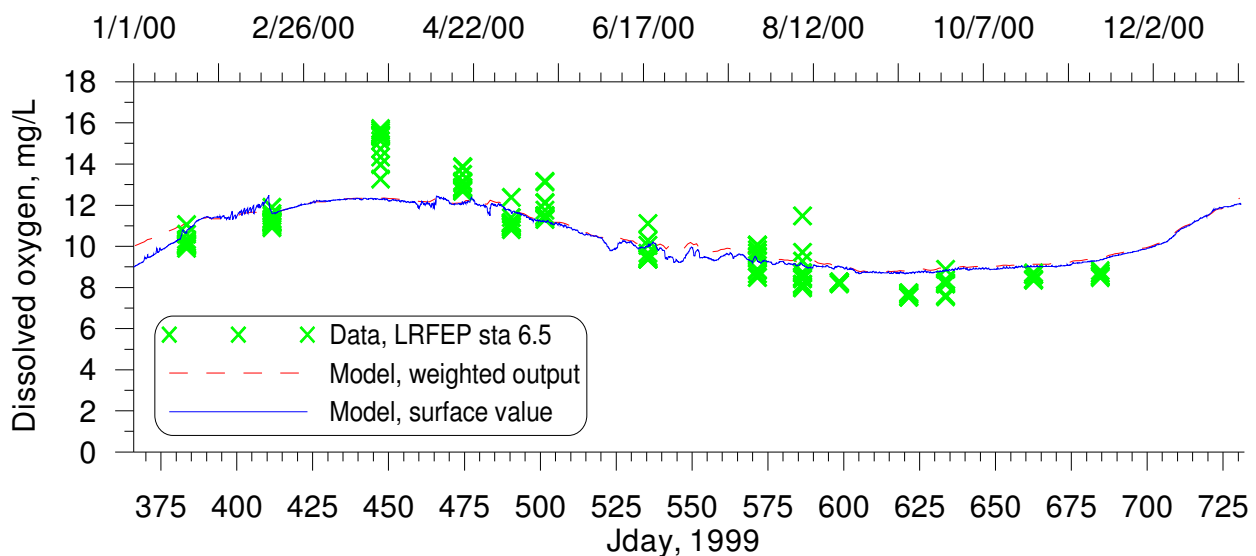


Figure 162. Model-data comparison of dissolved oxygen at LRFEP station 6.5.

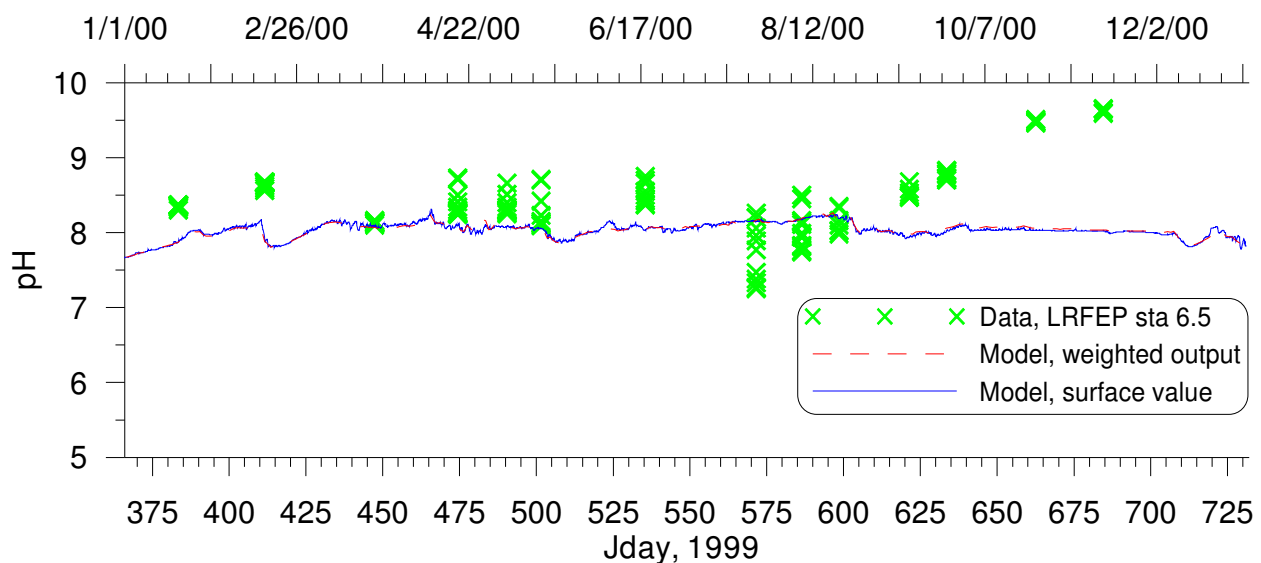


Figure 163. Model-data comparison of pH at LRFEP station 6.5.

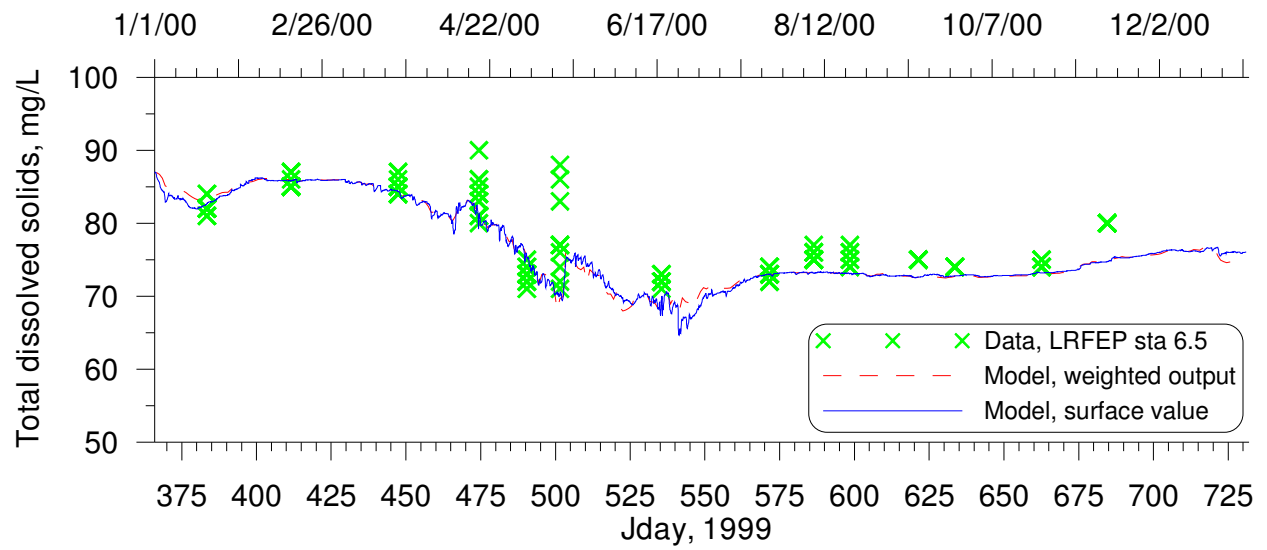


Figure 164. Model-data comparison of total dissolved solids at LRFEP station 6.5.

Station 7.0

The model exhibits some weak vertical gradients at station 7.0 during and after thermal stratification.

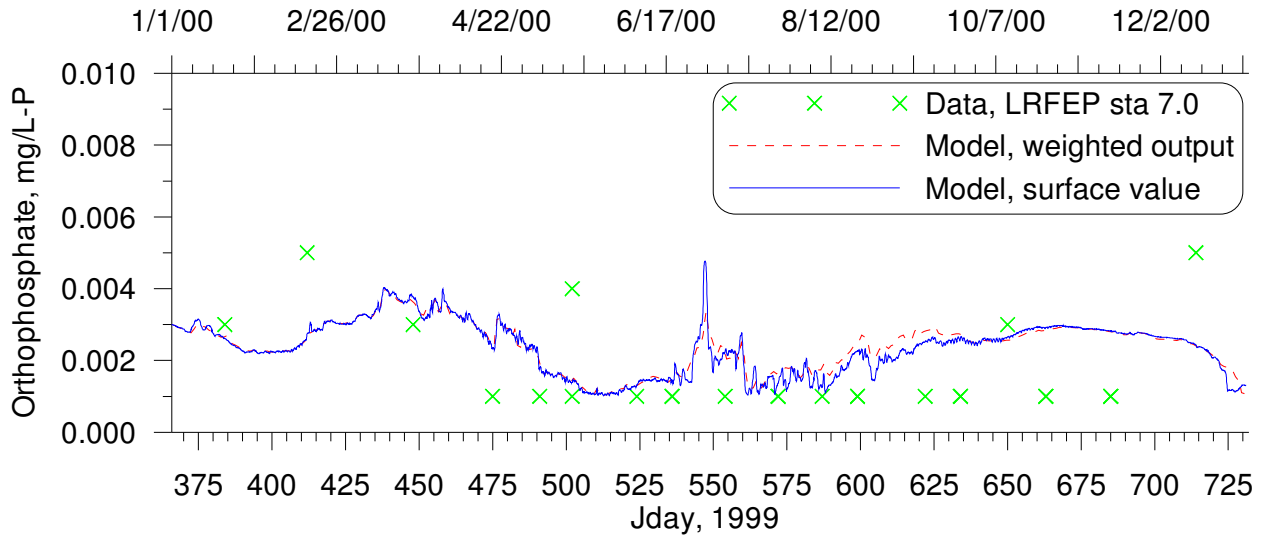


Figure 165. Model-data comparison of orthophosphate at LRFEP station 7.0.

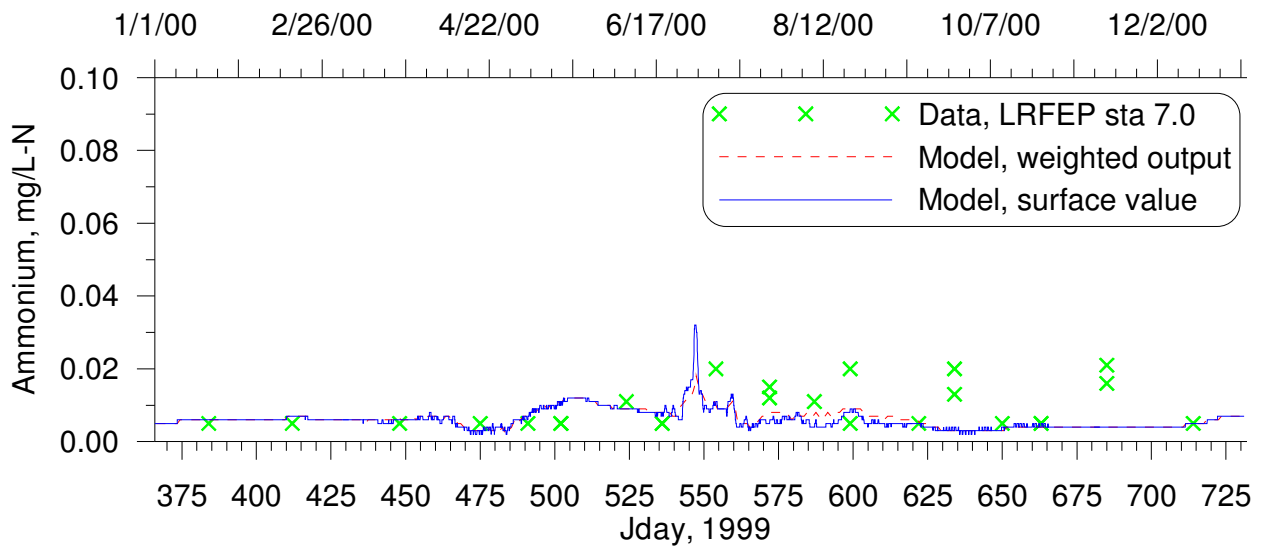


Figure 166. Model-data comparison of ammonium at LRFEP station 7.0.

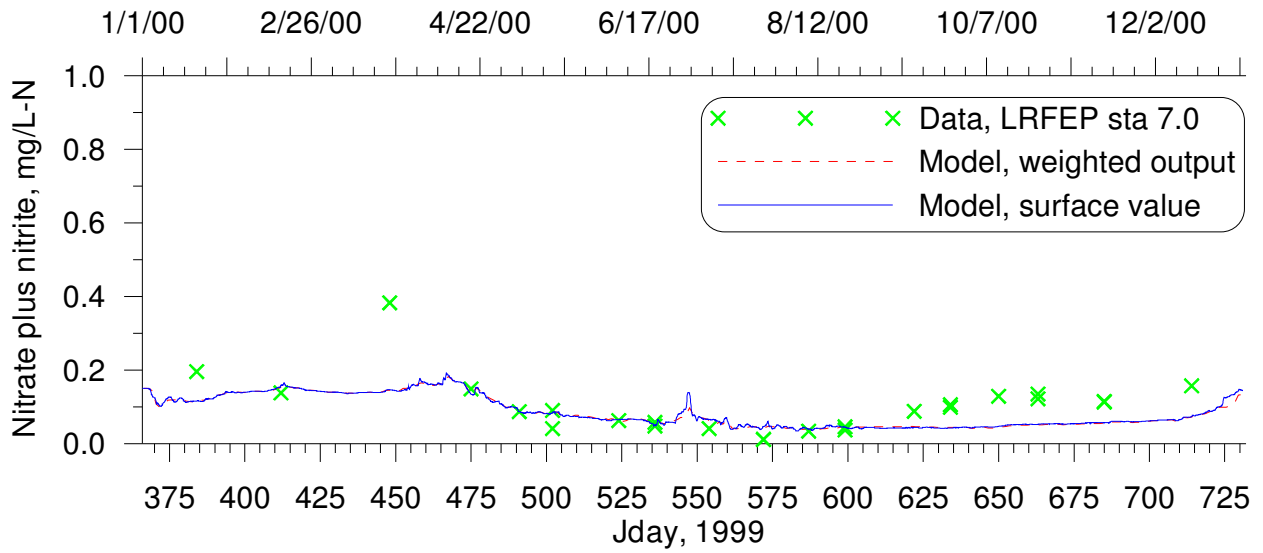


Figure 167. Model-data comparison of nitrate plus nitrite at LRFEP station 7.0.

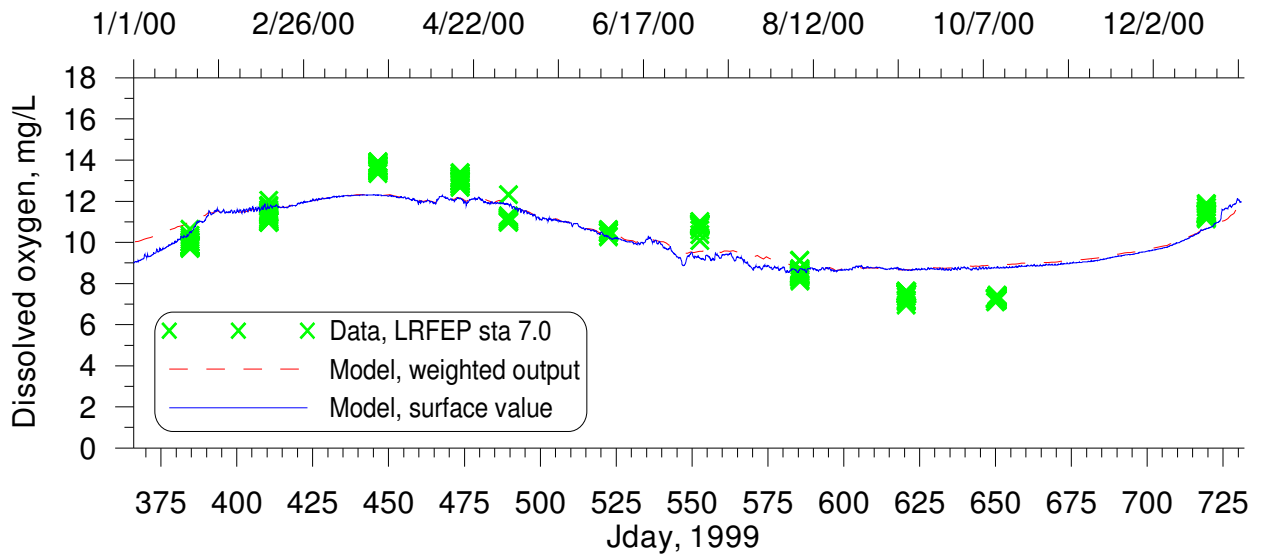


Figure 168. Model-data comparison of dissolved oxygen at LRFEP station 7.0.

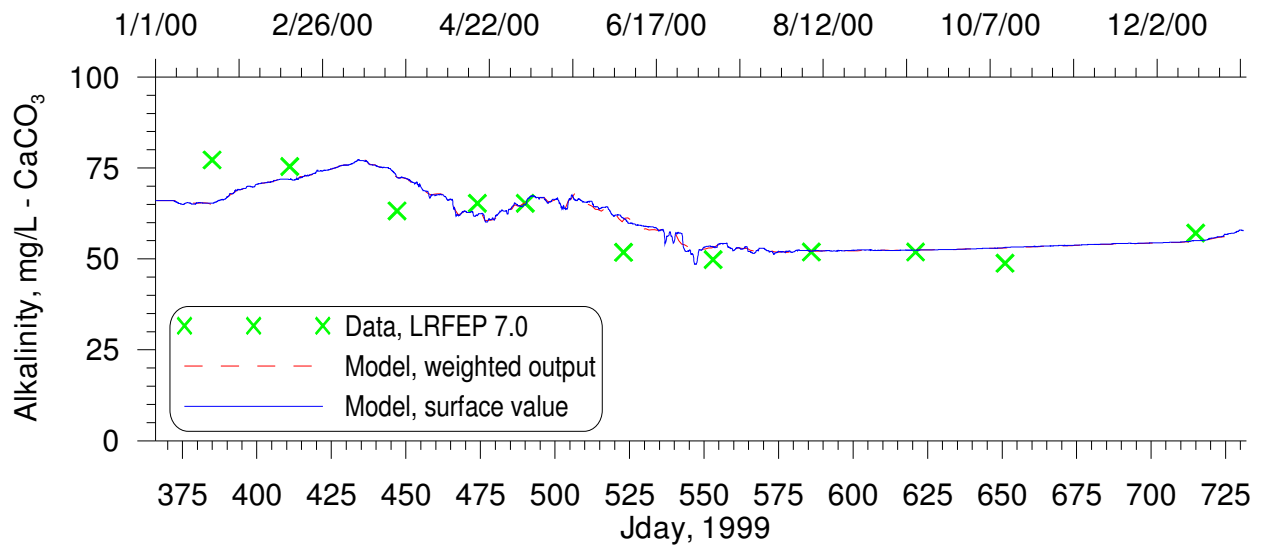


Figure 169. Model-data comparison of alkalinity at LRFEP station 7.0.

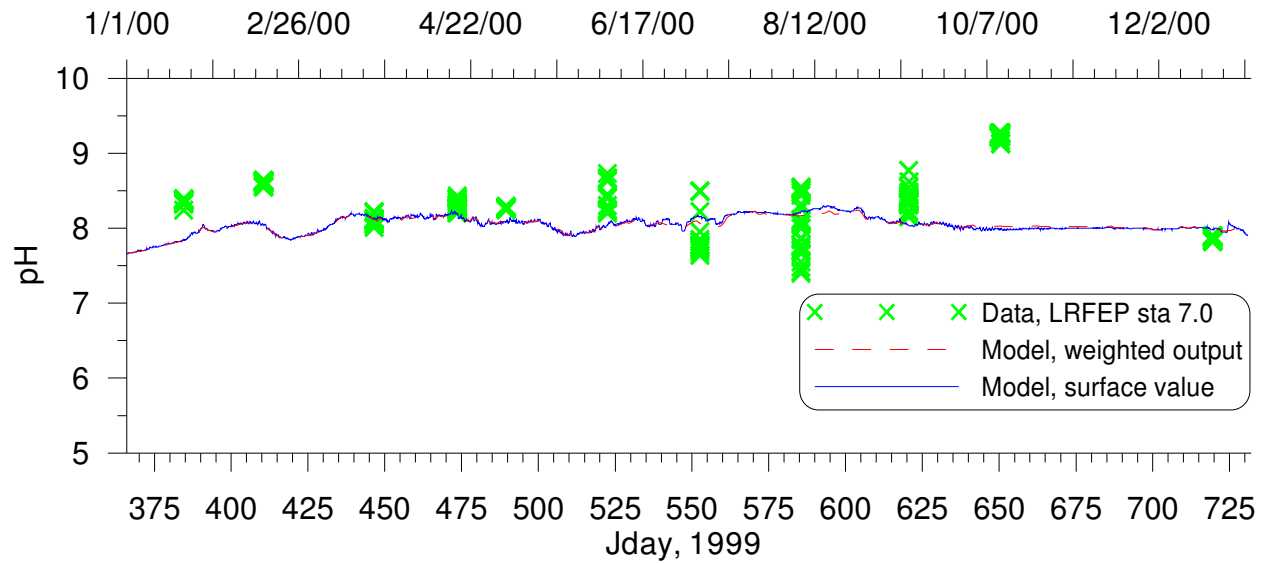


Figure 170. Model-data comparison of pH at LRFEP station 7.0.

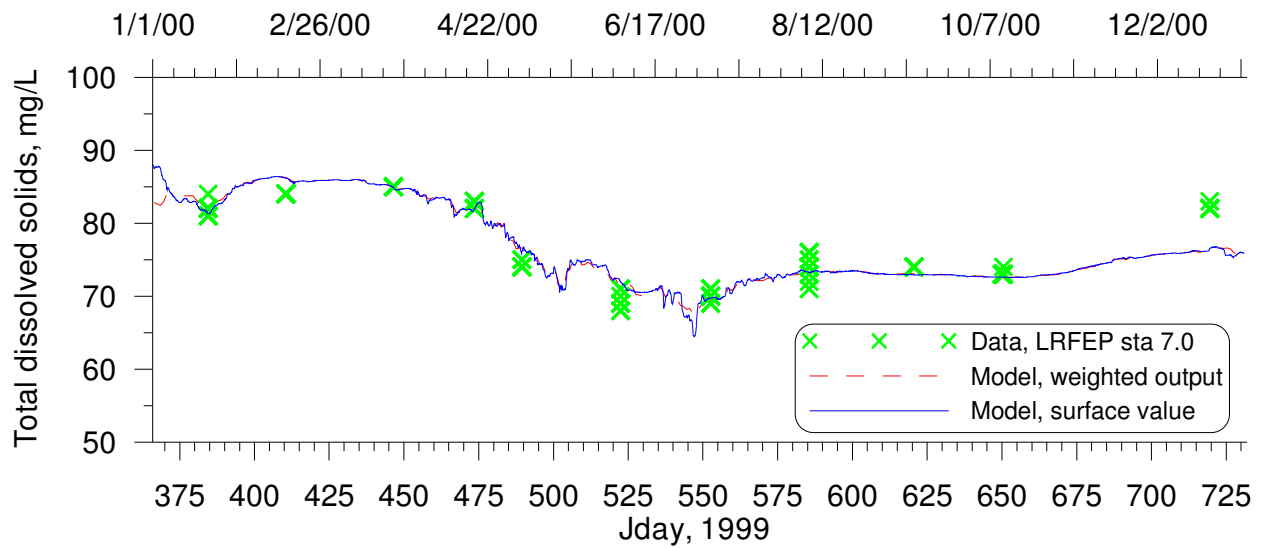


Figure 171. Model-data comparison of total dissolved solids at LRFEP station 7.0.

Station 8.0

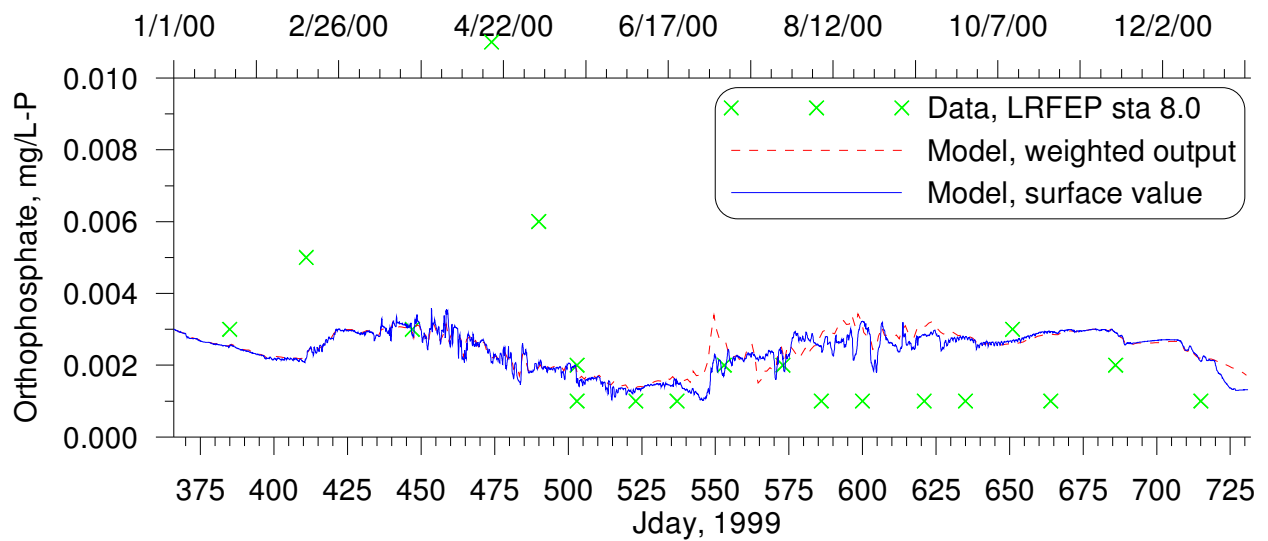


Figure 172. Model-data comparison of orthophosphate at LRFEP station 8.0.

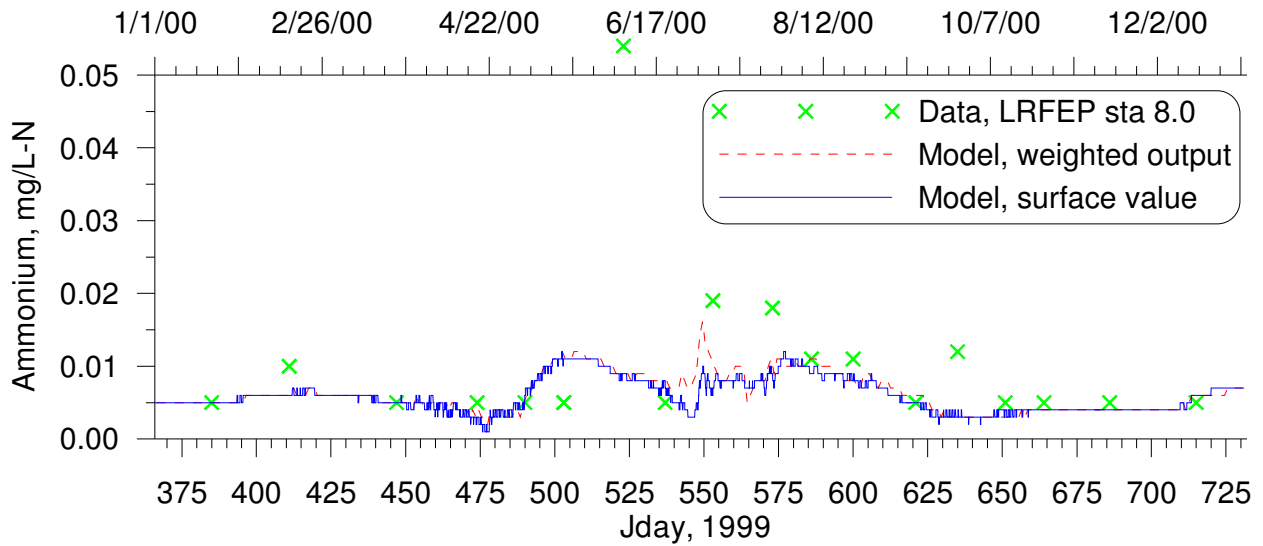


Figure 173. Model-data comparison of ammonium at LRFEP station 8.0.

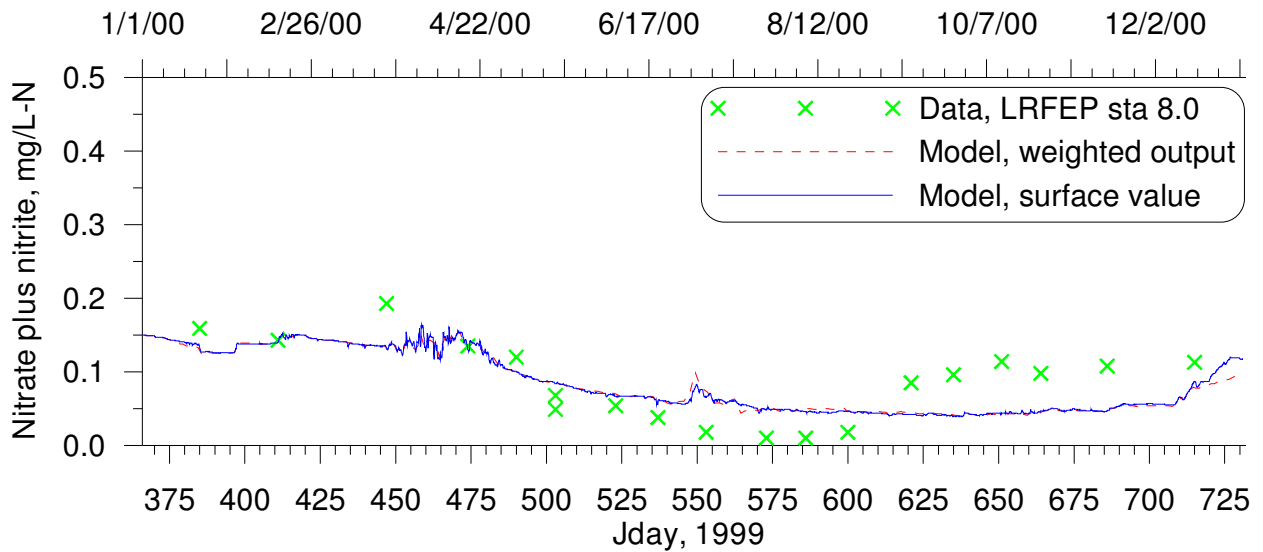


Figure 174. Model-data comparison of nitrate plus nitrite at LRFEP station 8.0.

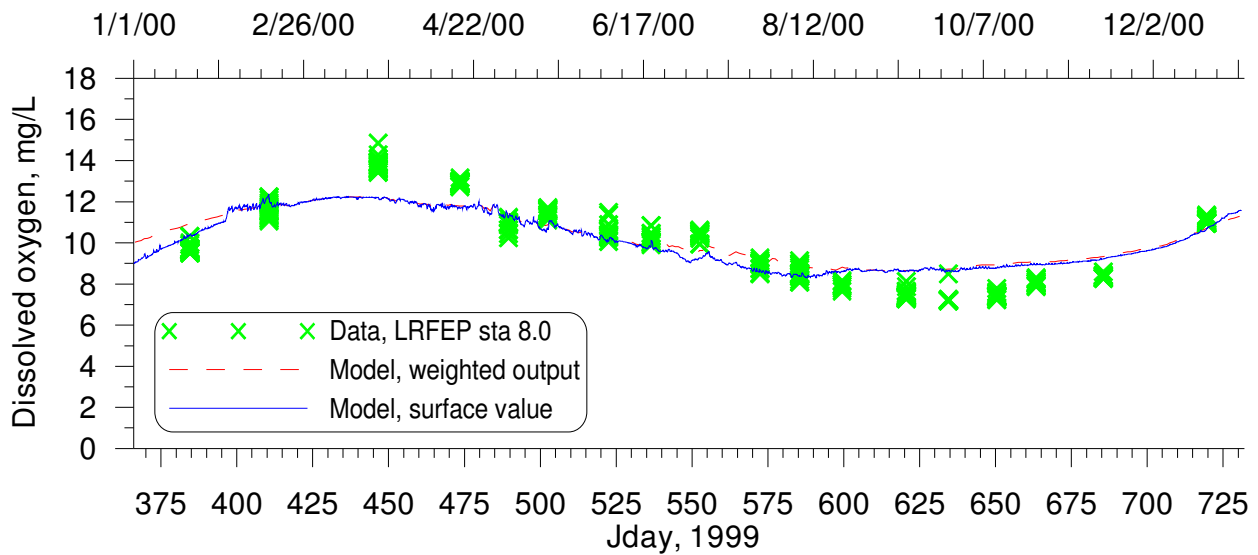


Figure 175. Model-data comparison of dissolved oxygen at LRFEP station 8.0.

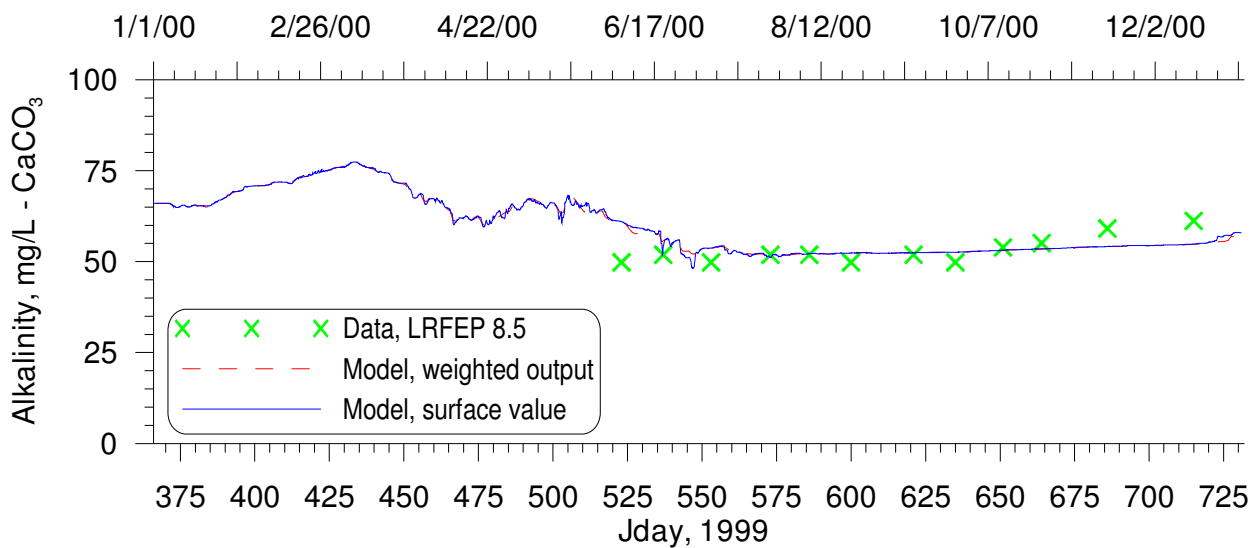


Figure 176. Model-data comparison of alkalinity at LRFEP station 8.0.

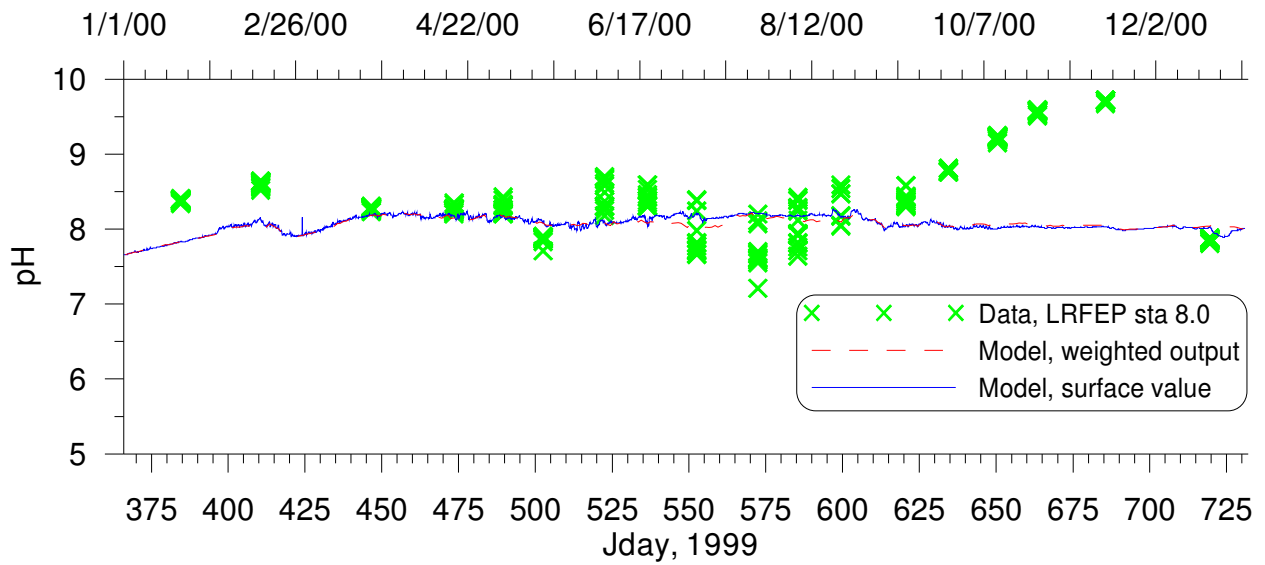


Figure 177. Model-data comparison of pH at LRFEP station 8.0.

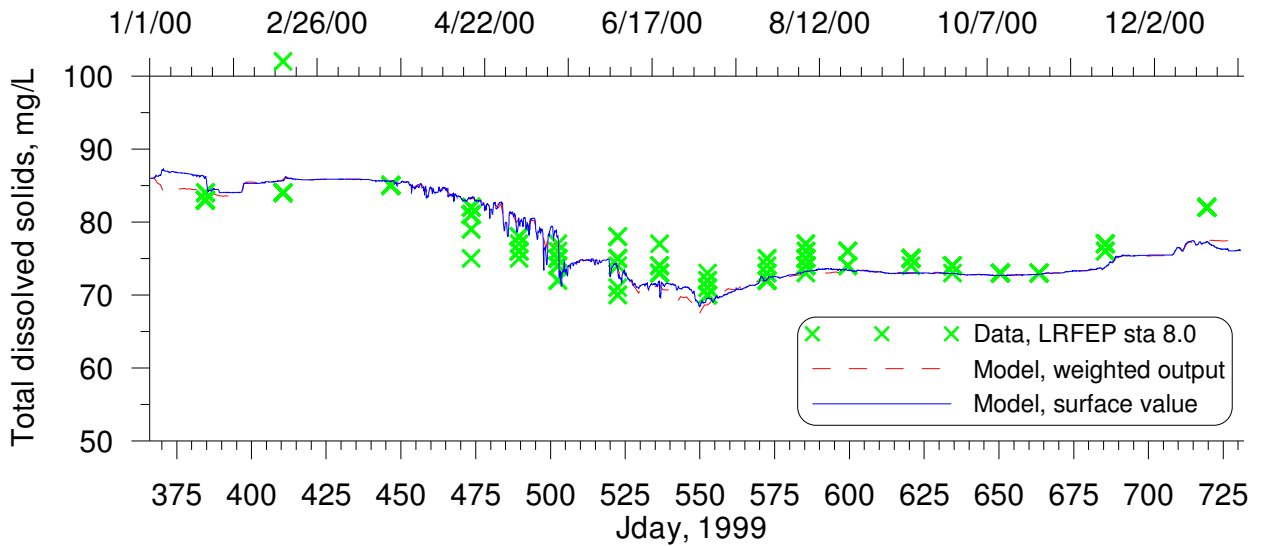


Figure 178. Model-data comparison of total dissolved solids at LRFEP station 8.0.

Station 8.5

The model exhibits some weak vertical gradients at station 8.5 during and after thermal stratification. The model does not exhibit the strong total dissolved solids gradient seen in the data (Figure 185).

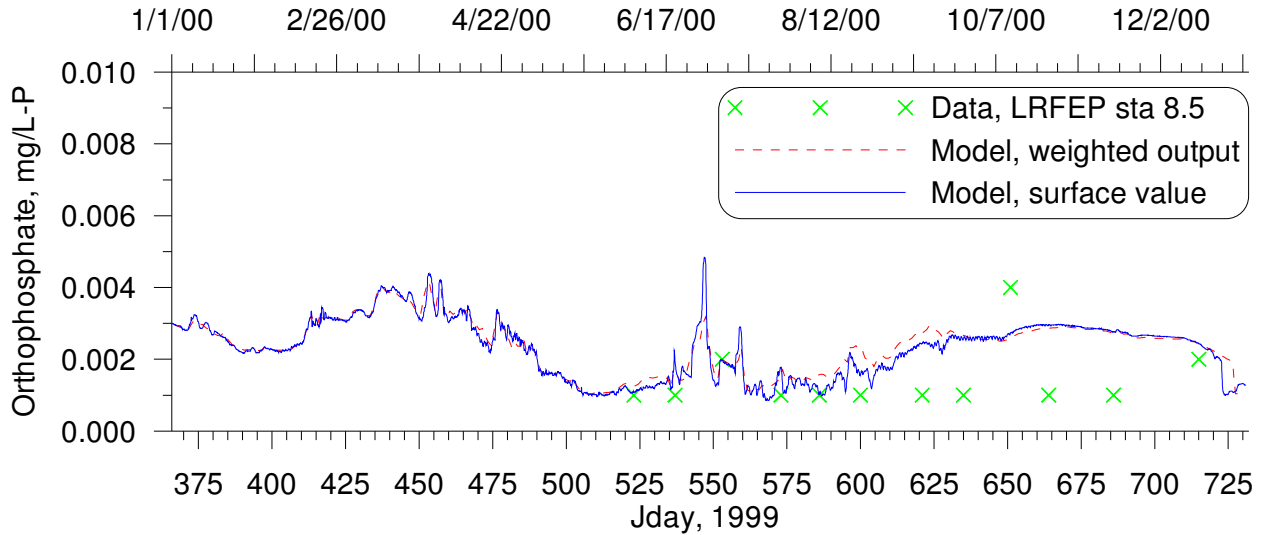


Figure 179. Model-data comparison of orthophosphate at LRFEP station 8.5.

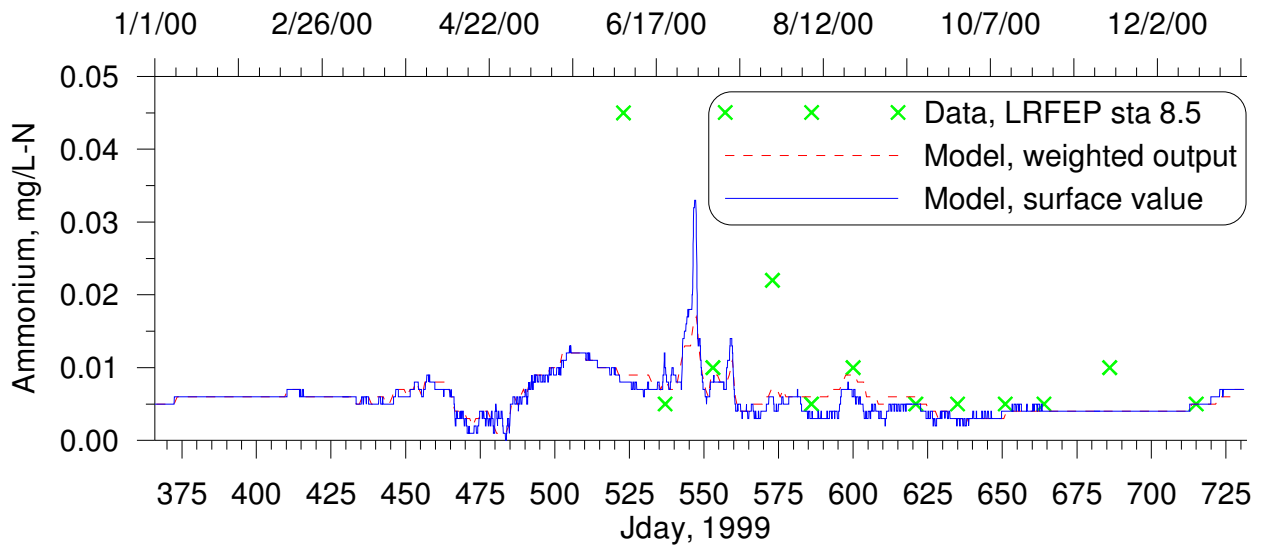


Figure 180. Model-data comparison of ammonium at LRFEP station 8.5.

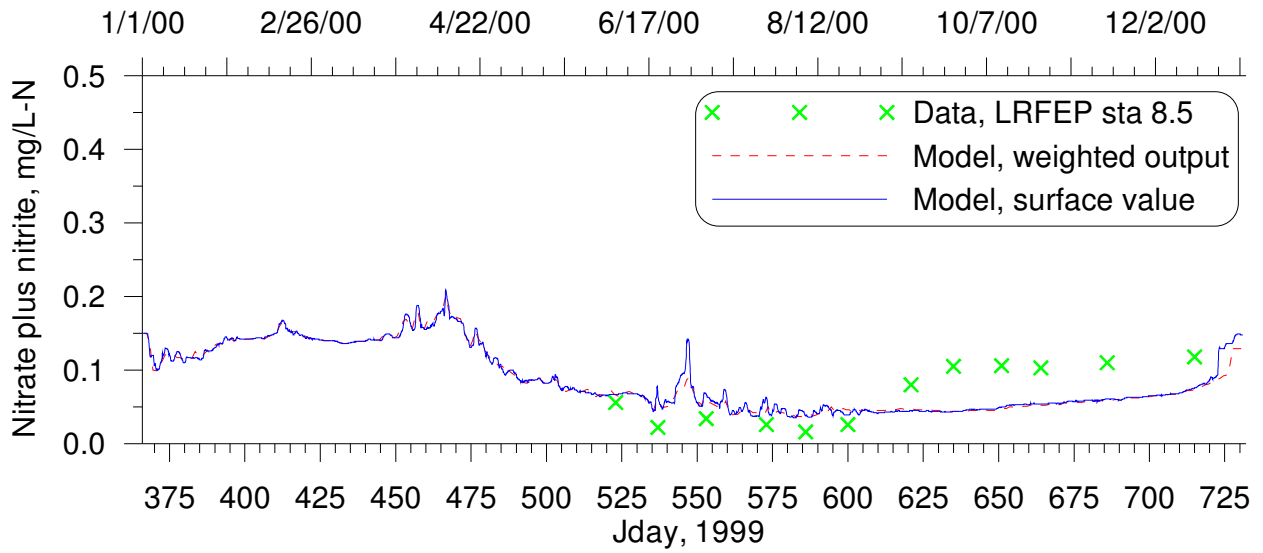


Figure 181. Model-data comparison of nitrate plus nitrite at LRFEP station 8.5.

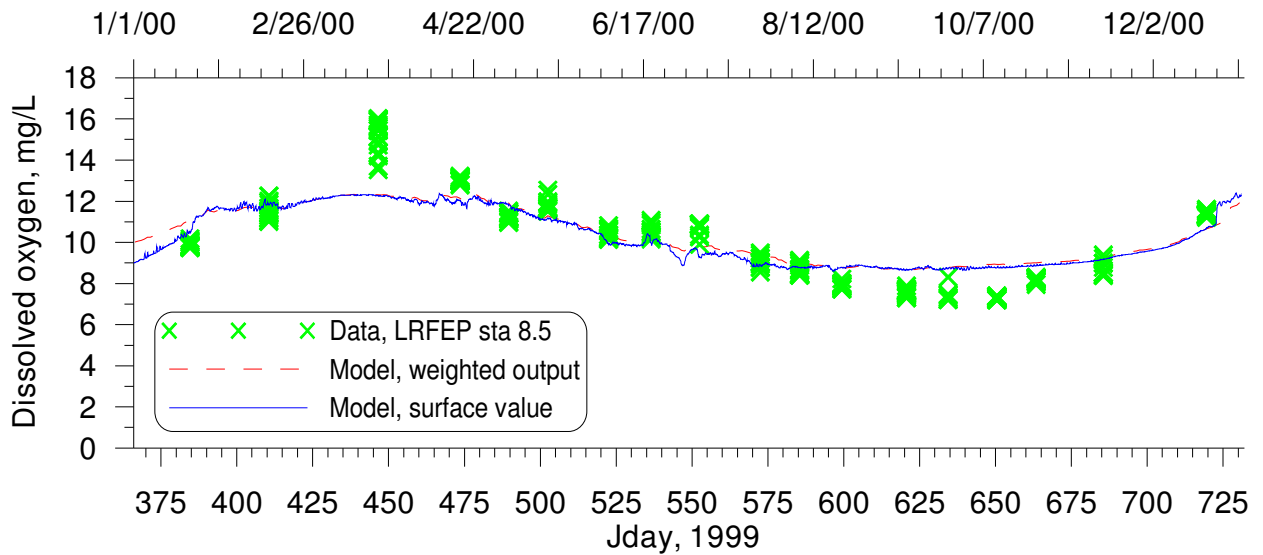


Figure 182. Model-data comparison of dissolved oxygen at LRFEP station 8.5.

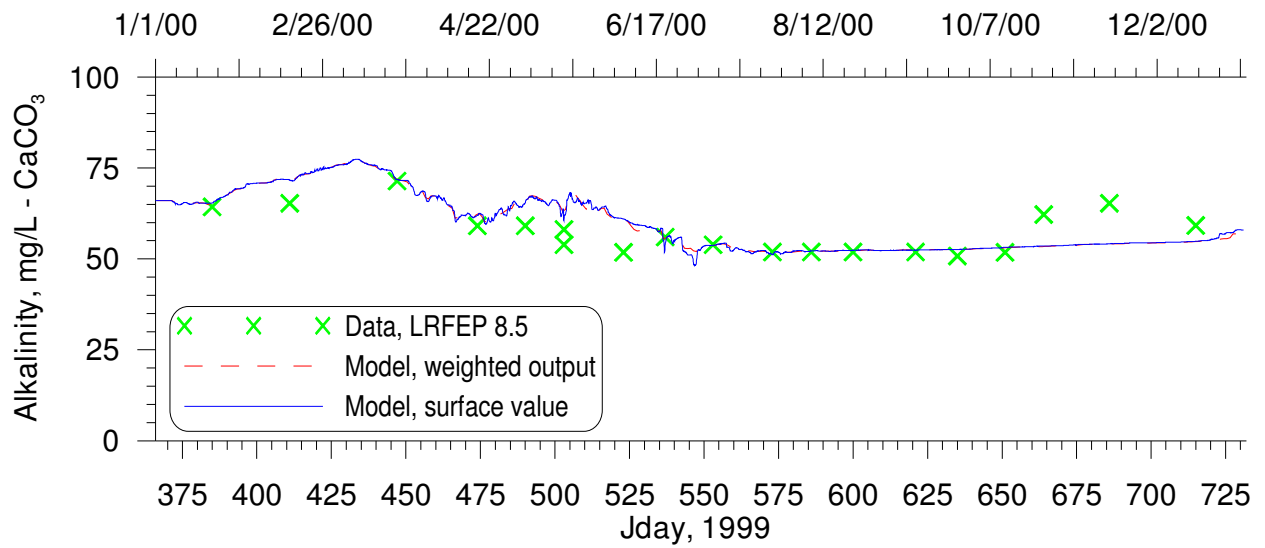


Figure 183. Model-data comparison of alkalinity at LRFEP station 8.5.

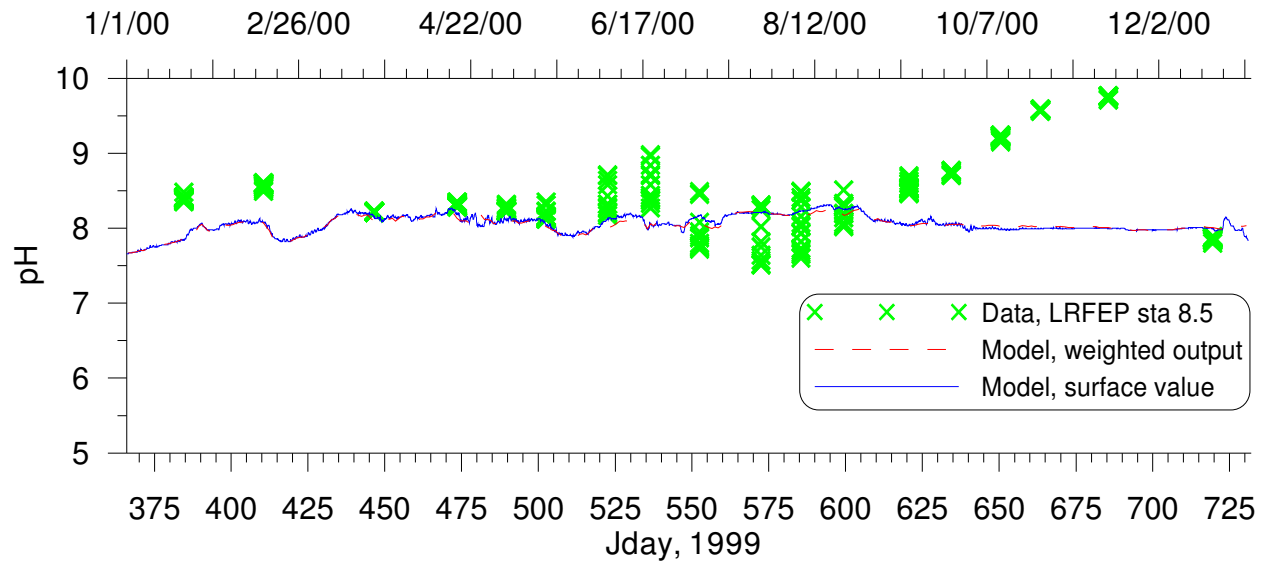


Figure 184. Model-data comparison of pH at LRFEP station 8.5.

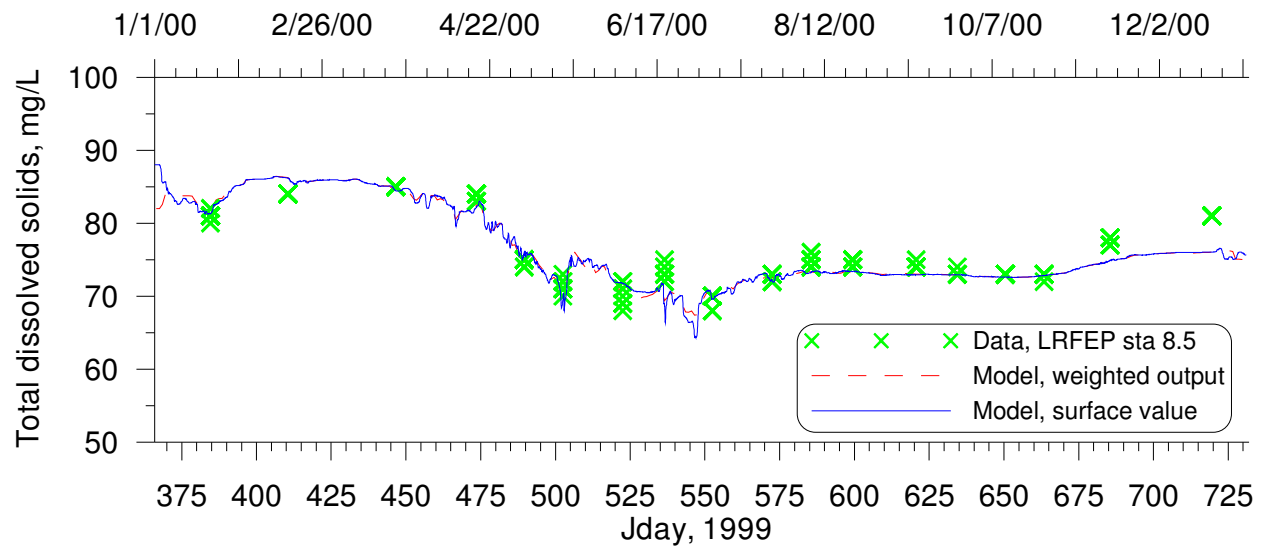


Figure 185. Model-data comparison of total dissolved solids at LRFEP station 8.5.

Station 9.0

The model exhibits a weak vertical gradient during and after thermal stratification. The dissolved oxygen (Figure 189) and total dissolved solids (Figure 192) data exhibit a stronger gradient.

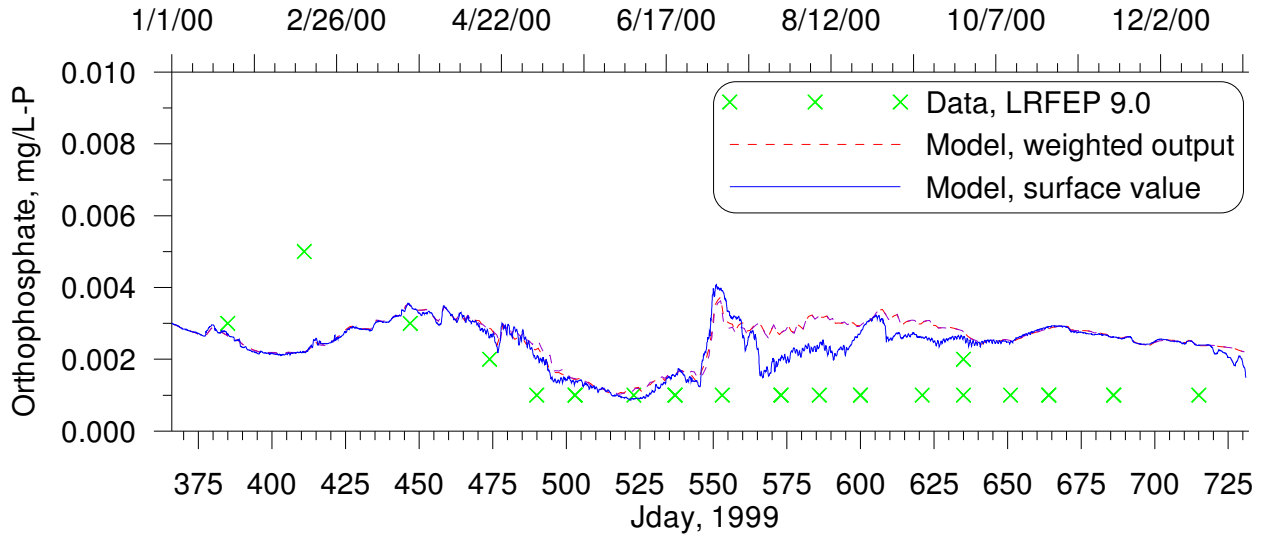


Figure 186. Model-data comparison of orthophosphate at LRFEP station 9.0.

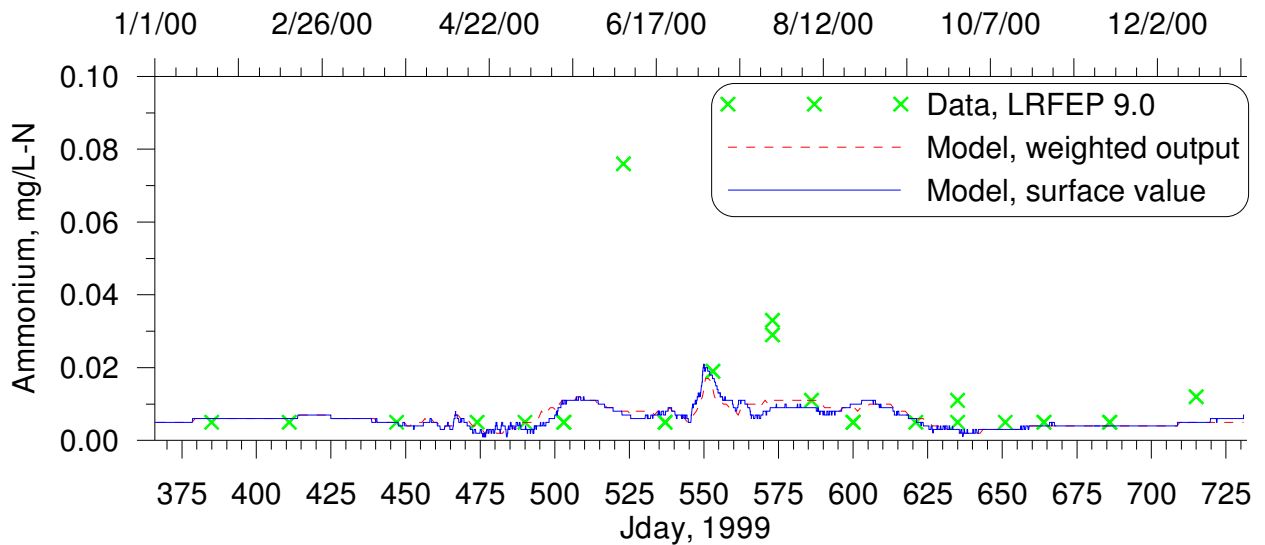


Figure 187. Model-data comparison of ammonium at LRFEP station 9.0.

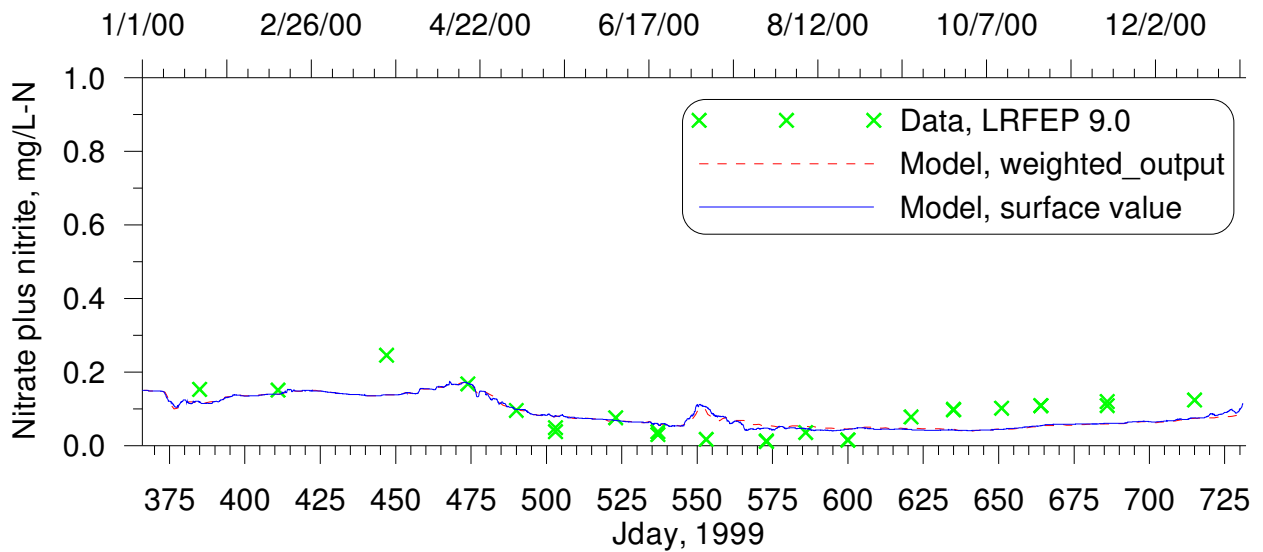


Figure 188. Model-data comparison of nitrate plus nitrite at LRFEP station 9.0.

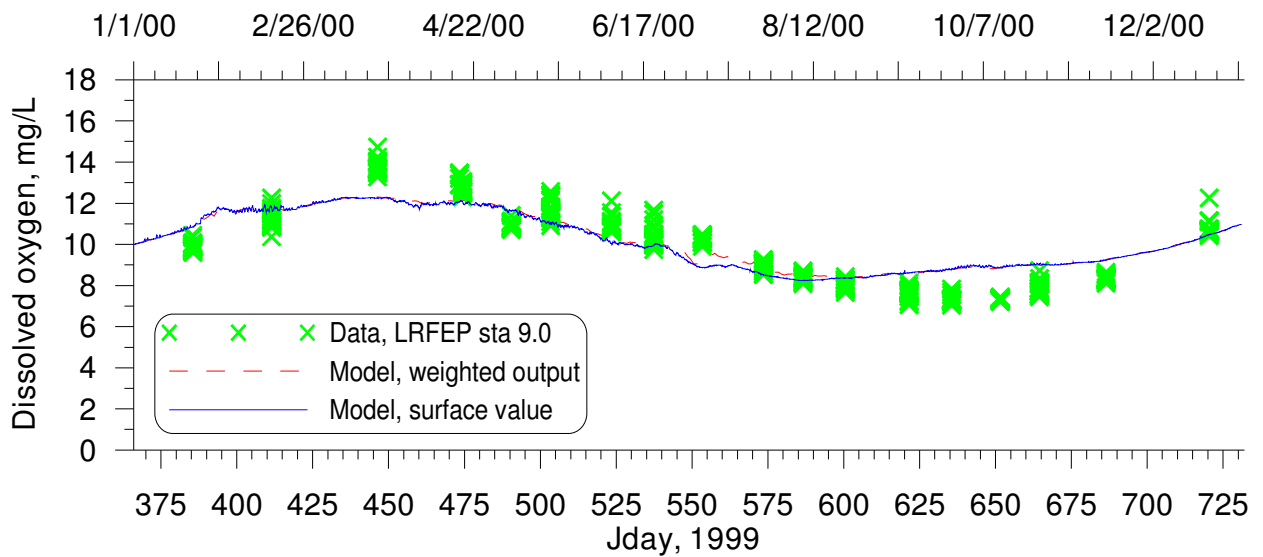


Figure 189. Model-data comparison of dissolved oxygen at LRFEP station 9.0.

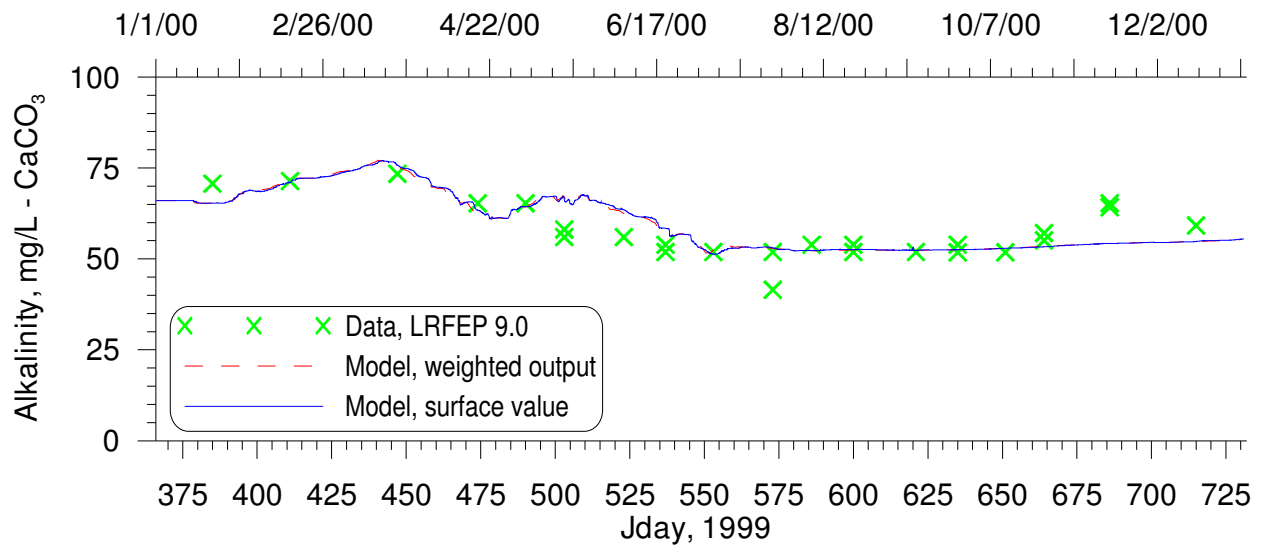


Figure 190. Model-data comparison of alkalinity at LRFEP station 9.0.

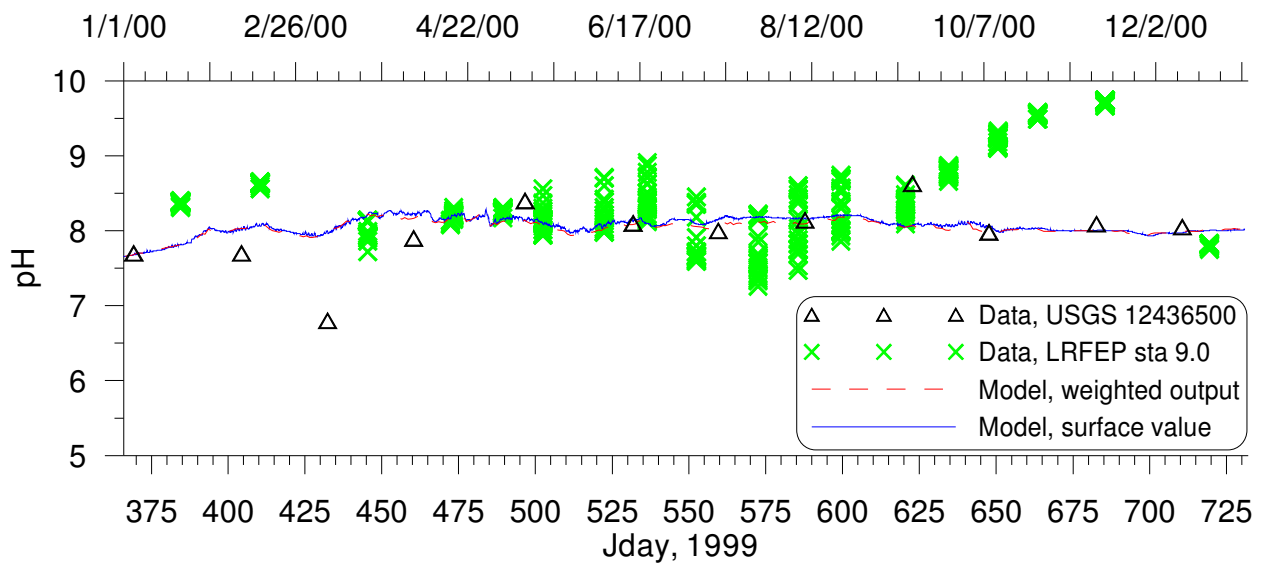


Figure 191. Model-data comparison of pH at LRFEP station 9.0.

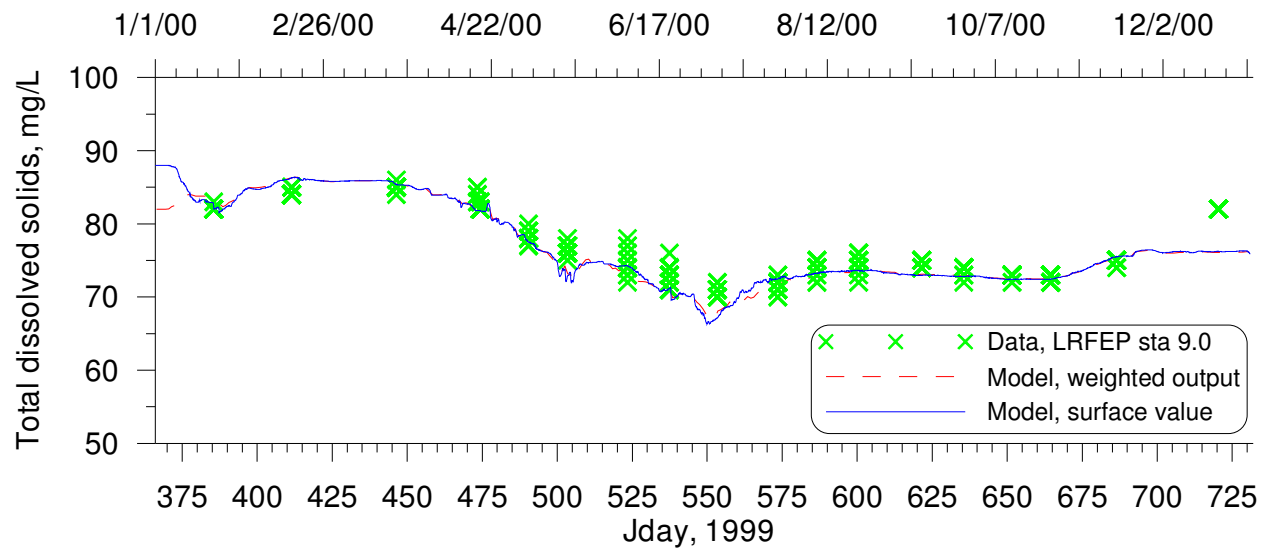


Figure 192. Model-data comparison of total dissolved solids at LRFEP station 9.0.

Appendix F: Vertical profile model-data comparison plots.

Vertical data exist for water temperature, dissolved oxygen, total dissolved solids, and pH. Model-data comparison statistics are reported in Table 31.

Table 31. Vertical profile calibration statistics, 2000.

| Station | Location | #days, (# points) | Statistic | TDS mg/L | DO mg/L | pH - | Temp °C |
|---------------|----------------|----------------------|-----------|-------------|------------|---------|------------|
| 0.0 | Columbia R. | 11 (80) | ME | -0.18 | -0.01 | -0.42 | -0.14 |
| | | | AME | 0.39 | 0.14 | 0.43 | 0.49 |
| | | | RMS | 0.44 | 0.16 | 0.43 | 0.50 |
| 1.0 | Columbia R. | 11 (111) | ME | 0.72 | 0.18 | -0.41 | -0.15 |
| | | | AME | 0.97 | 0.38 | 0.42 | 0.40 |
| | | | RMS | 1.03 | 0.40 | 0.42 | 0.41 |
| 2.0 | Columbia R. | 18 (194) | ME | -0.91 | 0.29 | -0.55 | -0.11 |
| | | | AME | 1.17 | 0.56 | 0.59 | 0.56 |
| | | | RMS | 1.21 | 0.59 | 0.59 | 0.60 |
| 3.0 | Columbia R. | 11 (121) | ME | -0.95 | 0.12 | -0.42 | -0.26 |
| | | | AME | 1.52 | 0.65 | 0.45 | 0.58 |
| | | | RMS | 1.56 | 0.69 | 0.45 | 0.61 |
| 5.5 | Columbia R. | 5 (55) | ME | -1.47 | 0.35 | -0.46 | 0.28 |
| | | | AME | 2.14 | 0.71 | 0.51 | 0.49 |
| | | | RMS | 2.29 | 0.75 | 0.52 | 0.51 |
| 7.0 | Columbia R. | 11 (163) | ME | -0.08 | 0.21 | -0.30 | 0.15 |
| | | | AME | 1.53 | 0.87 | 0.37 | 0.37 |
| | | | RMS | 1.62 | 0.90 | 0.38 | 0.42 |
| 9.0 | Columbia R. | 18 (284) | ME | -0.70 | 0.39 | -0.40 | 0.25 |
| | | | AME | 1.37 | 0.79 | 0.50 | 0.46 |
| | | | RMS | 1.48 | 0.85 | 0.51 | 0.55 |
| 4.0 | Spokane R. | 18 (179) | ME | -6.68 | 1.18 | -0.21 | -0.49 |
| | | | AME | 10.90 | 1.44 | 0.38 | 0.86 |
| | | | RMS | 11.80 | 1.51 | 0.40 | 0.94 |
| 8.0 | Sanpoil R. | 11 (121) | ME | -0.42 | 0.16 | -0.27 | 0.35 |
| | | | AME | 1.93 | 0.83 | 0.35 | 0.51 |
| | | | RMS | 2.09 | 0.87 | 0.36 | 0.55 |
| System | Average | 114 (1308) | ME | -0.18 | -0.01 | -0.42 | -0.14 |
| | | | AME | 0.39 | 0.14 | 0.43 | 0.49 |
| | | | RMS | 0.44 | 0.16 | 0.43 | 0.50 |

Total dissolved solids

For days when data were recorded, vertical profile plots of total dissolved solids are shown in Figure 193 through Figure 239.

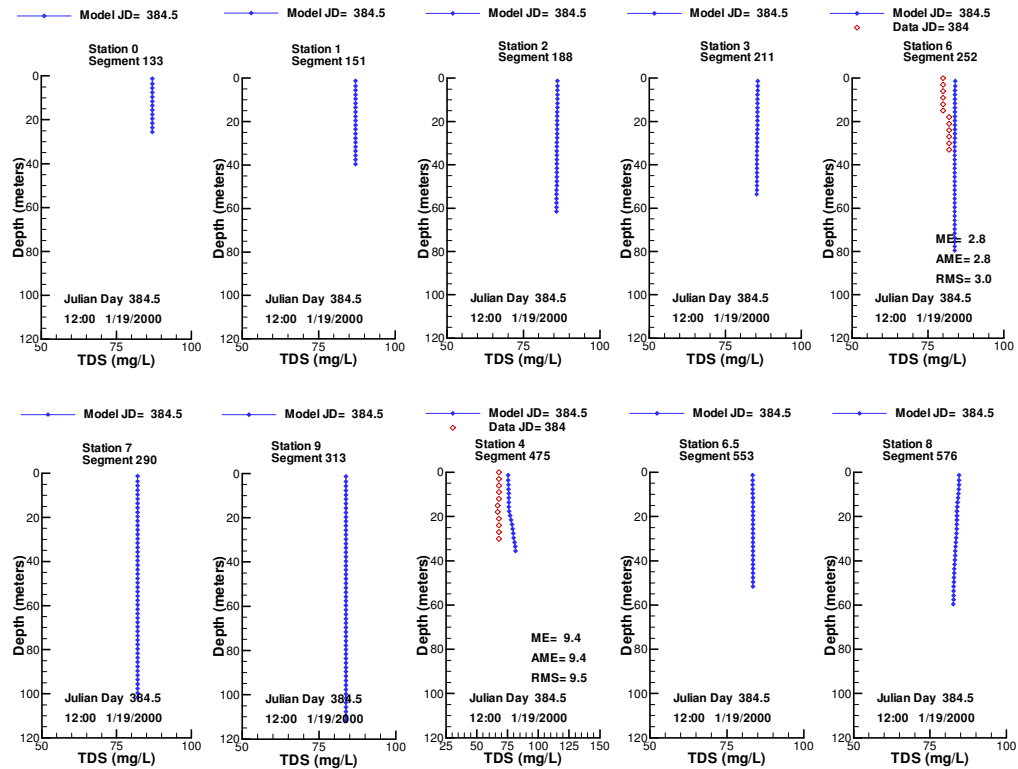


Figure 193. Vertical total dissolved solids model-data comparison, J384.

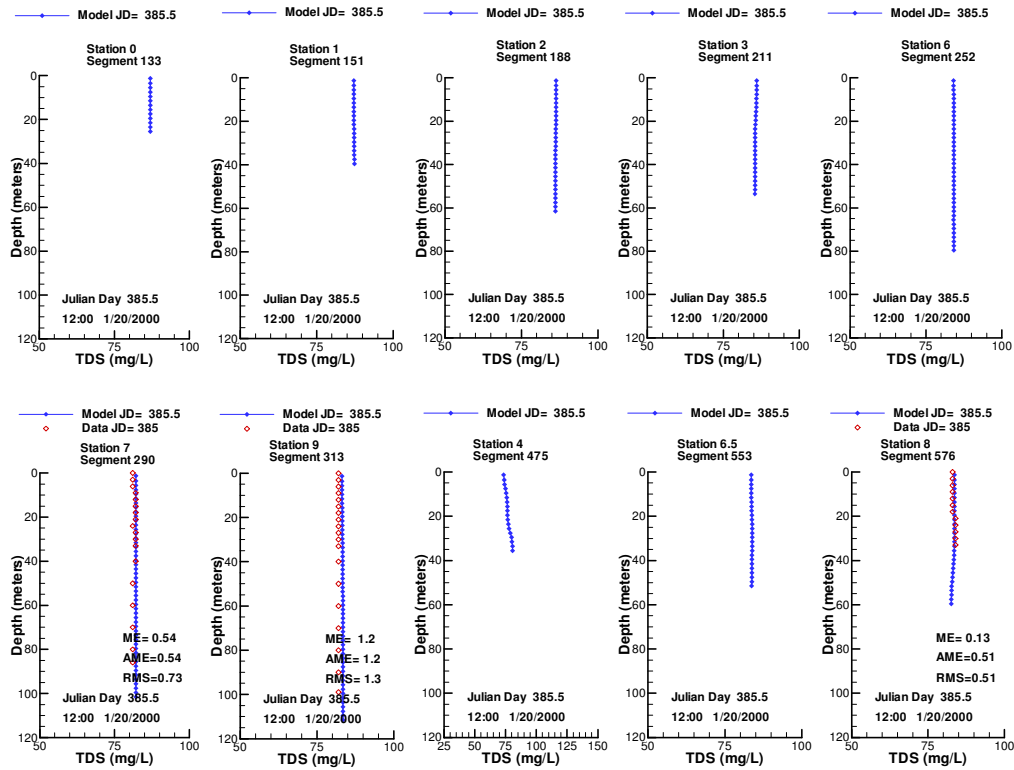


Figure 194. Vertical total dissolved solids model-data comparison, J385.

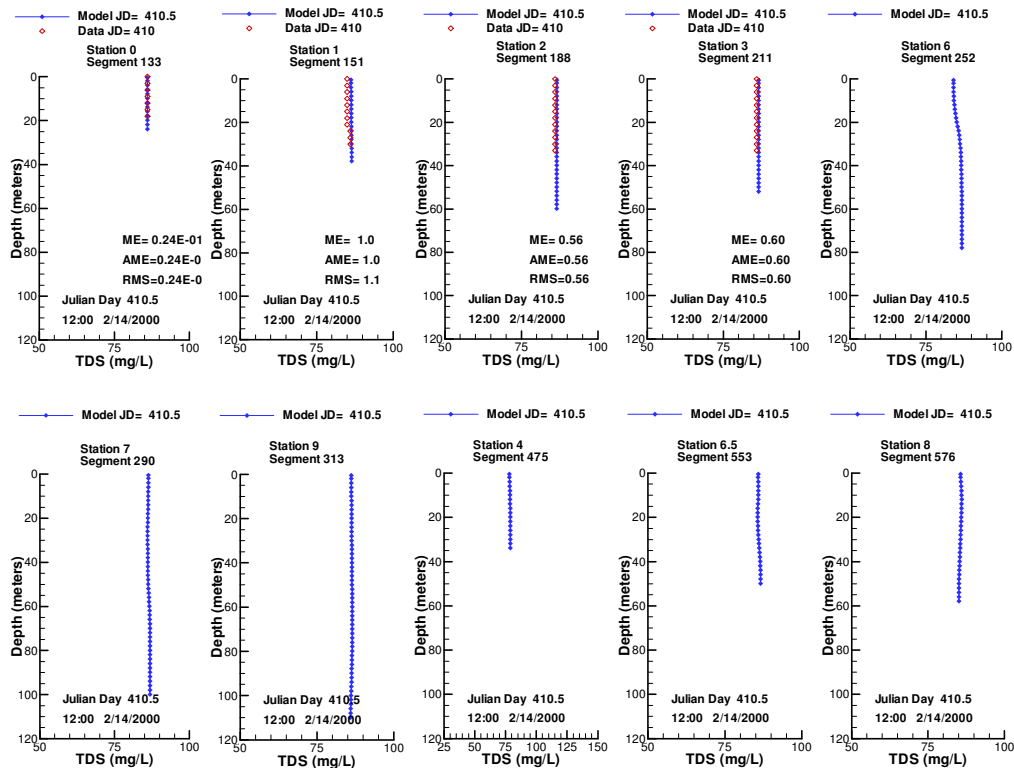


Figure 195. Vertical total dissolved solids model-data comparison, J410.

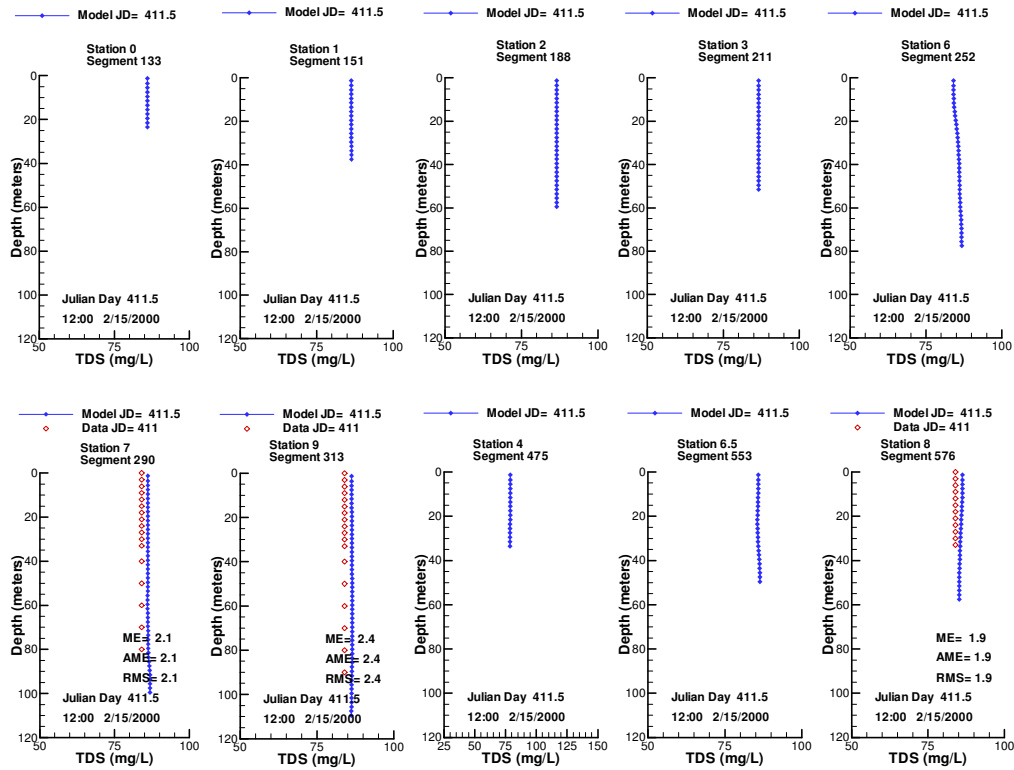


Figure 196. Vertical total dissolved solids model-data comparison, J411.

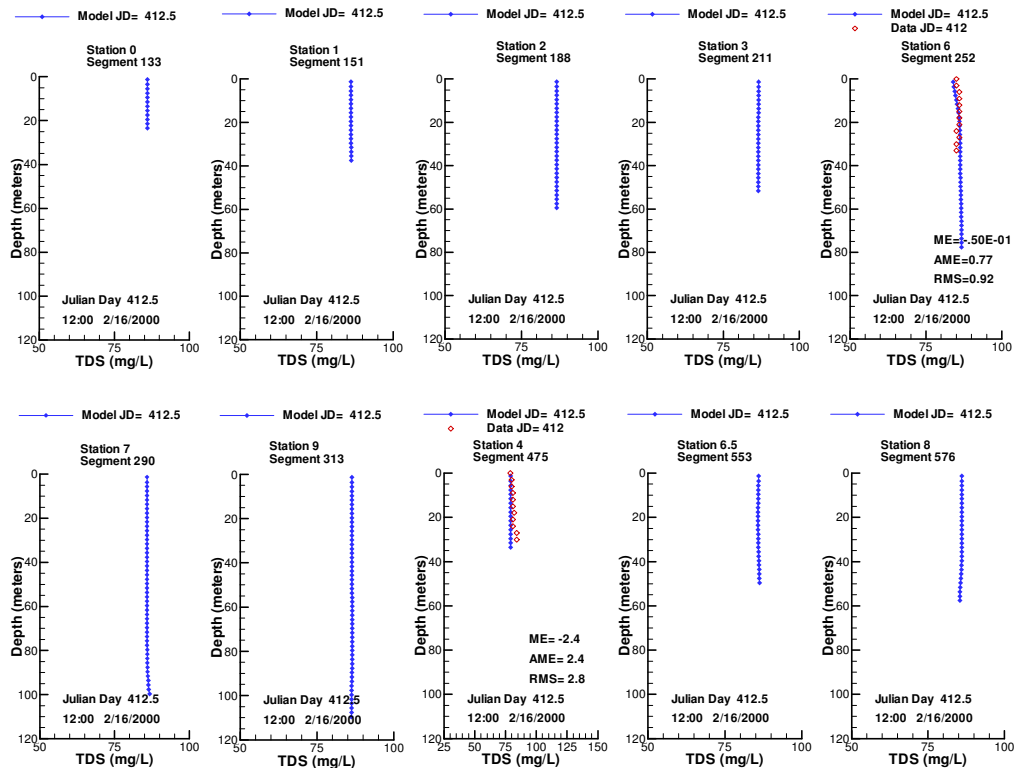


Figure 197. Vertical total dissolved solids model-data comparison, J412.

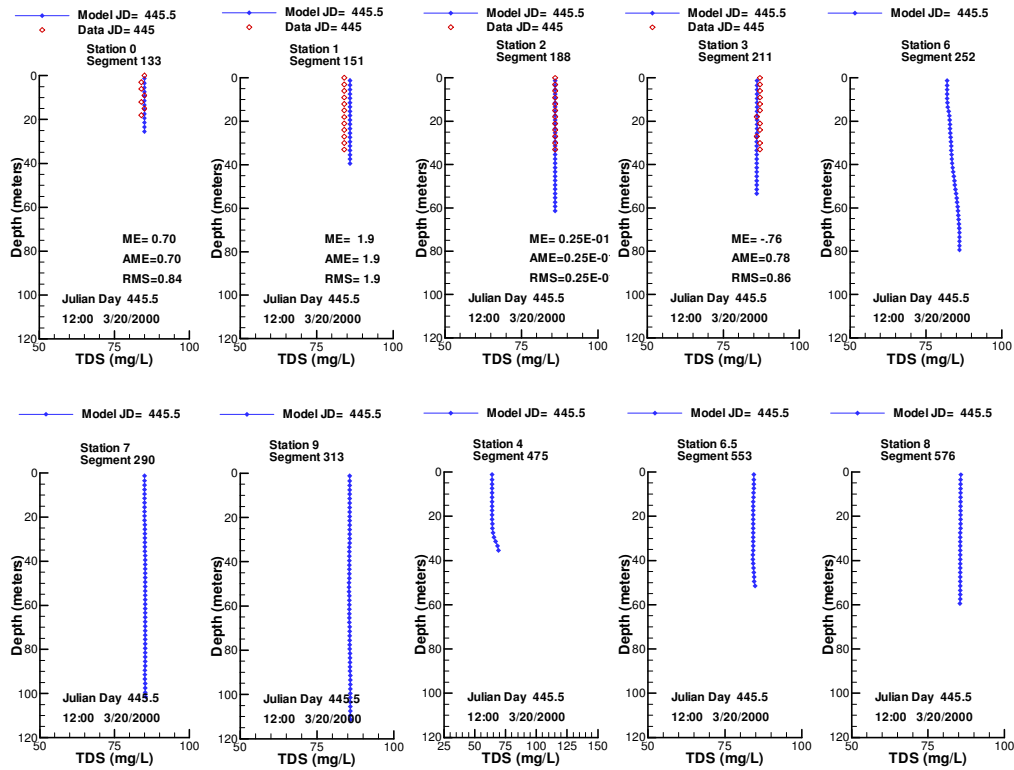


Figure 198. Vertical total dissolved solids model-data comparison, J445.

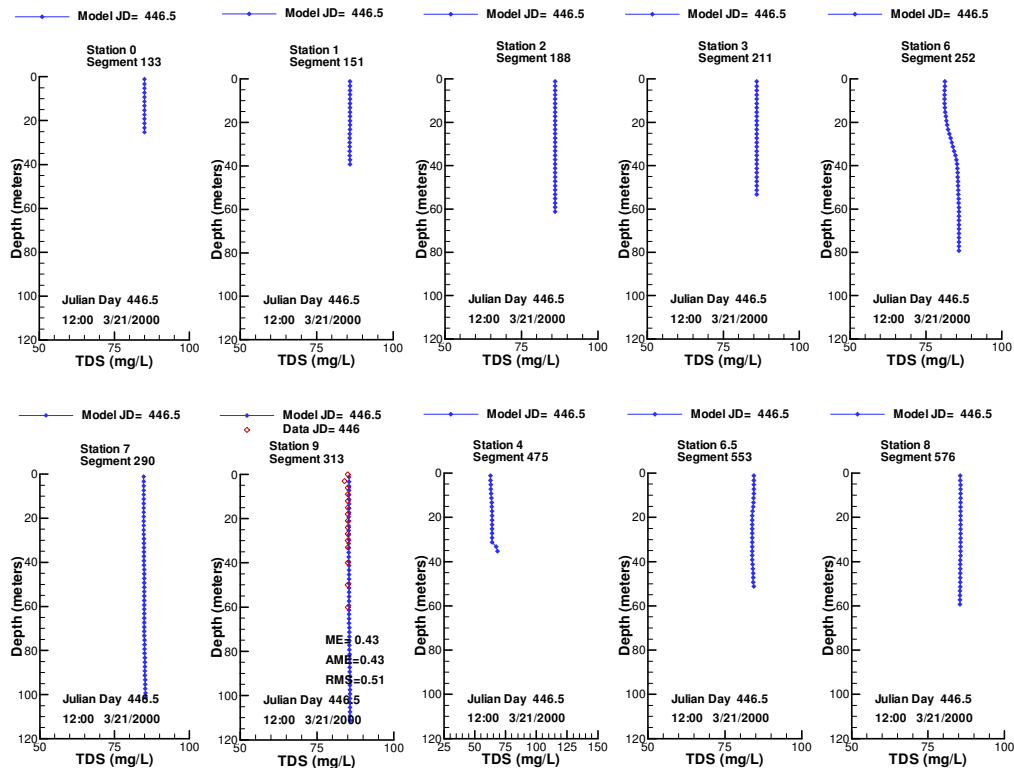


Figure 199. Vertical total dissolved solids model-data comparison, J446.

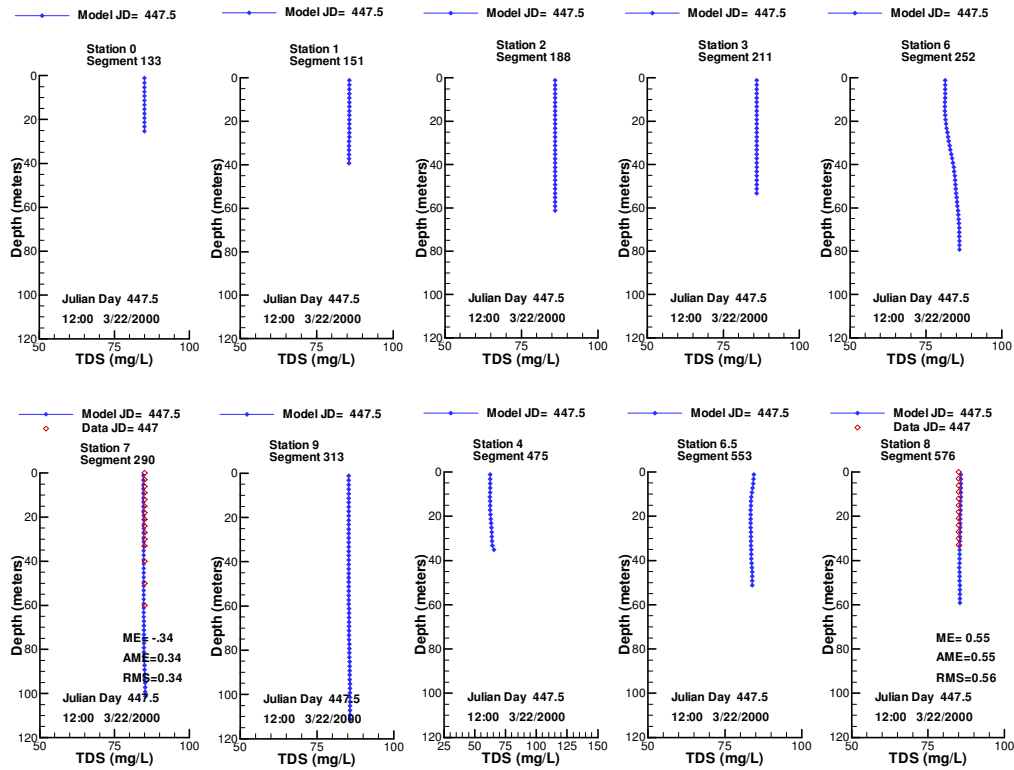


Figure 200. Vertical total dissolved solids model-data comparison, J447.

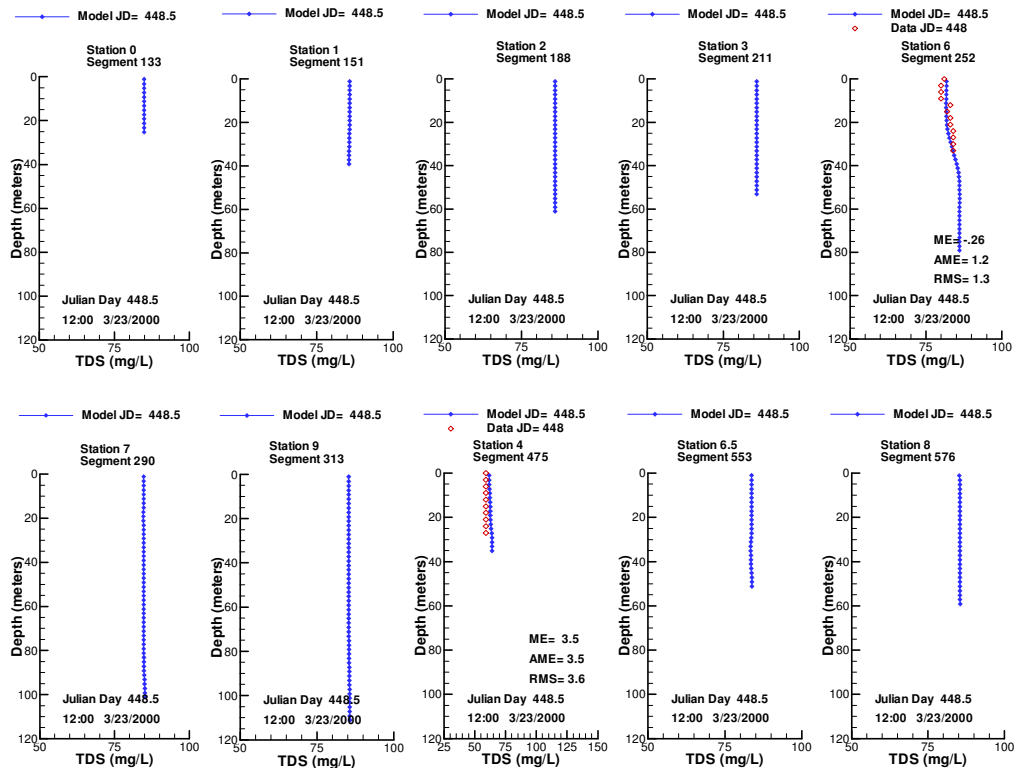


Figure 201. Vertical total dissolved solids model-data comparison, J448.

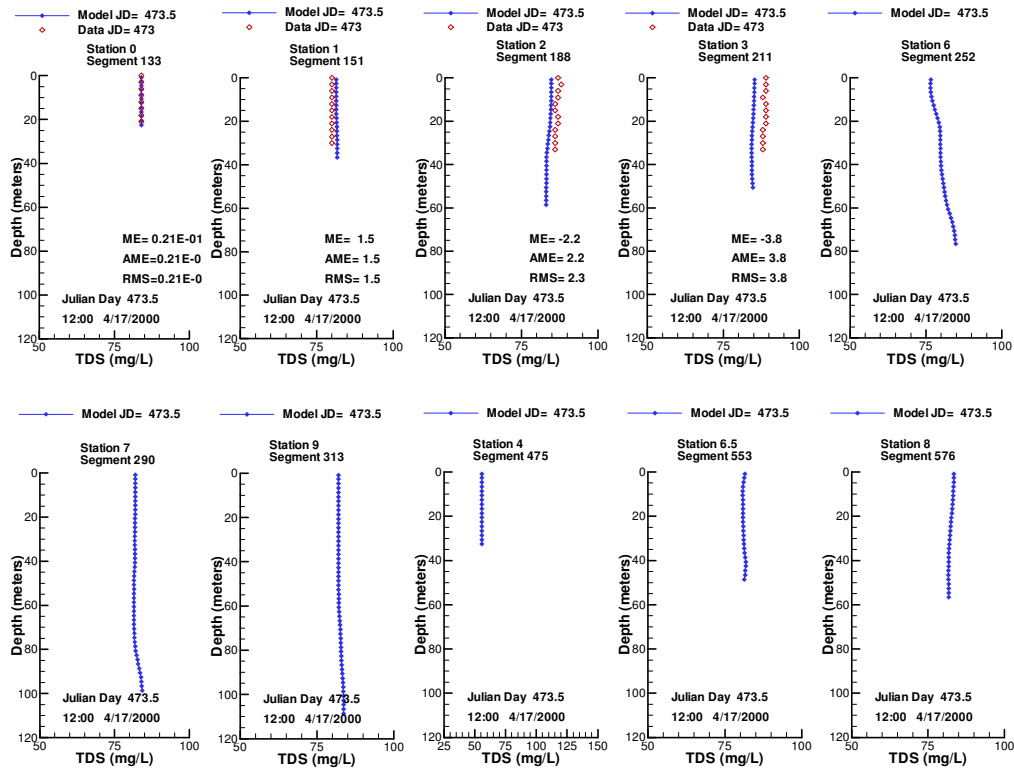


Figure 202. Vertical total dissolved solids model-data comparison, J473.

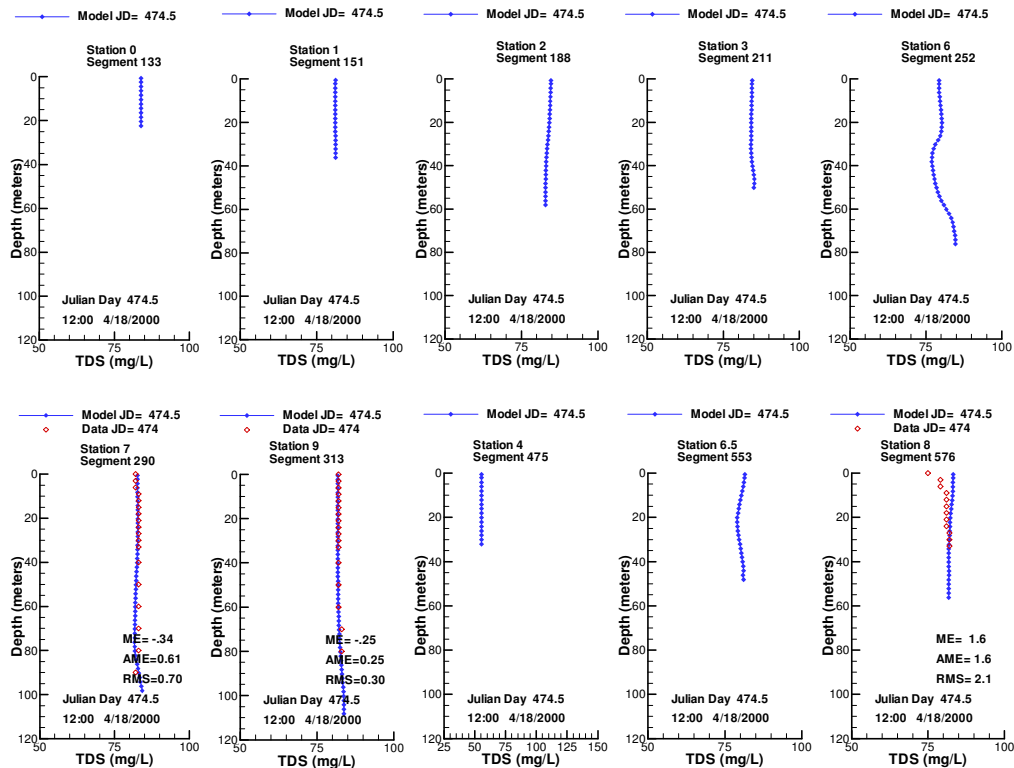


Figure 203. Vertical total dissolved solids model-data comparison, J474.

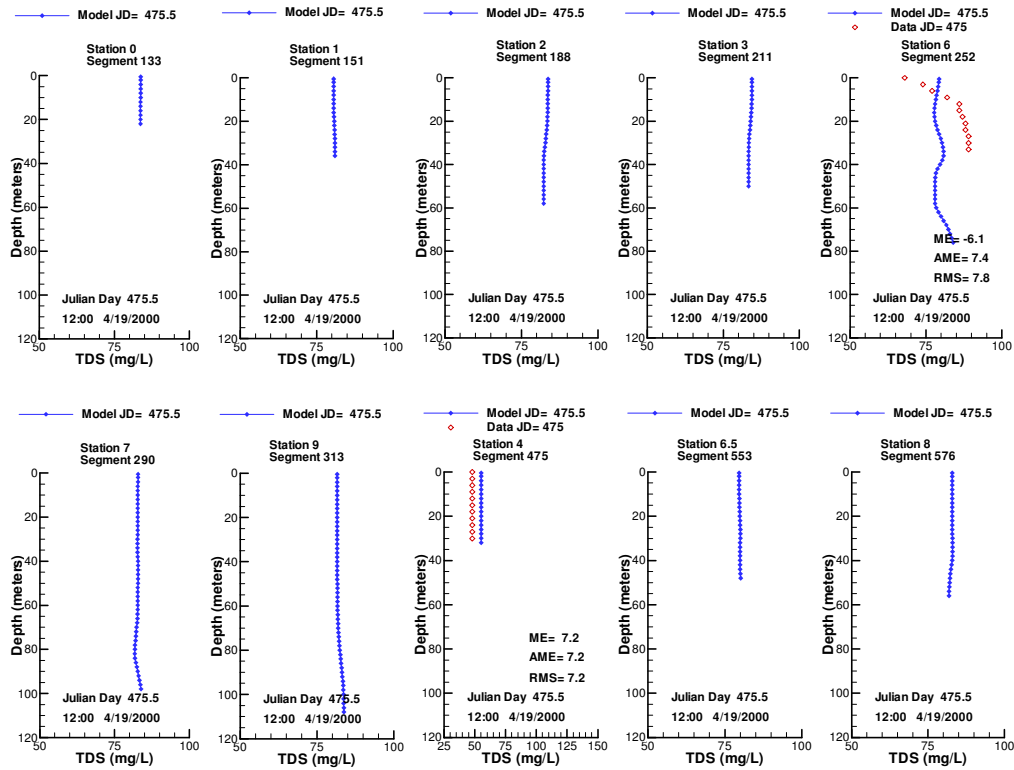


Figure 204. Vertical total dissolved solids model-data comparison, J475.

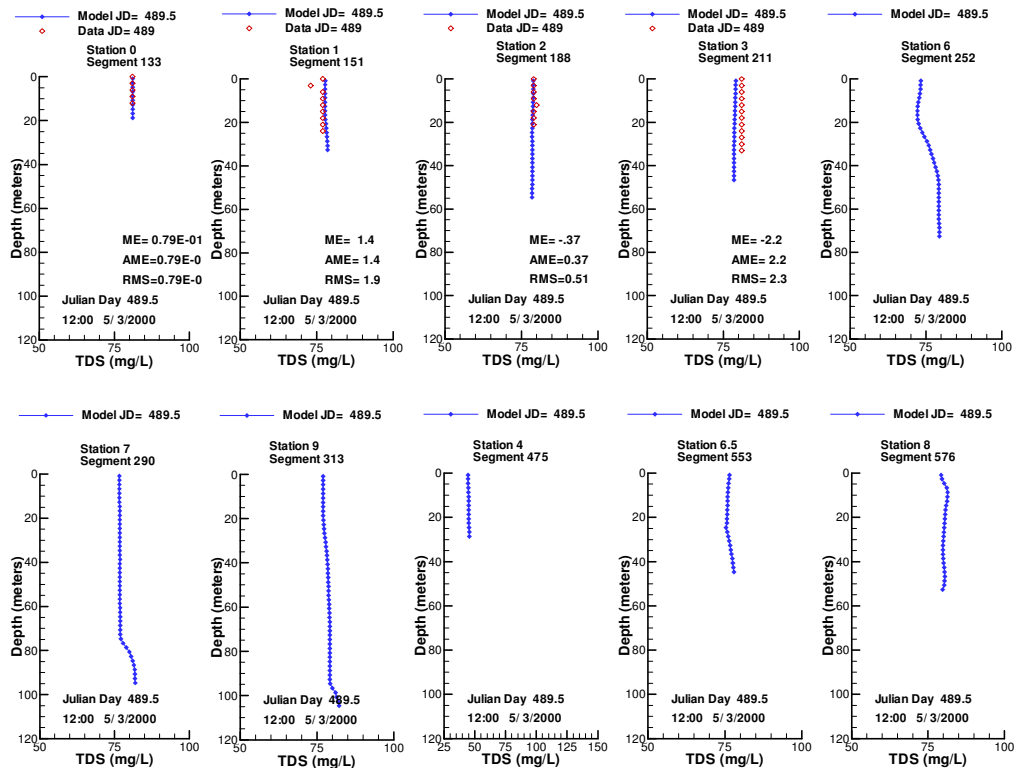


Figure 205. Vertical total dissolved solids model-data comparison, J489.

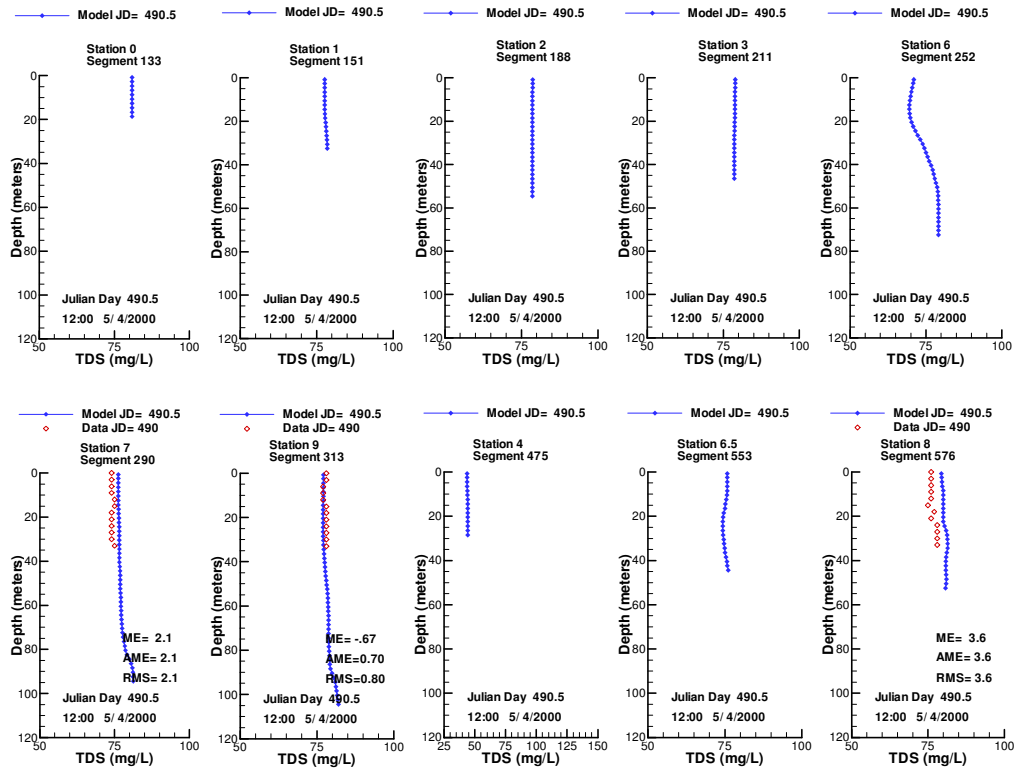


Figure 206. Vertical total dissolved solids model-data comparison, J490.

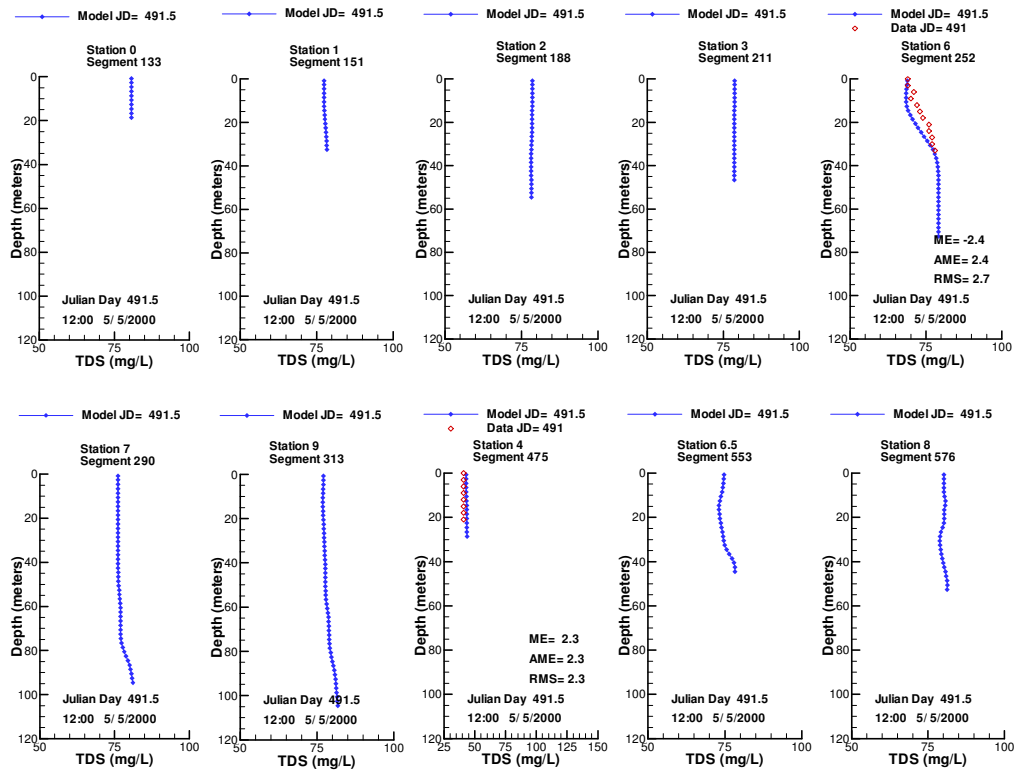


Figure 207. Vertical total dissolved solids model-data comparison, J491.

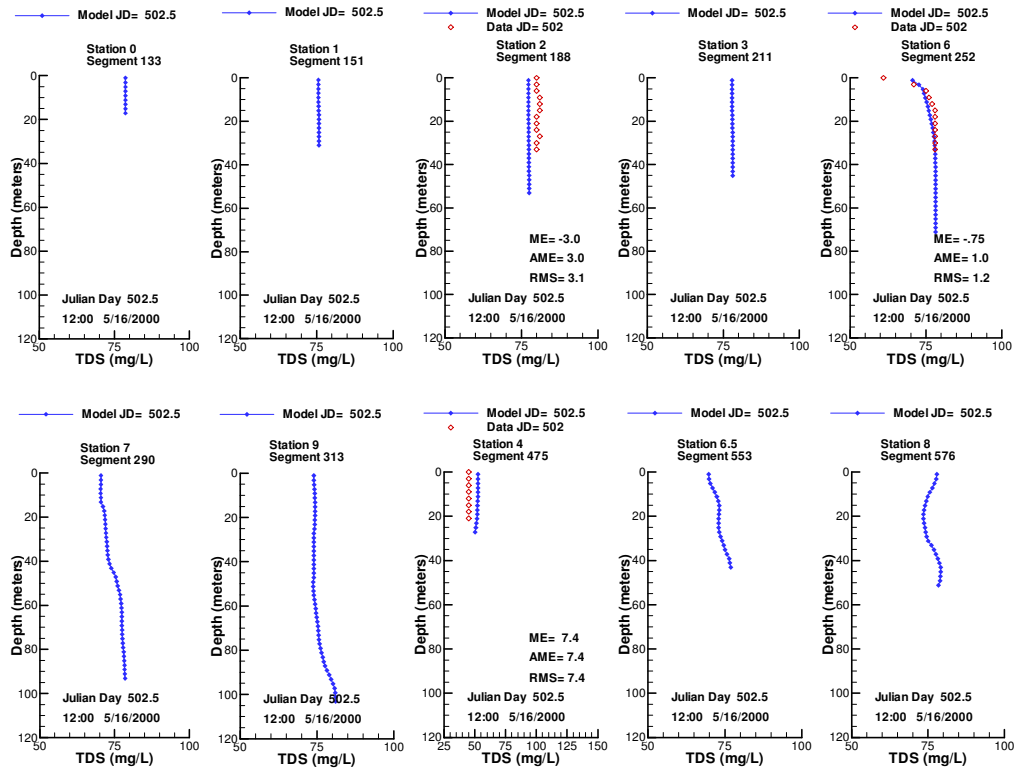


Figure 208. Vertical total dissolved solids model-data comparison, J502.

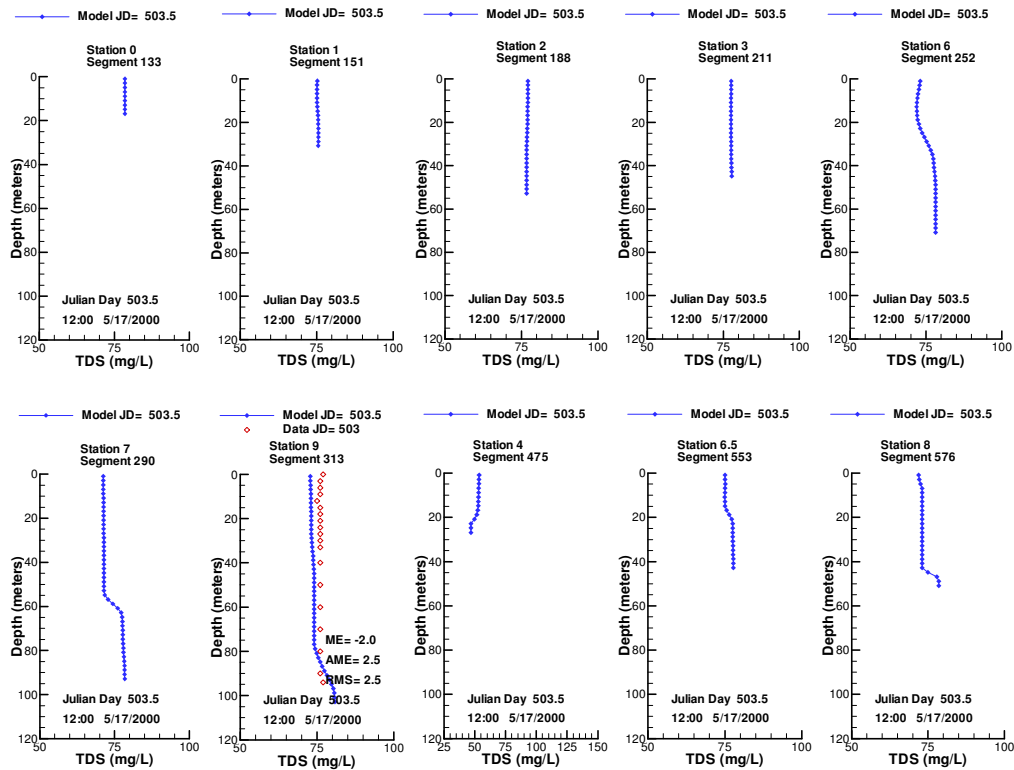


Figure 209. Vertical total dissolved solids model-data comparison, J503.

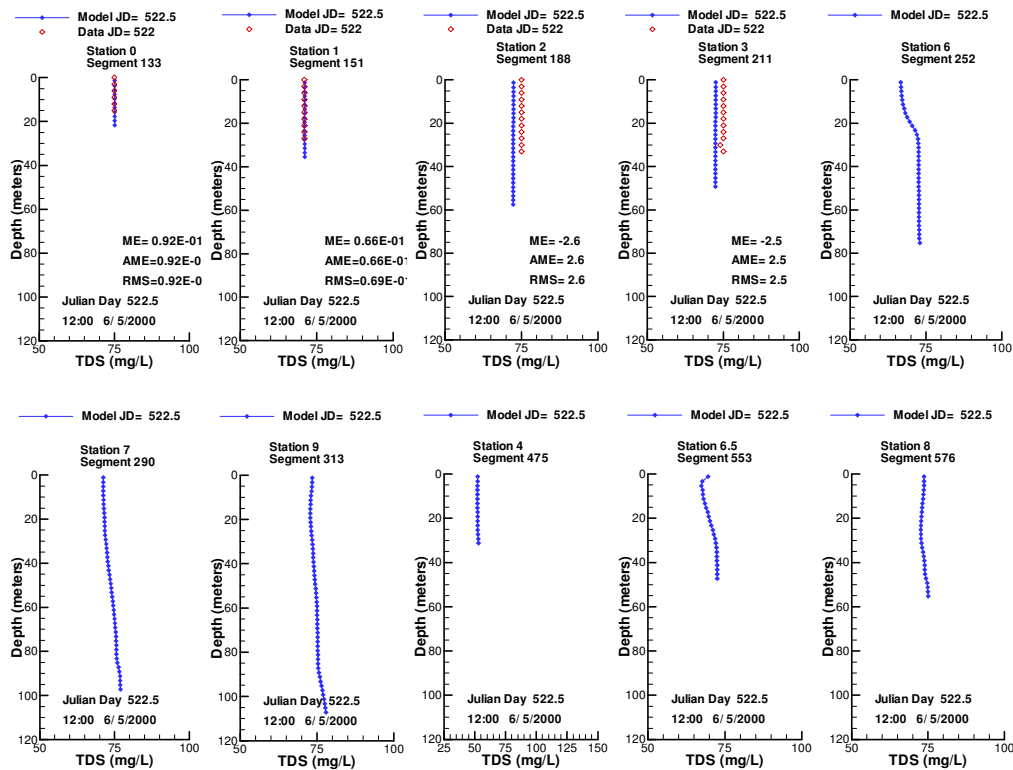


Figure 210. Vertical total dissolved solids model-data comparison, J522.

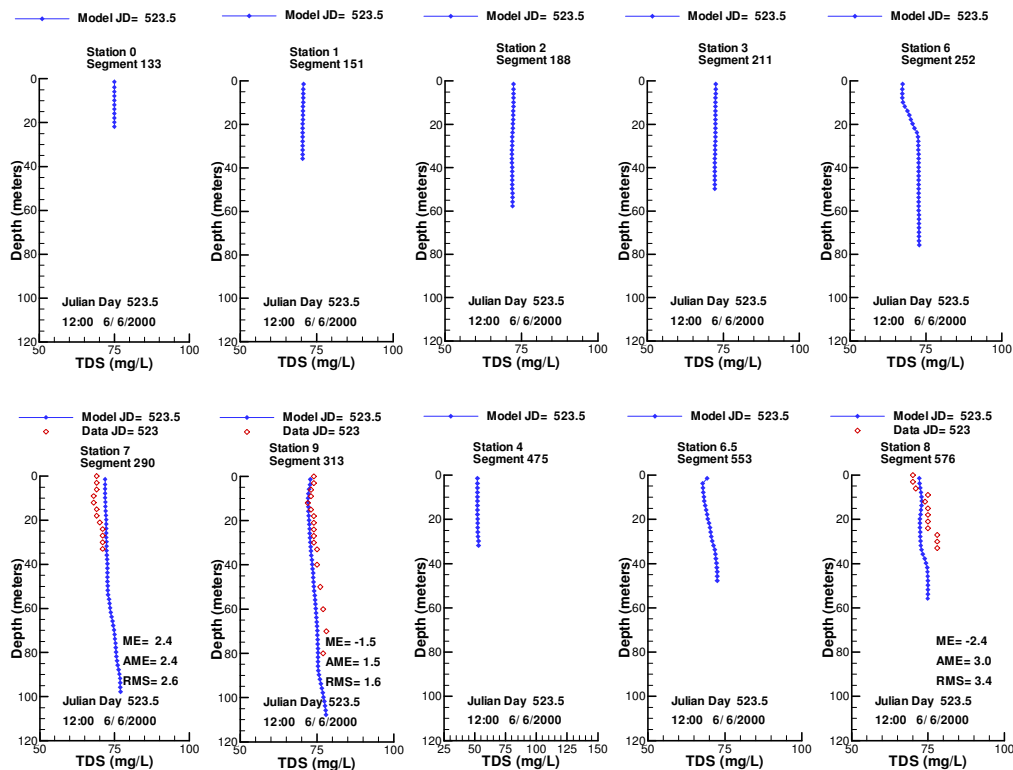


Figure 211. Vertical total dissolved solids model-data comparison, J523.

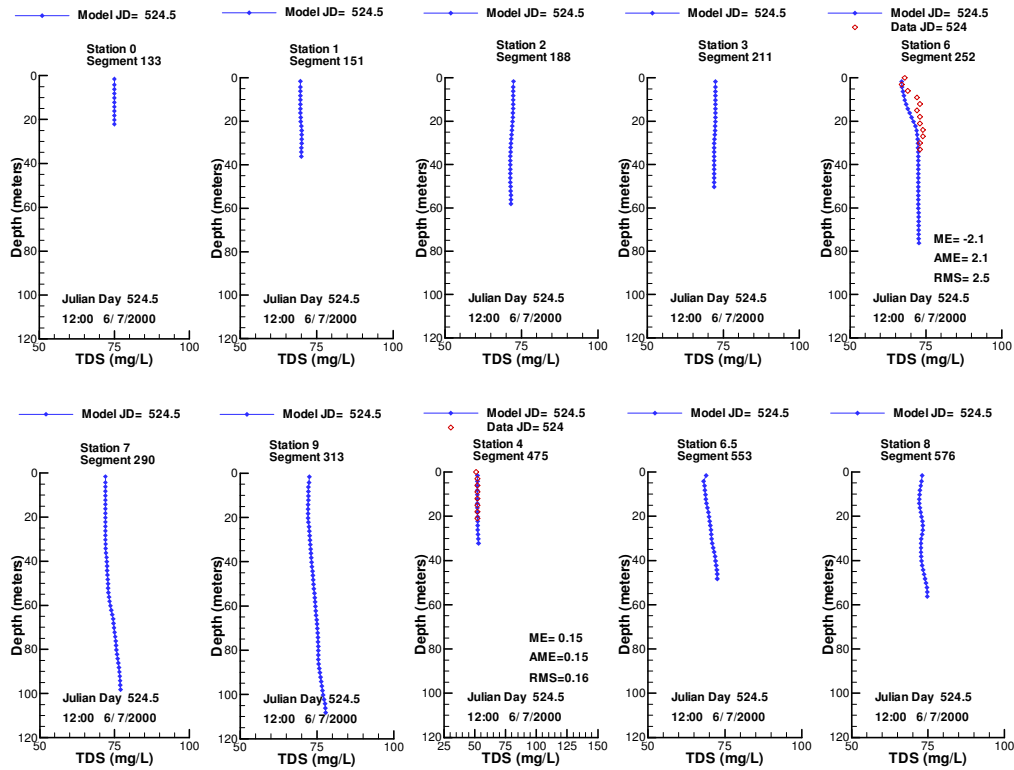


Figure 212. Vertical total dissolved solids model-data comparison, J524.

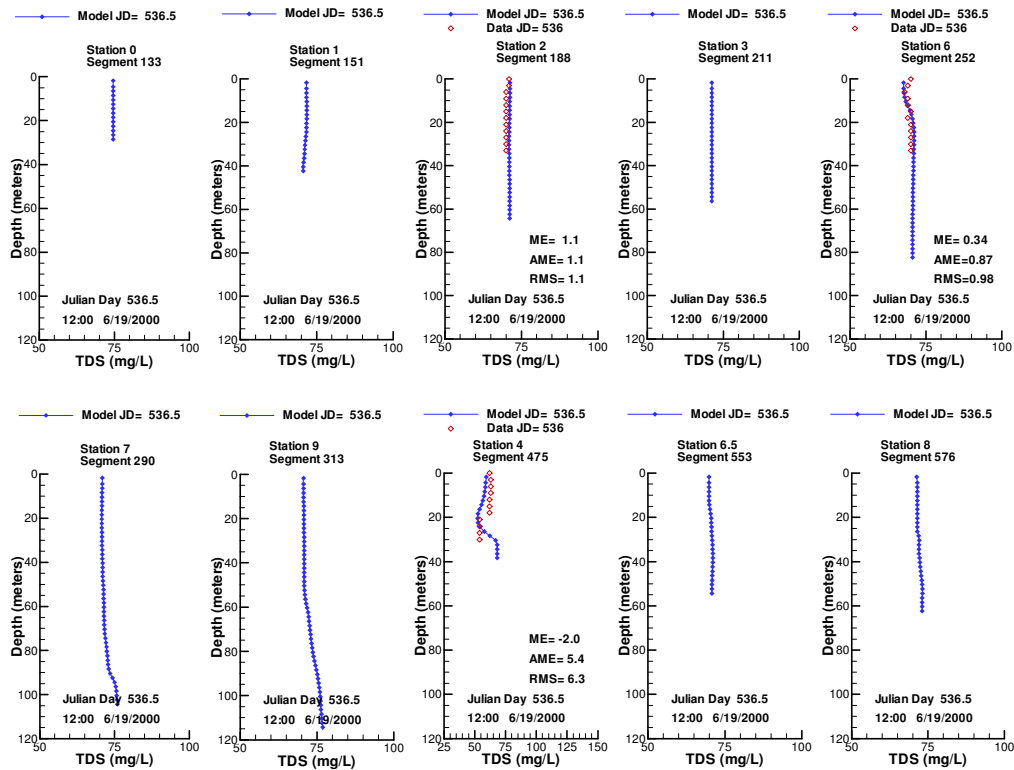


Figure 213. Vertical total dissolved solids model-data comparison, J536.

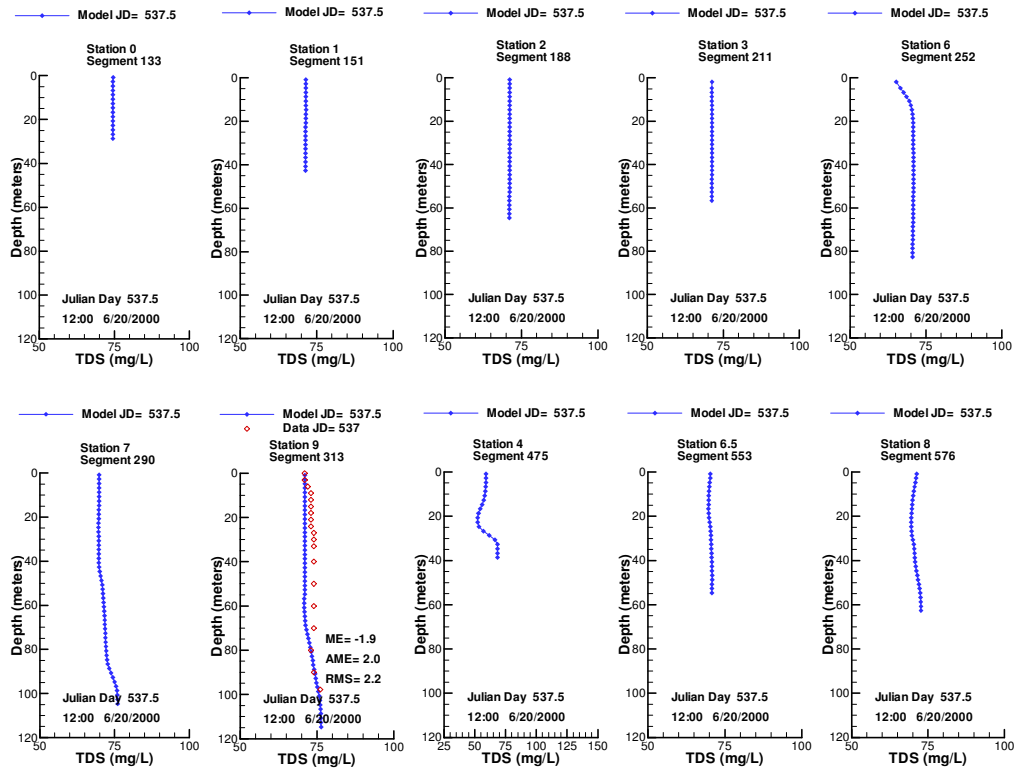


Figure 214. Vertical total dissolved solids model-data comparison, J537.

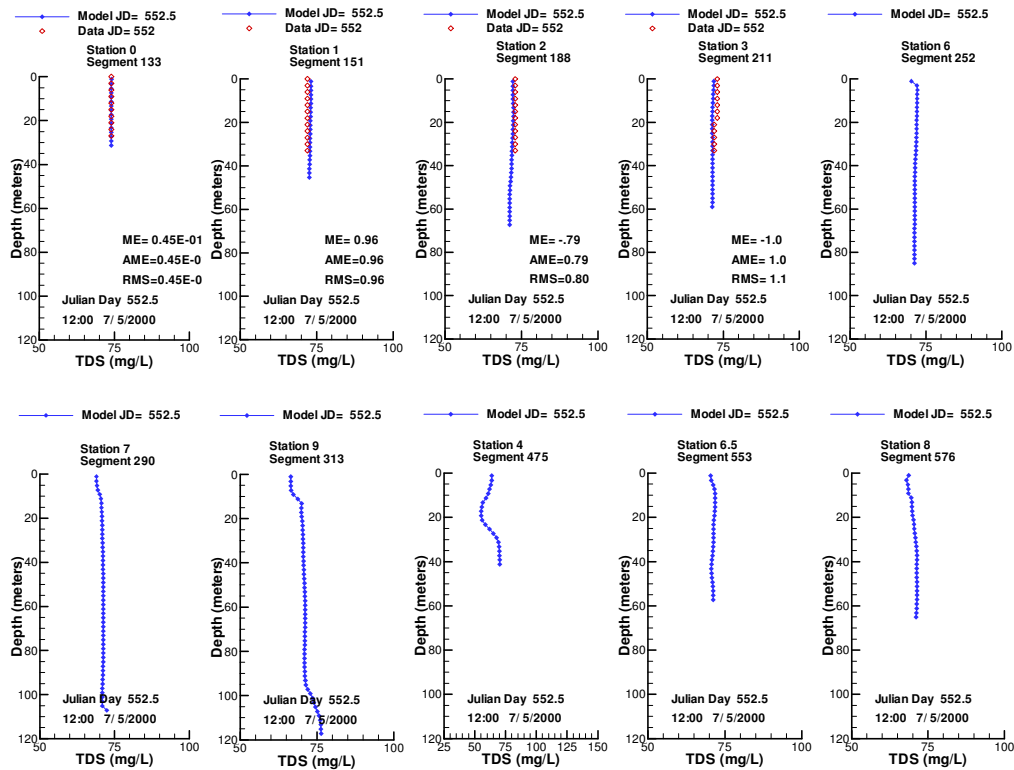


Figure 215. Vertical total dissolved solids model-data comparison, J552.

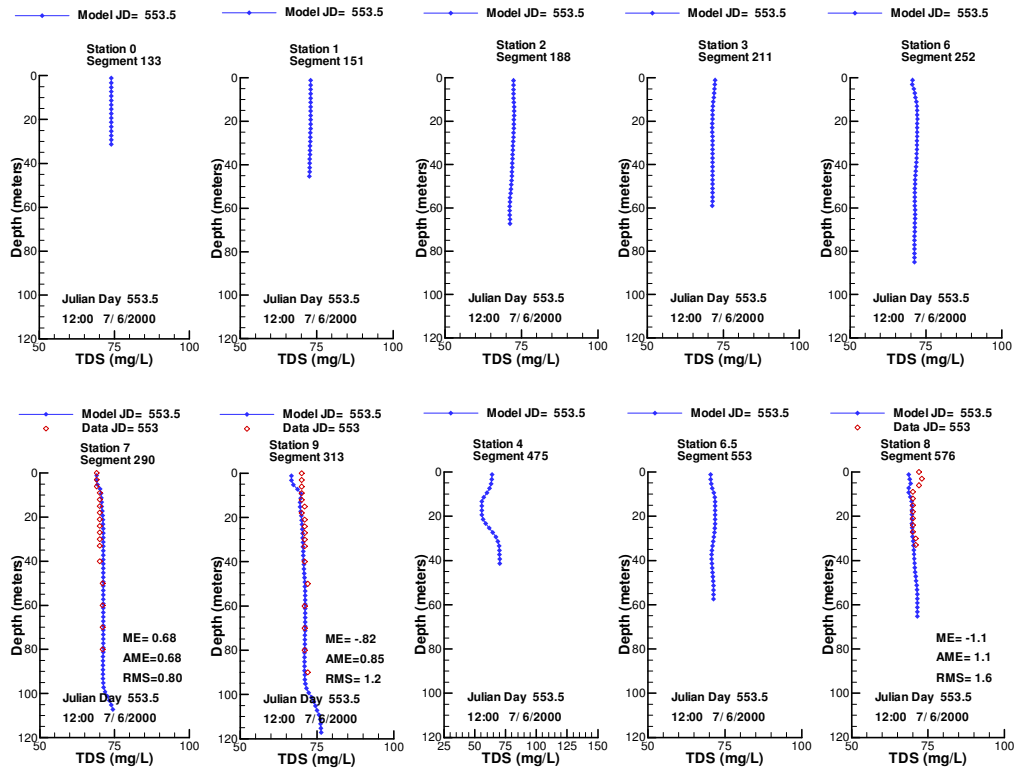


Figure 216. Vertical total dissolved solids model-data comparison, J553.

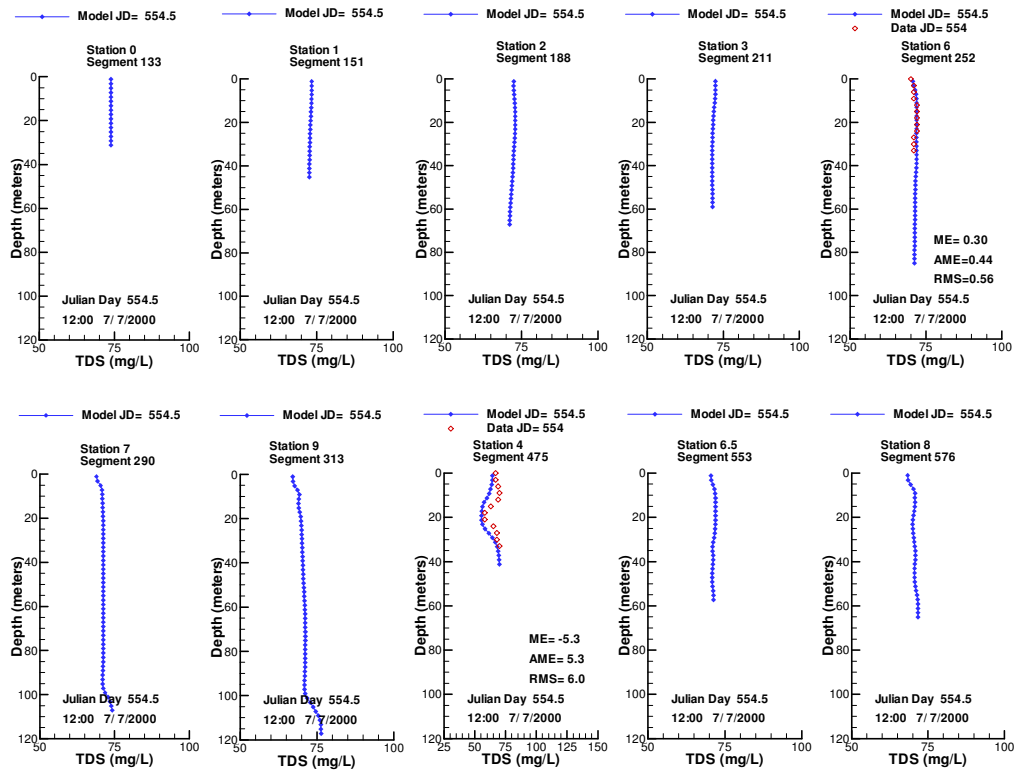


Figure 217. Vertical total dissolved solids model-data comparison, J554.

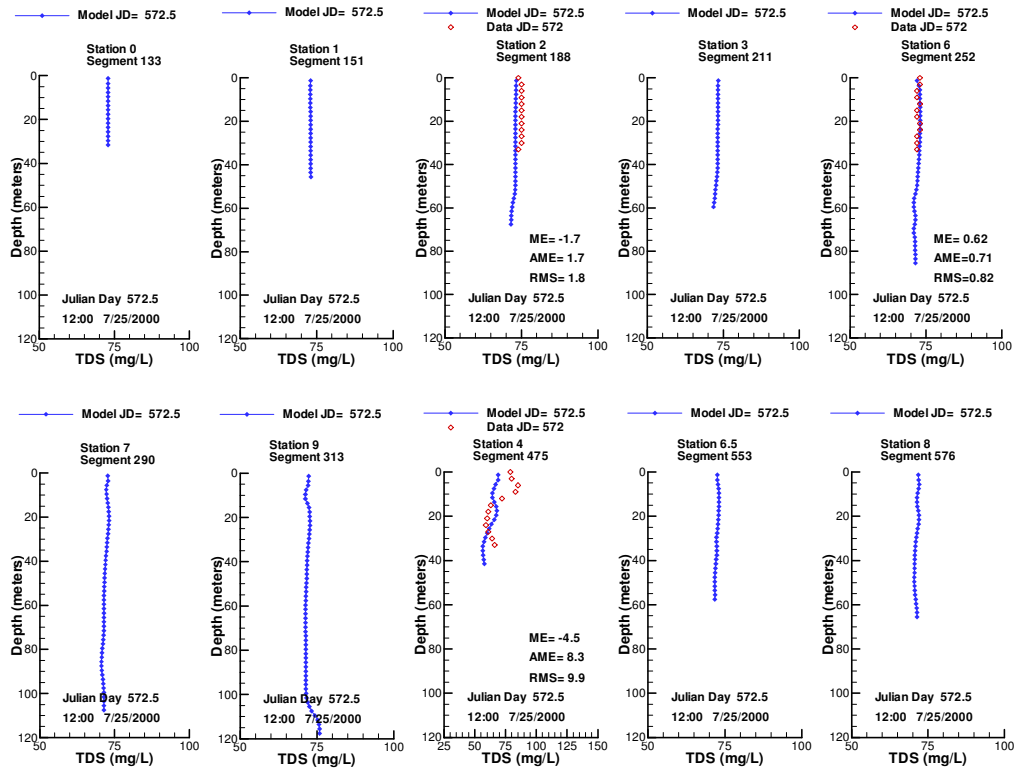


Figure 218. Vertical total dissolved solids model-data comparison, J572.

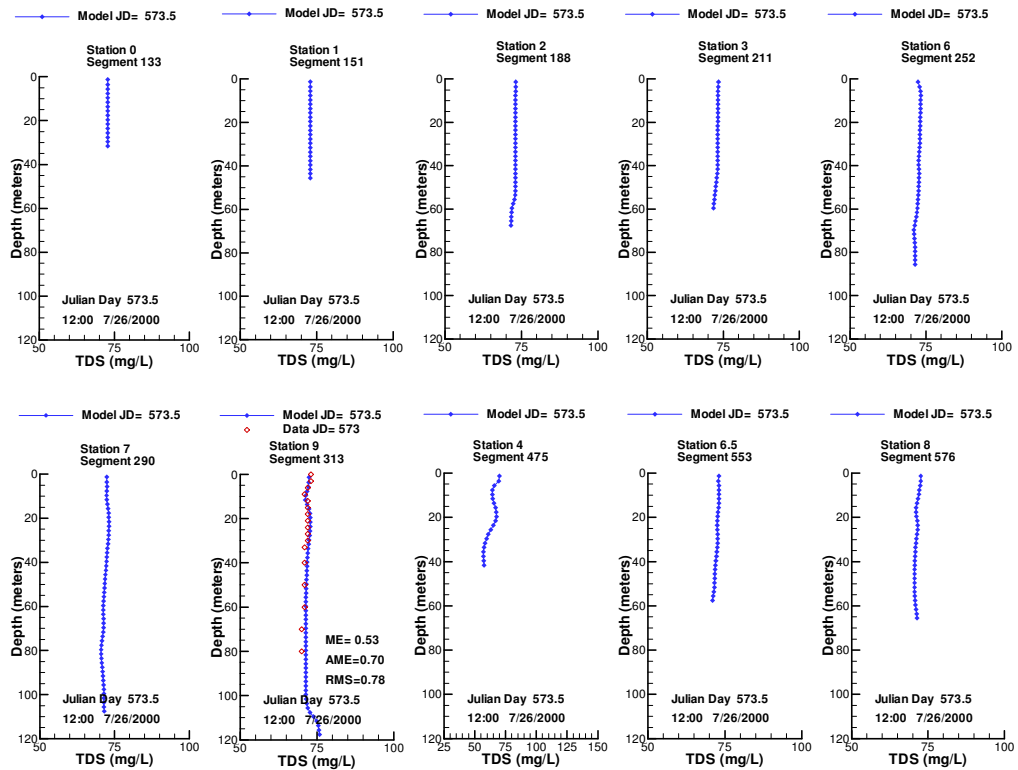


Figure 219. Vertical total dissolved solids model-data comparison, J573.

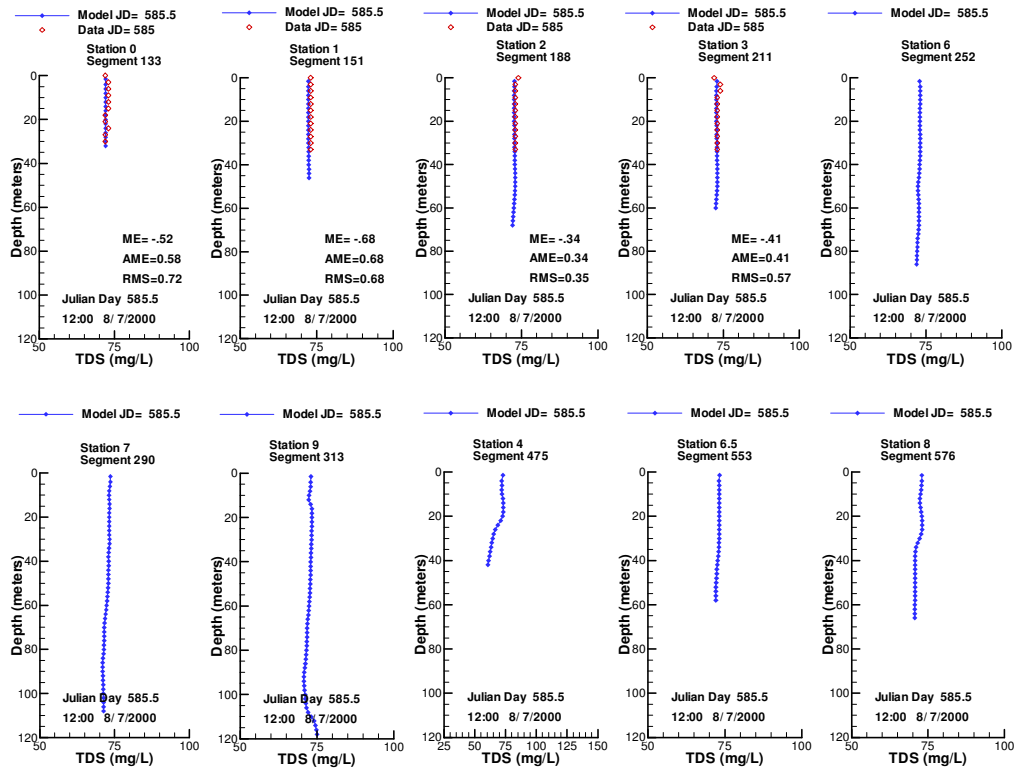


Figure 220. Vertical total dissolved solids model-data comparison, J585.

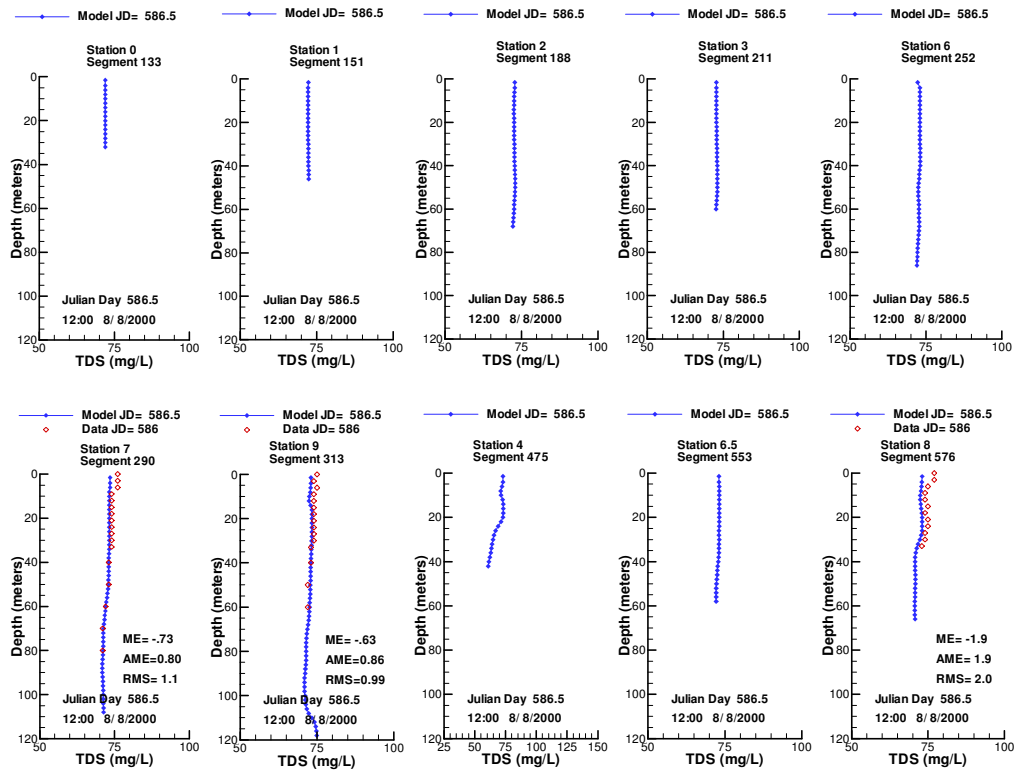


Figure 221. Vertical total dissolved solids model-data comparison, J586.

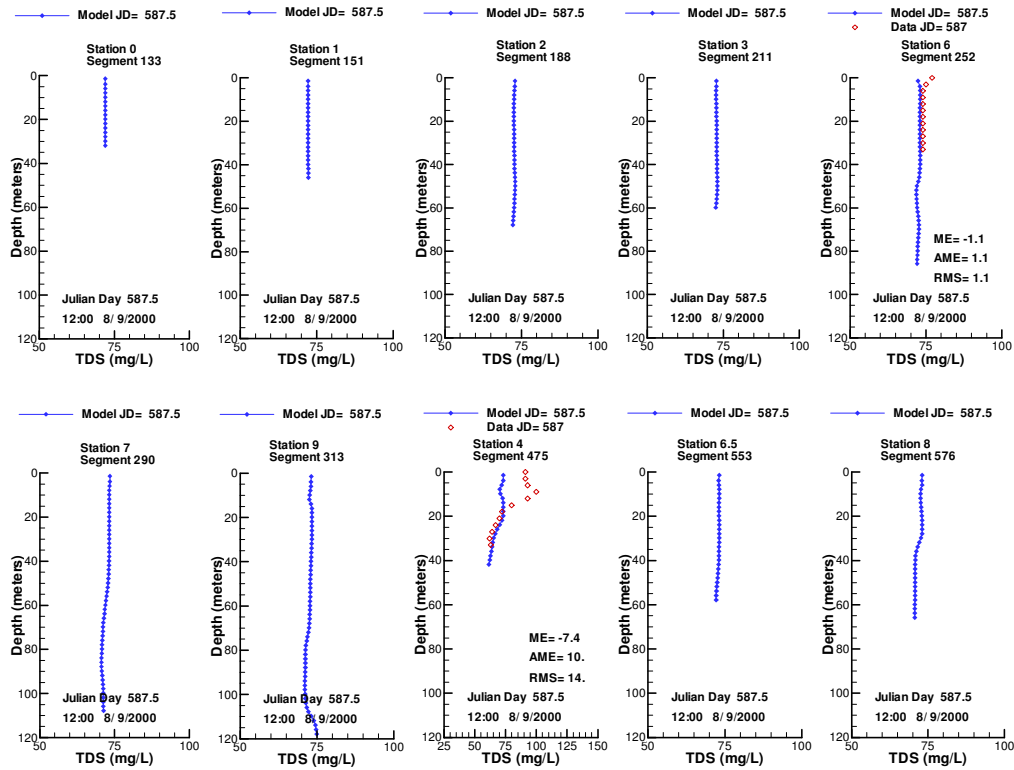


Figure 222. Vertical total dissolved solids model-data comparison, J587.

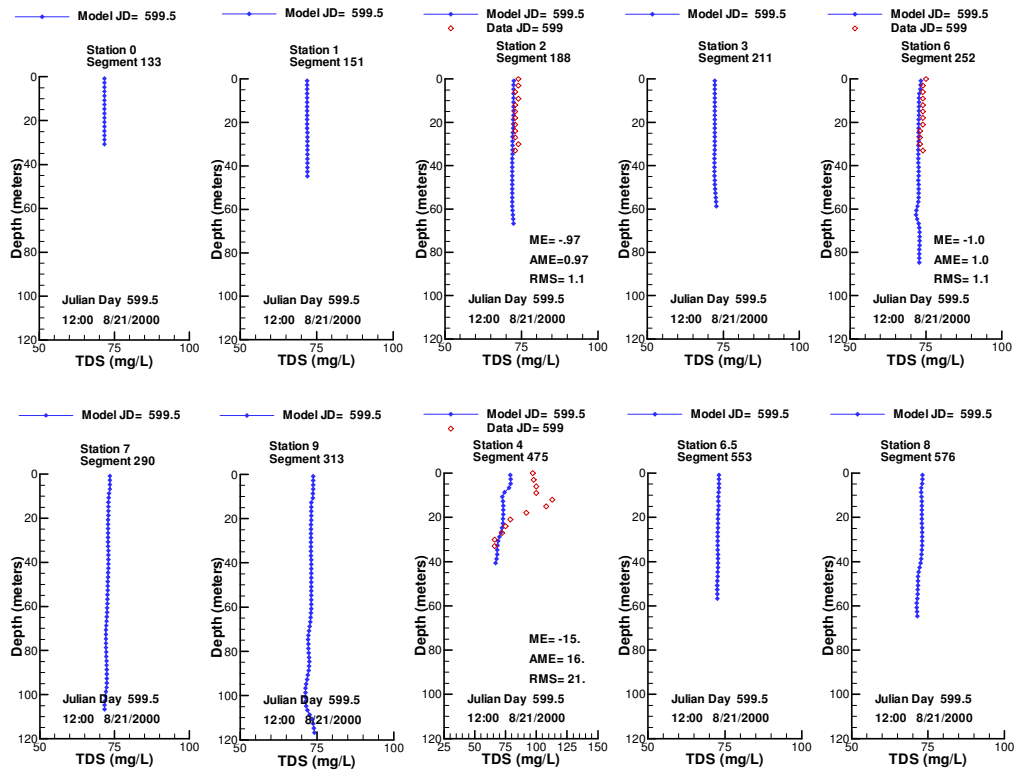


Figure 223. Vertical total dissolved solids model-data comparison, J599.

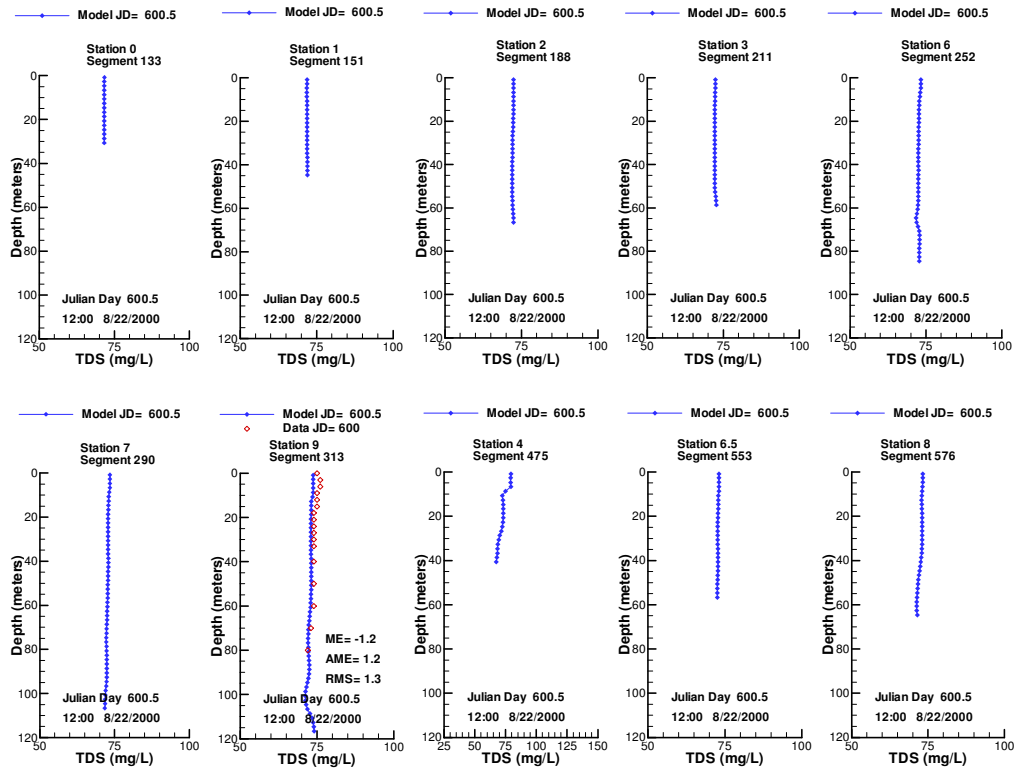


Figure 224. Vertical total dissolved solids model-data comparison, J600.

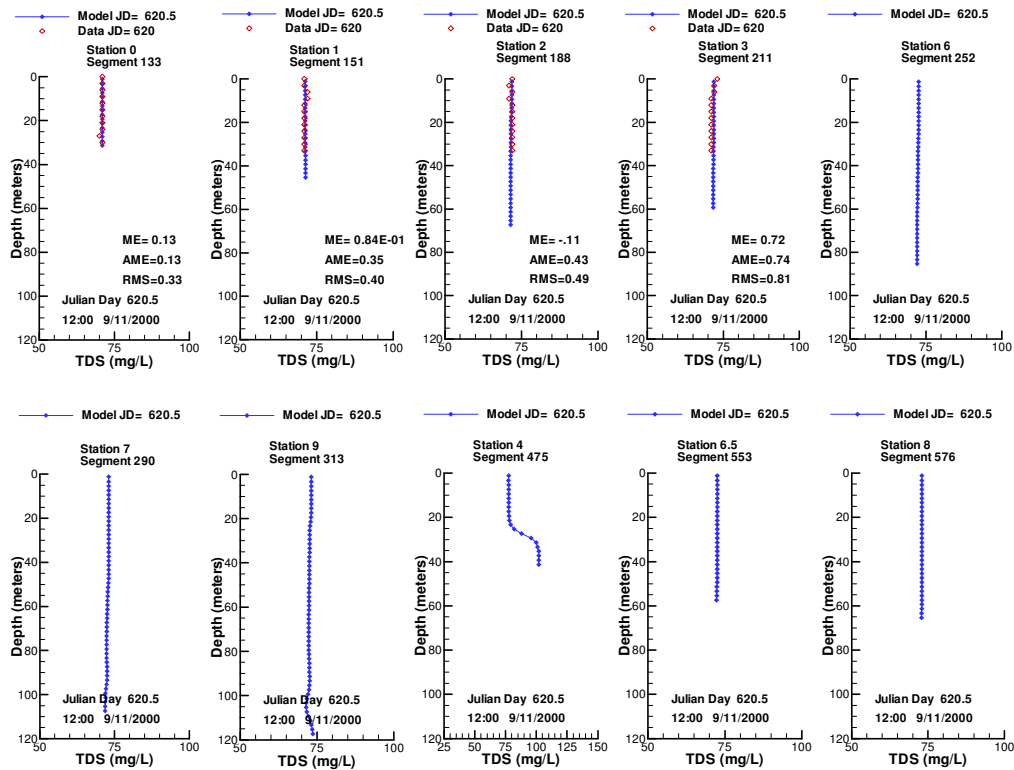


Figure 225. Vertical total dissolved solids model-data comparison, J620.

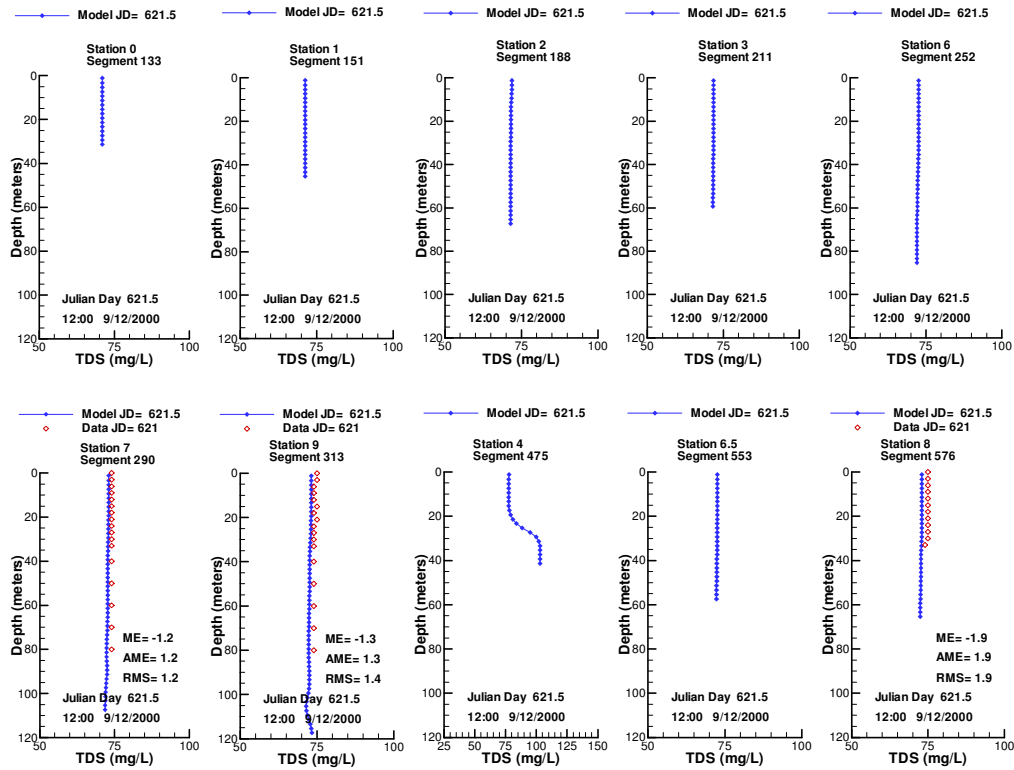


Figure 226. Vertical total dissolved solids model-data comparison, J621.

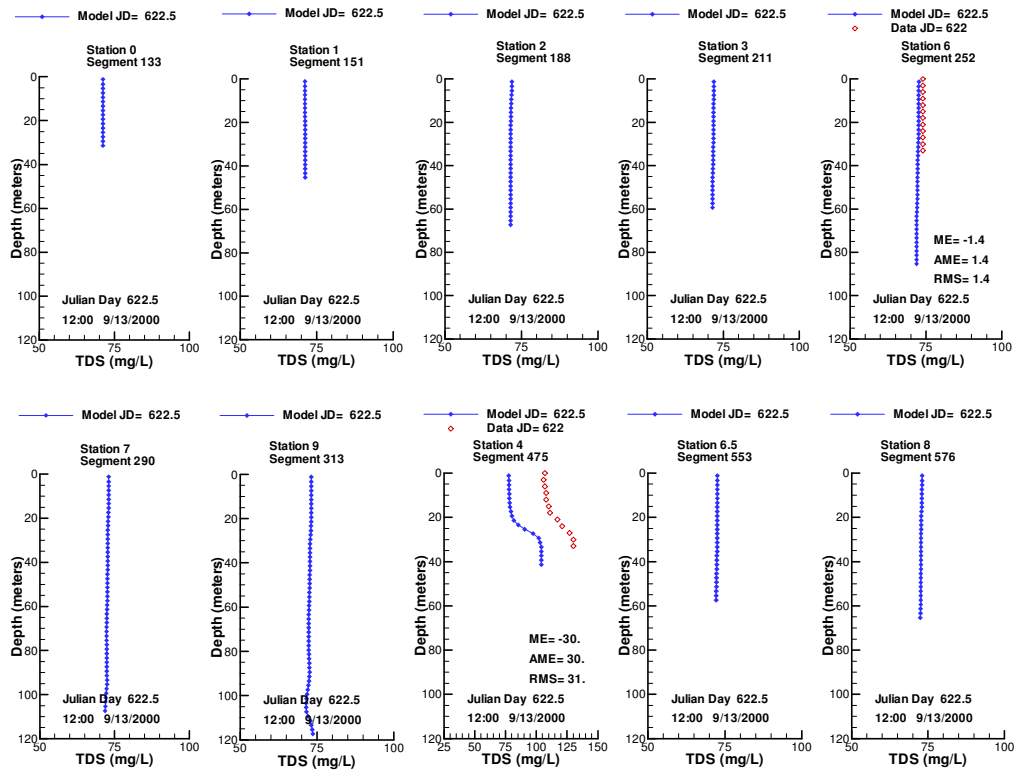


Figure 227. Vertical total dissolved solids model-data comparison, J622.

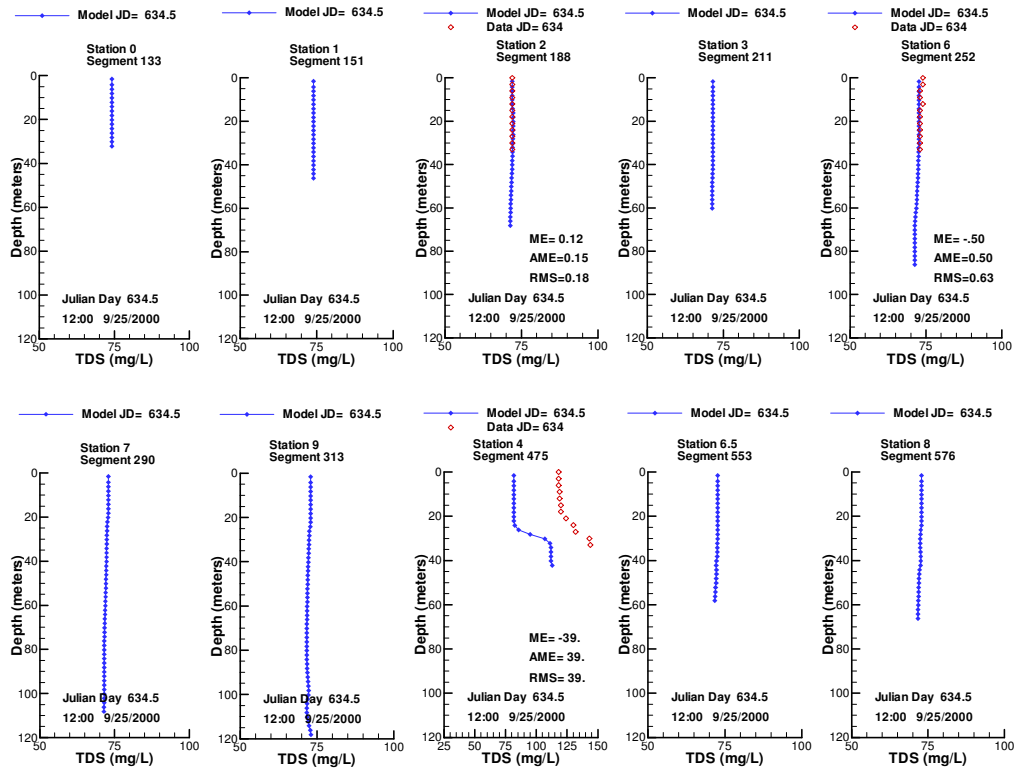


Figure 228. Vertical total dissolved solids model-data comparison, J634.

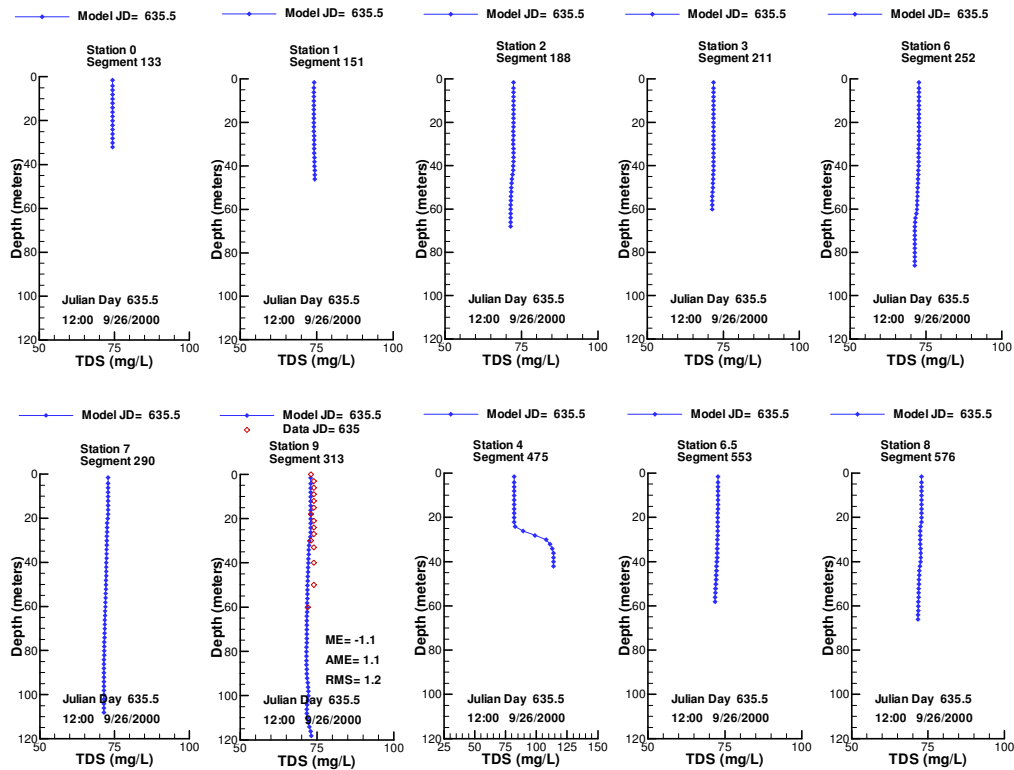


Figure 229. Vertical total dissolved solids model-data comparison, J635.

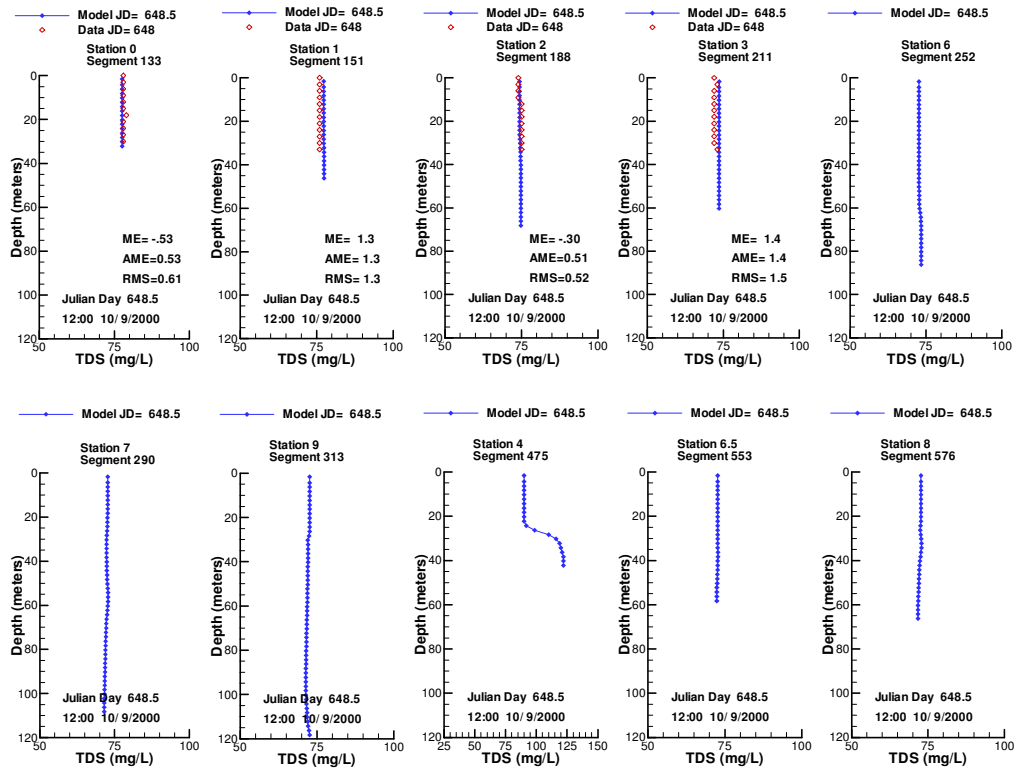


Figure 230. Vertical total dissolved solids model-data comparison, J648.

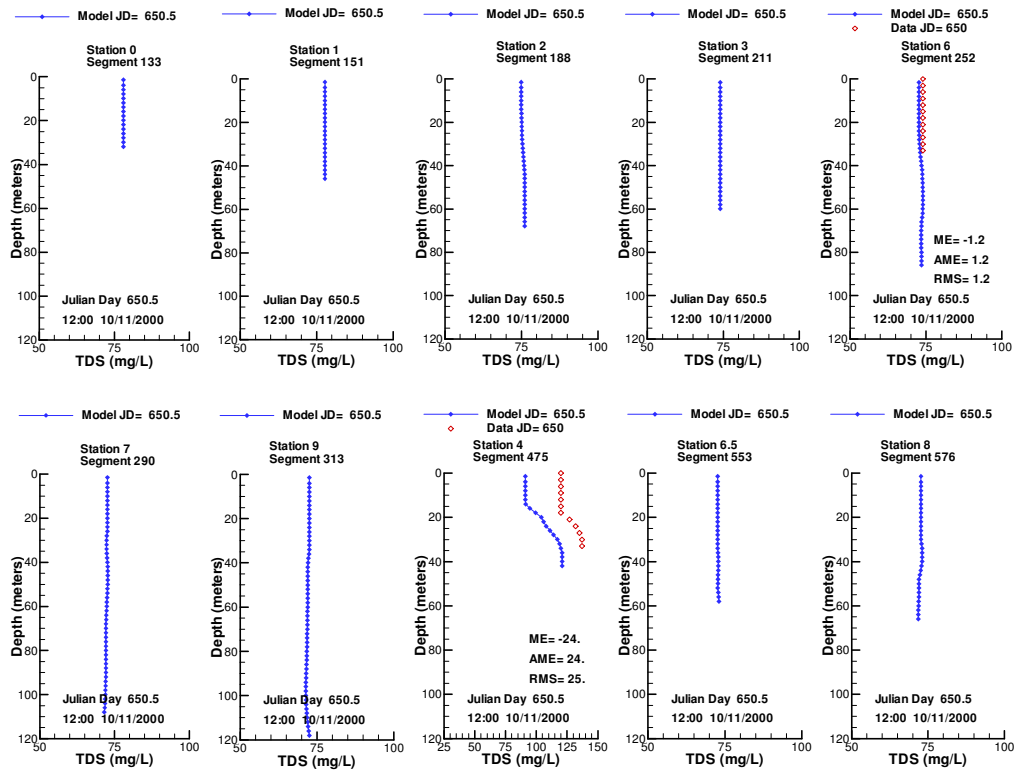


Figure 231. Vertical total dissolved solids model-data comparison, J650.

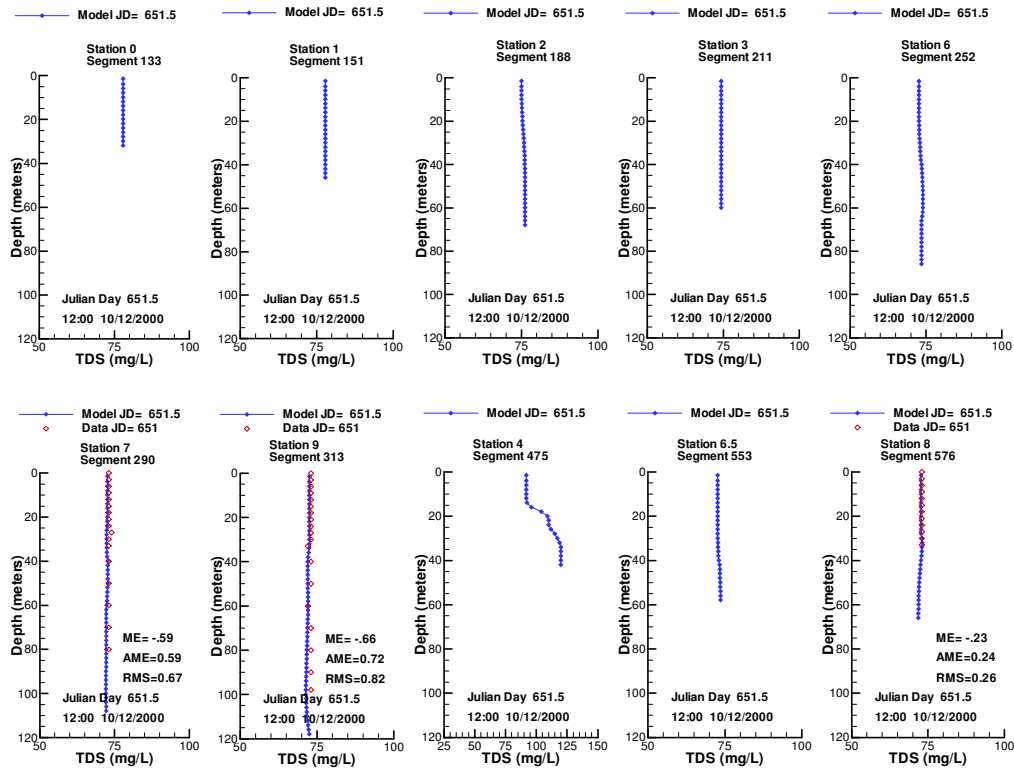


Figure 232. Vertical total dissolved solids model-data comparison, J651.

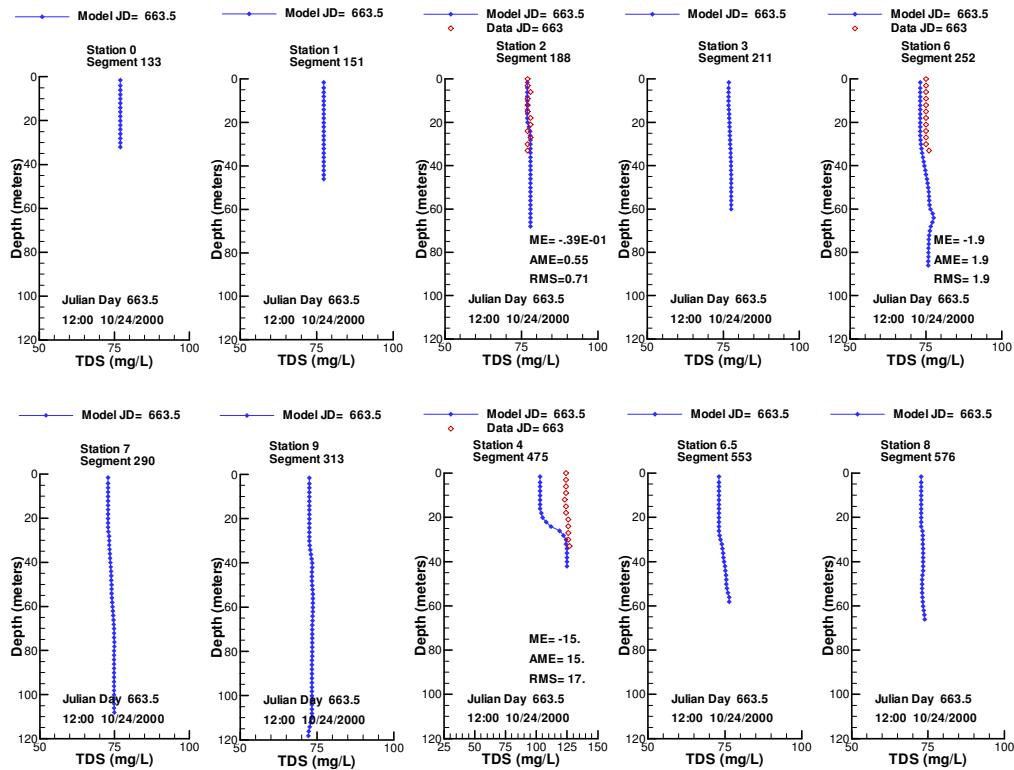


Figure 233. Vertical total dissolved solids model-data comparison, J663.

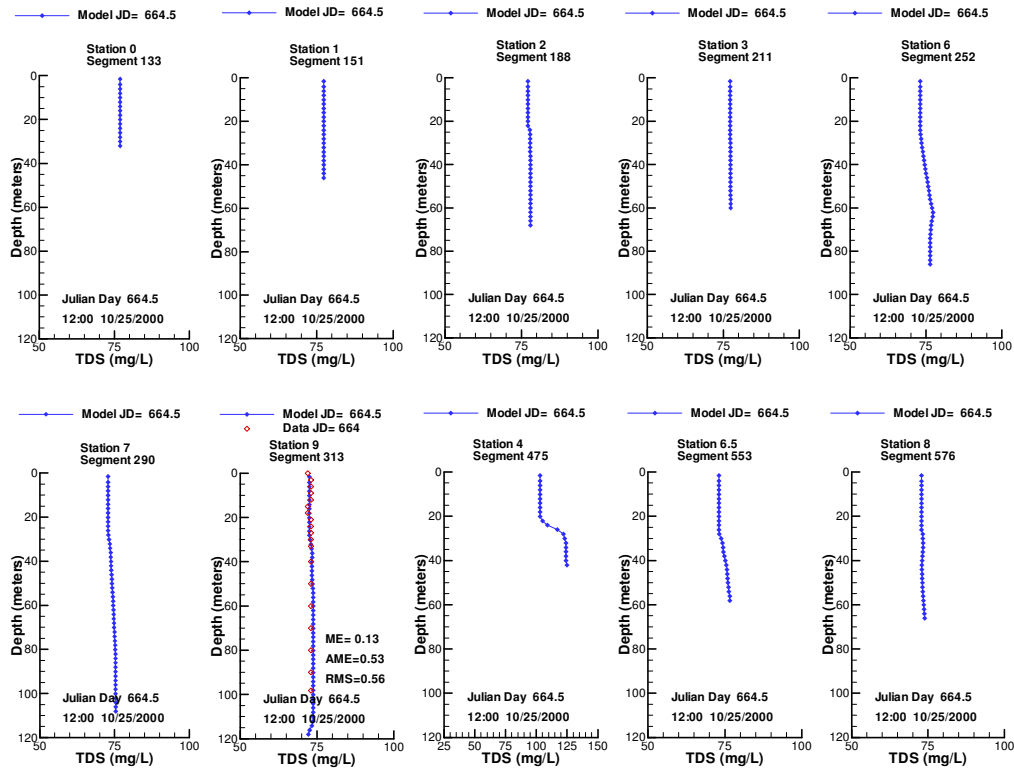


Figure 234. Vertical total dissolved solids model-data comparison, J664.

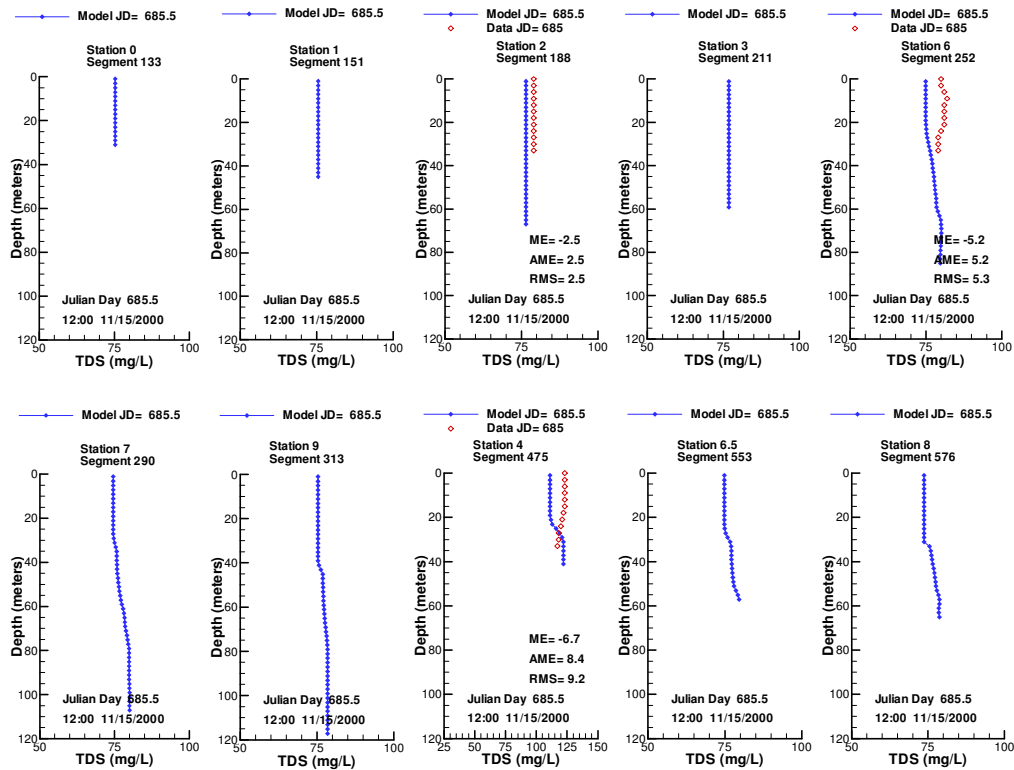


Figure 235. Vertical total dissolved solids model-data comparison, J685.

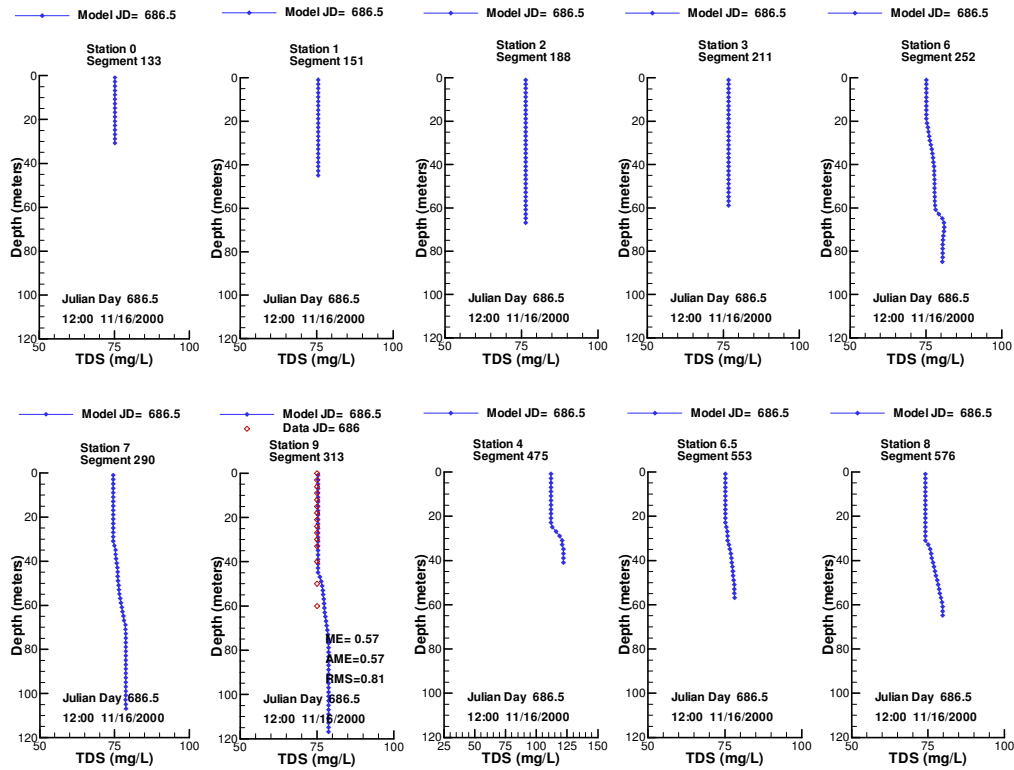


Figure 236. Vertical total dissolved solids model-data comparison, J686.

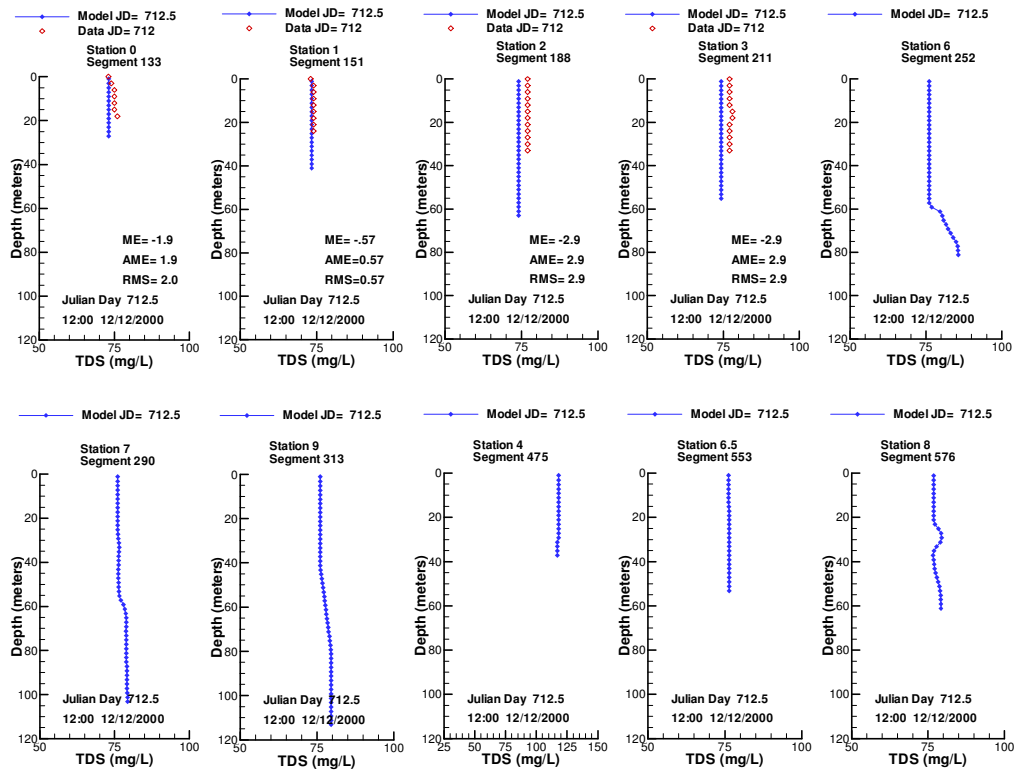


Figure 237. Vertical total dissolved solids model-data comparison, J712.

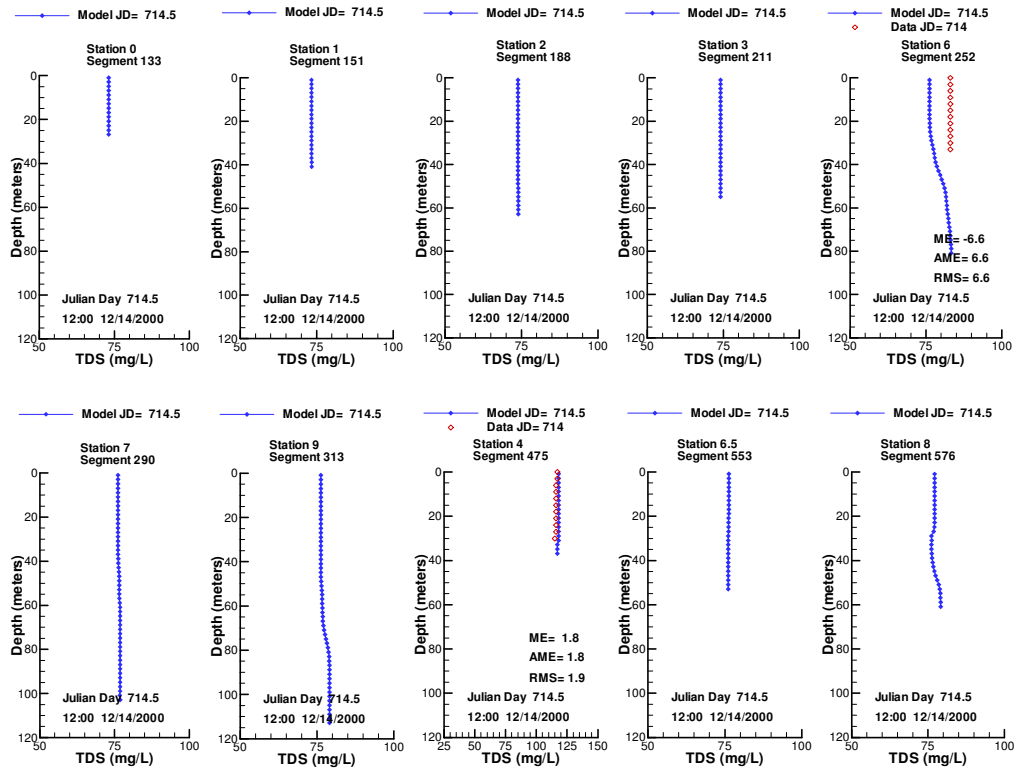


Figure 238. Vertical total dissolved solids model-data comparison, J714.

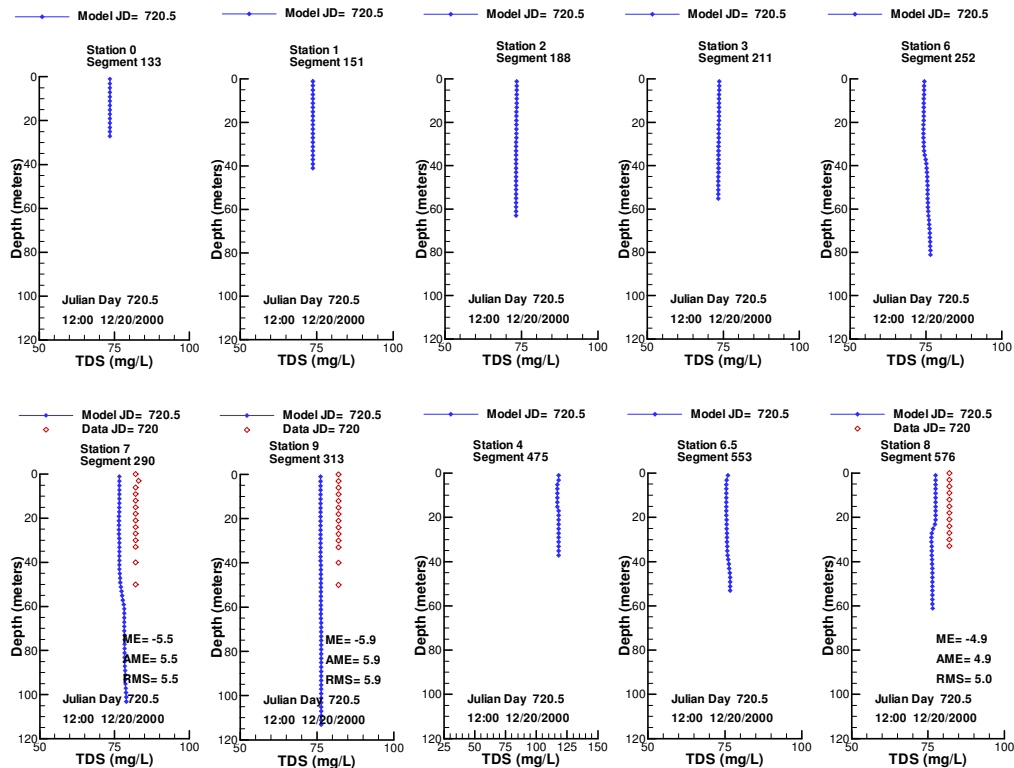


Figure 239. Vertical total dissolved solids model-data comparison, J720.

Dissolved oxygen

For days when data were recorded, vertical profile plots of dissolved oxygen are shown in Figure 240 through Figure 286.

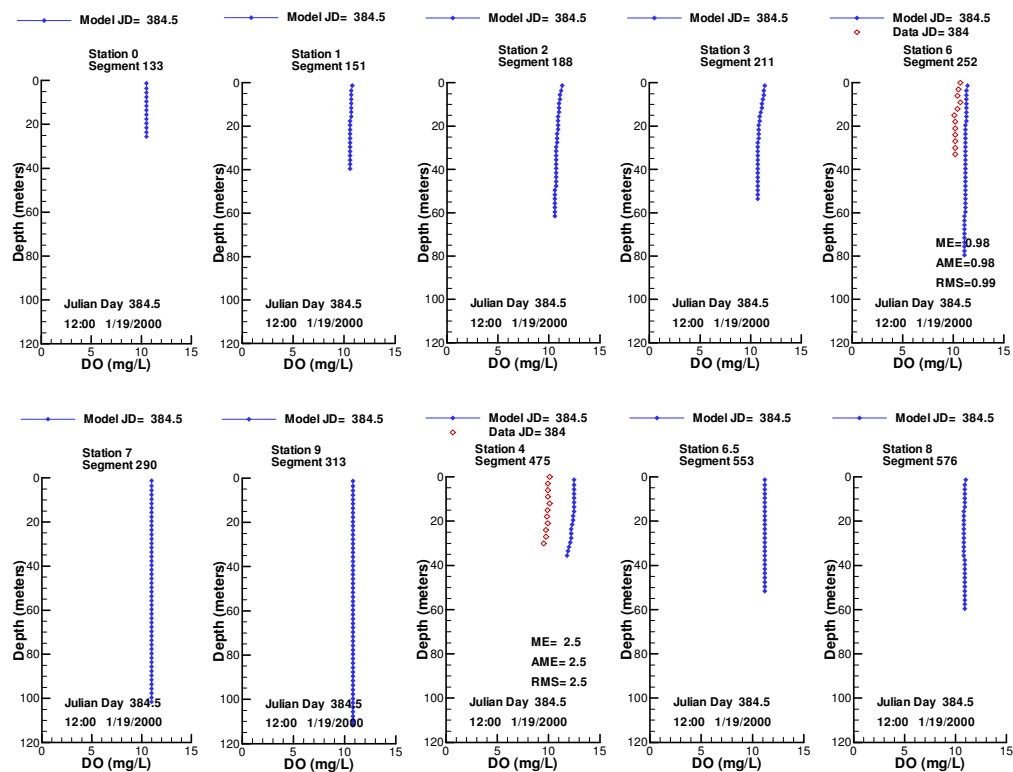


Figure 240. Vertical dissolved oxygen profile model-data comparison, J384.

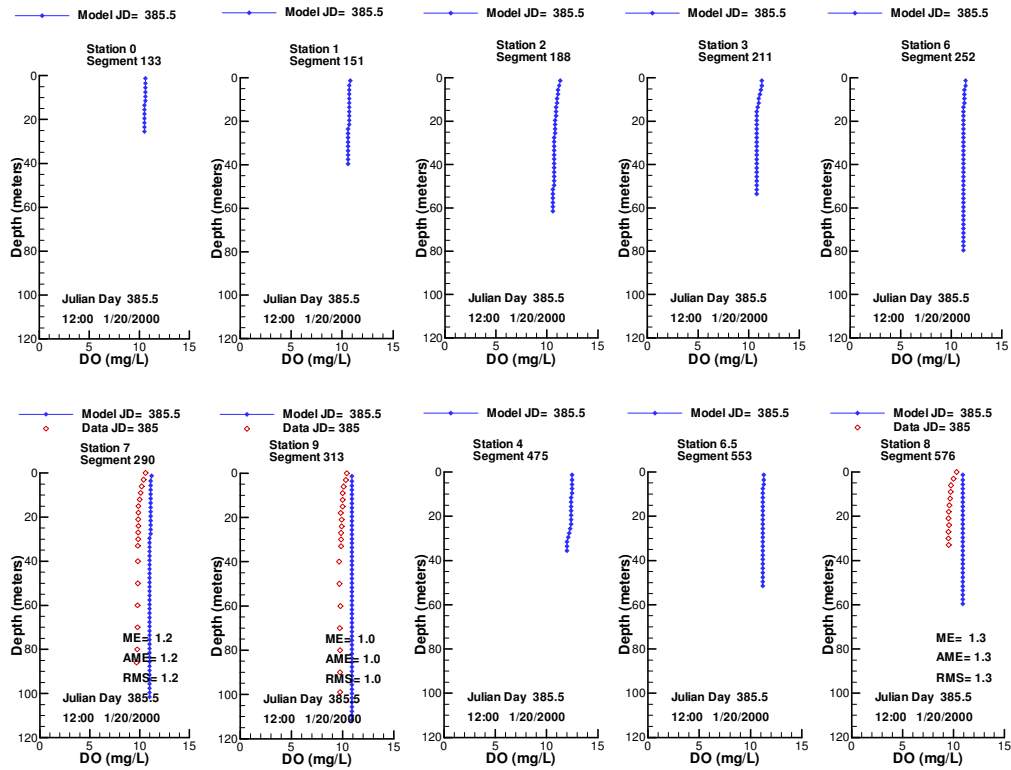


Figure 241. Vertical dissolved oxygen profile model-data comparison, J385.

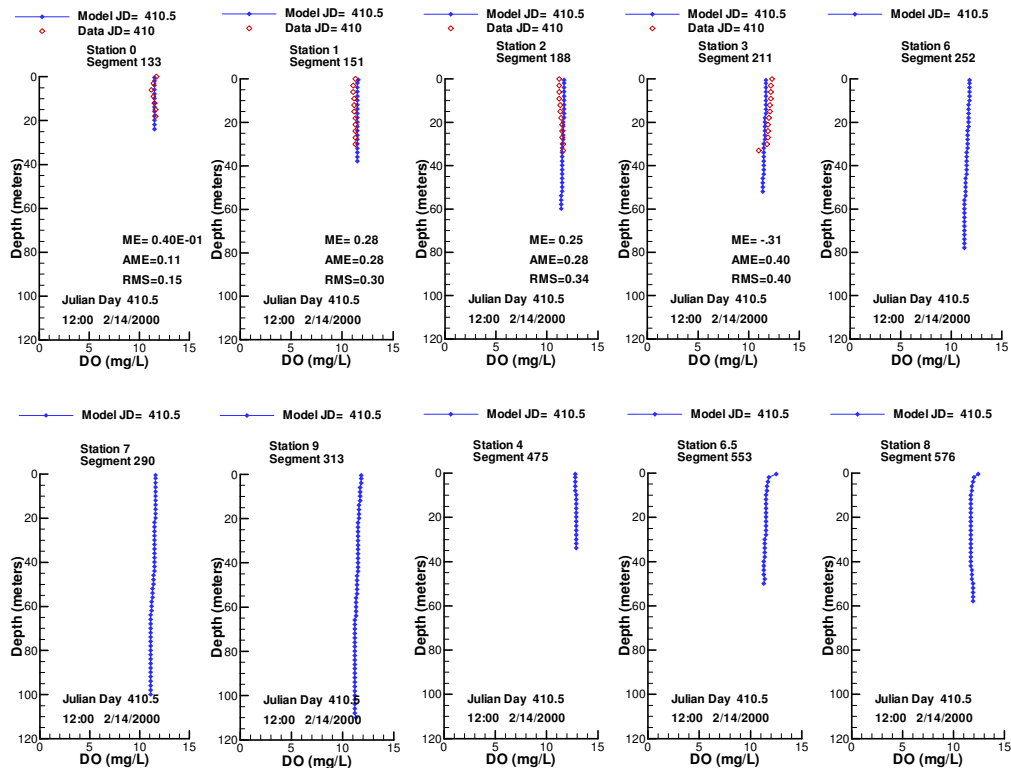


Figure 242. Vertical dissolved oxygen profile model-data comparison, J410.

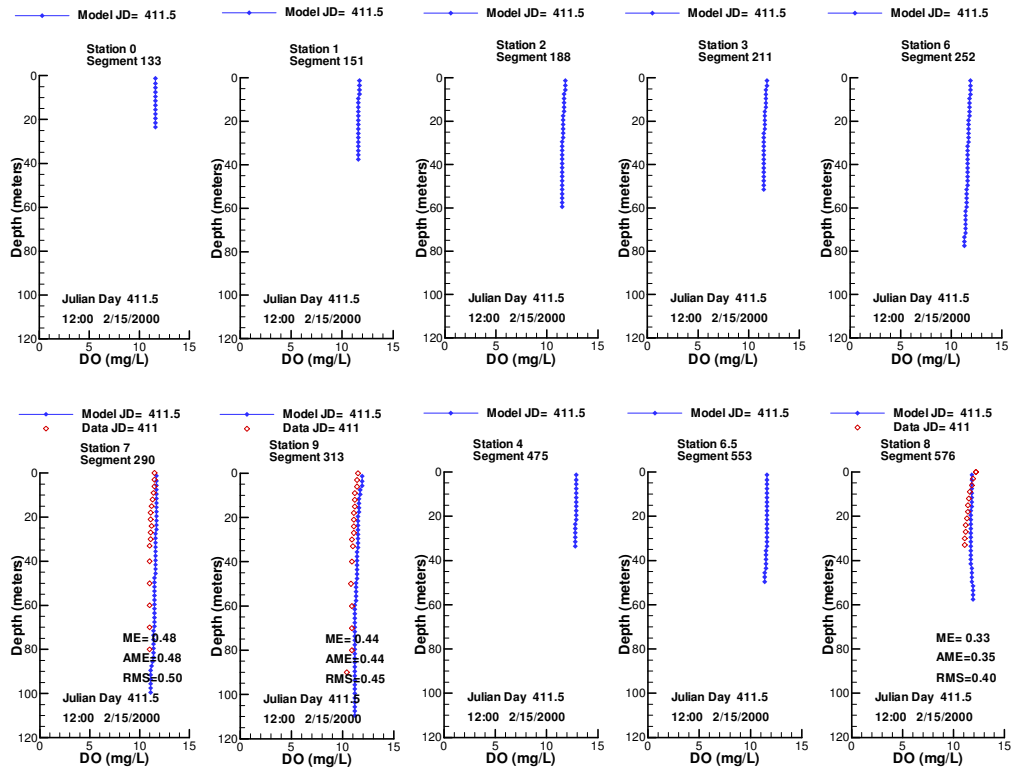


Figure 243. Vertical dissolved oxygen profile model-data comparison, J411.

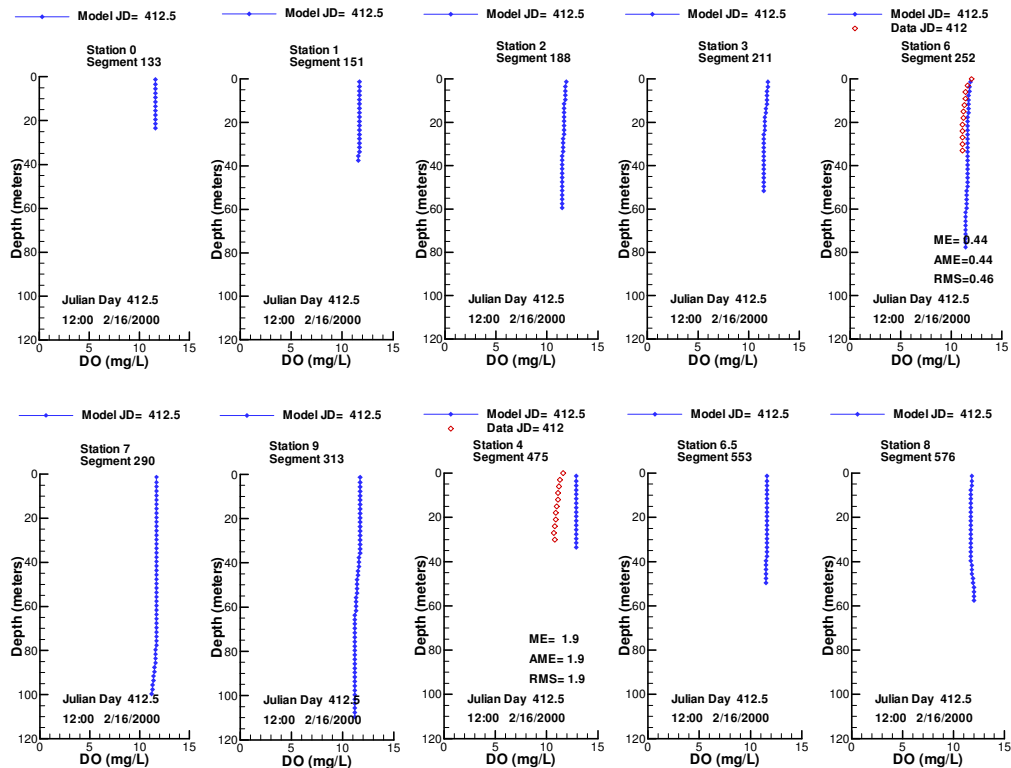


Figure 244. Vertical dissolved oxygen profile model-data comparison, J412.

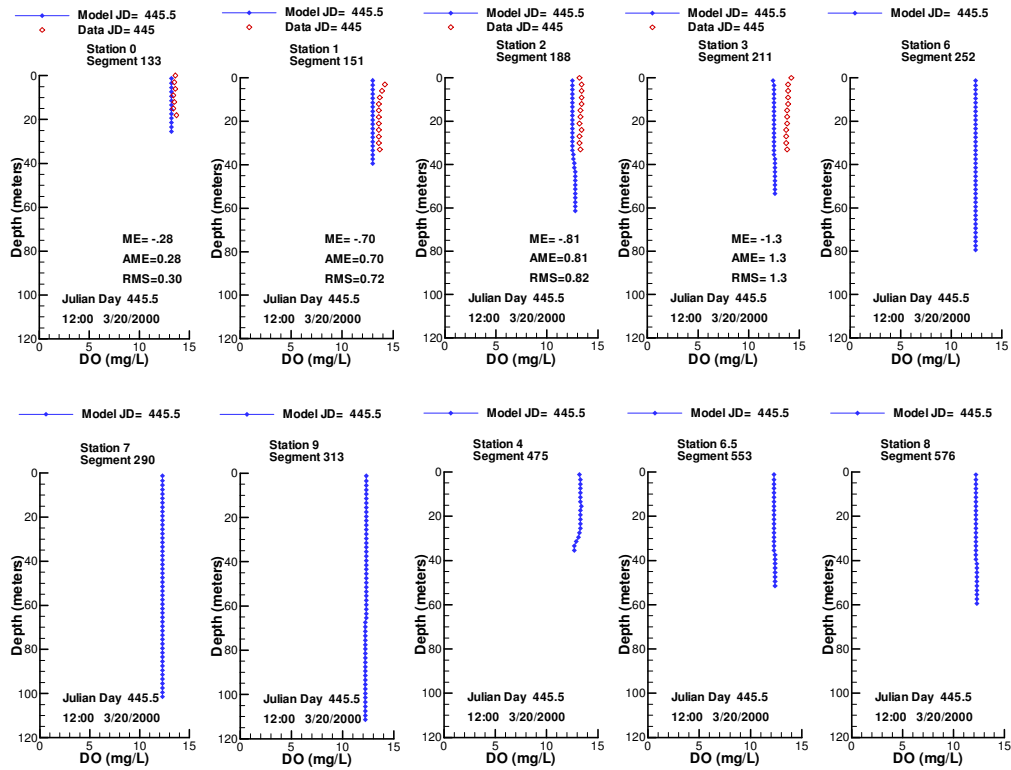


Figure 245. Vertical dissolved oxygen profile model-data comparison, J445.

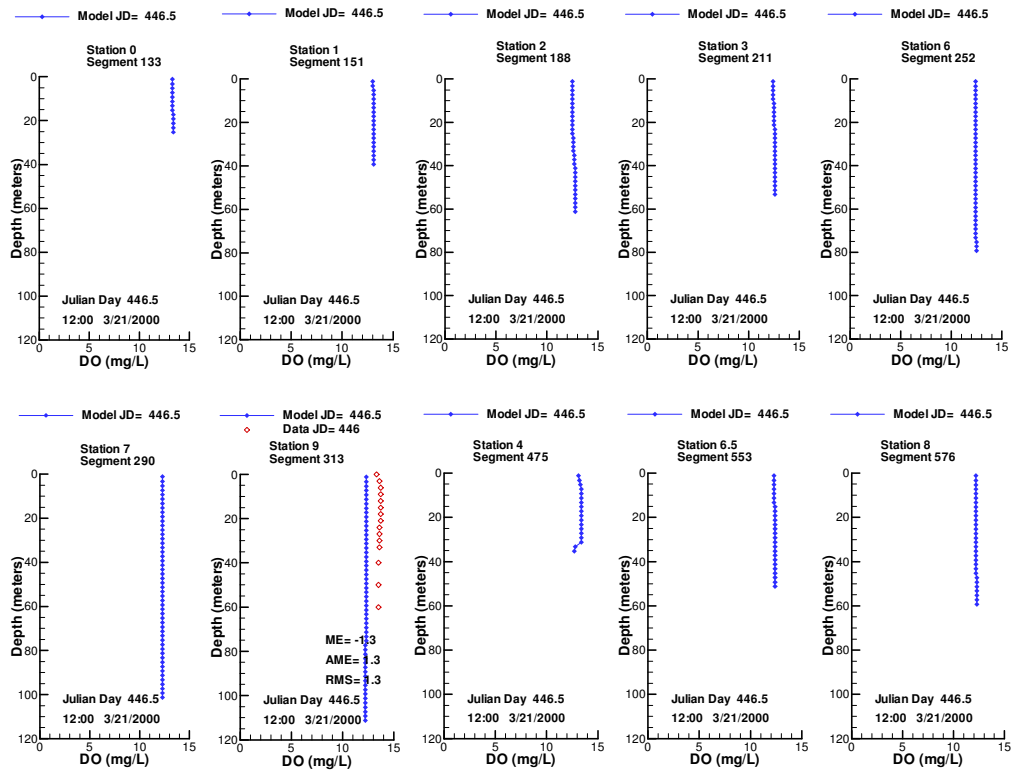


Figure 246. Vertical dissolved oxygen profile model-data comparison, J446.

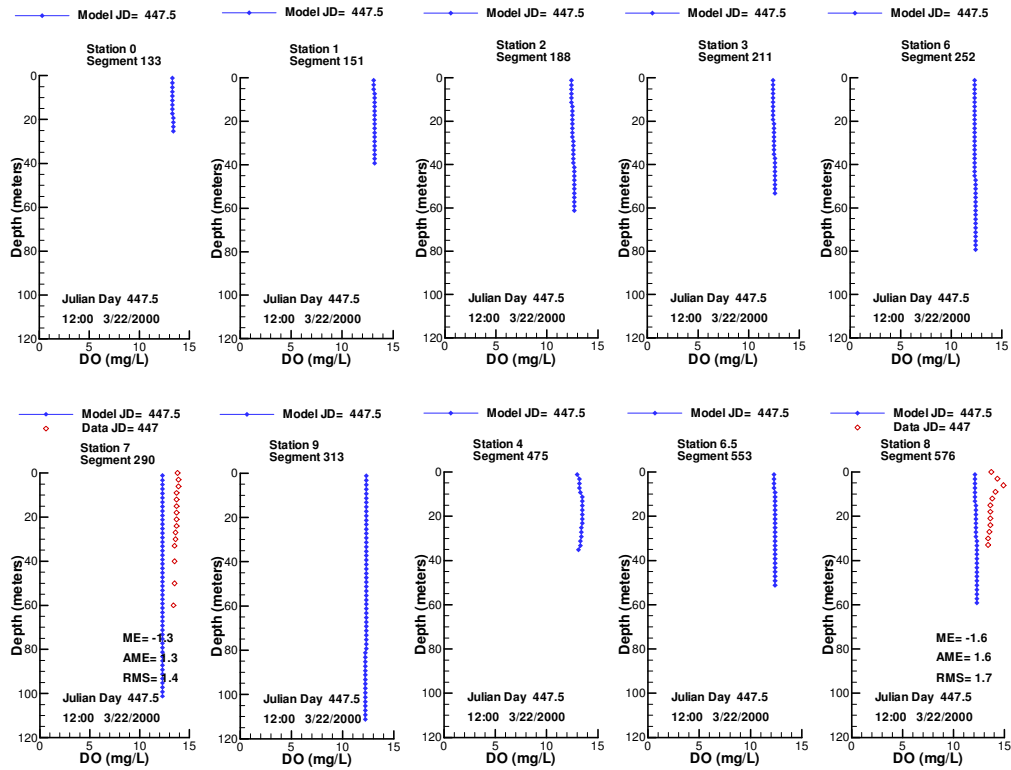


Figure 247. Vertical dissolved oxygen profile model-data comparison, J447.

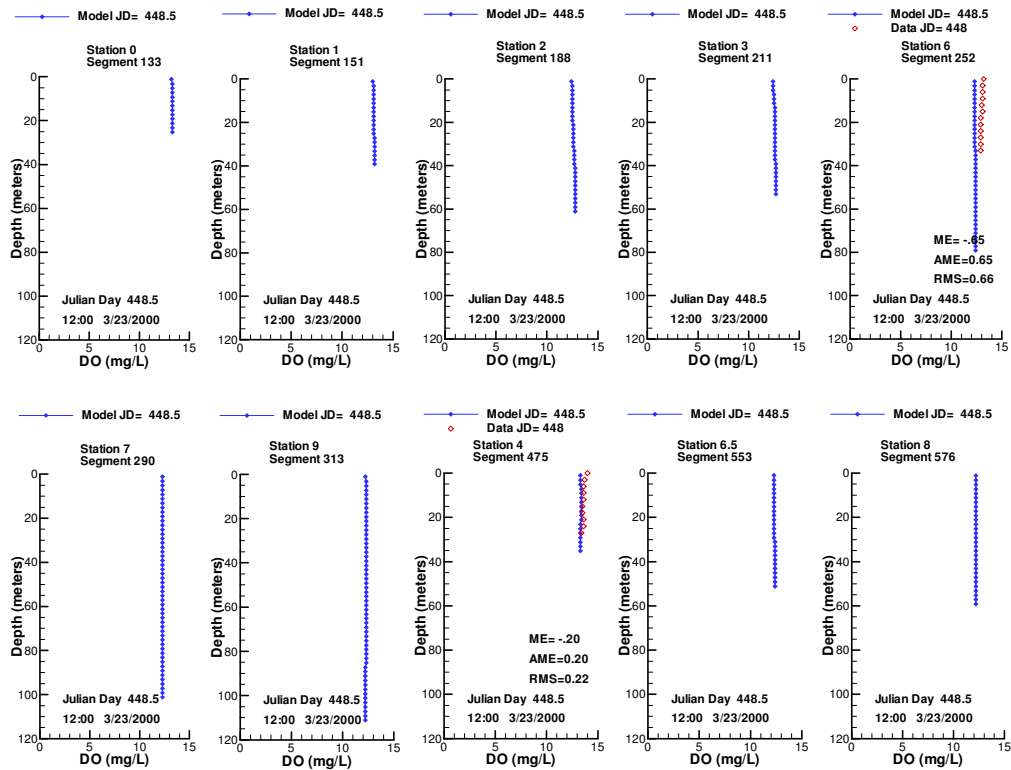


Figure 248. Vertical dissolved oxygen profile model-data comparison, J448.

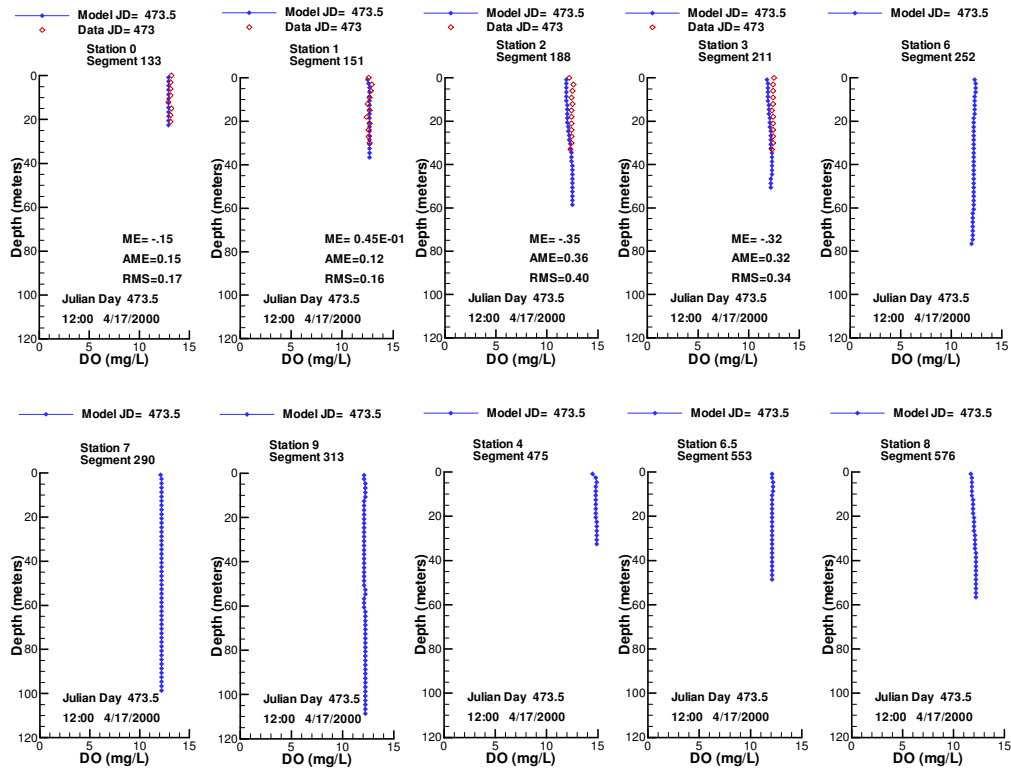


Figure 249. Vertical dissolved oxygen profile model-data comparison, J473.

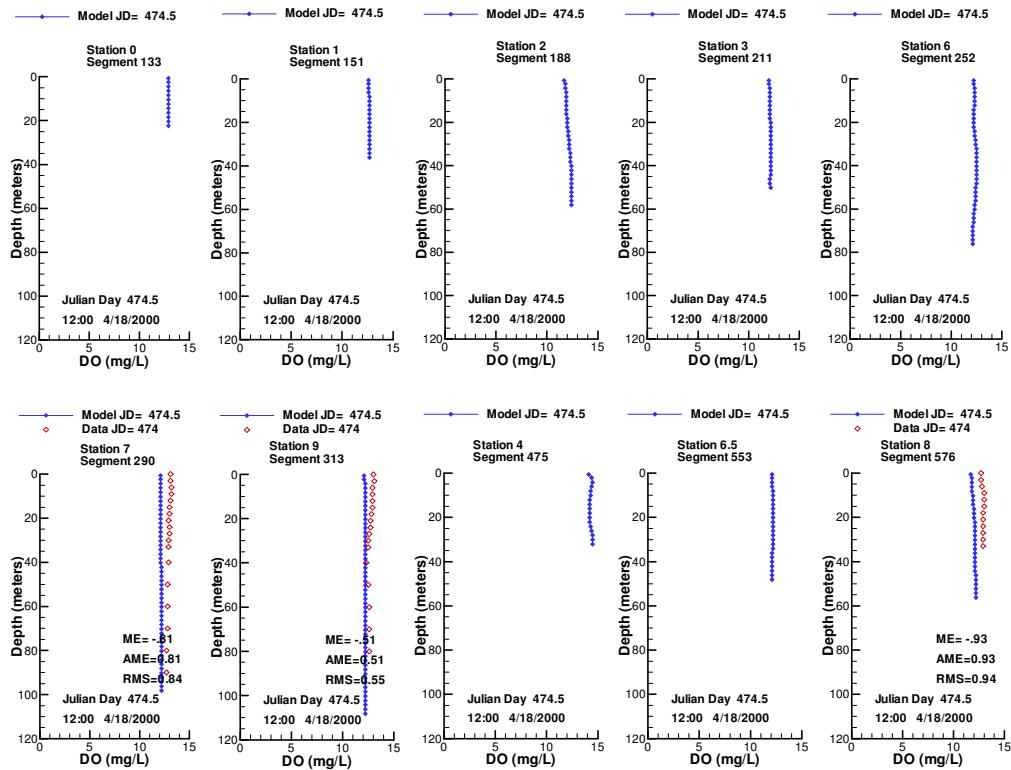


Figure 250. Vertical dissolved oxygen profile model-data comparison, J474.

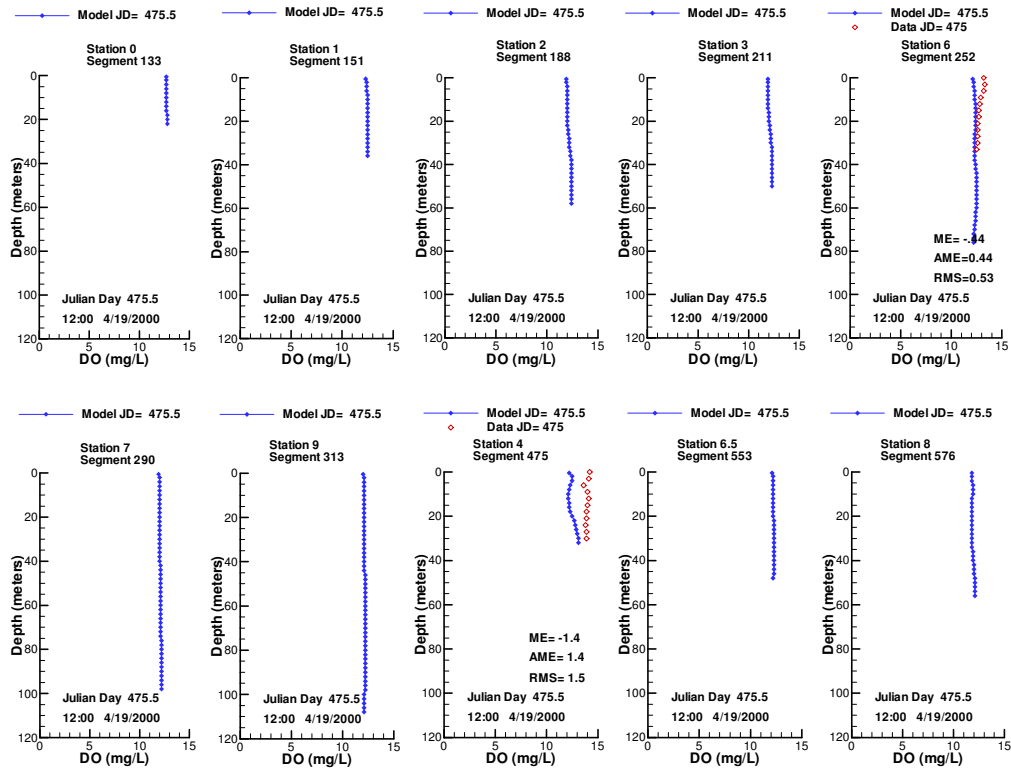


Figure 251. Vertical dissolved oxygen profile model-data comparison, J475.

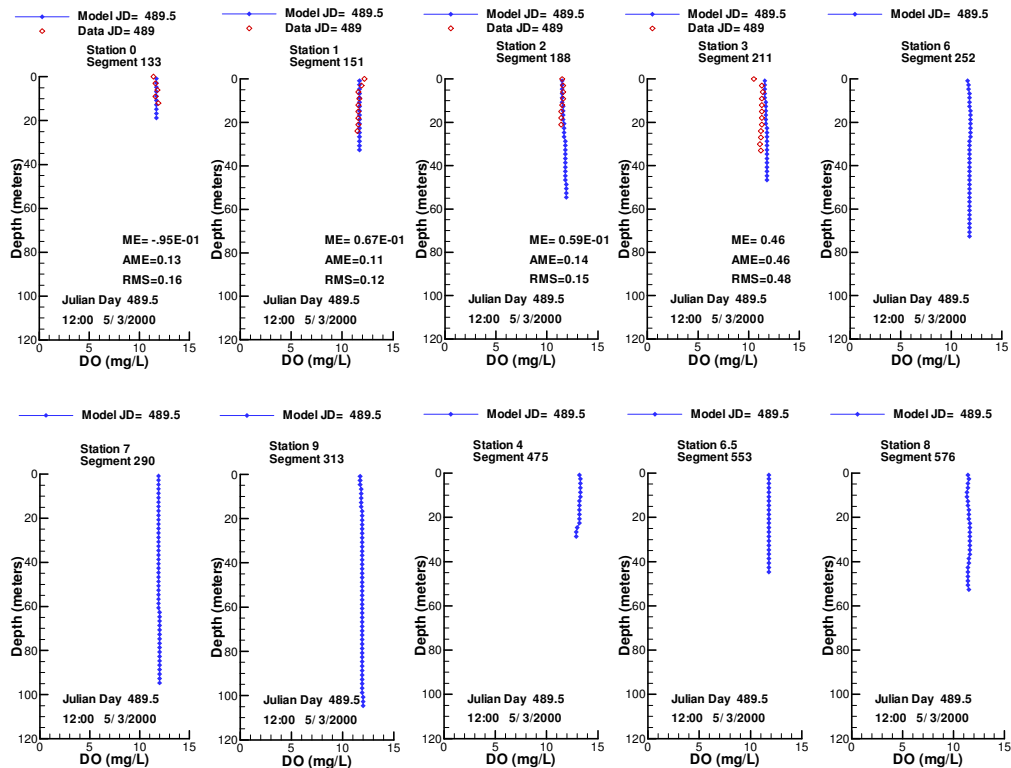


Figure 252. Vertical dissolved oxygen profile model-data comparison, J489.

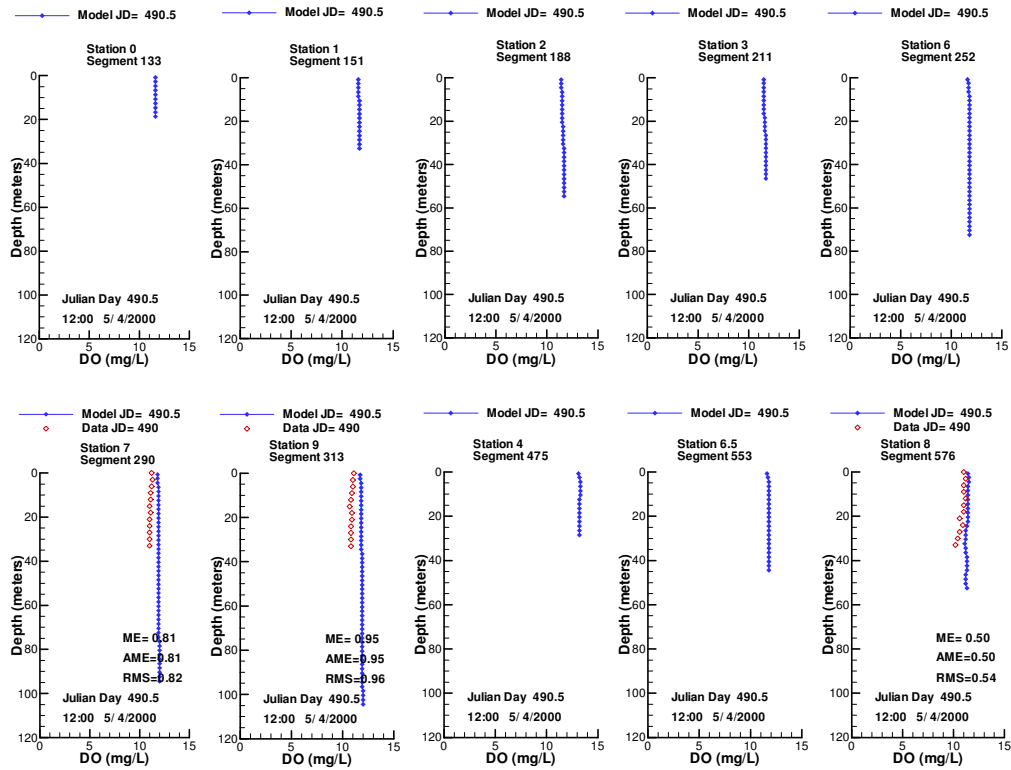


Figure 253. Vertical dissolved oxygen profile model-data comparison, J490.

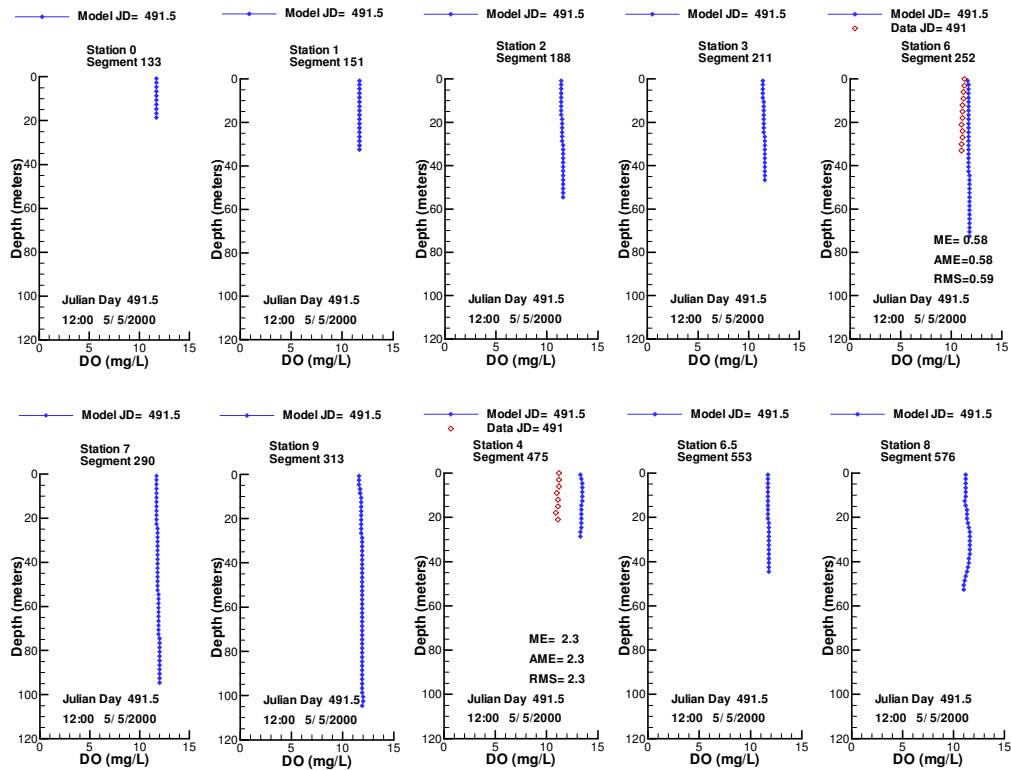


Figure 254. Vertical dissolved oxygen profile model-data comparison, J491.

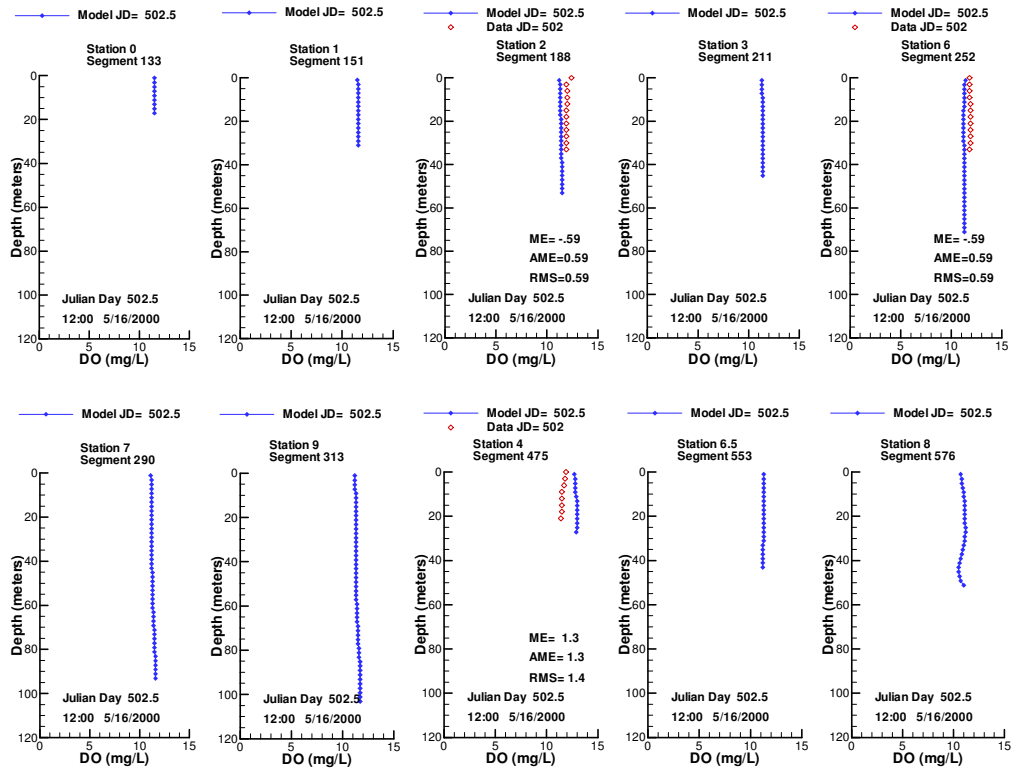


Figure 255. Vertical dissolved oxygen profile model-data comparison, J502.

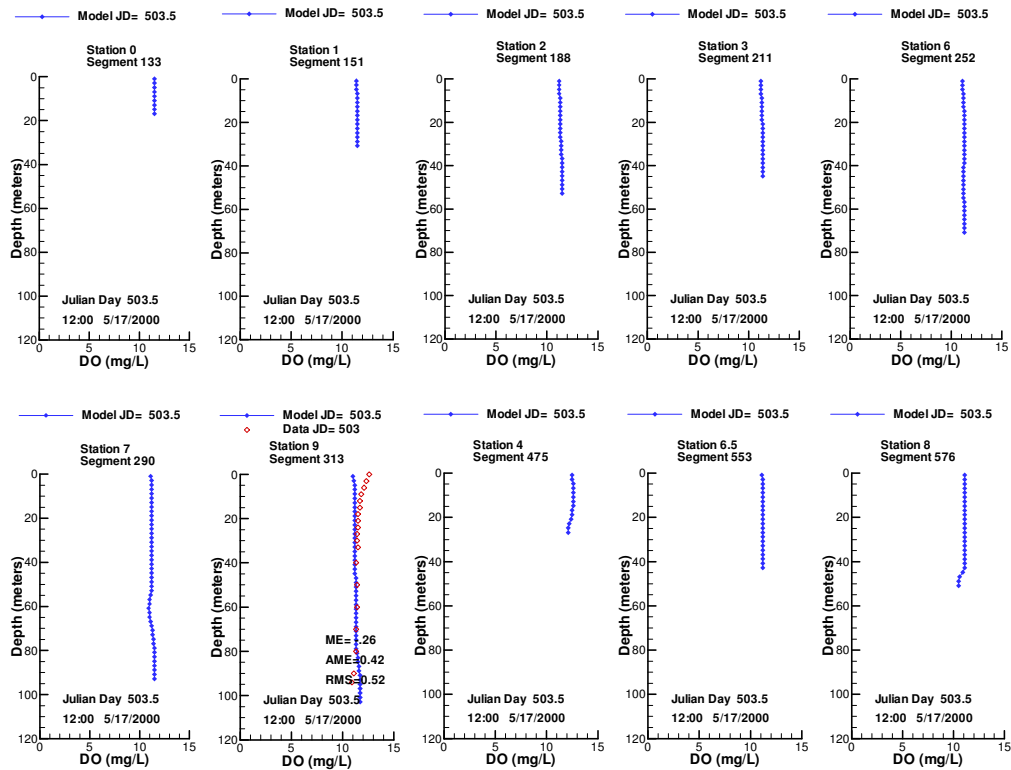


Figure 256. Vertical dissolved oxygen profile model-data comparison, J503.

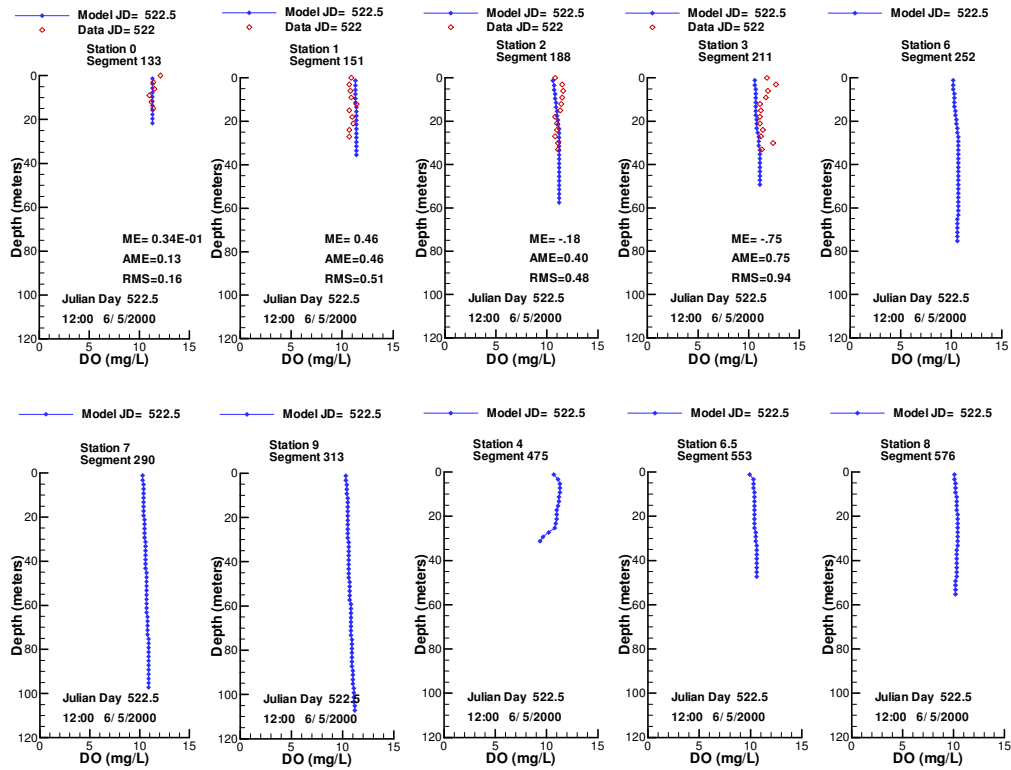


Figure 257. Vertical dissolved oxygen profile model-data comparison, J522.

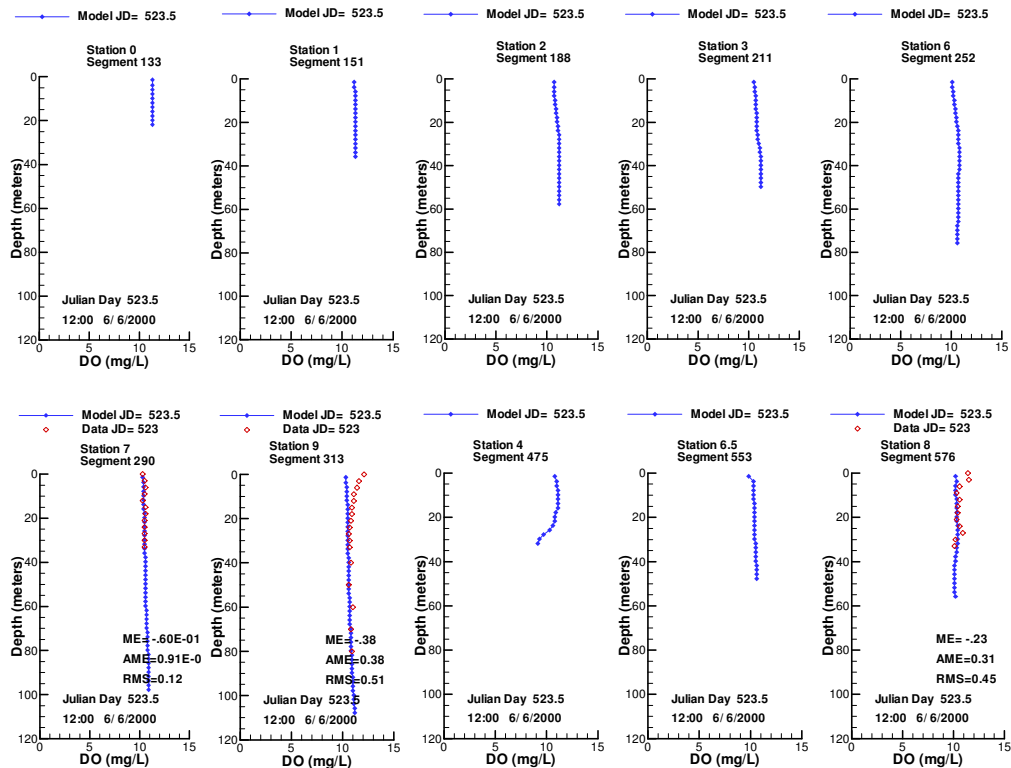


Figure 258. Vertical dissolved oxygen profile model-data comparison, J523.

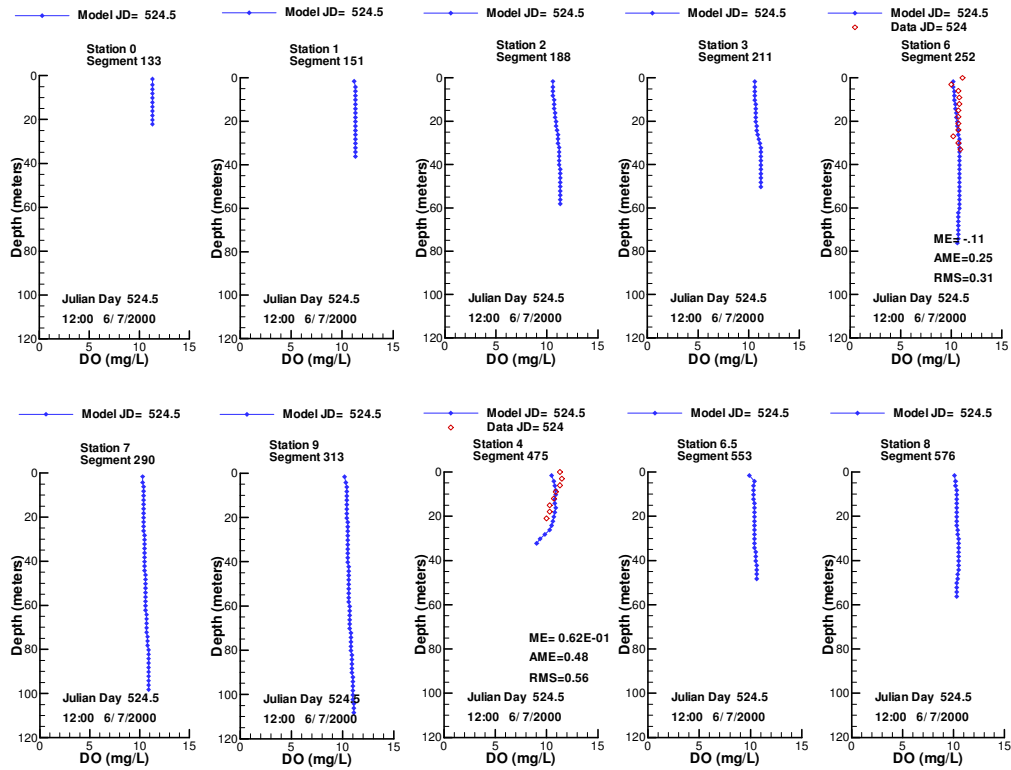


Figure 259. Vertical dissolved oxygen profile model-data comparison, J524.

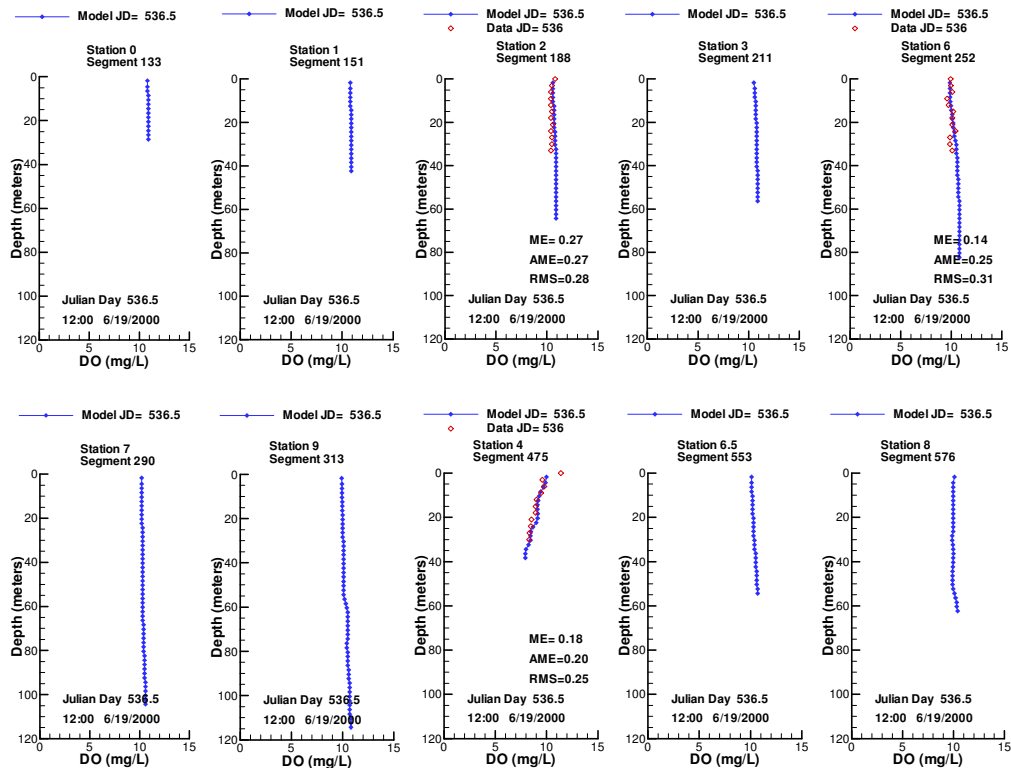


Figure 260. Vertical dissolved oxygen profile model-data comparison, J536.

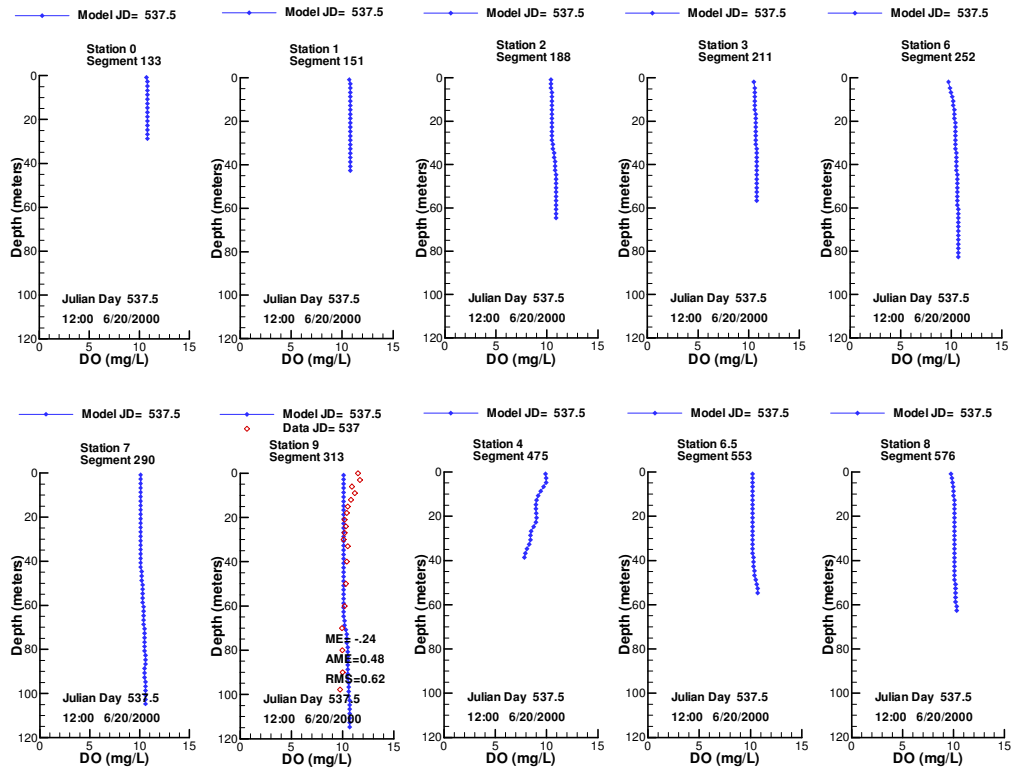


Figure 261. Vertical dissolved oxygen profile model-data comparison, J537.

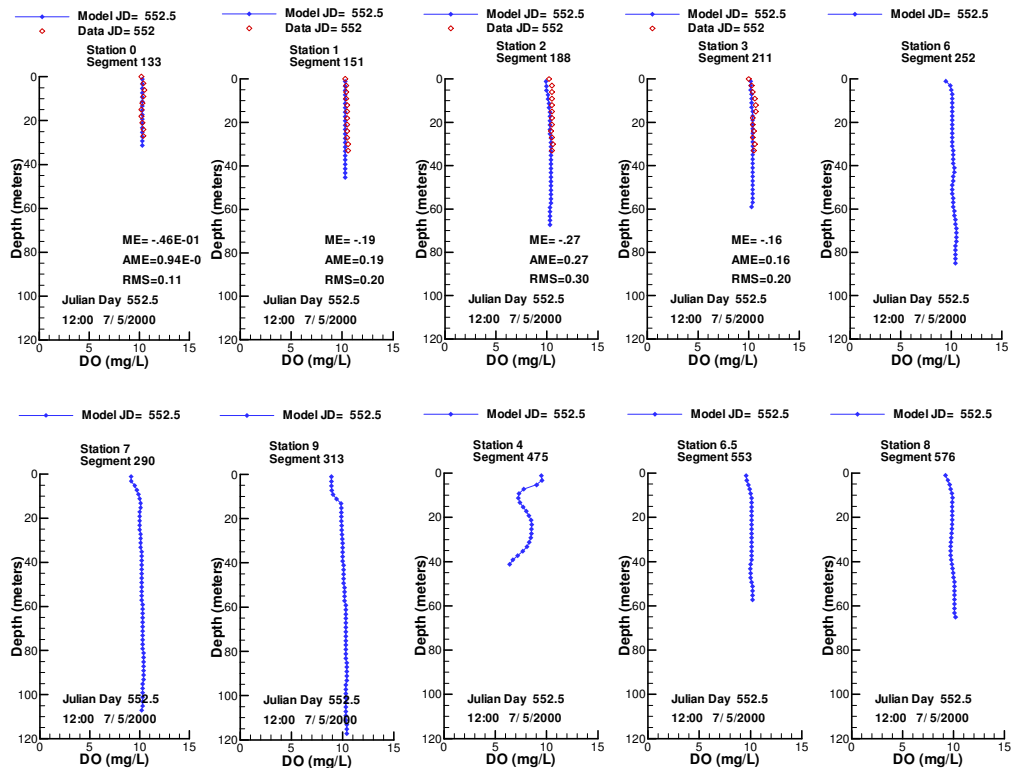


Figure 262. Vertical dissolved oxygen profile model-data comparison, J552.

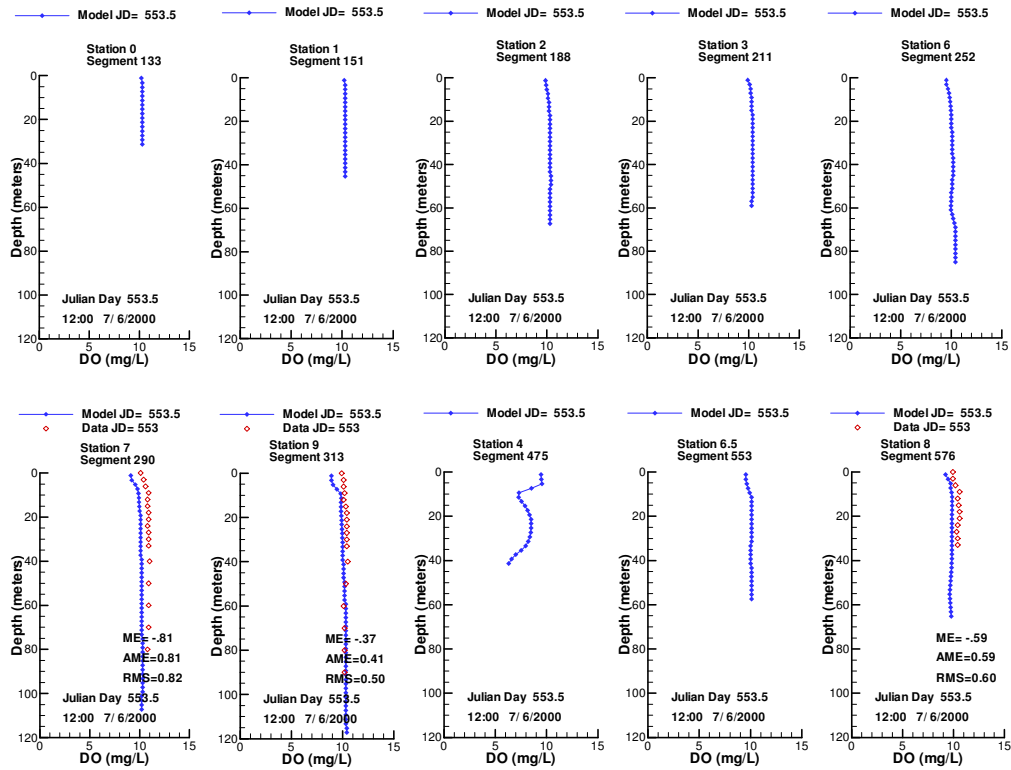


Figure 263. Vertical dissolved oxygen profile model-data comparison, J553.

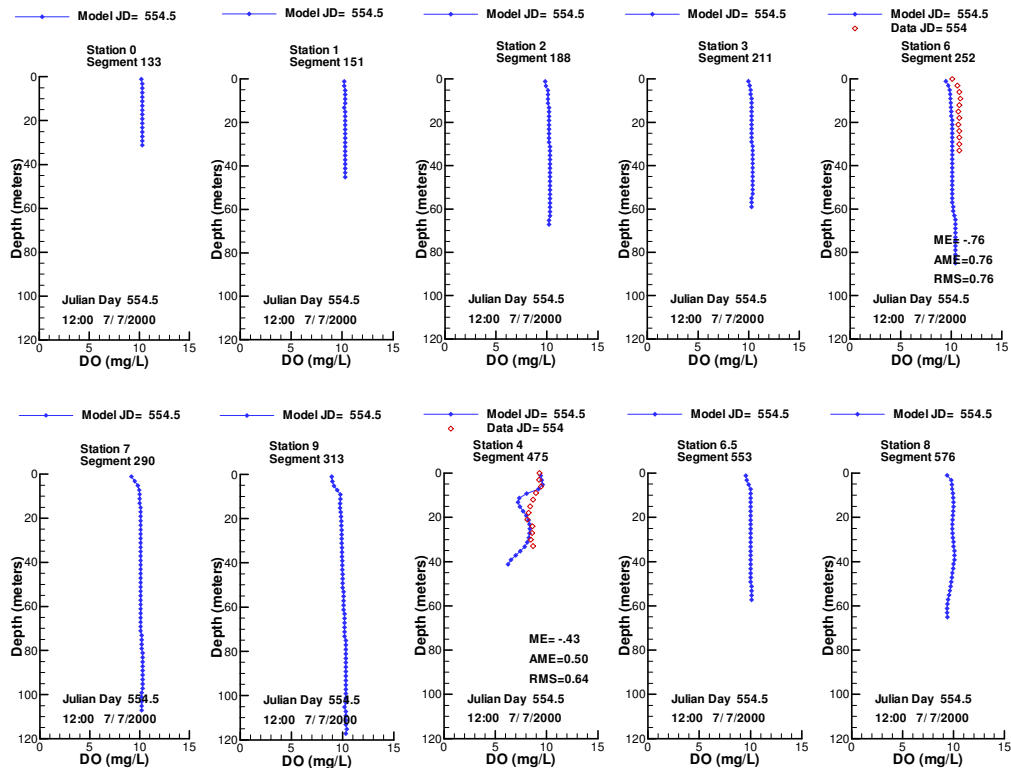


Figure 264. Vertical dissolved oxygen profile model-data comparison, J554.

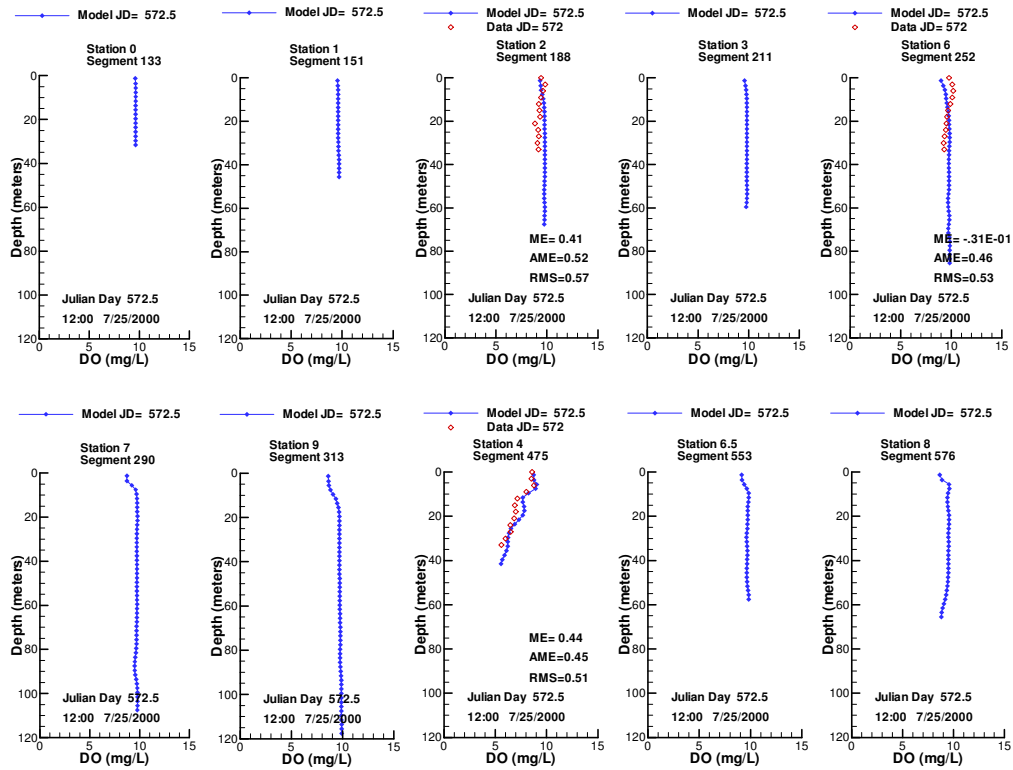


Figure 265. Vertical dissolved oxygen profile model-data comparison, J572.

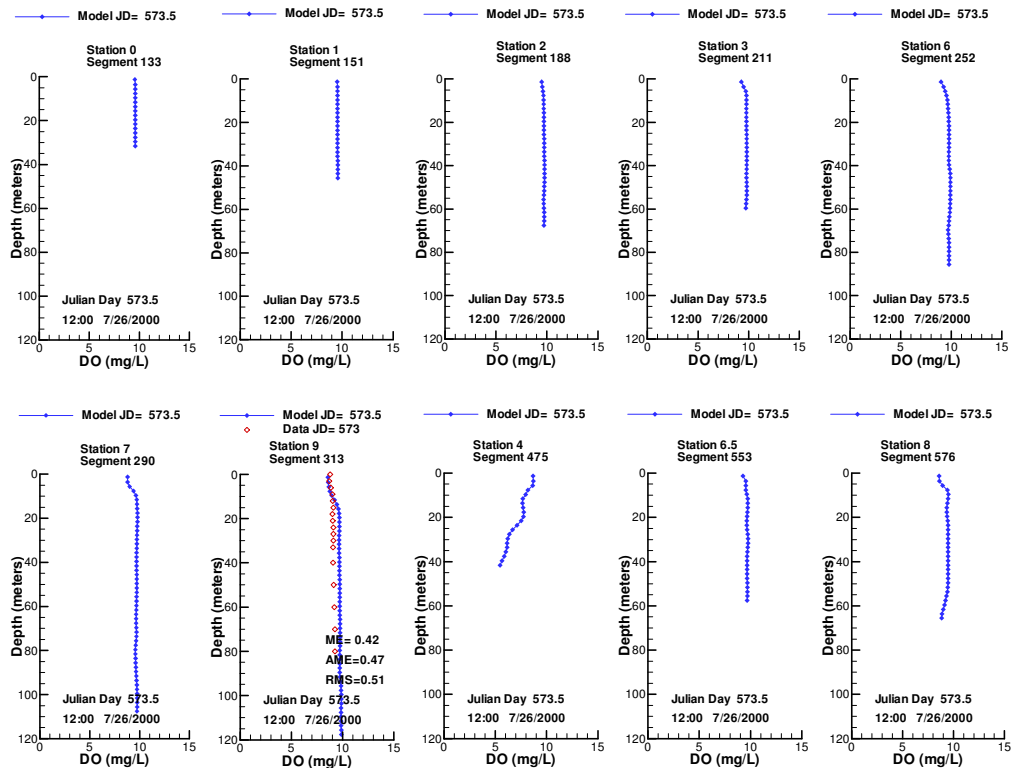


Figure 266. Vertical dissolved oxygen profile model-data comparison, J573.

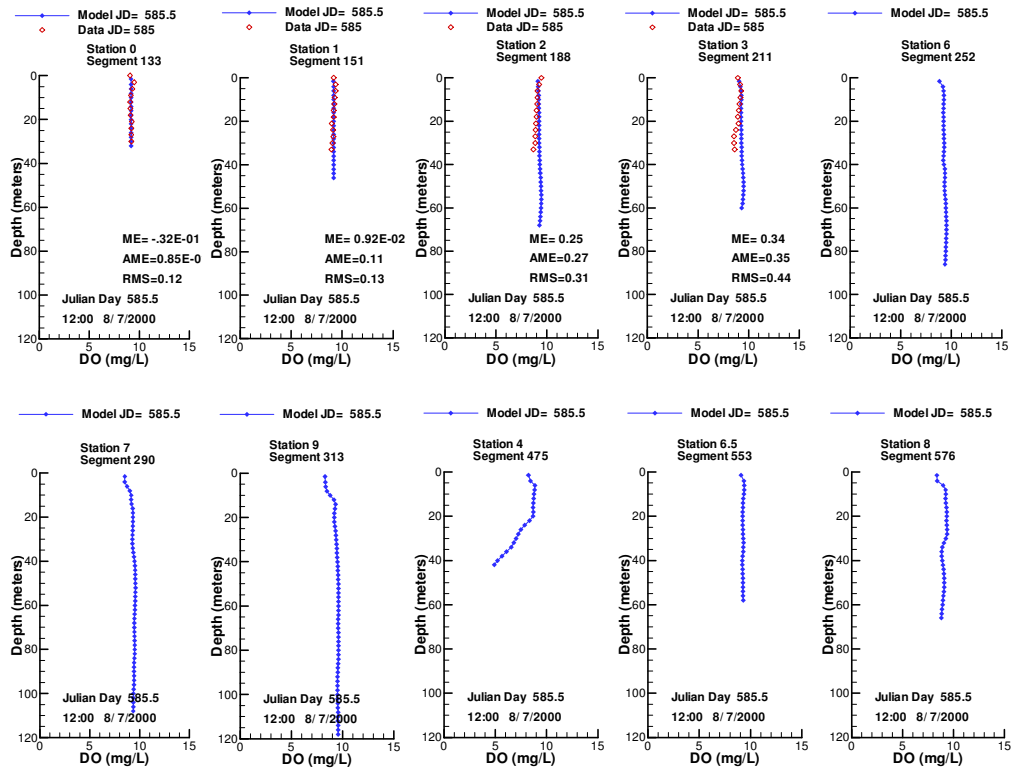


Figure 267. Vertical dissolved oxygen profile model-data comparison, J585.

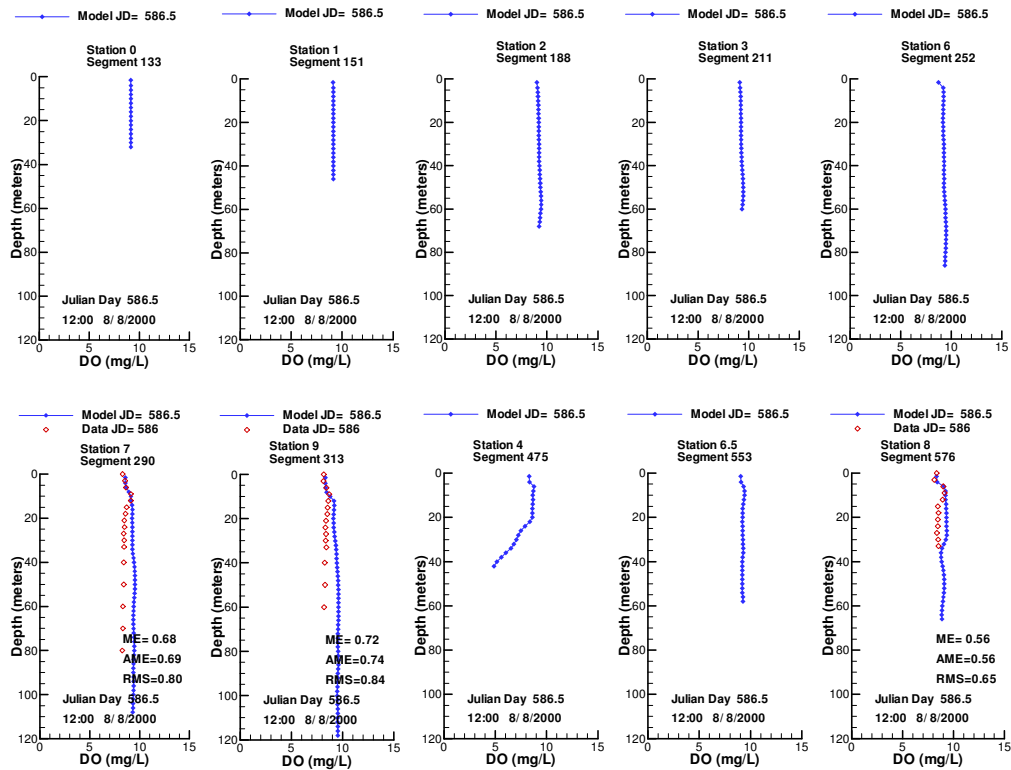


Figure 268. Vertical dissolved oxygen profile model-data comparison, J586.

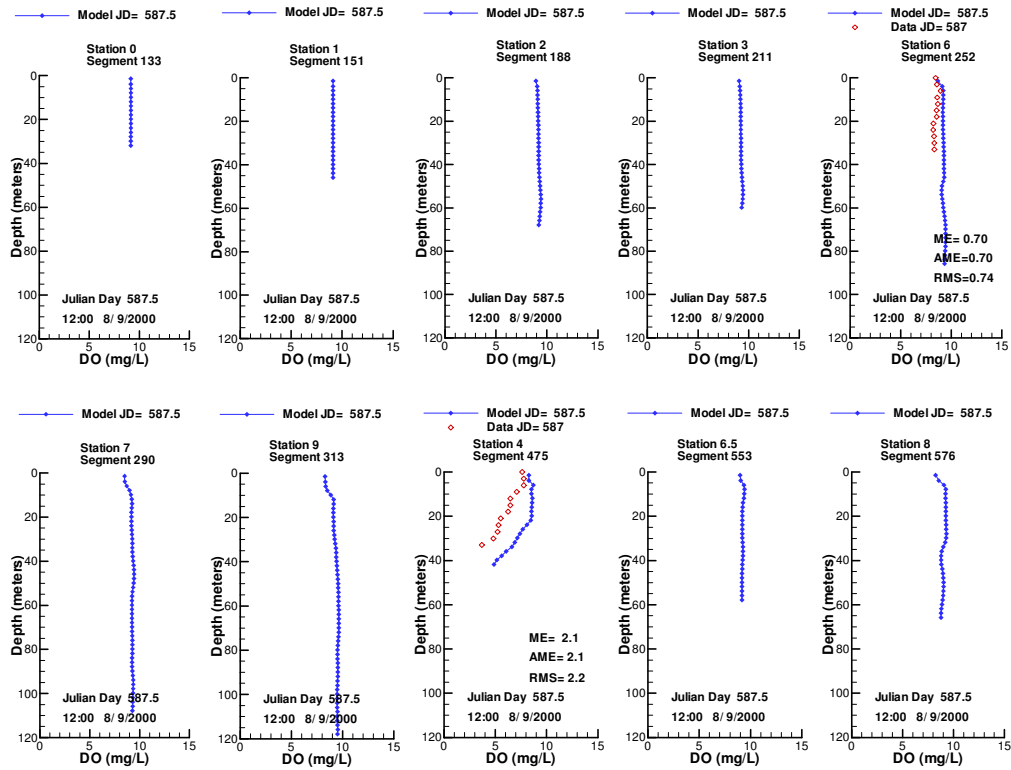


Figure 269. Vertical dissolved oxygen profile model-data comparison, J587.

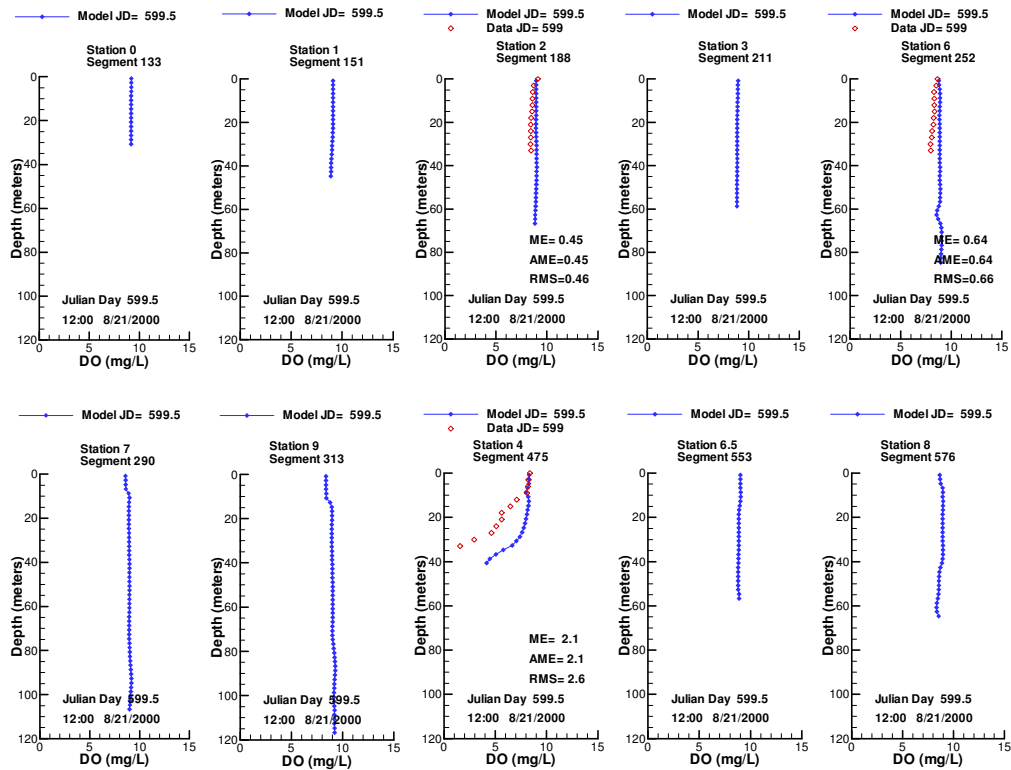


Figure 270. Vertical dissolved oxygen profile model-data comparison, J599.

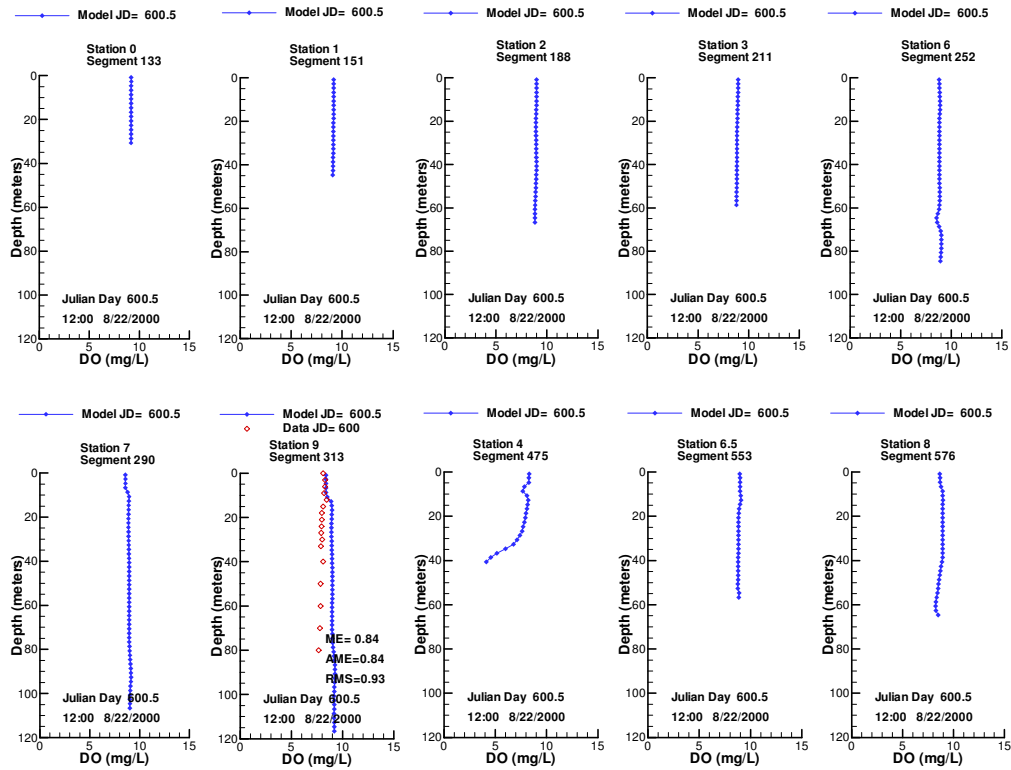


Figure 271. Vertical dissolved oxygen profile model-data comparison, J600.

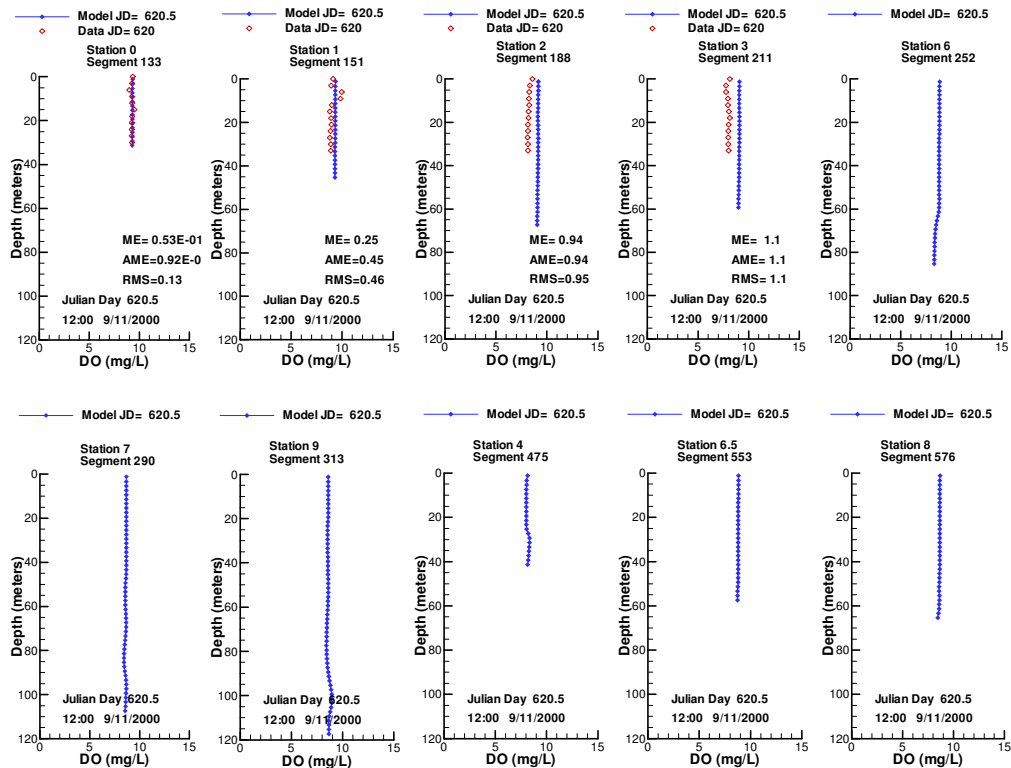


Figure 272. Vertical dissolved oxygen profile model-data comparison, J620.

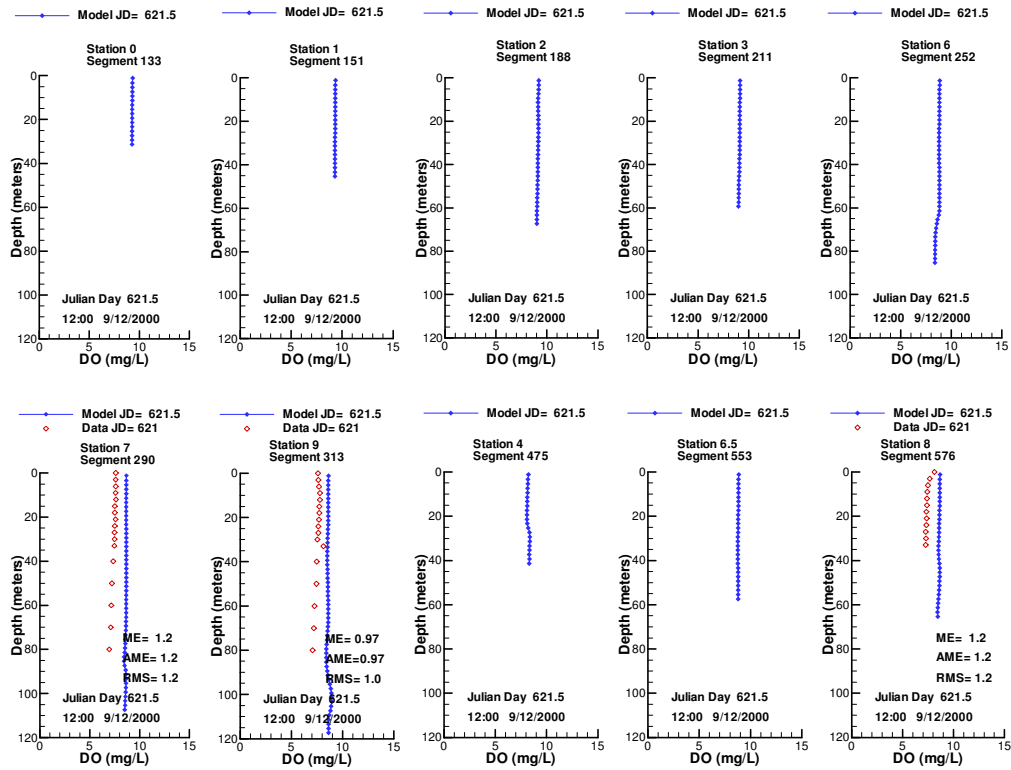


Figure 273. Vertical dissolved oxygen profile model-data comparison, J621.

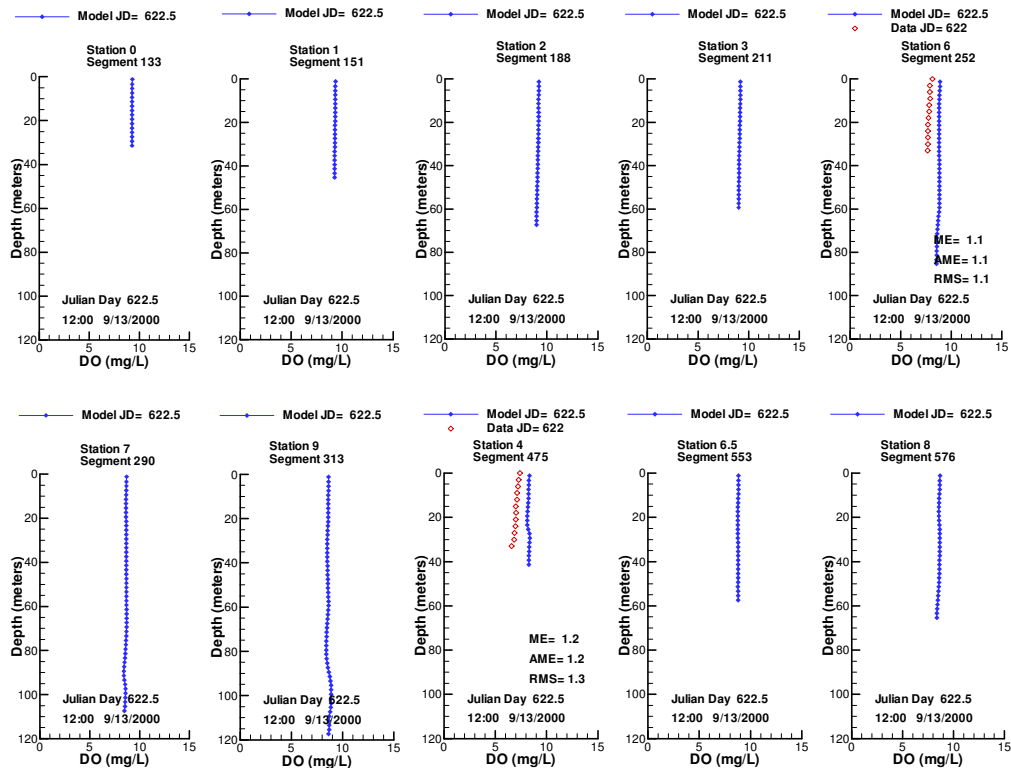


Figure 274. Vertical dissolved oxygen profile model-data comparison, J622.

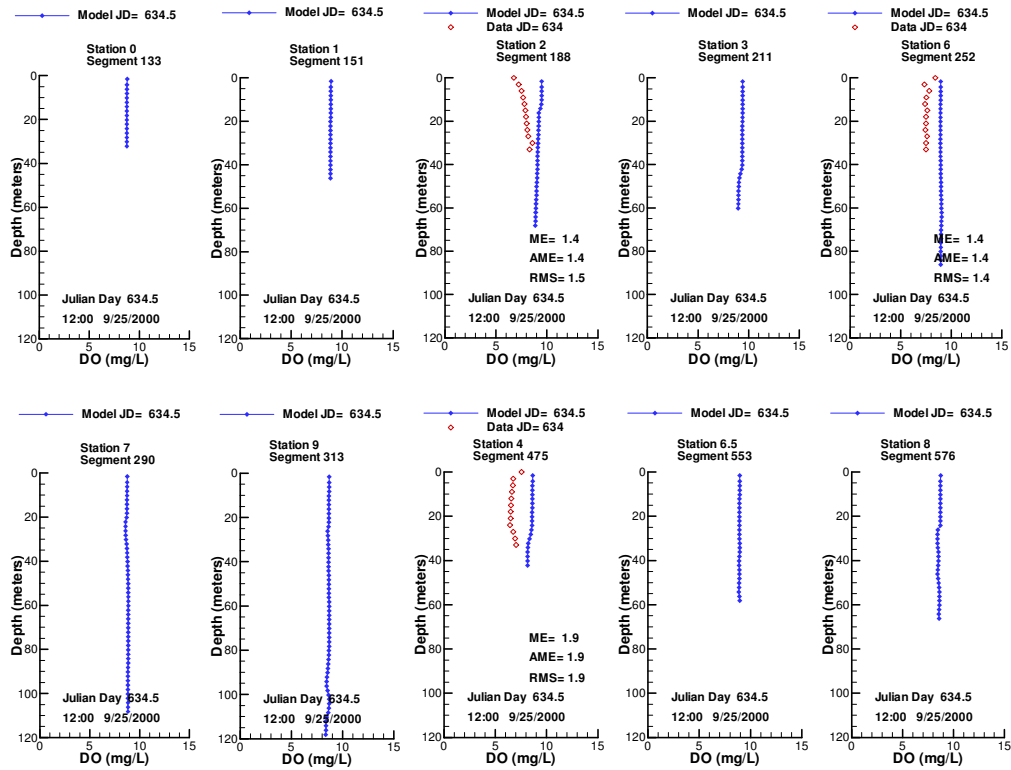


Figure 275. Vertical dissolved oxygen profile model-data comparison, J634.

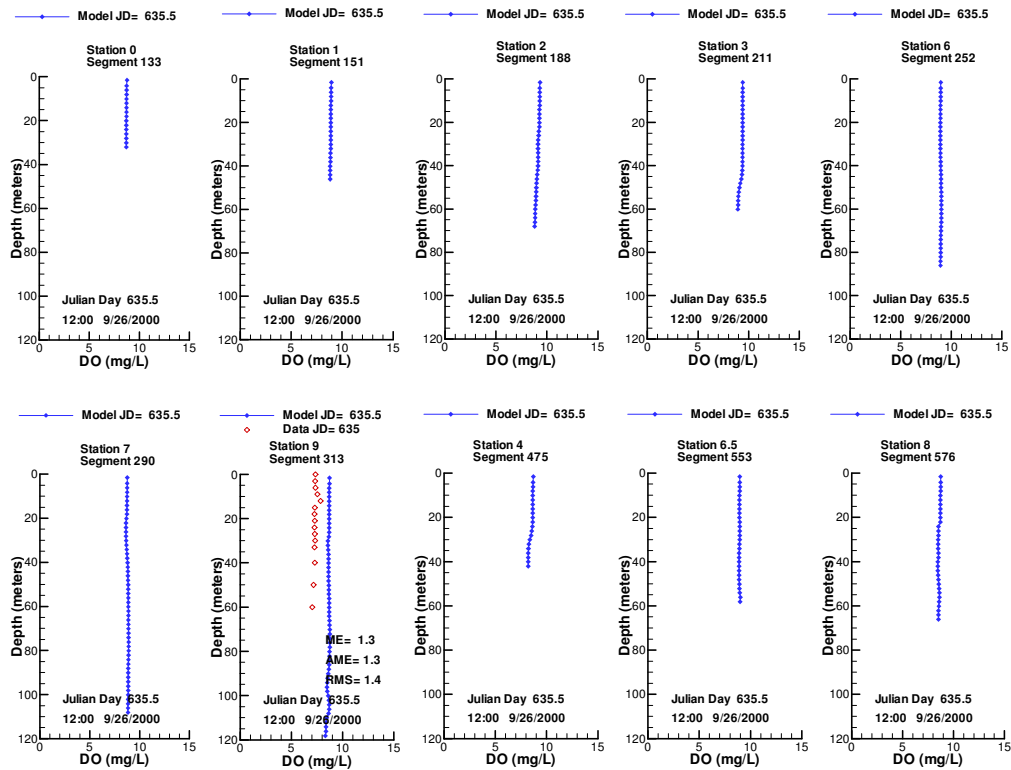


Figure 276. Vertical dissolved oxygen profile model-data comparison, J635.

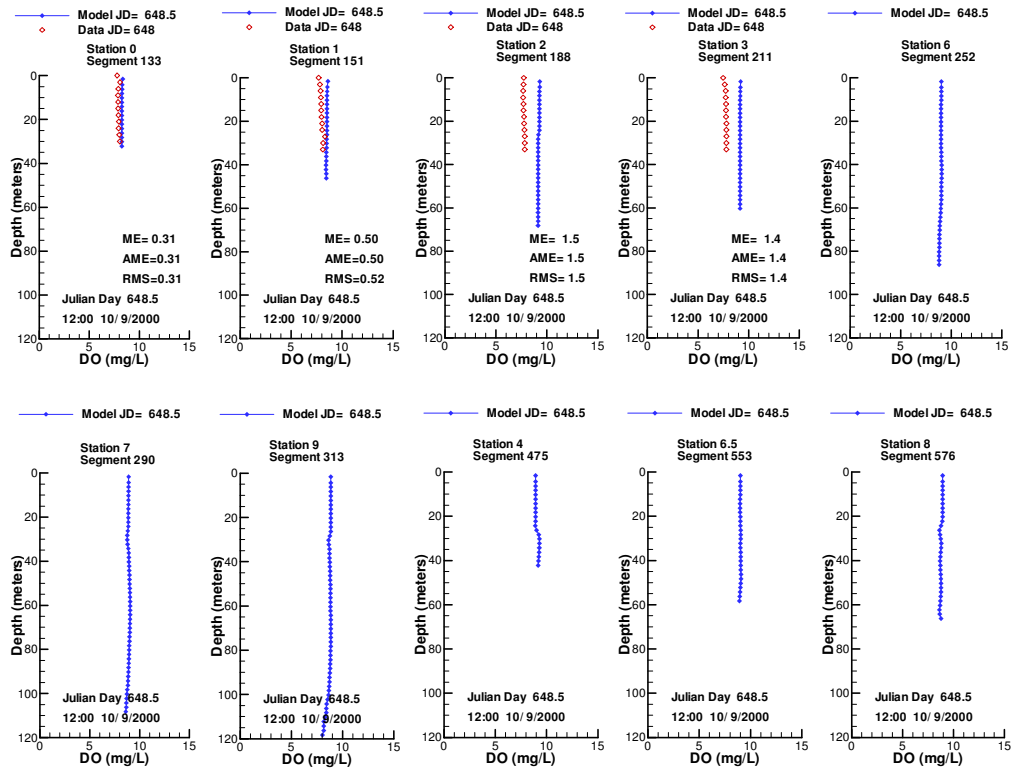


Figure 277. Vertical dissolved oxygen profile model-data comparison, J648.

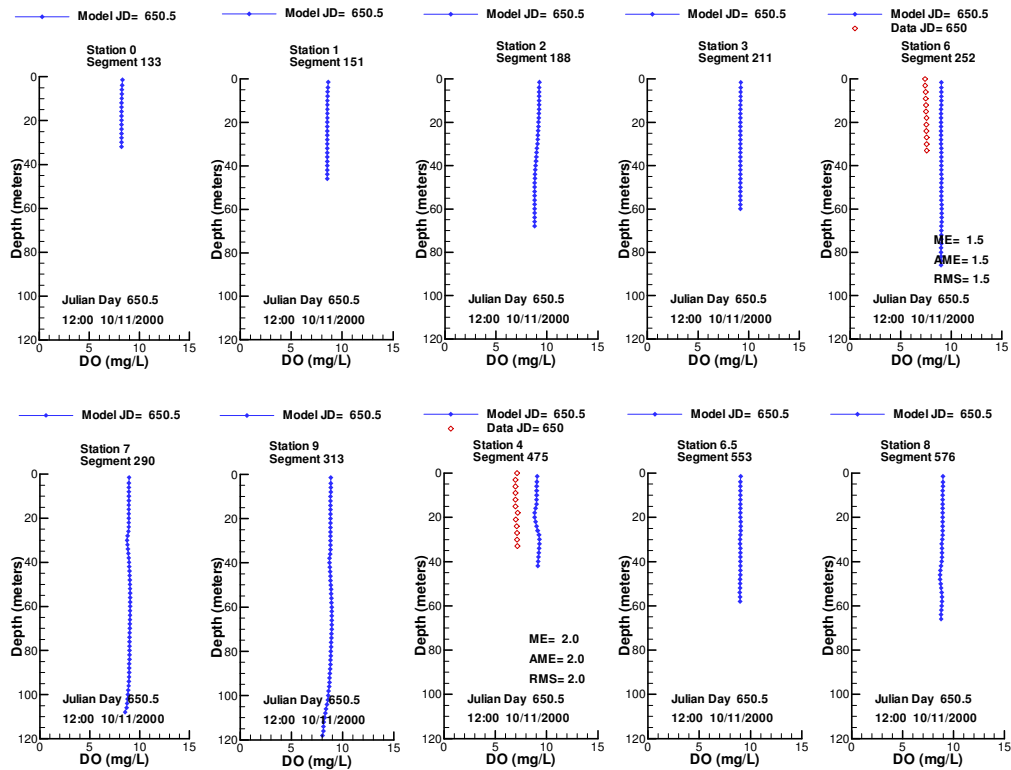


Figure 278. Vertical dissolved oxygen profile model-data comparison, J650.

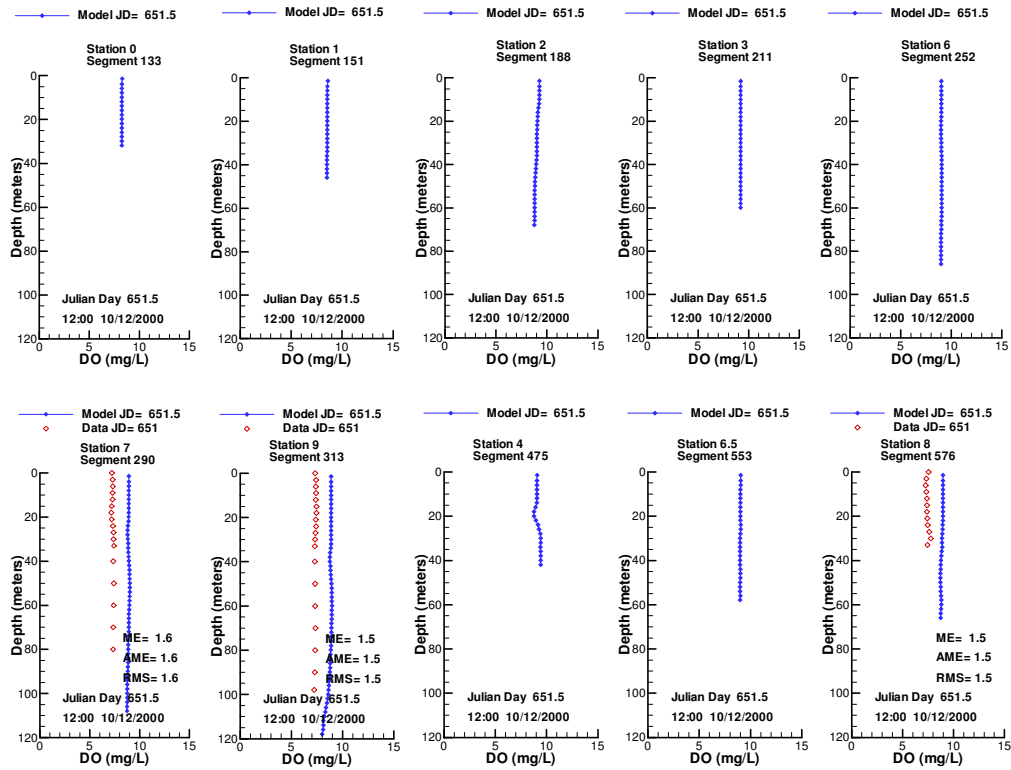


Figure 279. Vertical dissolved oxygen profile model-data comparison, J651.

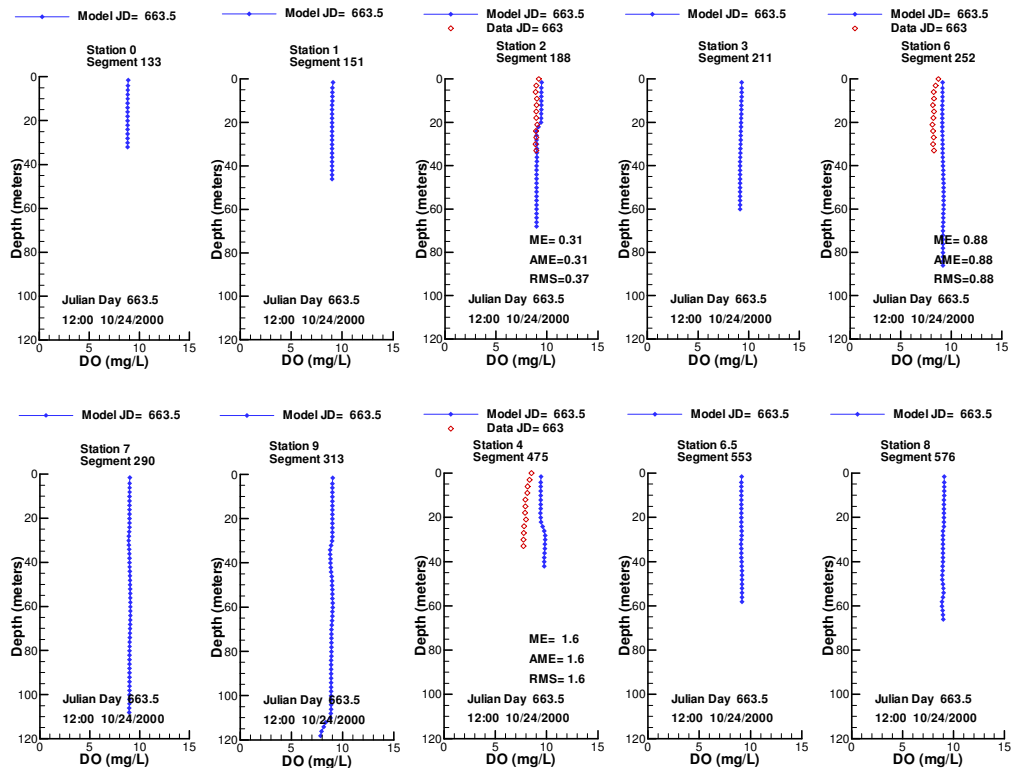


Figure 280. Vertical dissolved oxygen profile model-data comparison, J663.

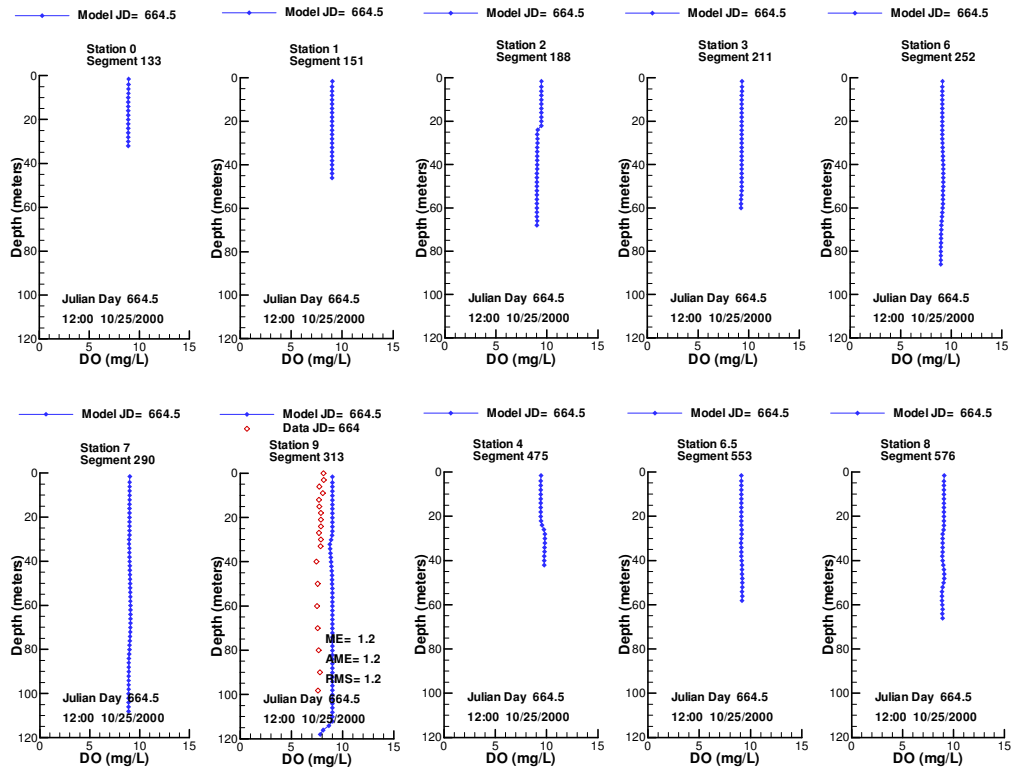


Figure 281. Vertical dissolved oxygen profile model-data comparison, J664.

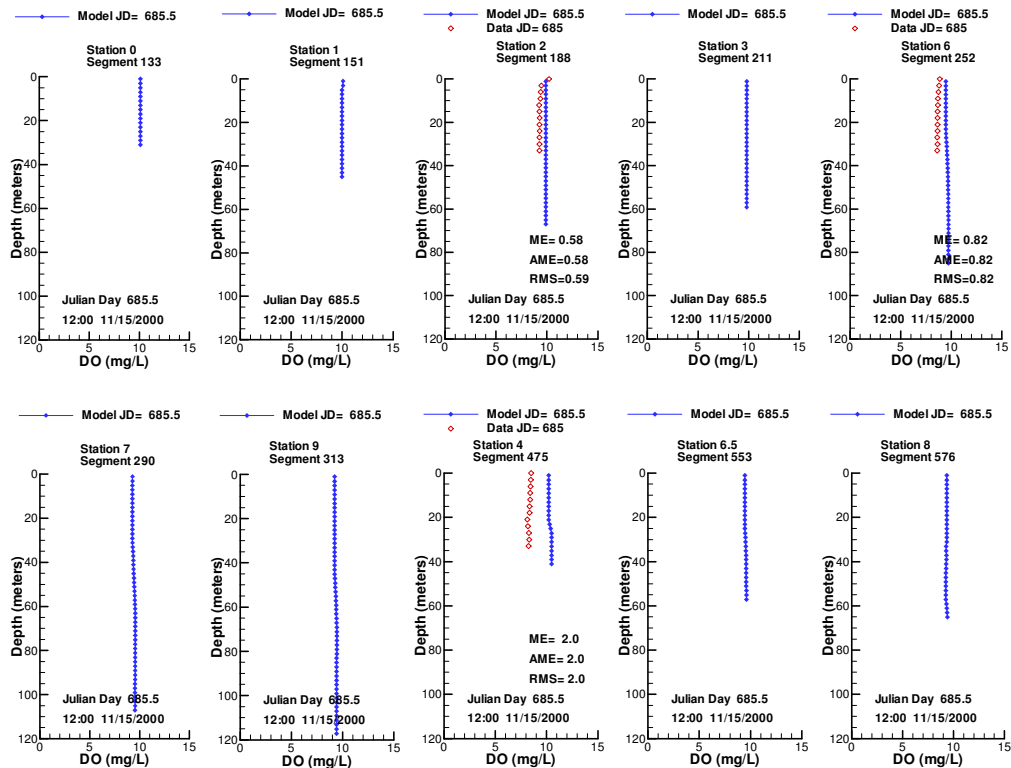


Figure 282. Vertical dissolved oxygen profile model-data comparison, J685.

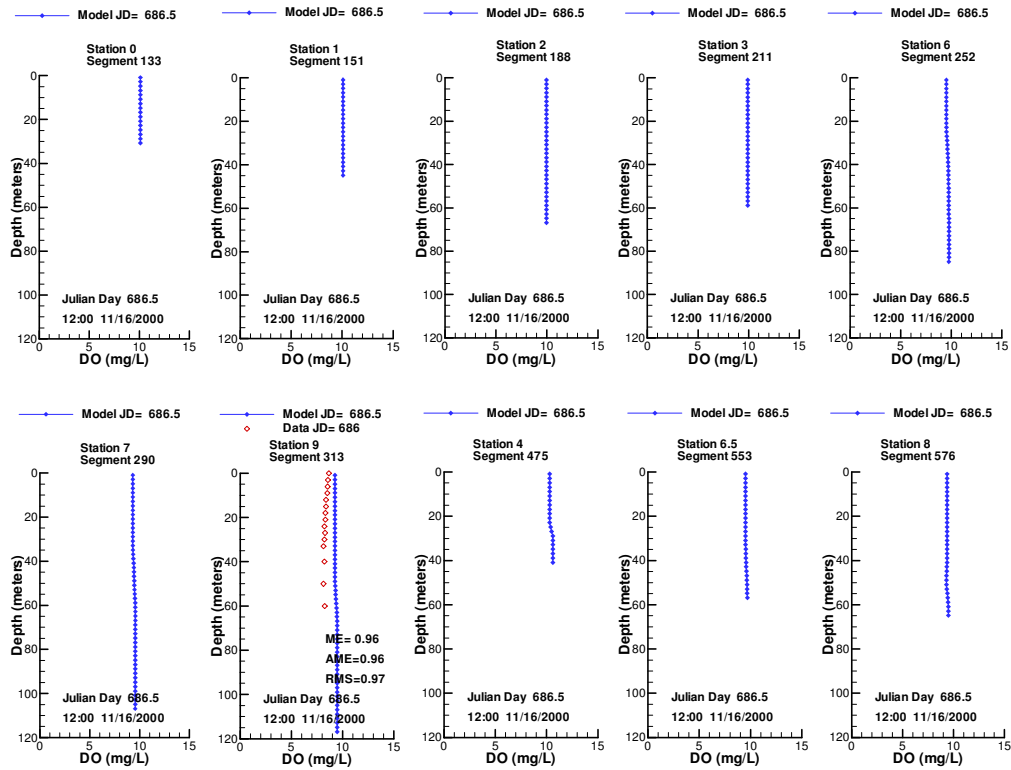


Figure 283. Vertical dissolved oxygen profile model-data comparison, J686.

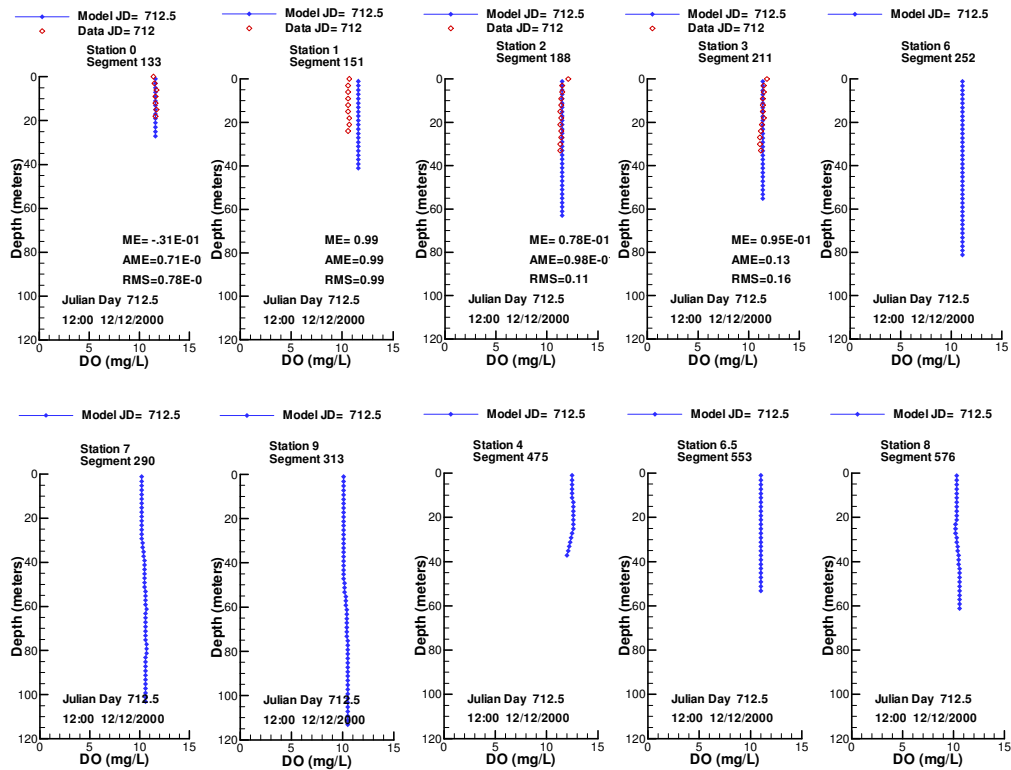


Figure 284. Vertical dissolved oxygen profile model-data comparison, J712.

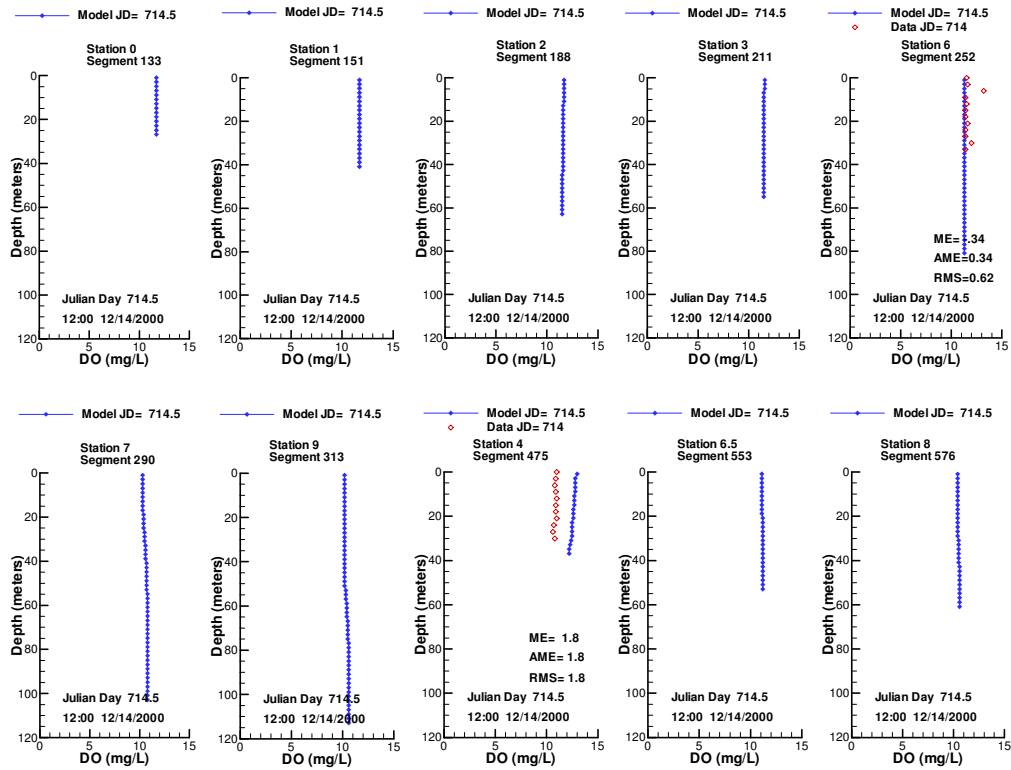


Figure 285. Vertical dissolved oxygen profile model-data comparison, J714.

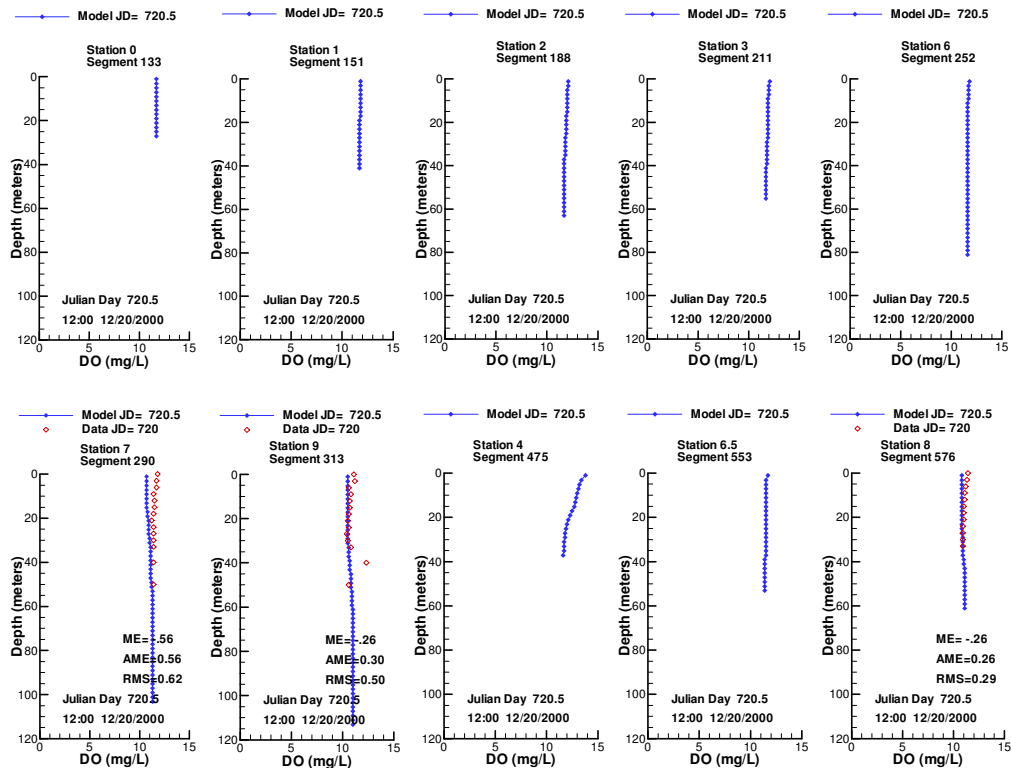


Figure 286. Vertical dissolved oxygen profile model-data comparison, J720.

Temperature

For days when data were recorded, vertical profile plots of water temperature are shown in Figure 287 through Figure 333.

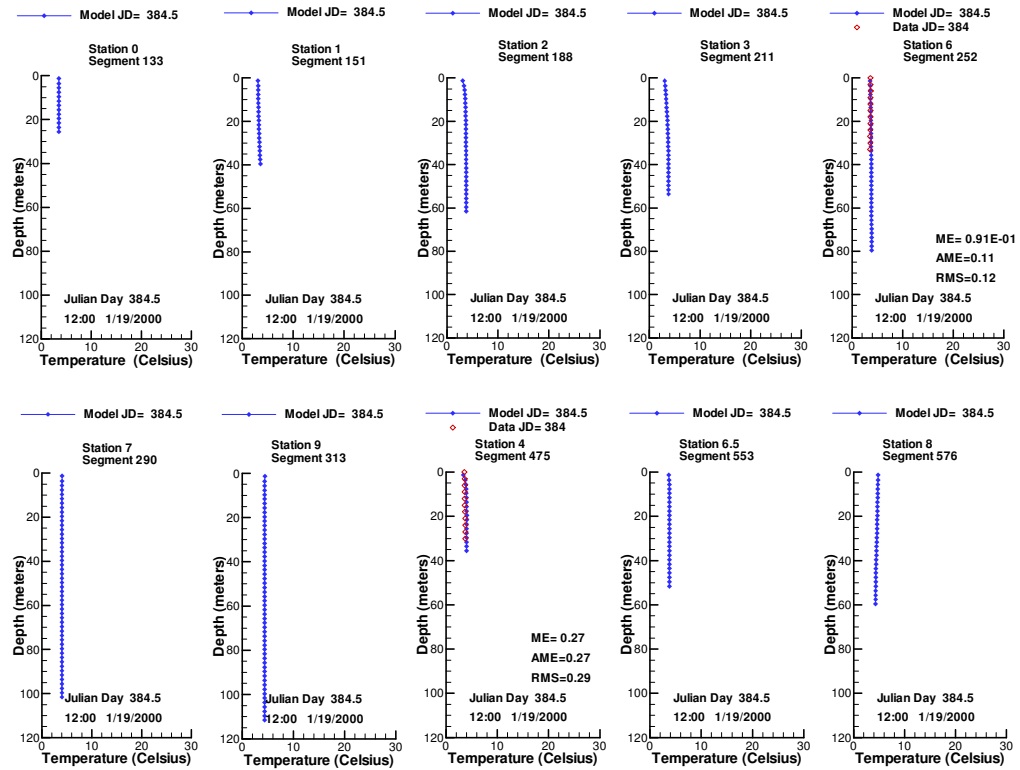


Figure 287. Vertical temperature profile model-data comparison, J384.

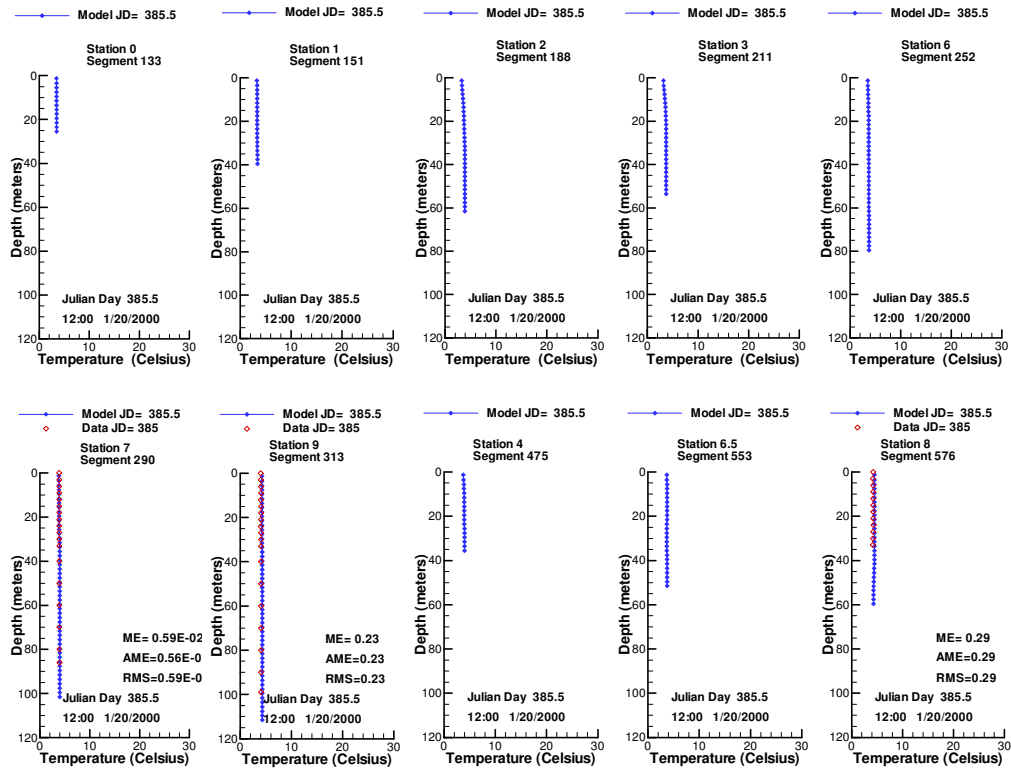


Figure 288. Vertical temperature profile model-data comparison, J385.

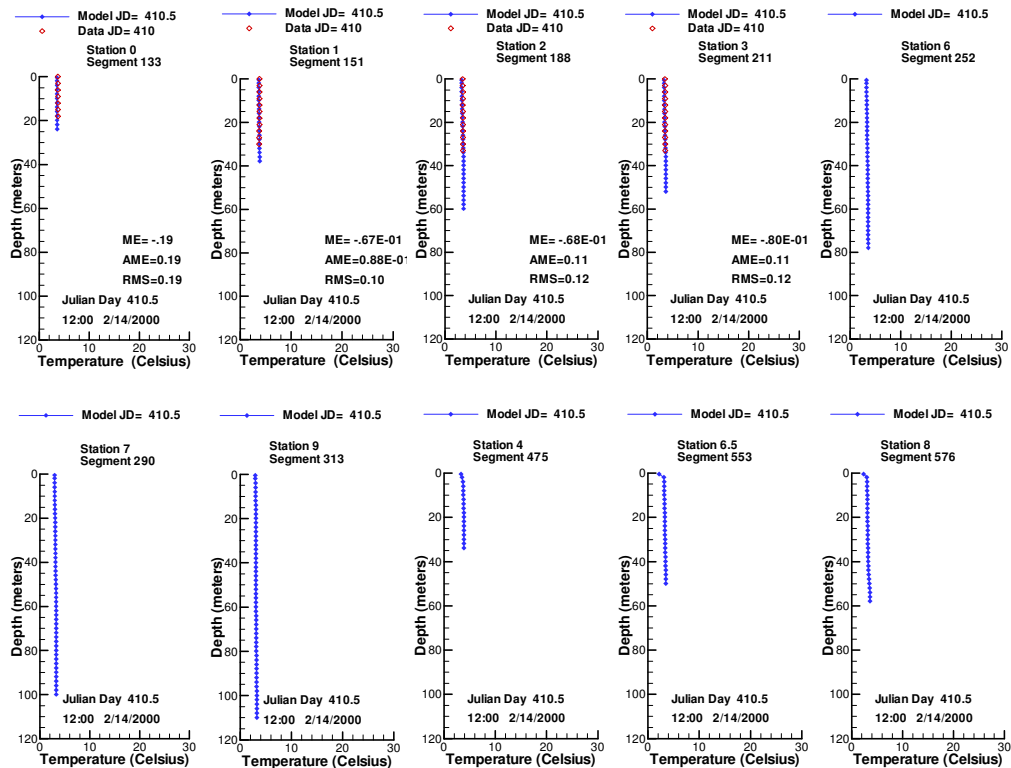


Figure 289. Vertical temperature profile model-data comparison, J410.

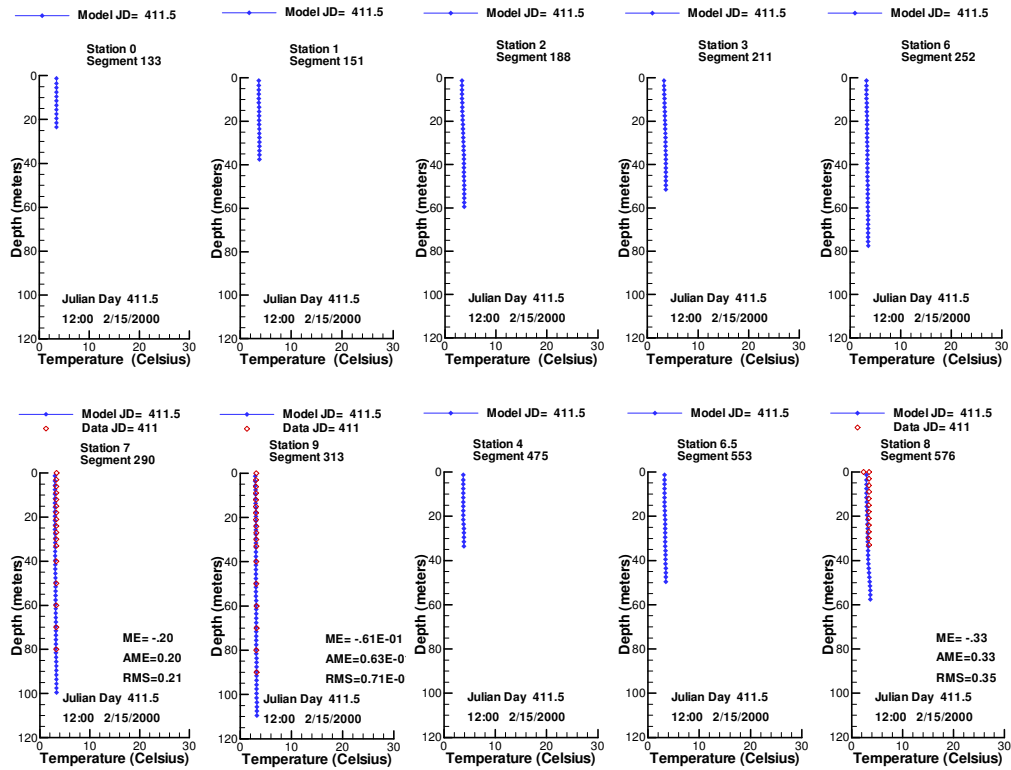


Figure 290. Vertical temperature profile model-data comparison, J411.

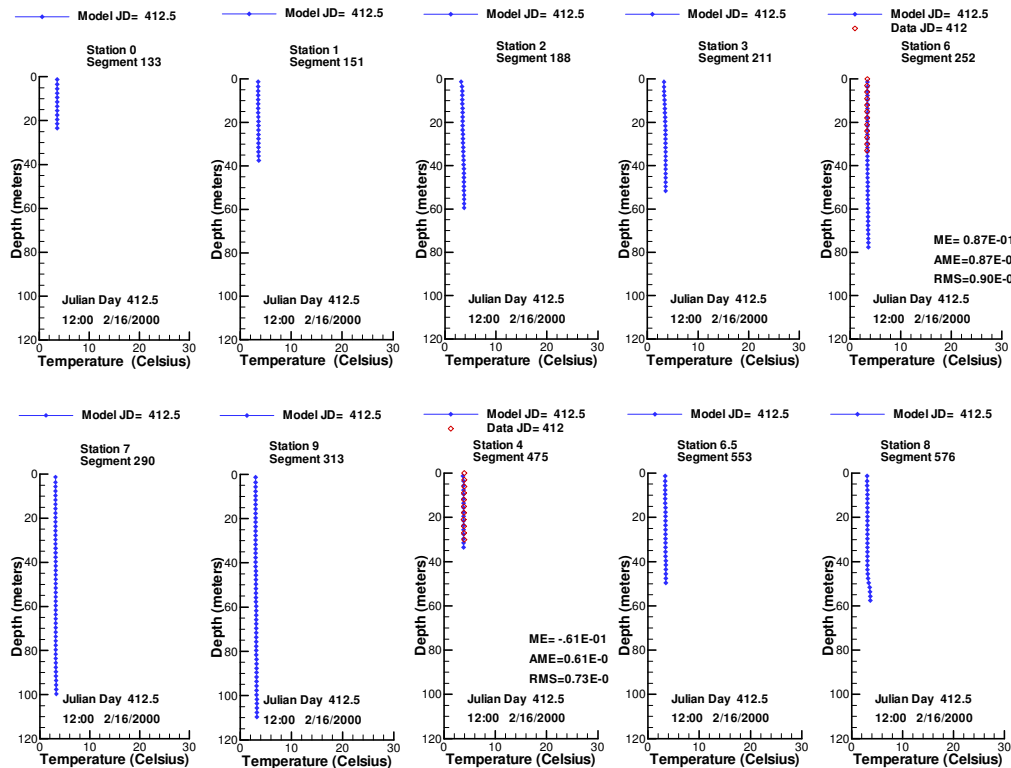


Figure 291. Vertical temperature profile model-data comparison, J412.

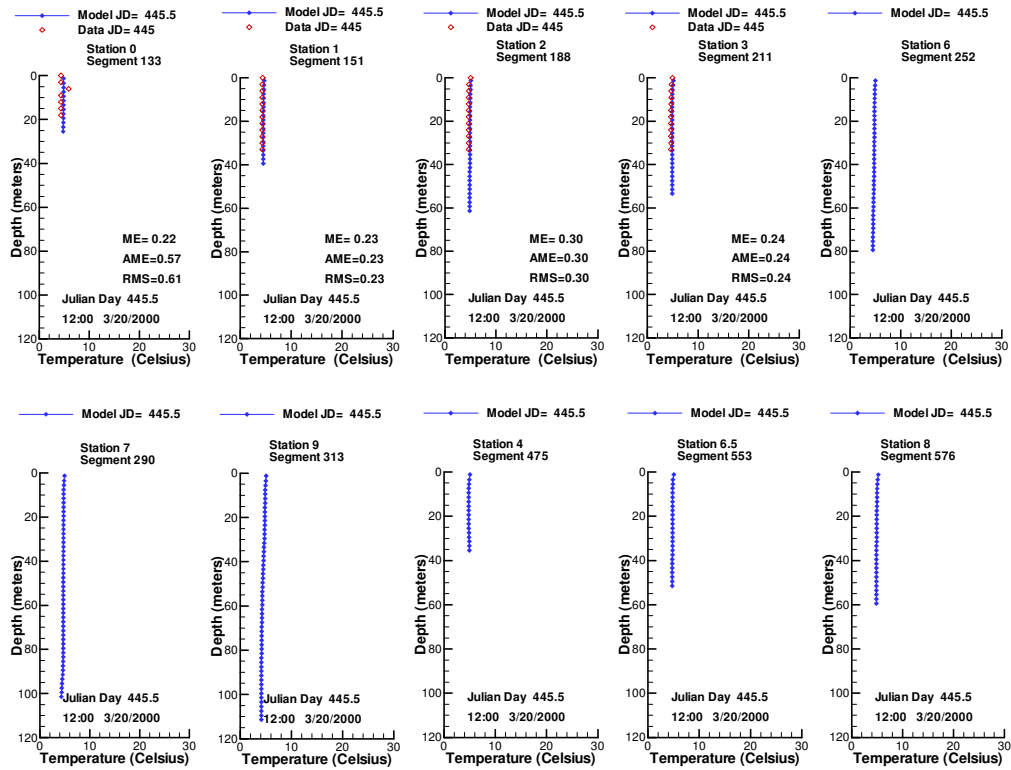


Figure 292. Vertical temperature profile model-data comparison, J445.

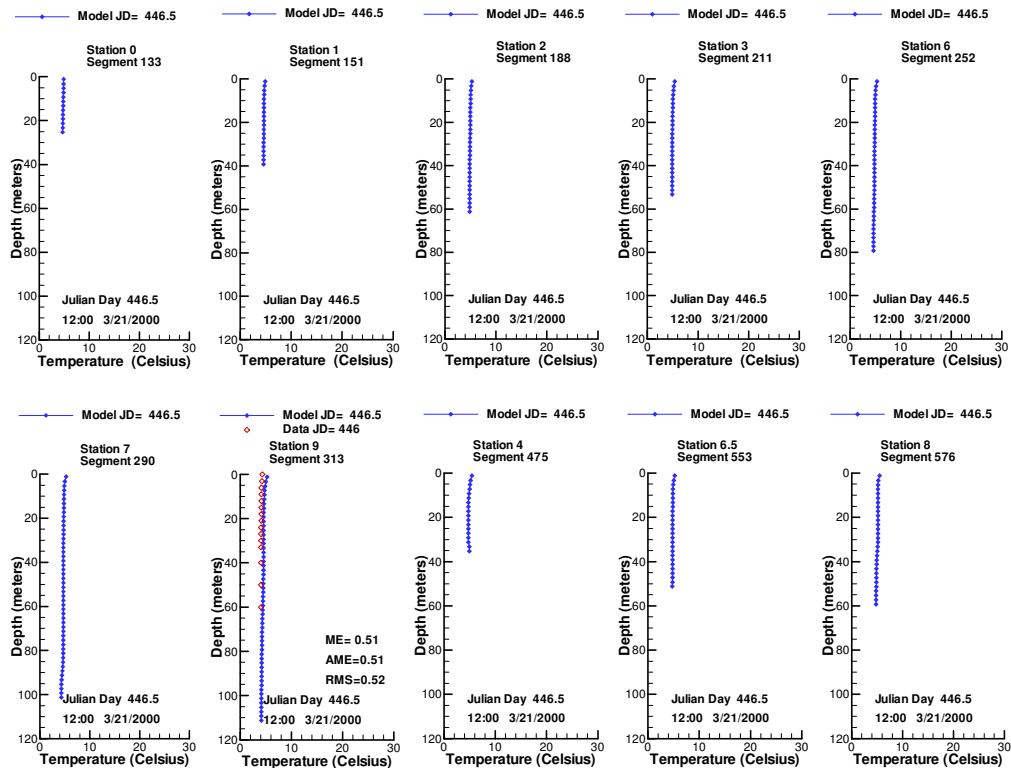


Figure 293. Vertical temperature profile model-data comparison, J446.

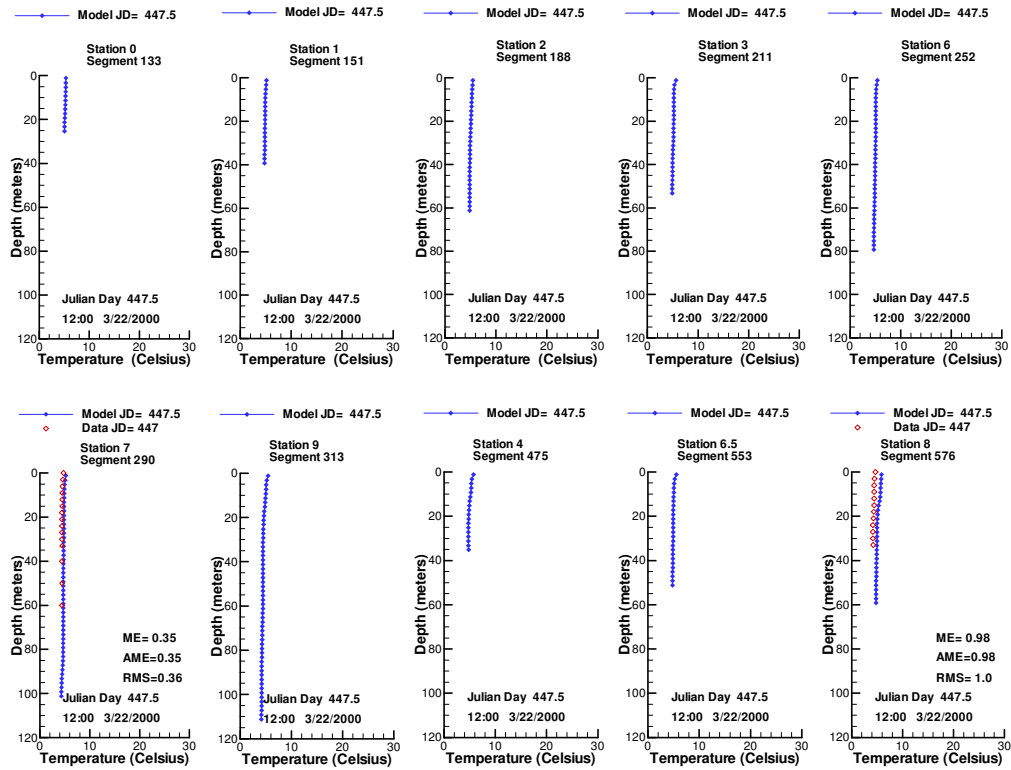


Figure 294. Vertical temperature profile model-data comparison, J447.

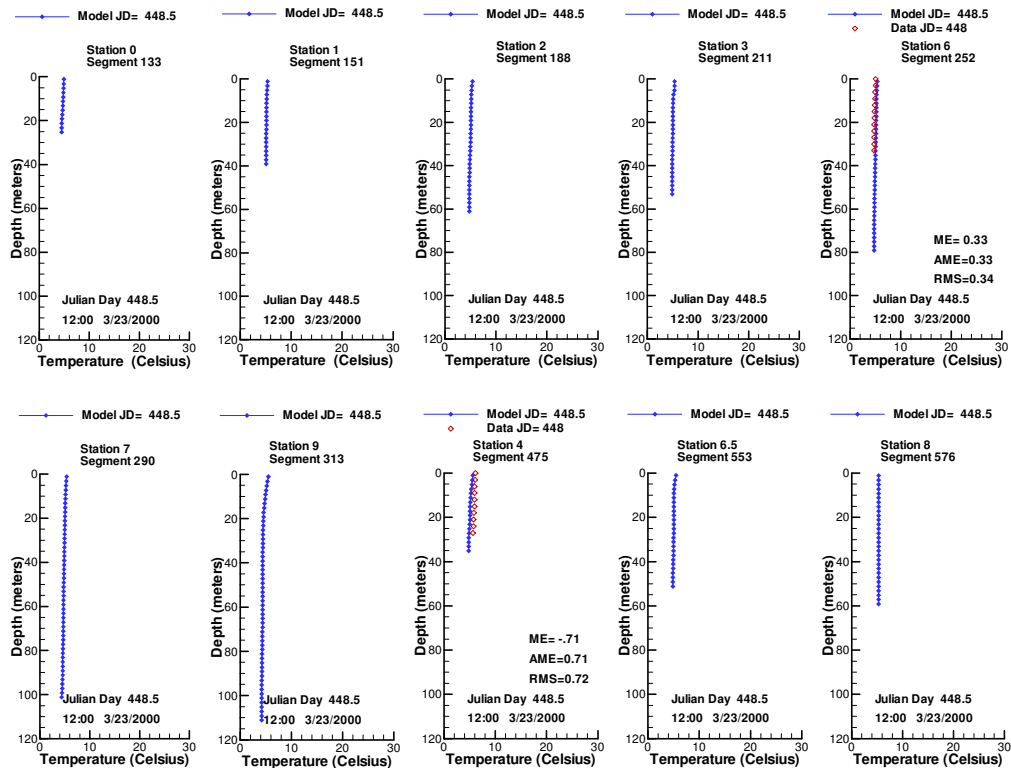


Figure 295. Vertical temperature profile model-data comparison, J448.

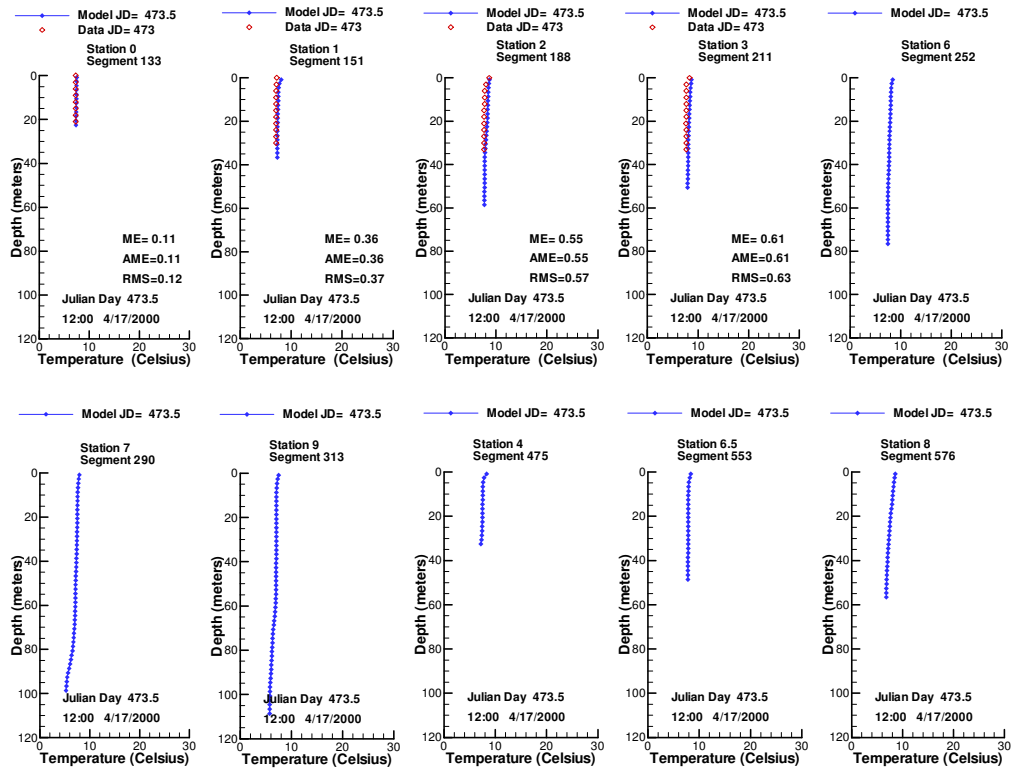


Figure 296. Vertical temperature profile model-data comparison, J473.

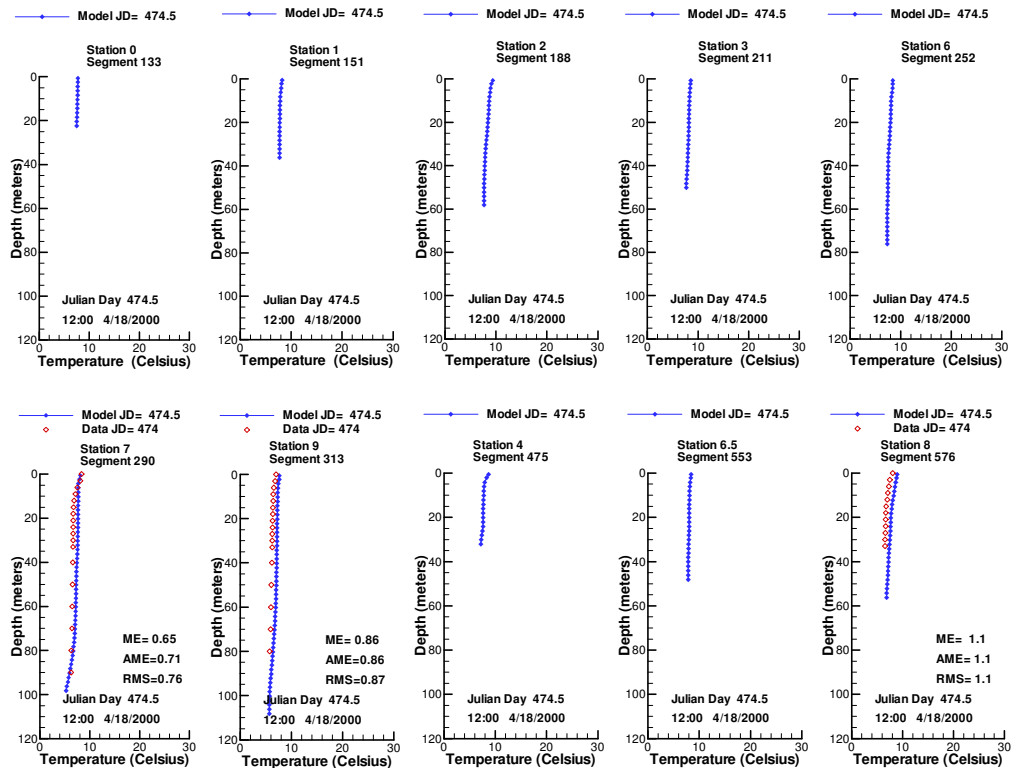


Figure 297. Vertical temperature profile model-data comparison, J474.

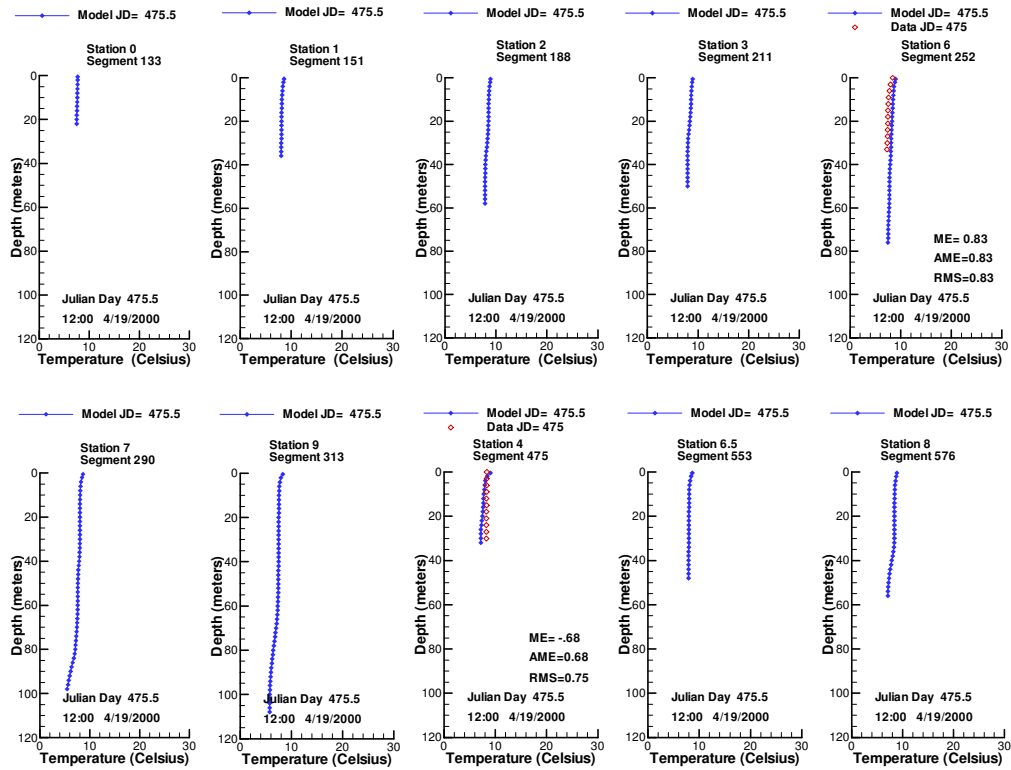


Figure 298. Vertical temperature profile model-data comparison, J475.

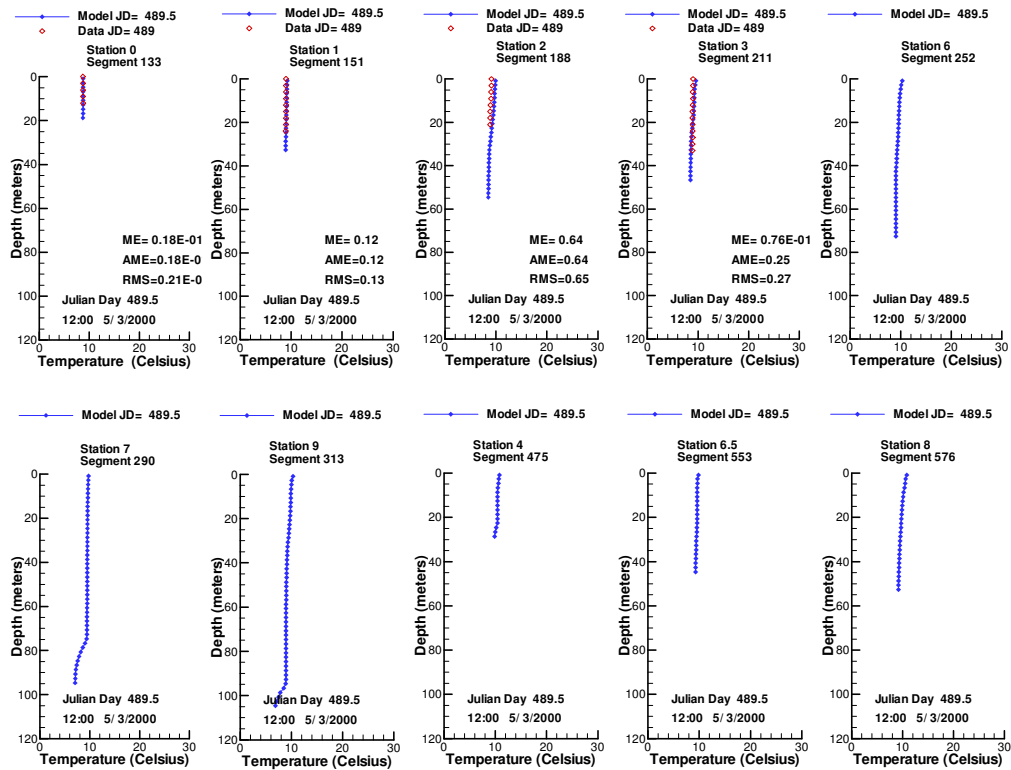


Figure 299. Vertical temperature profile model-data comparison, J489.

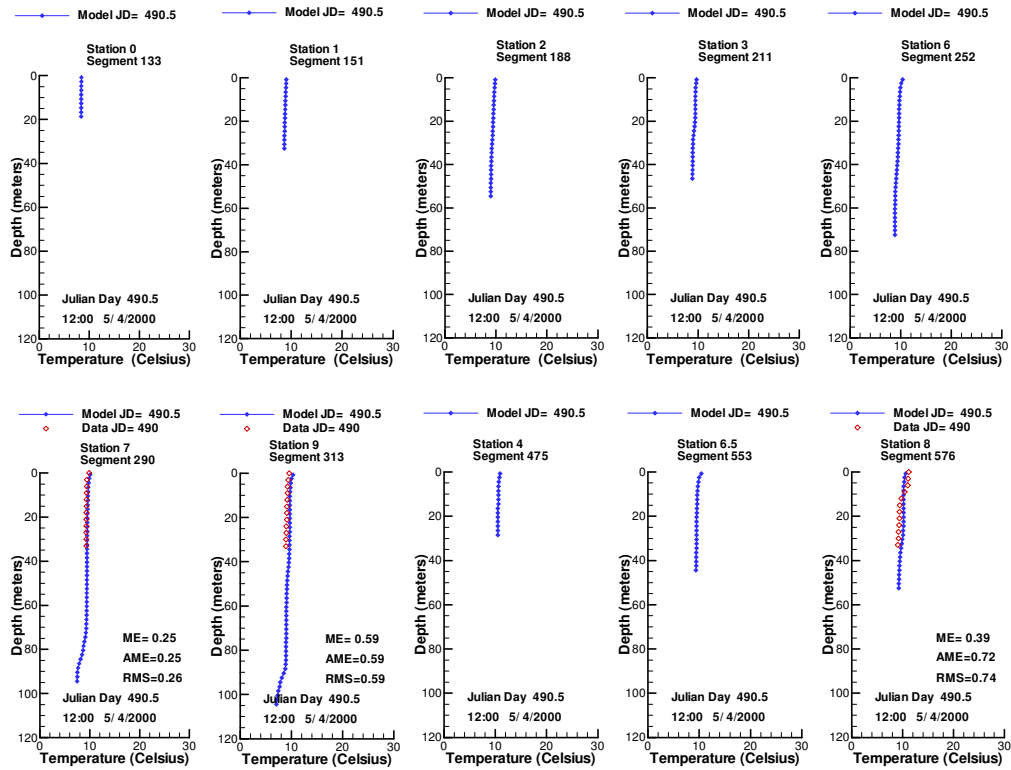


Figure 300. Vertical temperature profile model-data comparison, J490.

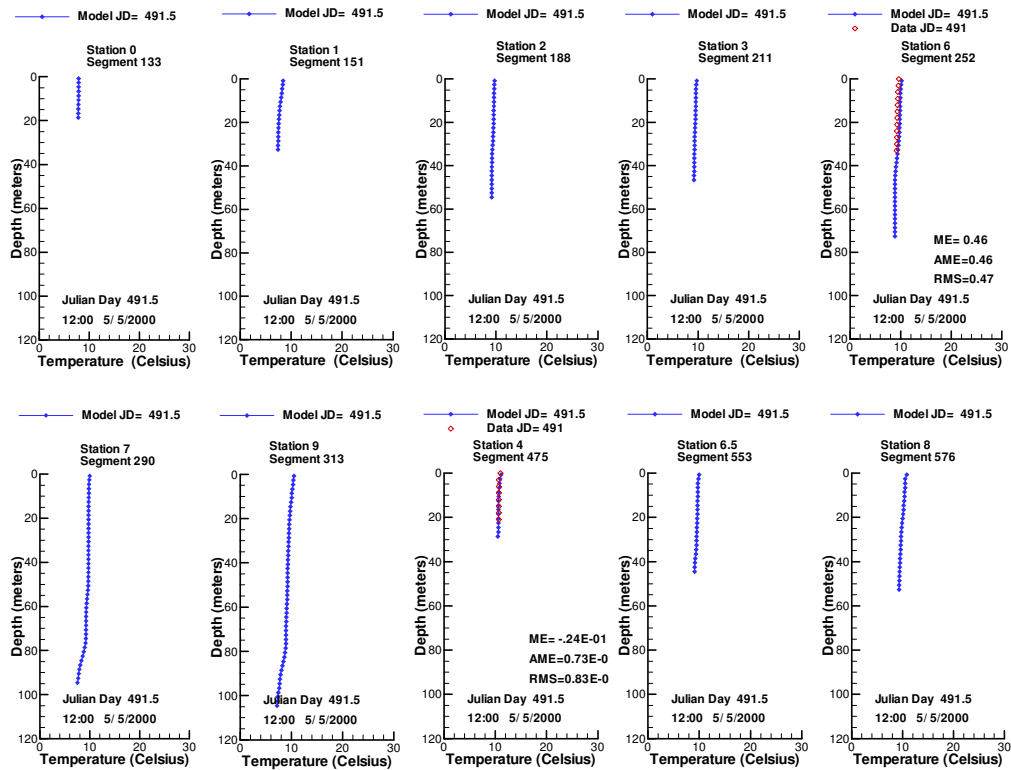


Figure 301. Vertical temperature profile model-data comparison, J491.

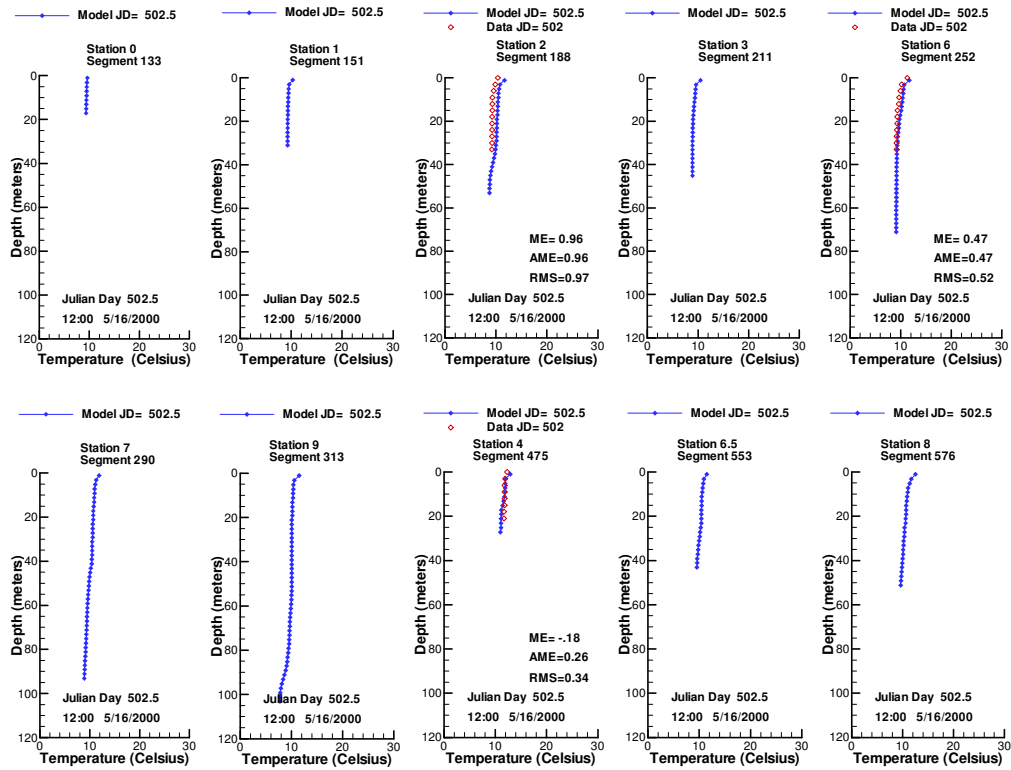


Figure 302. Vertical temperature profile model-data comparison, J502.

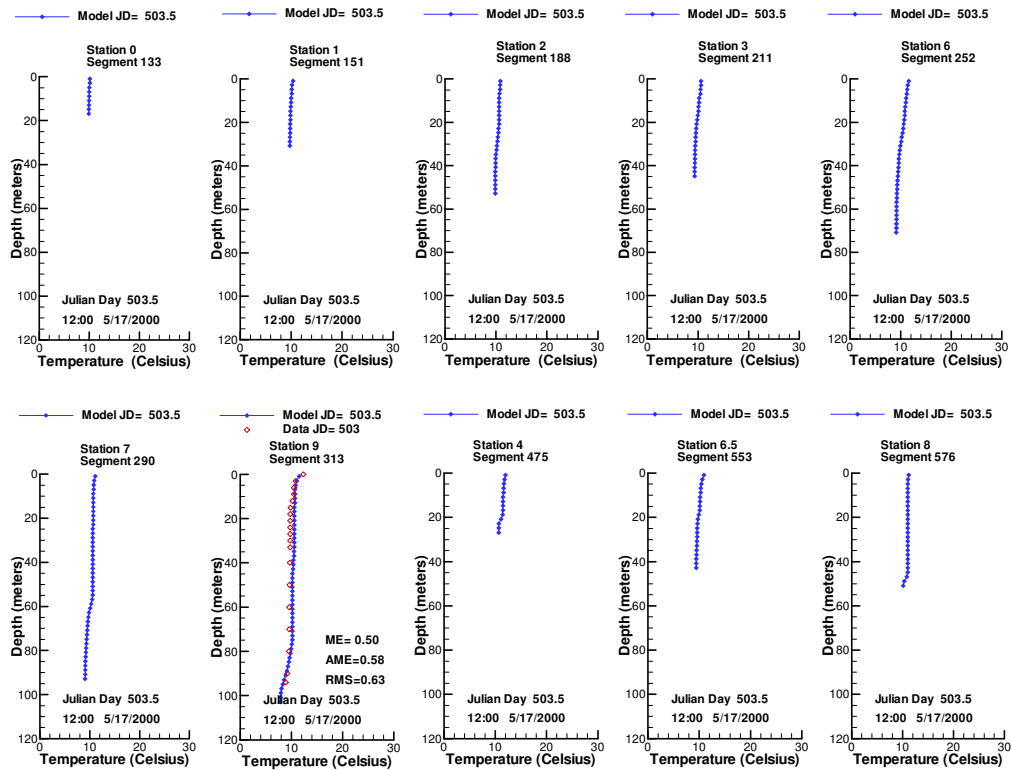


Figure 303. Vertical temperature profile model-data comparison, J503.

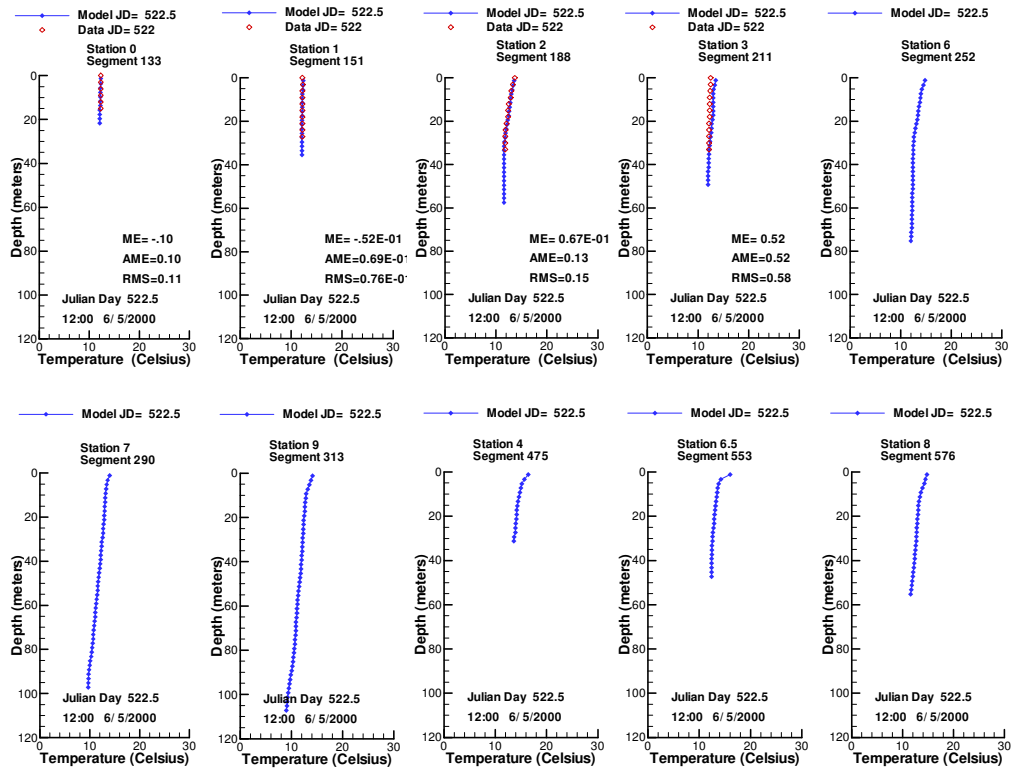


Figure 304. Vertical temperature profile model-data comparison, J522.

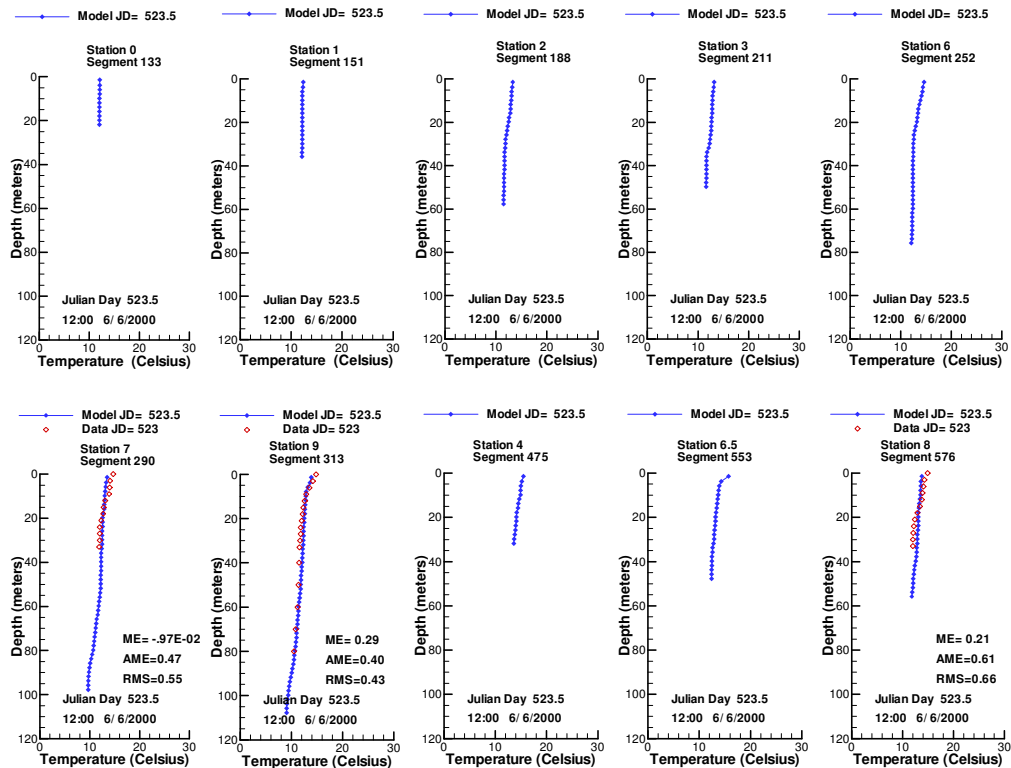


Figure 305. Vertical temperature profile model-data comparison, J523.

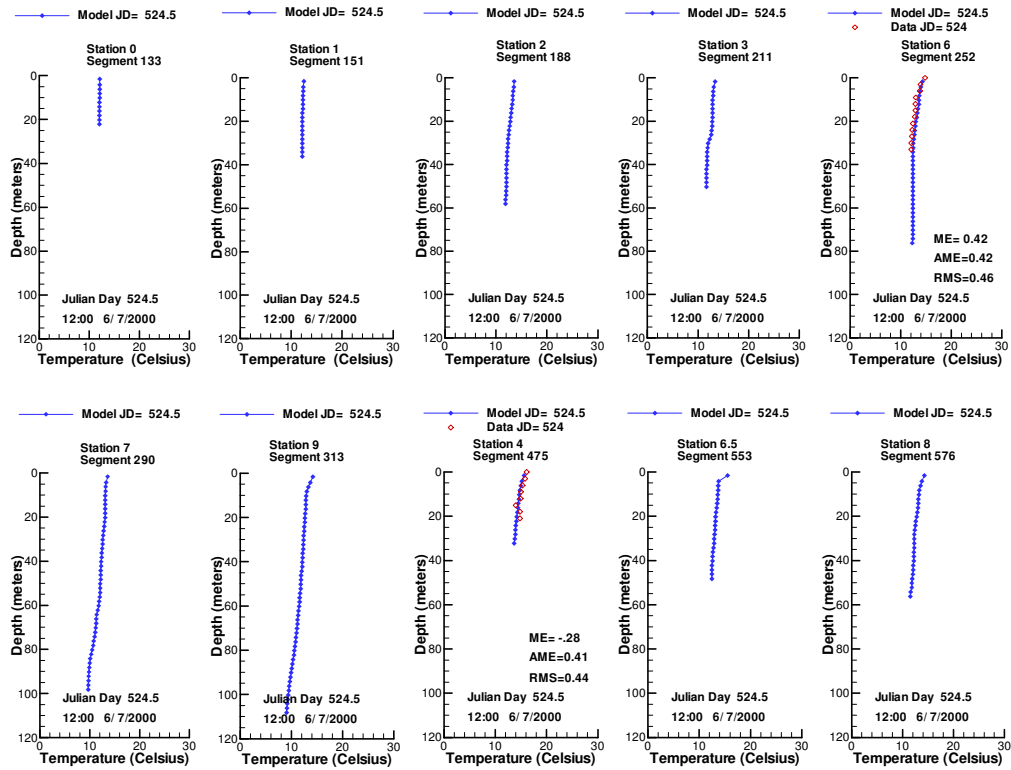


Figure 306. Vertical temperature profile model-data comparison, J524.

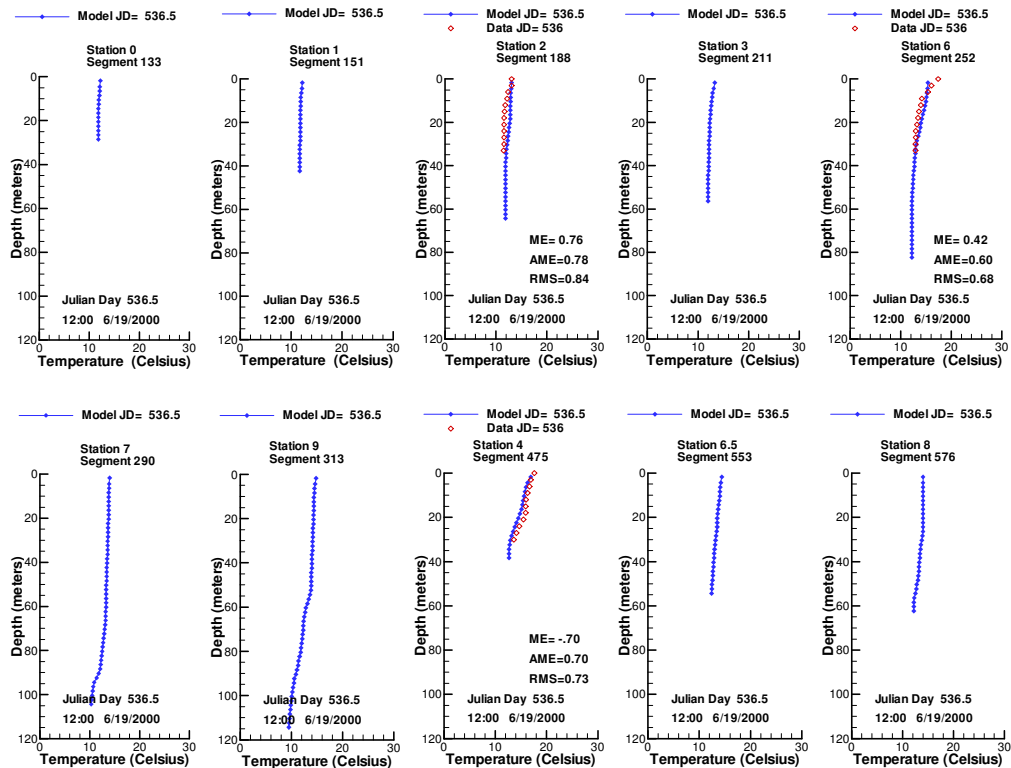


Figure 307. Vertical temperature profile model-data comparison, J536.

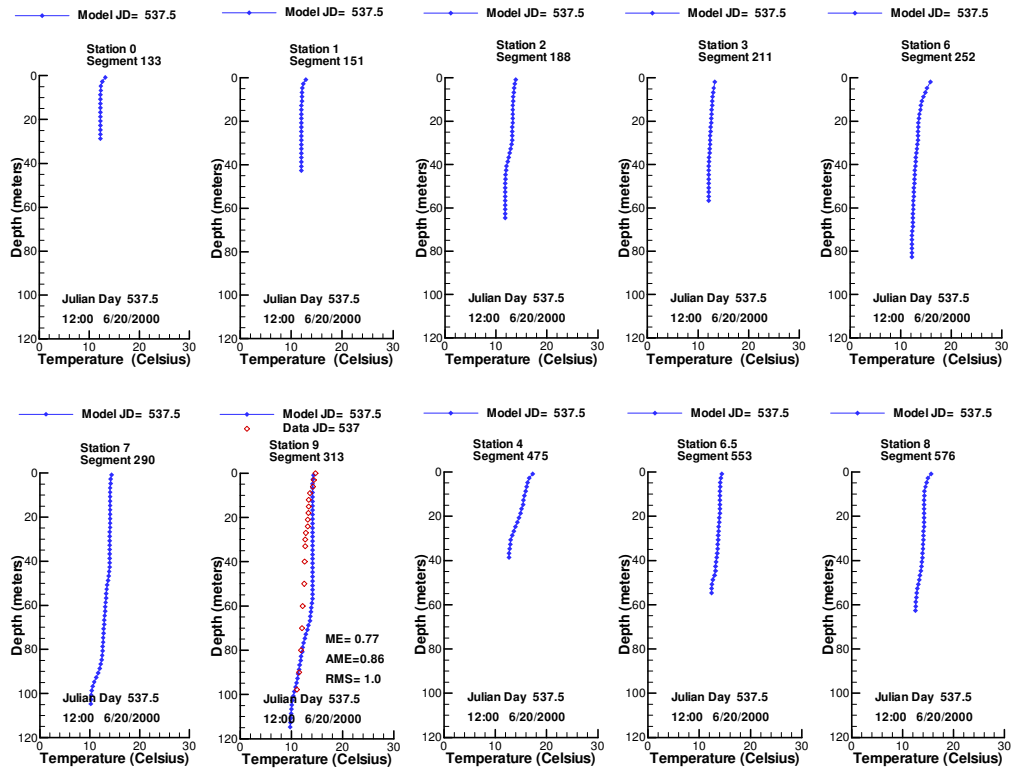


Figure 308. Vertical temperature profile model-data comparison, J537.

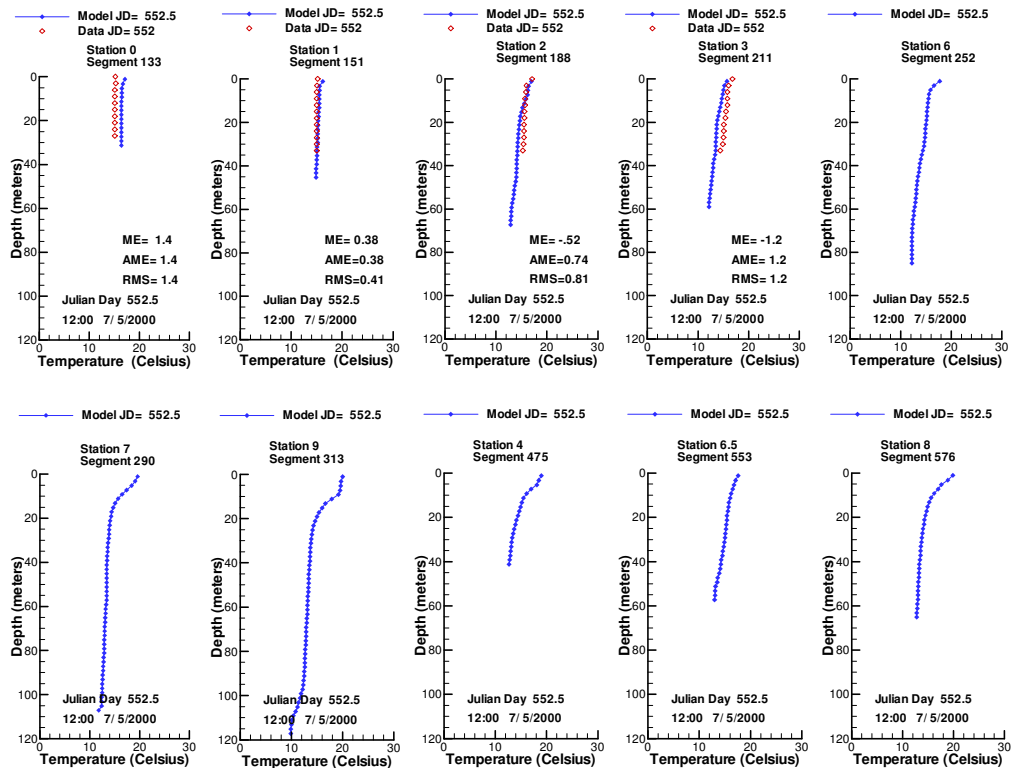


Figure 309. Vertical temperature profile model-data comparison, J552.

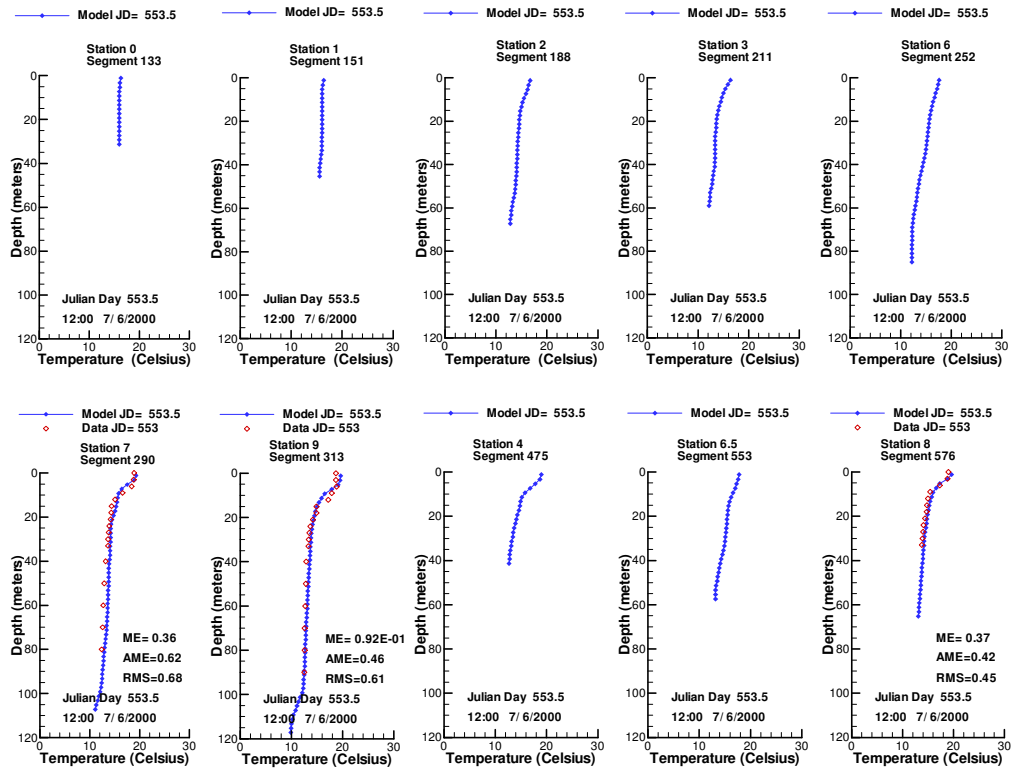


Figure 310. Vertical temperature profile model-data comparison, J553.

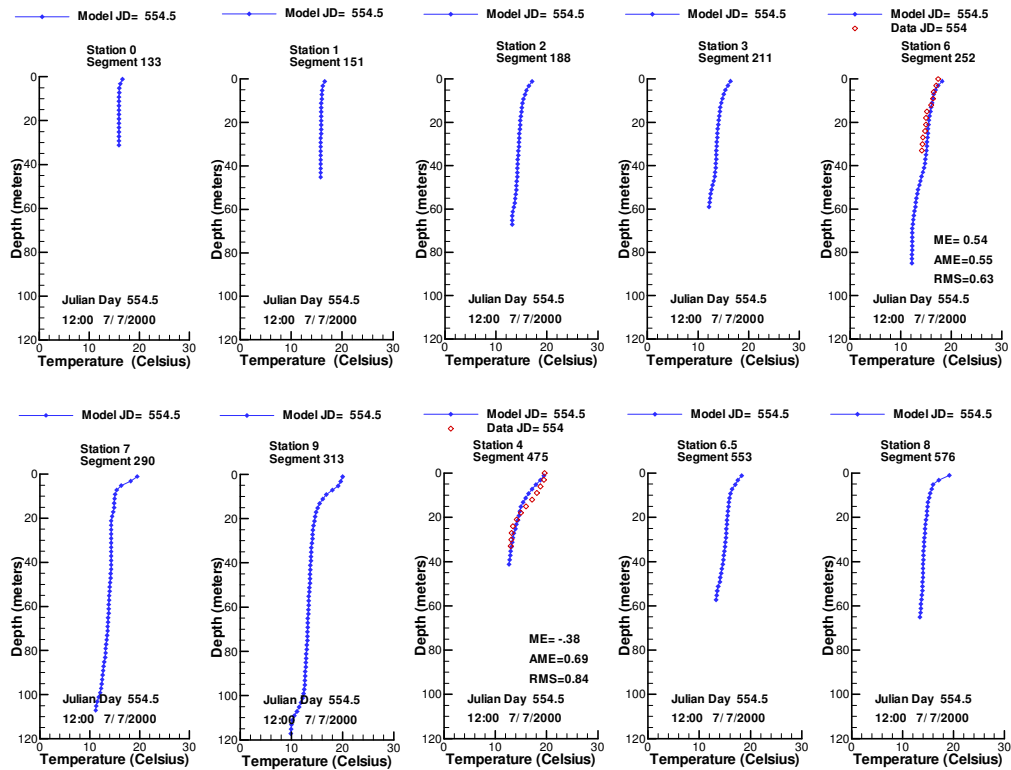


Figure 311. Vertical temperature profile model-data comparison, J554.

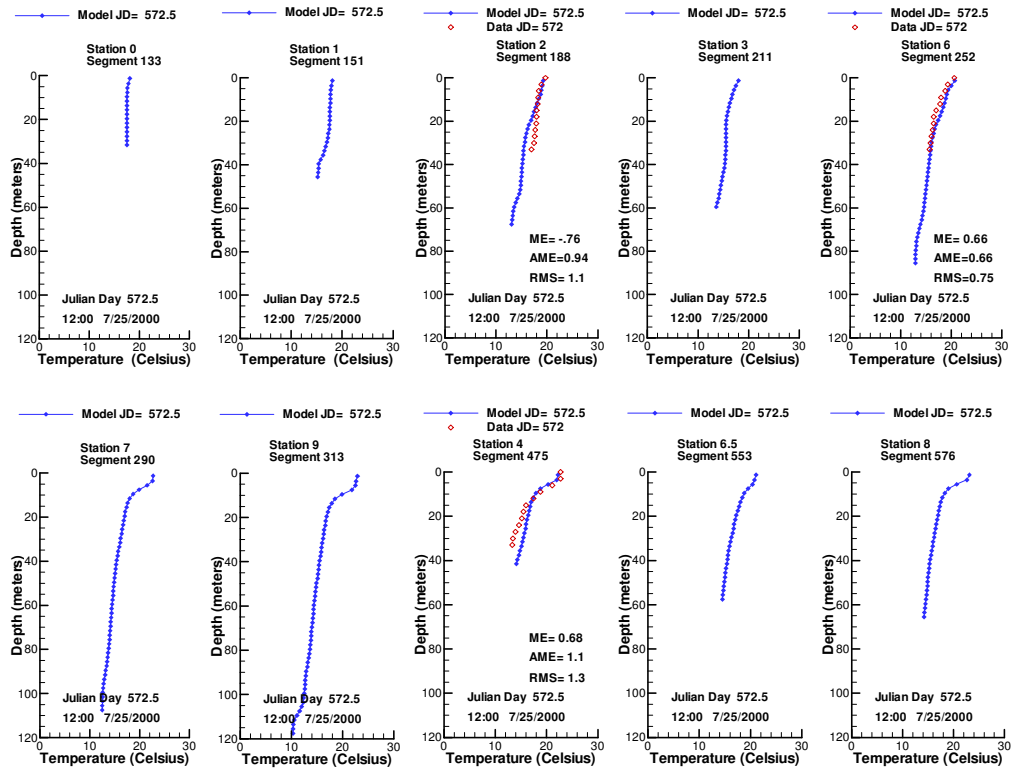


Figure 312. Vertical temperature profile model-data comparison, J572.

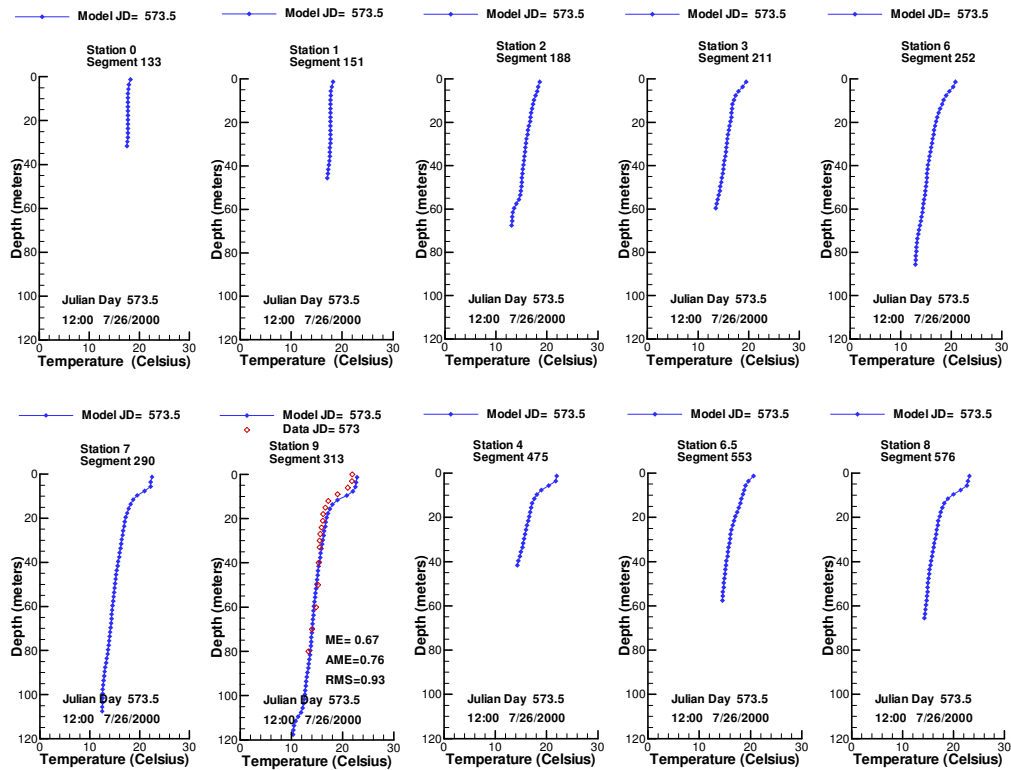


Figure 313. Vertical temperature profile model-data comparison, J573.

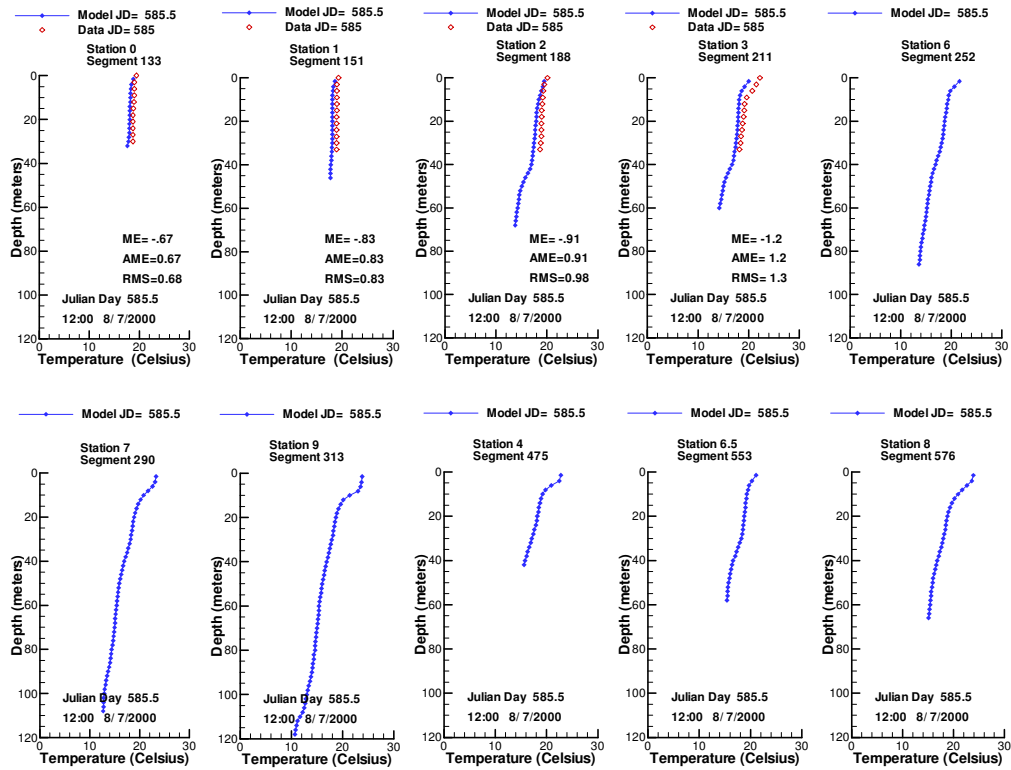


Figure 314. Vertical temperature profile model-data comparison, J585.

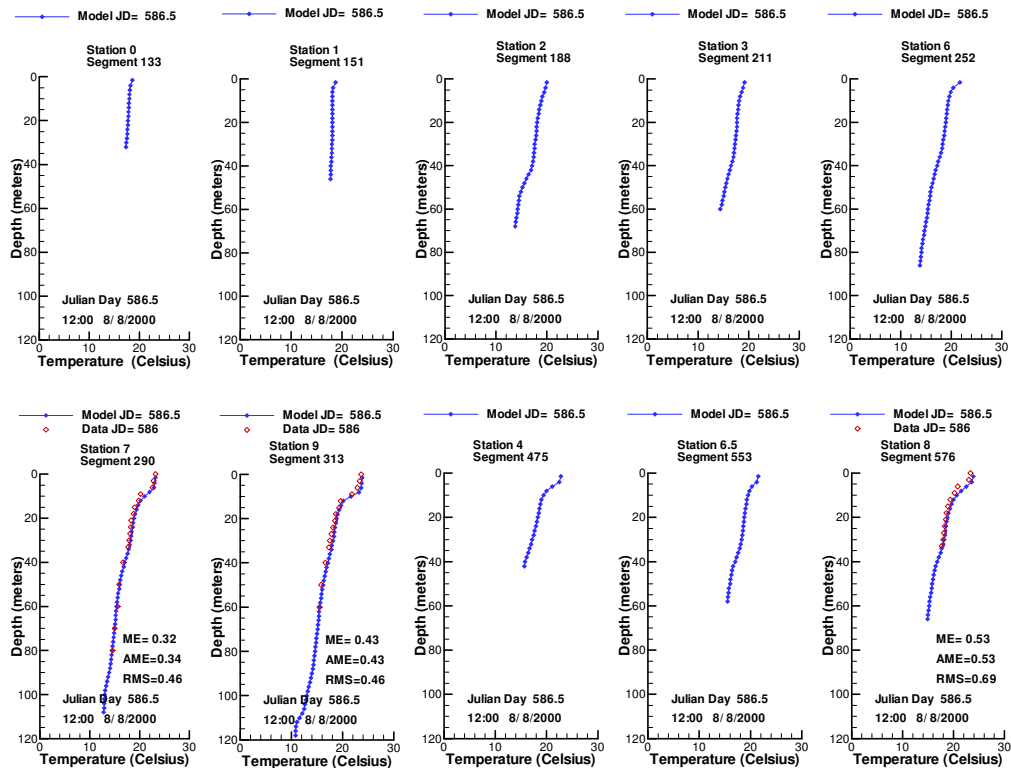


Figure 315. Vertical temperature profile model-data comparison, J586.

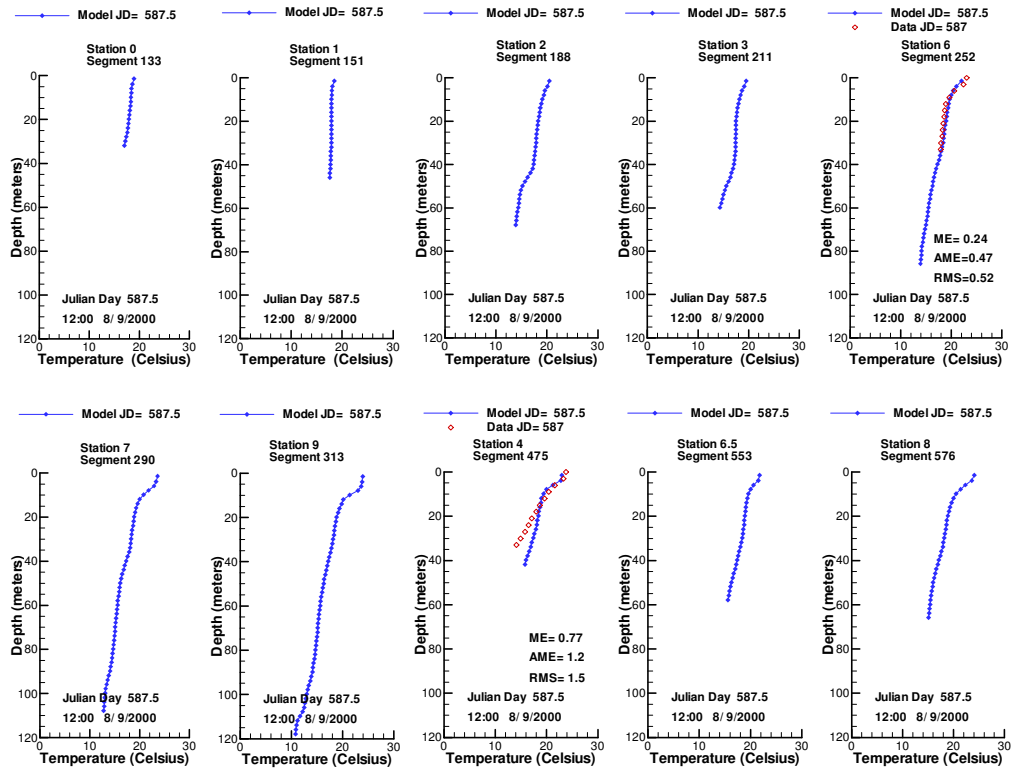


Figure 316. Vertical temperature profile model-data comparison, J587.

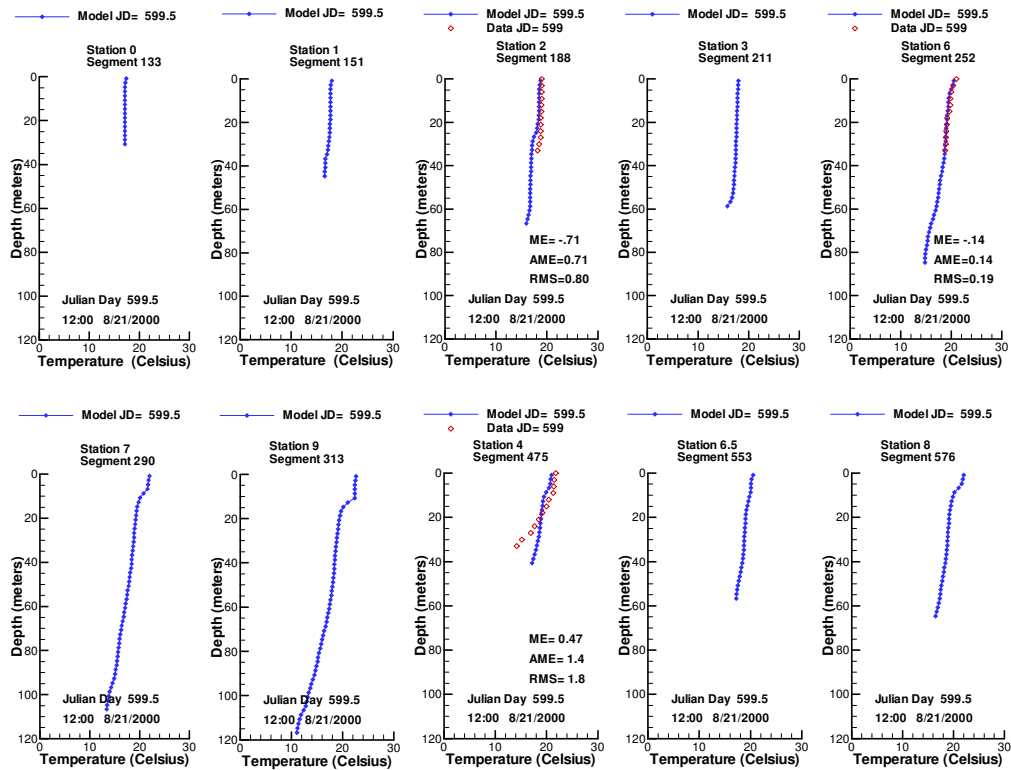


Figure 317. Vertical temperature profile model-data comparison, J599.

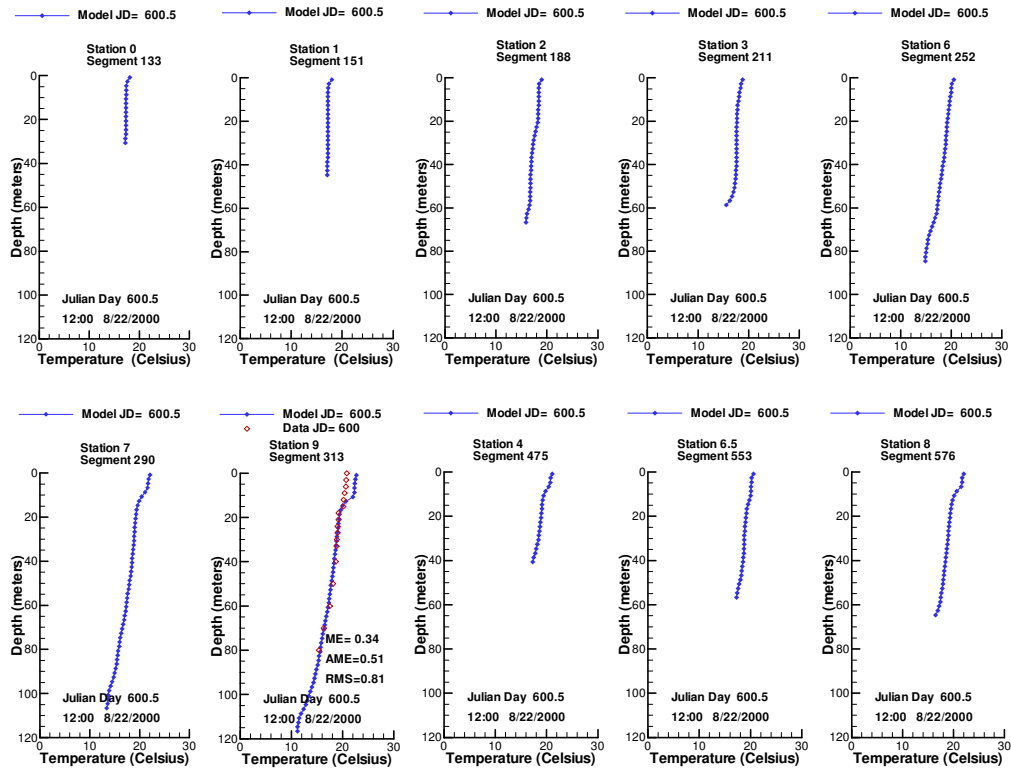


Figure 318. Vertical temperature profile model-data comparison, J600.

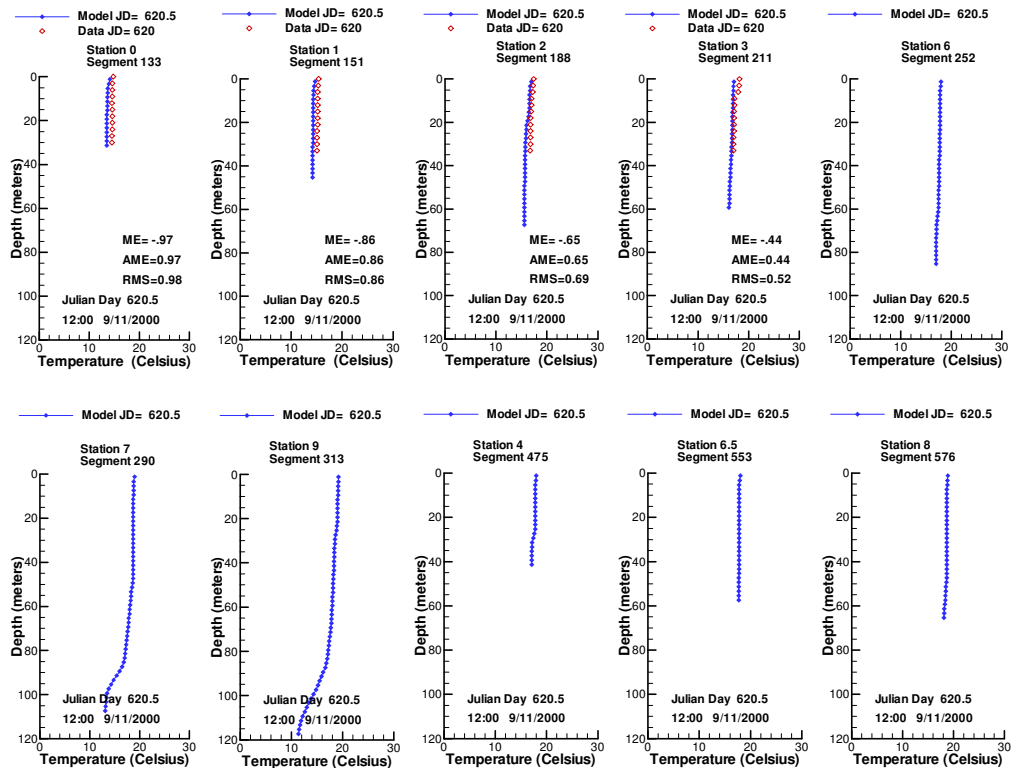


Figure 319. Vertical temperature profile model-data comparison, J620.

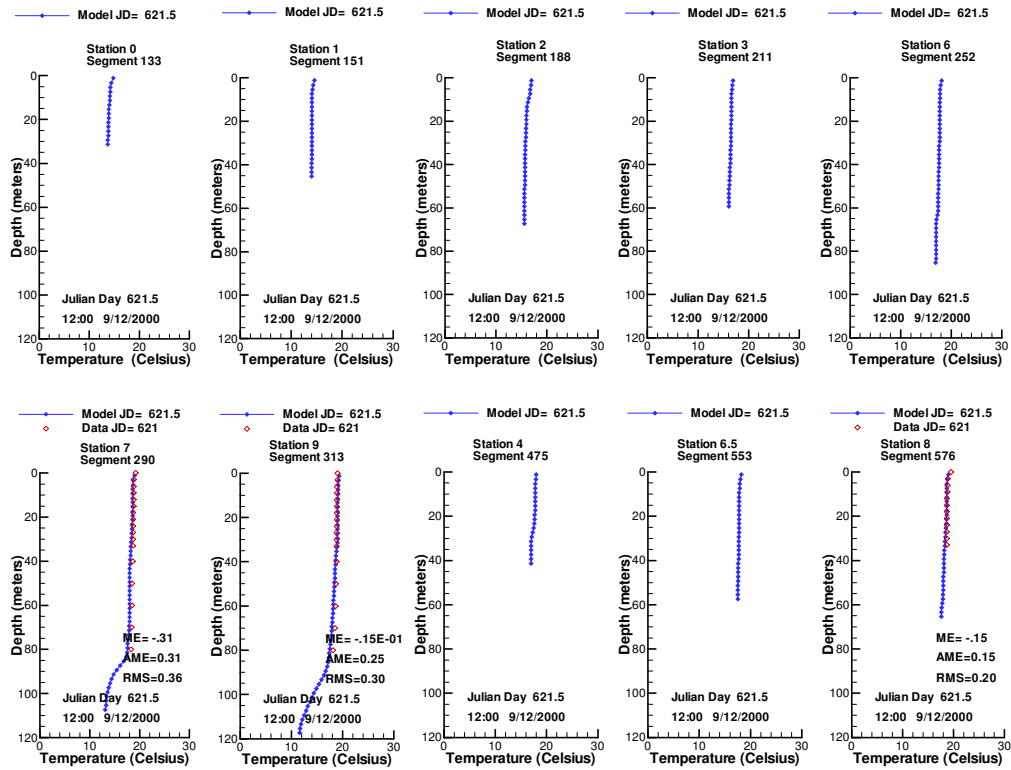


Figure 320. Vertical temperature profile model-data comparison, J621.

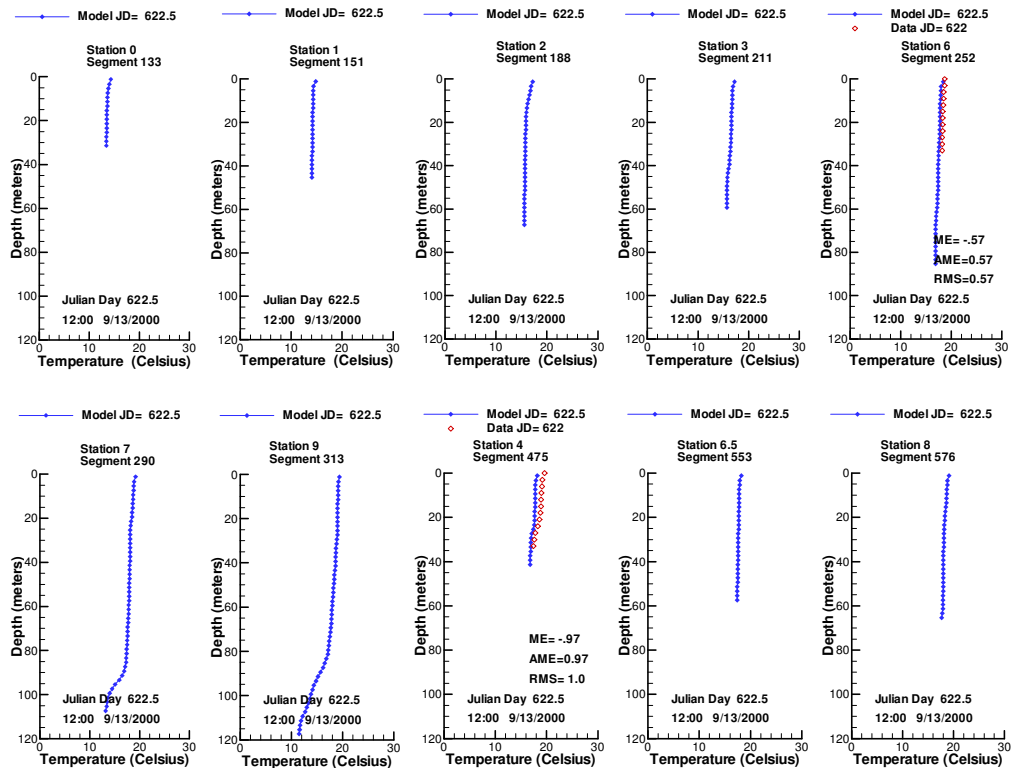


Figure 321. Vertical temperature profile model-data comparison, J622.

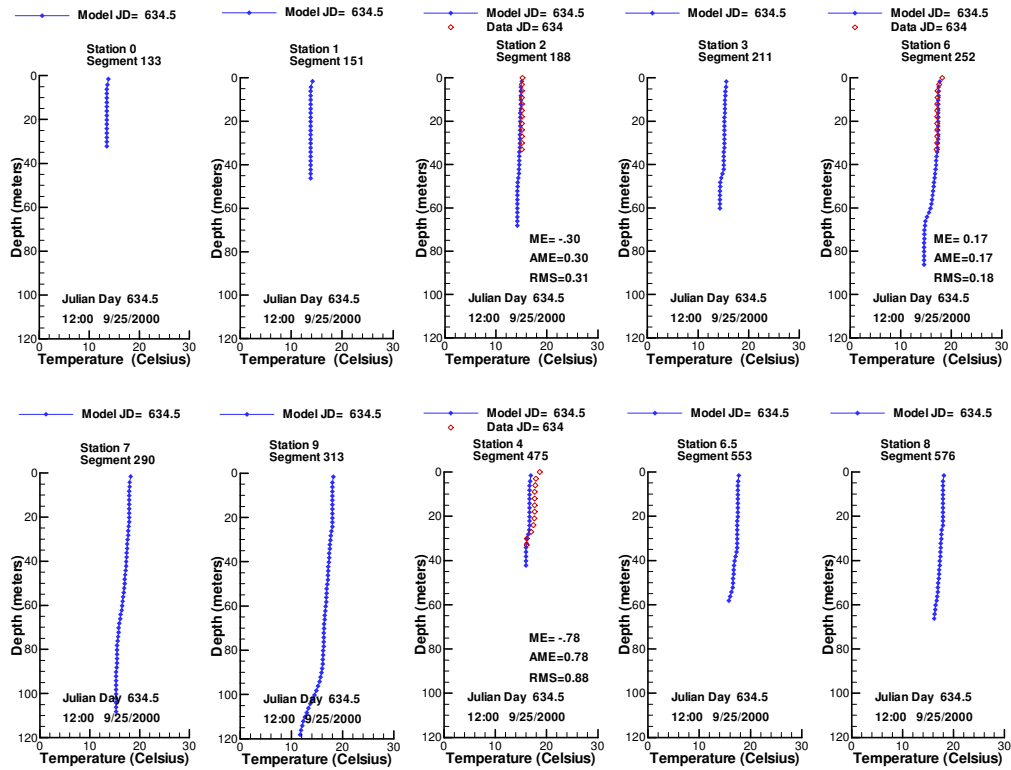


Figure 322. Vertical temperature profile model-data comparison, J634.

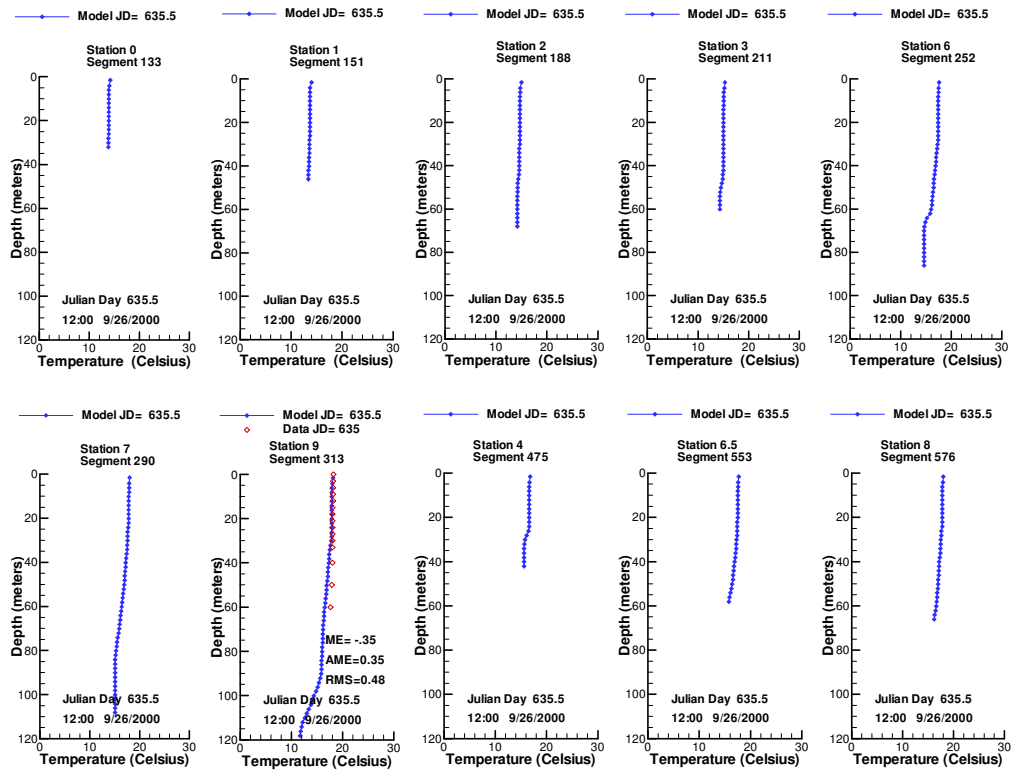


Figure 323. Vertical temperature profile model-data comparison, J635.

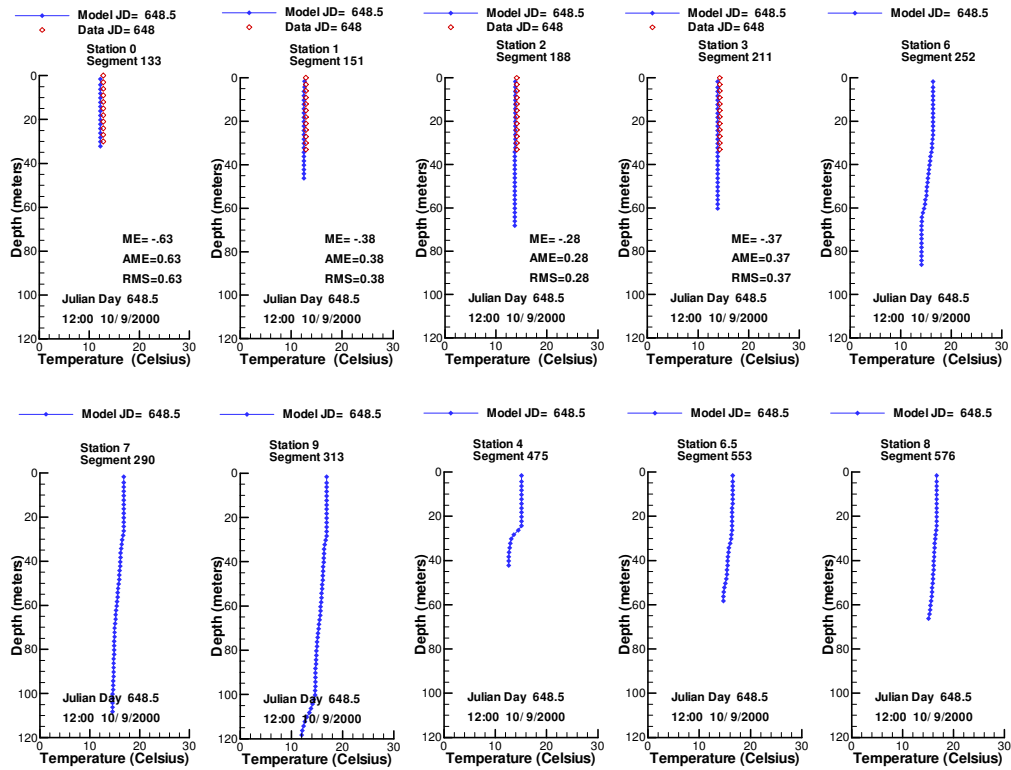


Figure 324. Vertical temperature profile model-data comparison, J648.

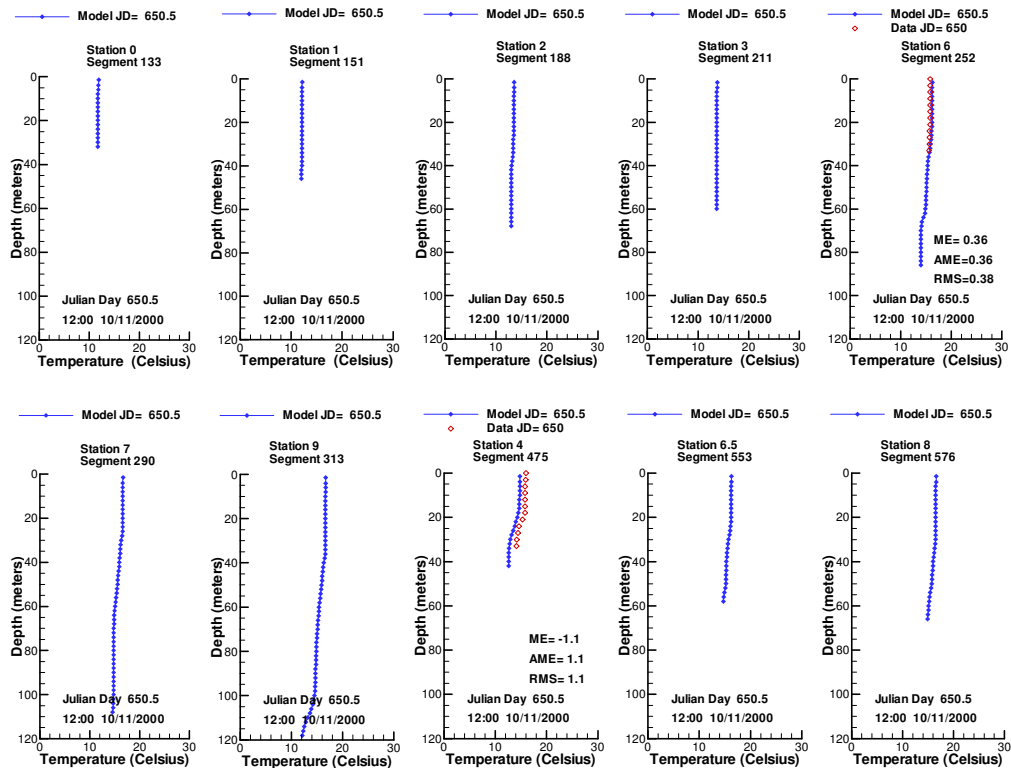


Figure 325. Vertical temperature profile model-data comparison, J650.

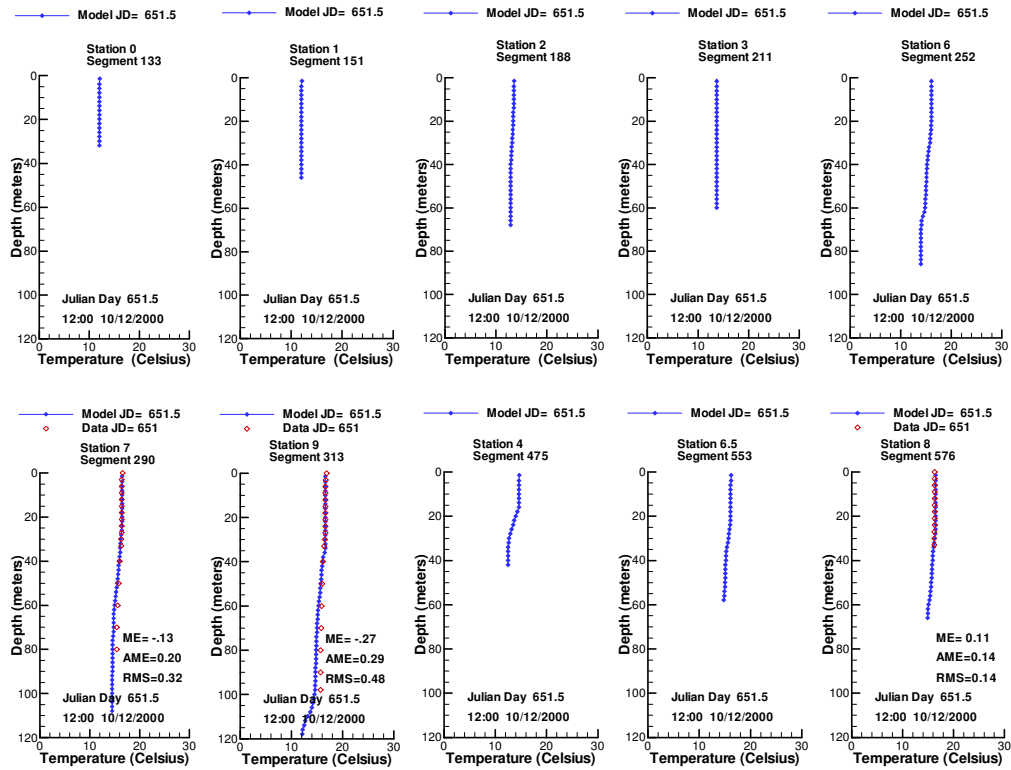


Figure 326. Vertical temperature profile model-data comparison, J651.

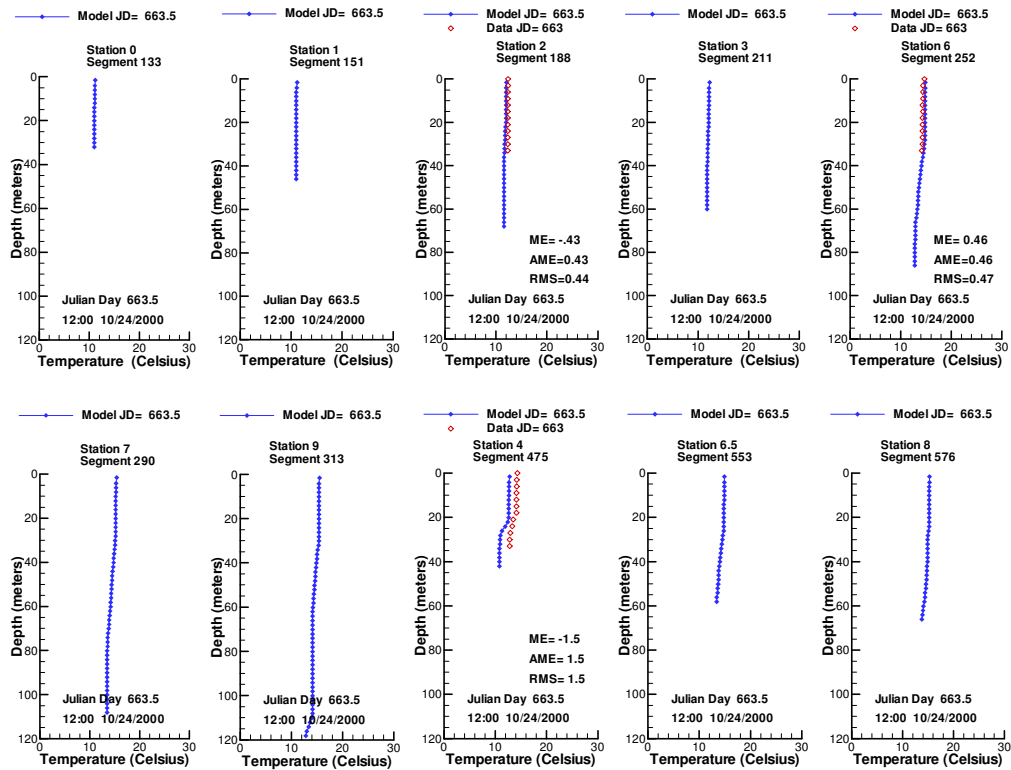


Figure 327. Vertical temperature profile model-data comparison, J663.

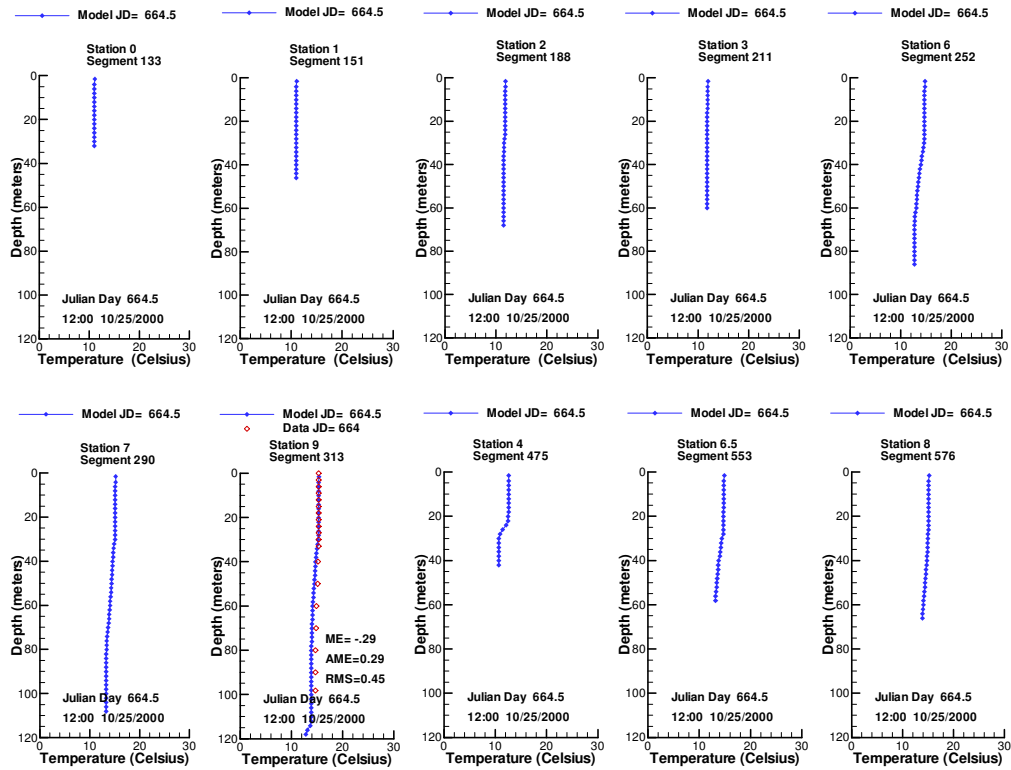


Figure 328. Vertical temperature profile model-data comparison, J664.

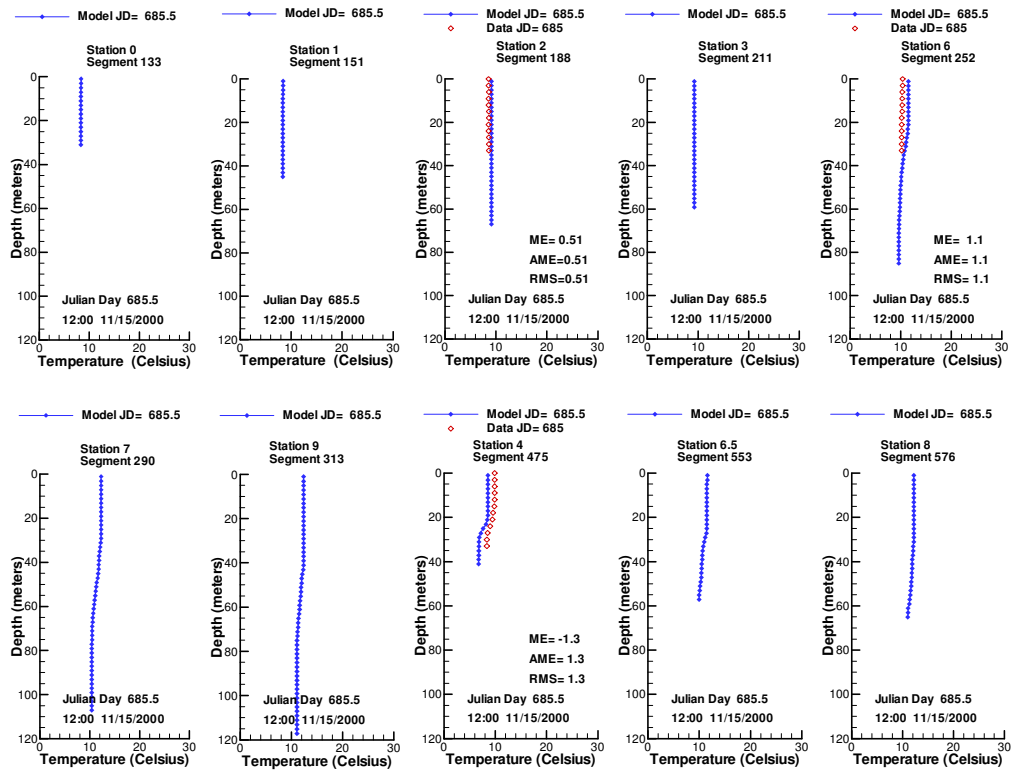


Figure 329. Vertical temperature profile model-data comparison, J685.

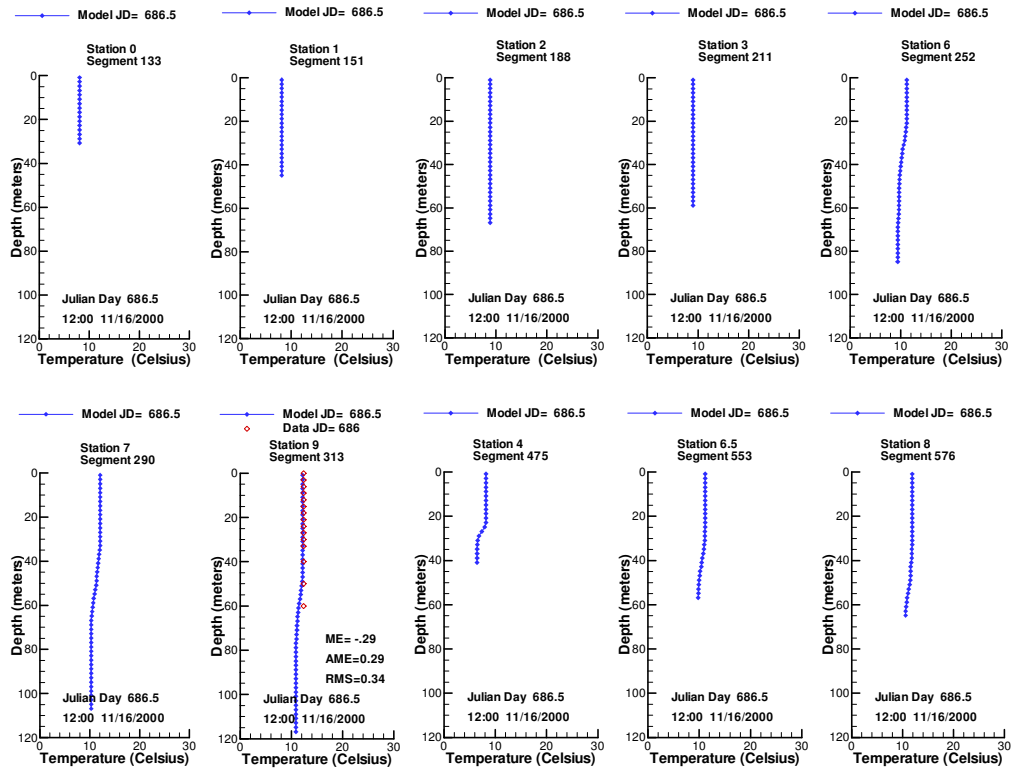


Figure 330. Vertical temperature profile model-data comparison, J686.

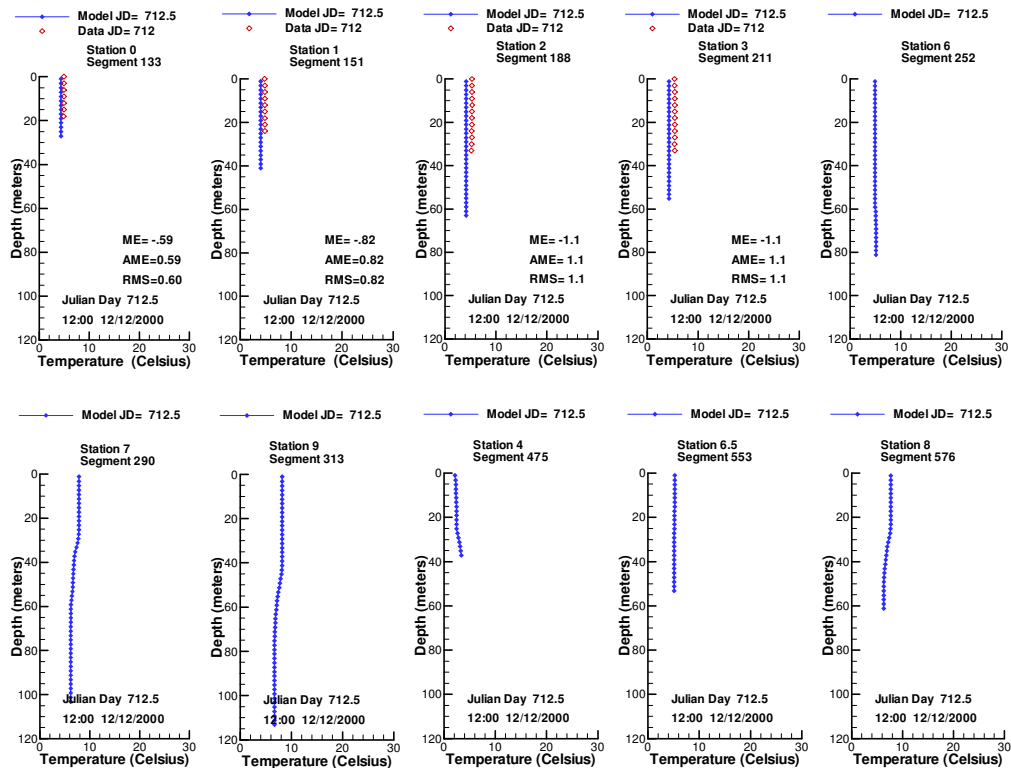


Figure 331. Vertical temperature profile model-data comparison, J712.

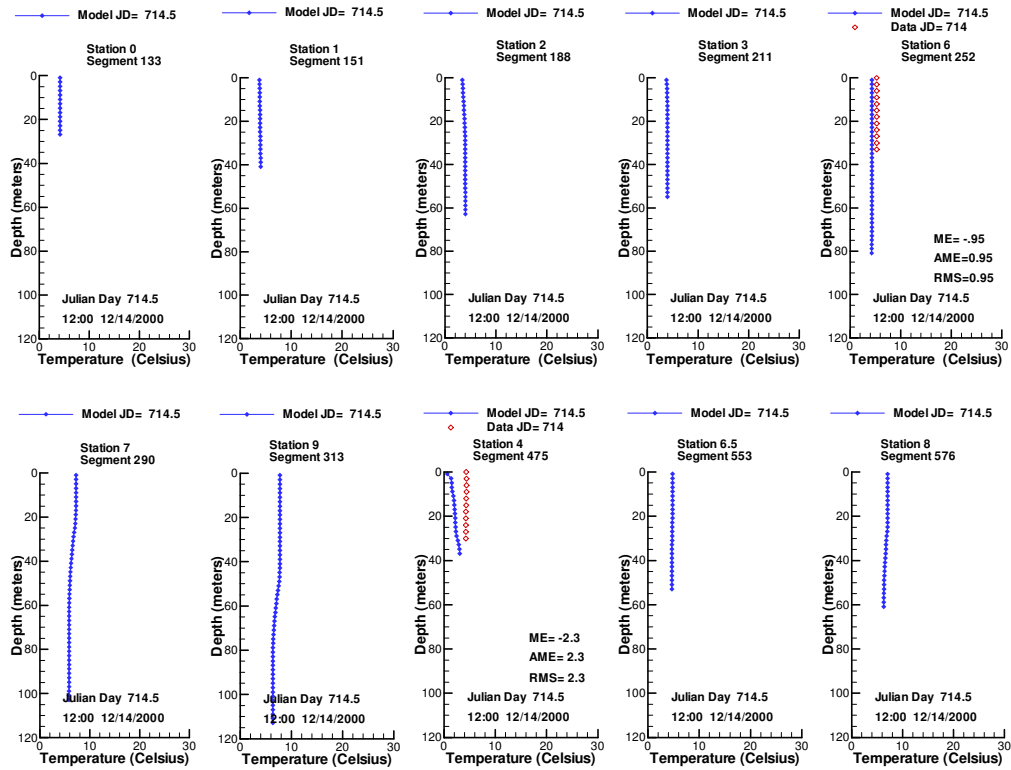


Figure 332. Vertical temperature profile model-data comparison, J714.

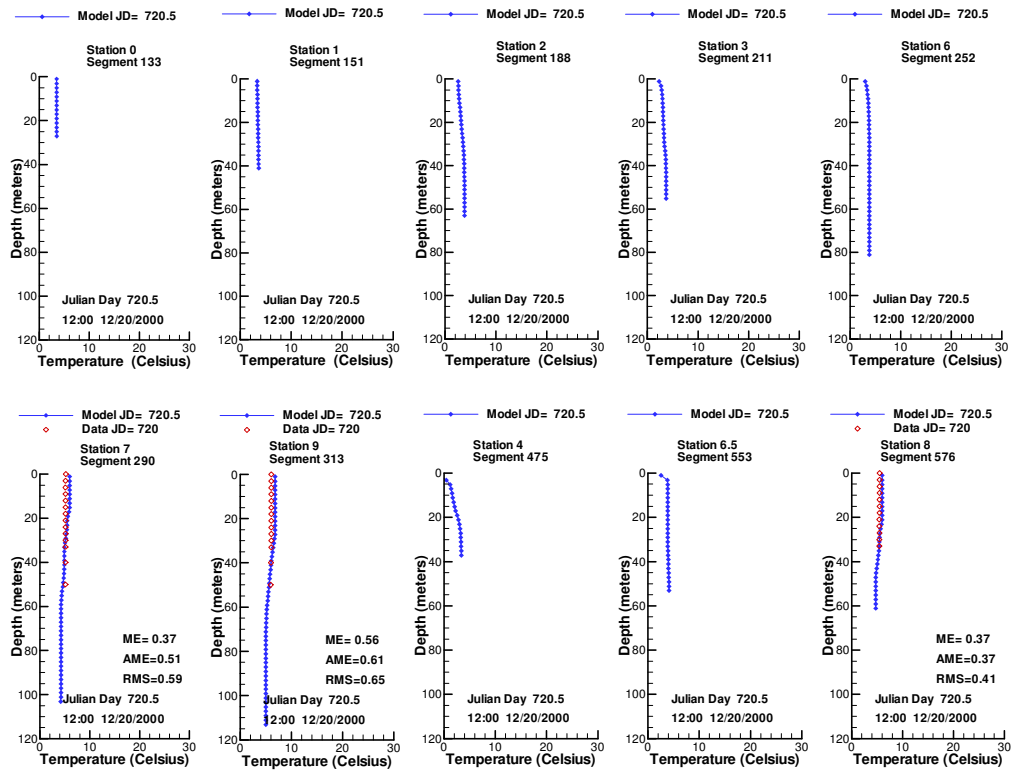


Figure 333. Vertical temperature profile model-data comparison, J720.

pH

For days when data were recorded, vertical profile plots of pH are shown in Figure 334 through Figure 380. The model-data comparison for pH is poor due to a disagreement between the boundary condition and in-stream data. Refer to the section Alternate Boundary Conditions section for a discussion.

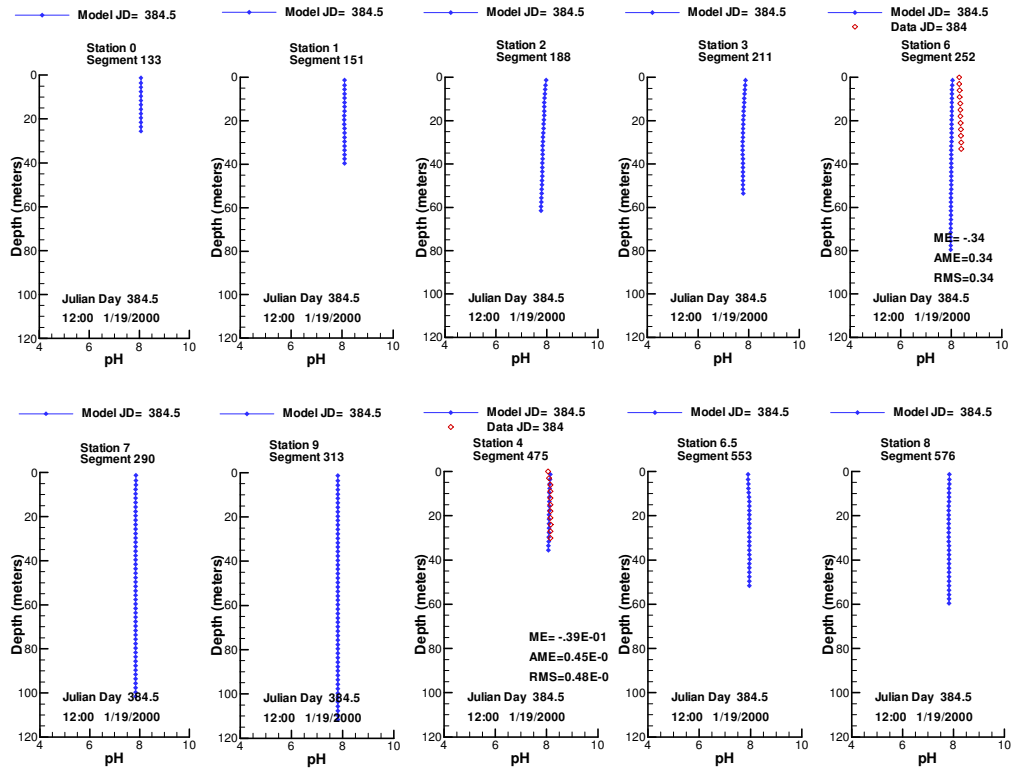


Figure 334. Vertical pH model-data comparison, J384.

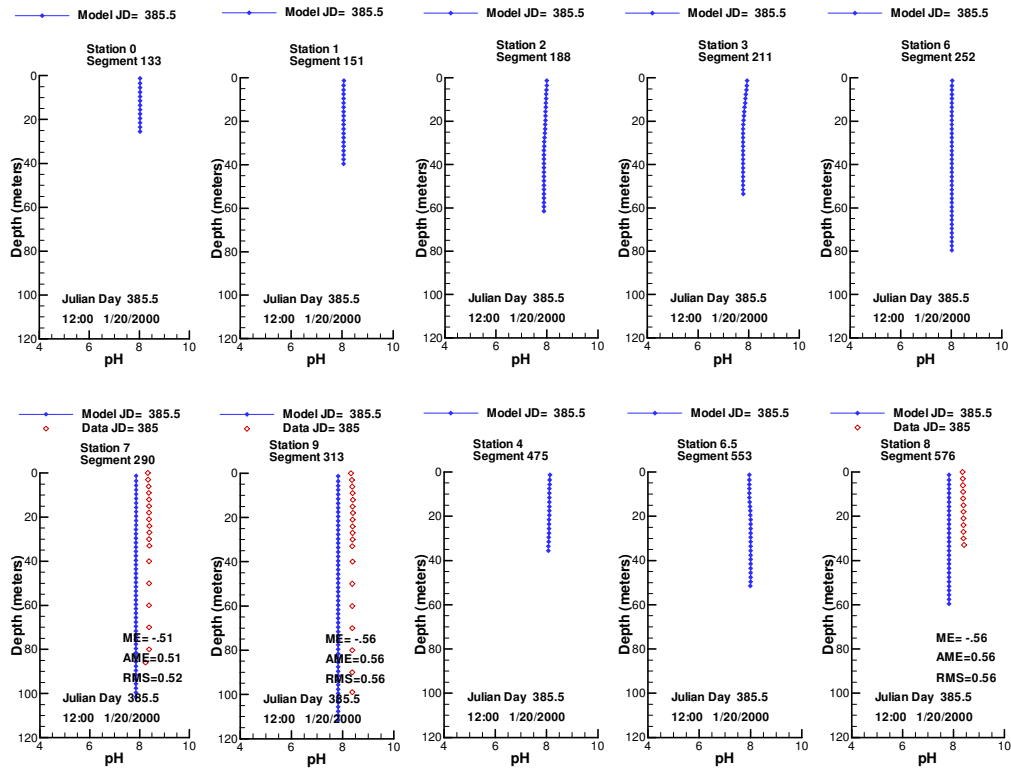


Figure 335. Vertical pH model-data comparison, J385.

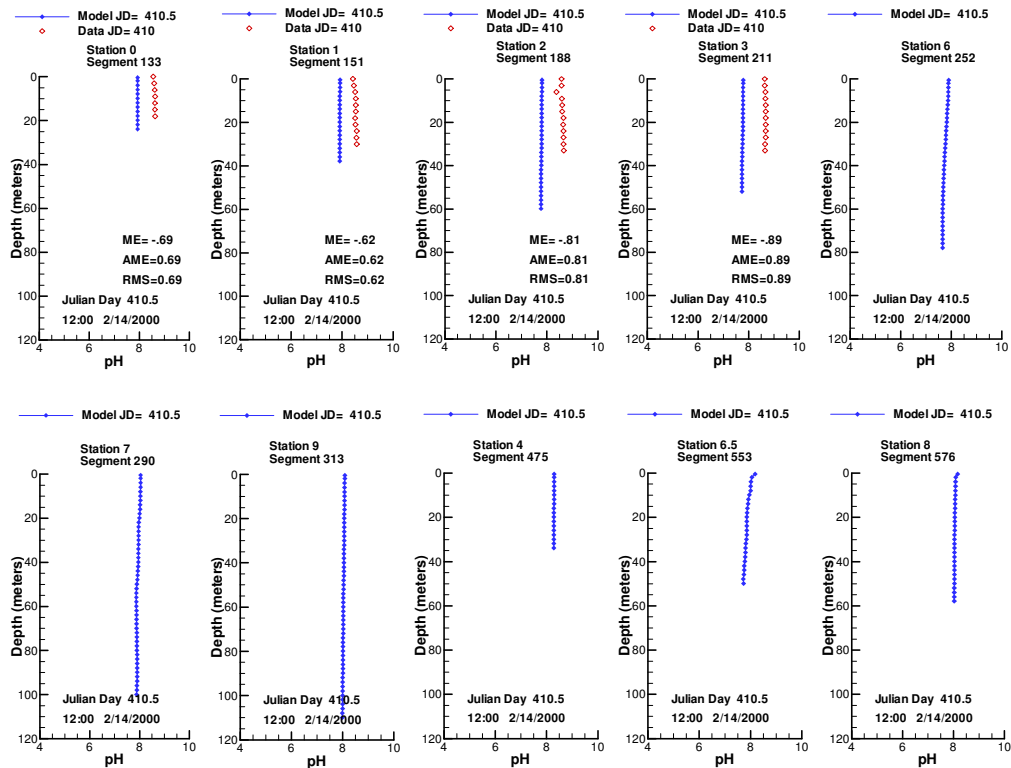


Figure 336. Vertical pH model-data comparison, J410.

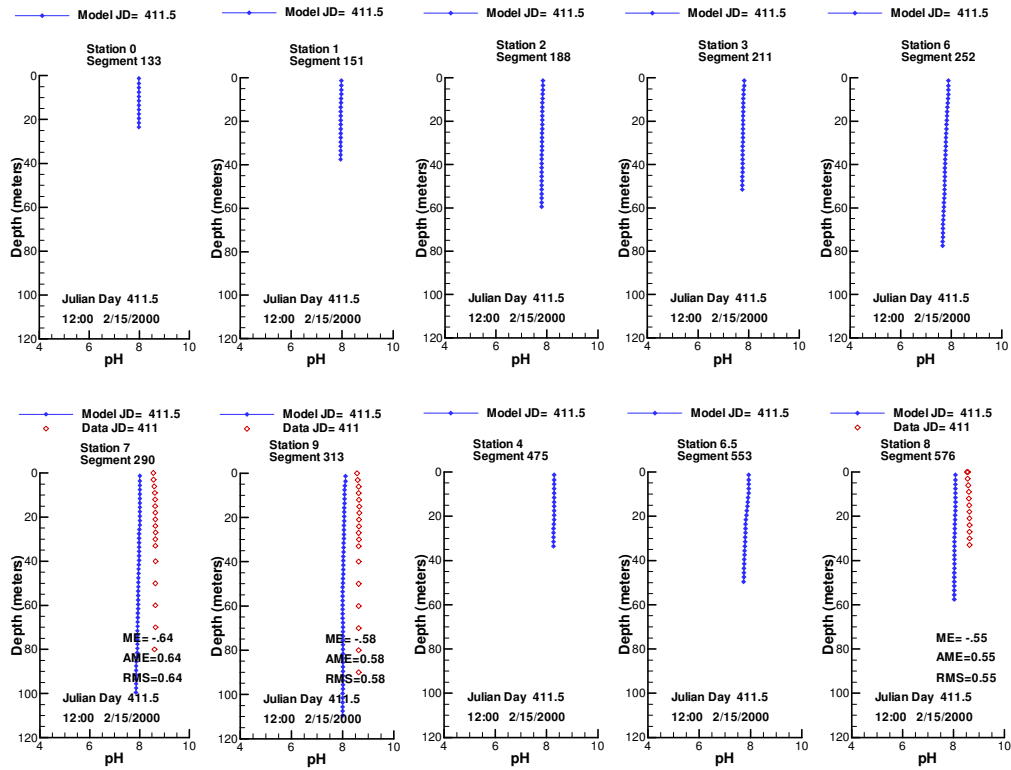


Figure 337. Vertical pH model-data comparison, J411.

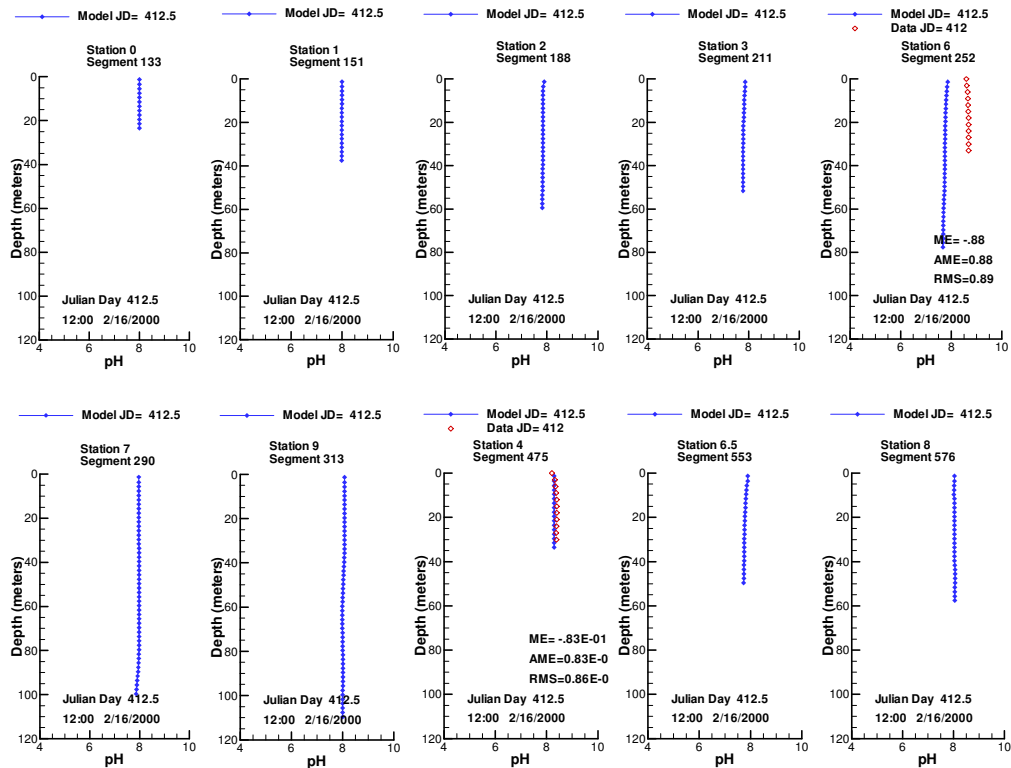


Figure 338. Vertical pH model-data comparison, J412.

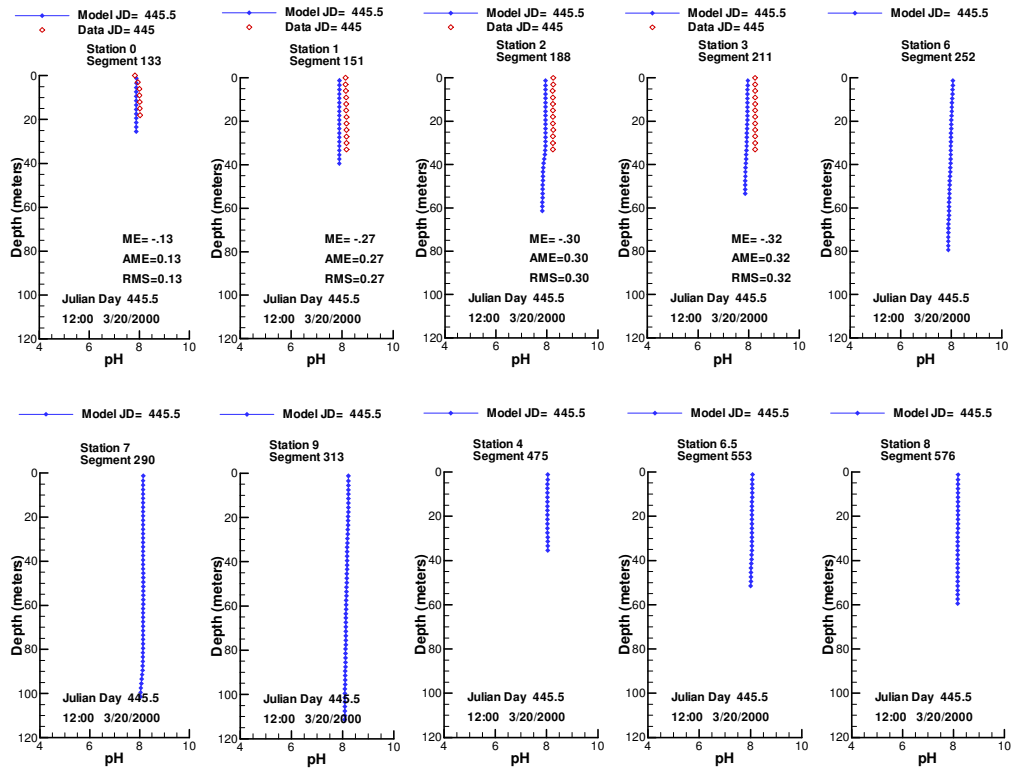


Figure 339. Vertical pH model-data comparison, J445.

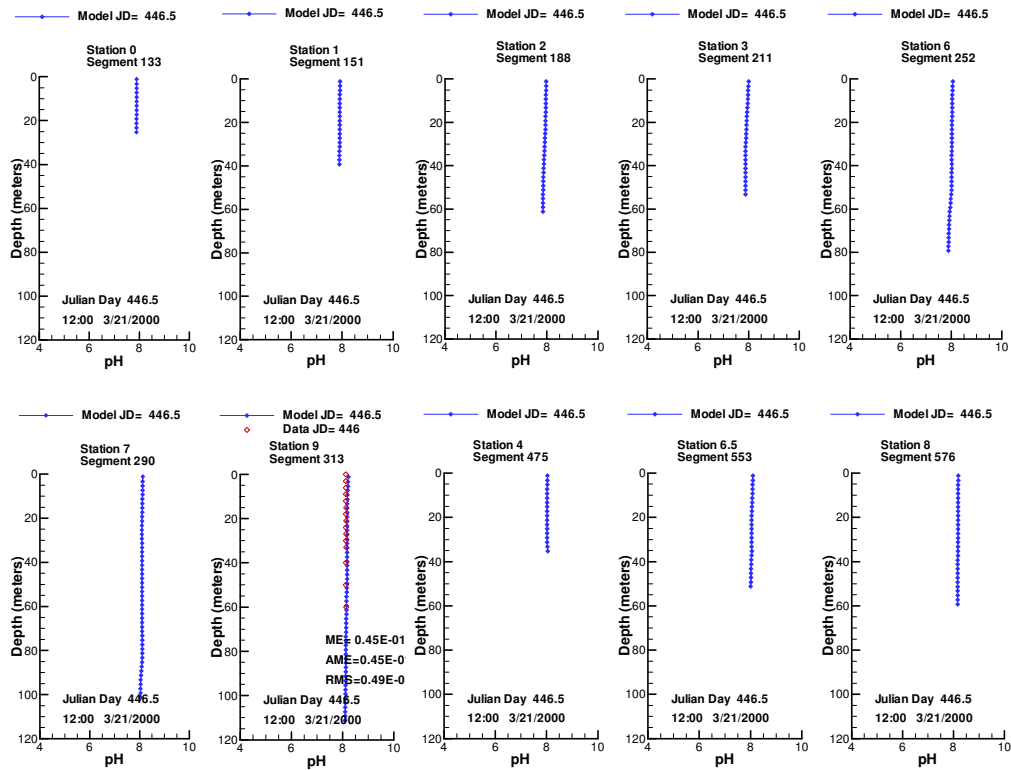


Figure 340. Vertical pH model-data comparison, J446.

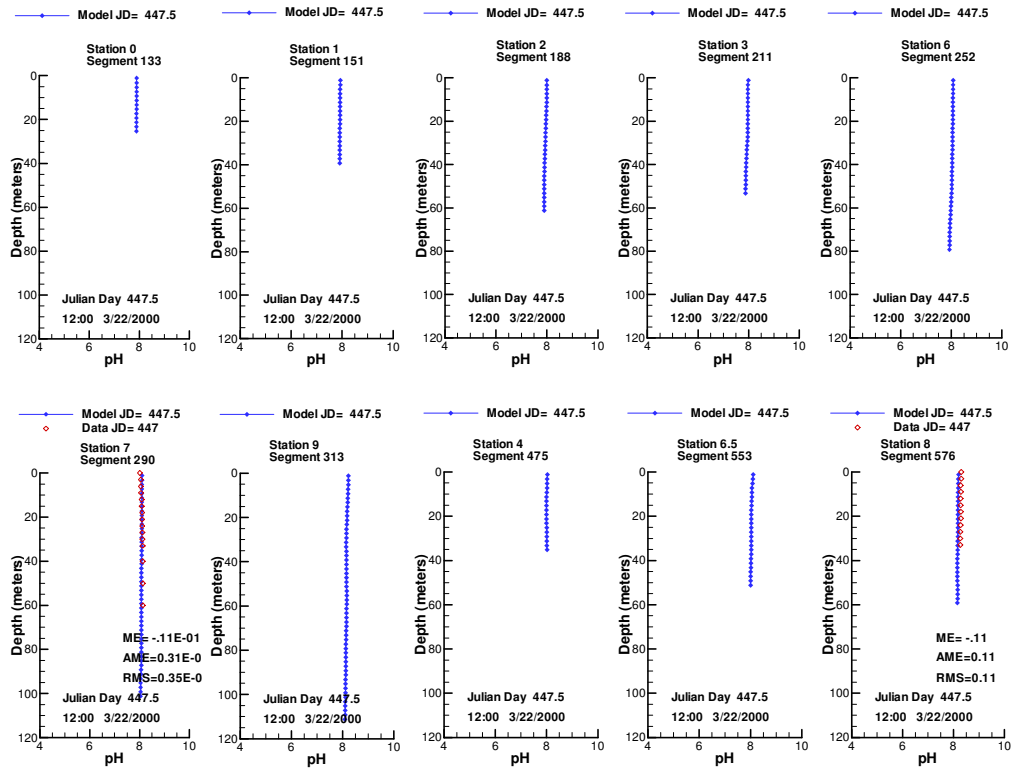


Figure 341. Vertical pH model-data comparison, J447.

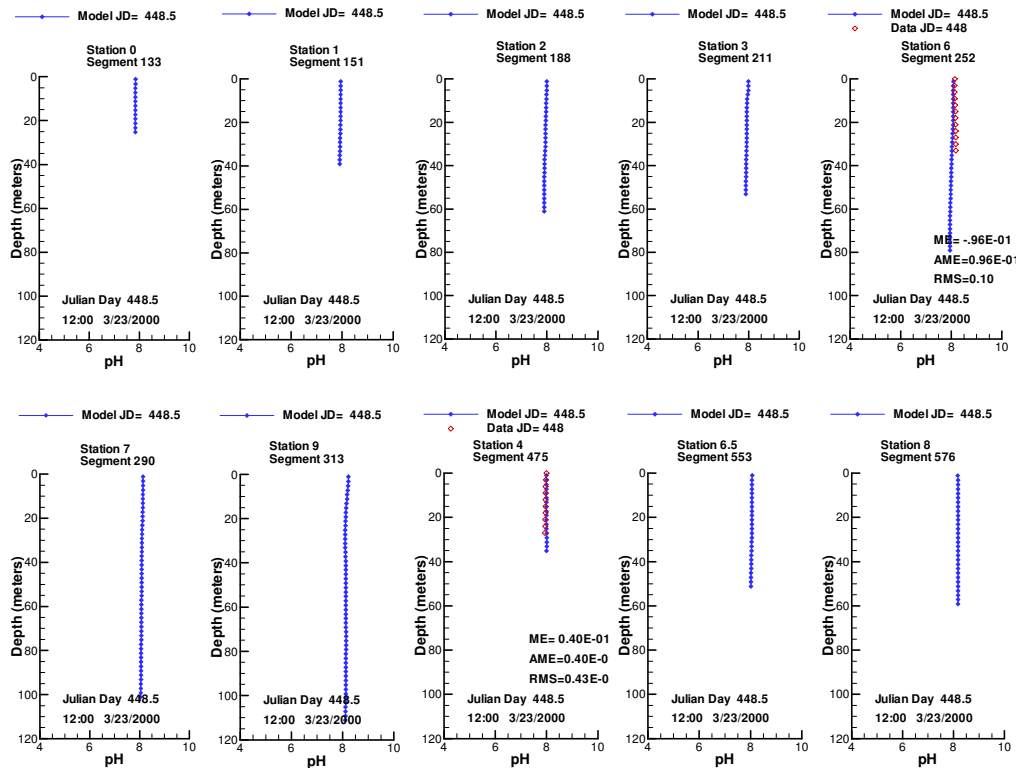


Figure 342. Vertical pH model-data comparison, J448.

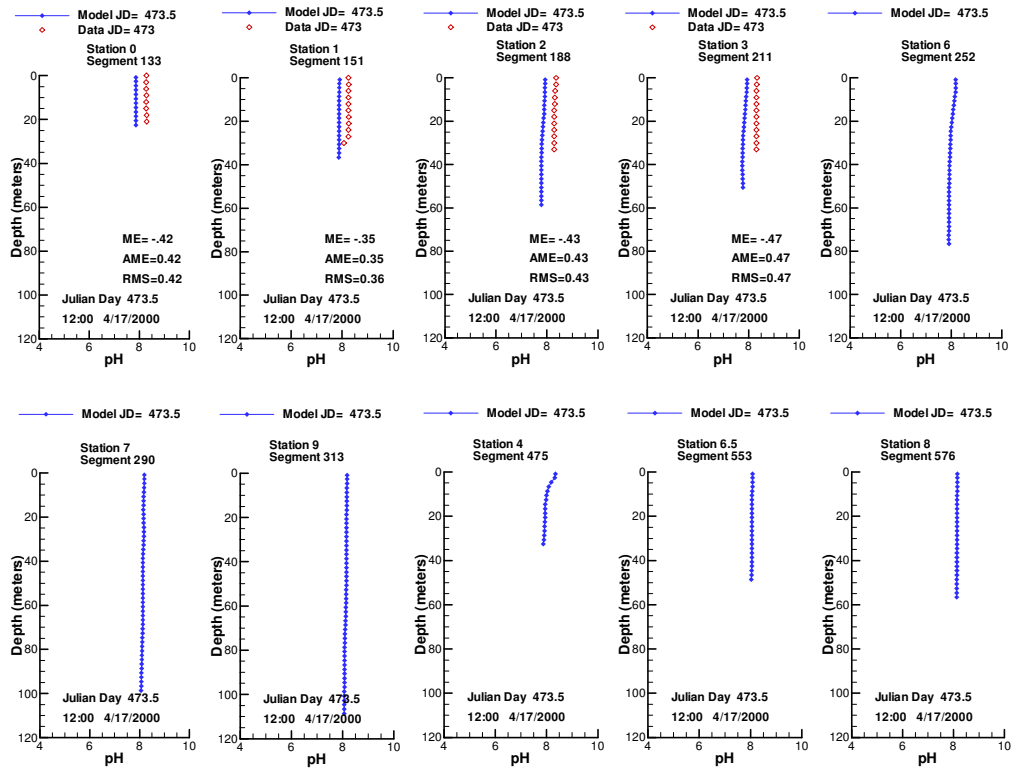


Figure 343. Vertical pH model-data comparison, J473.

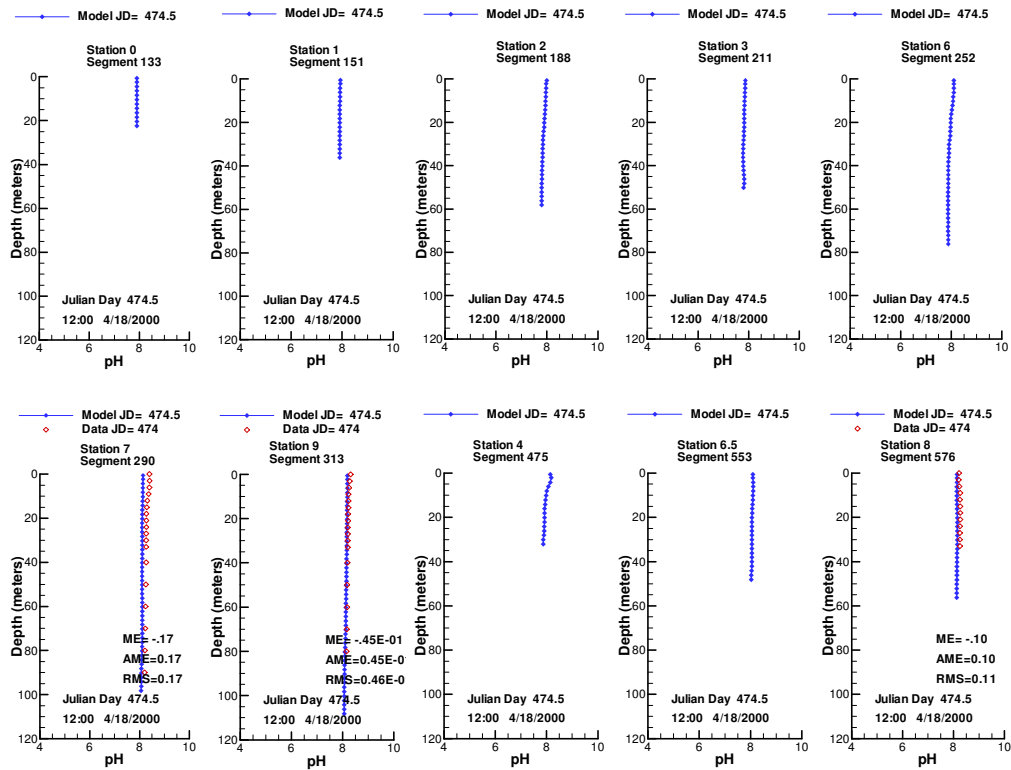


Figure 344. Vertical pH model-data comparison, J474.

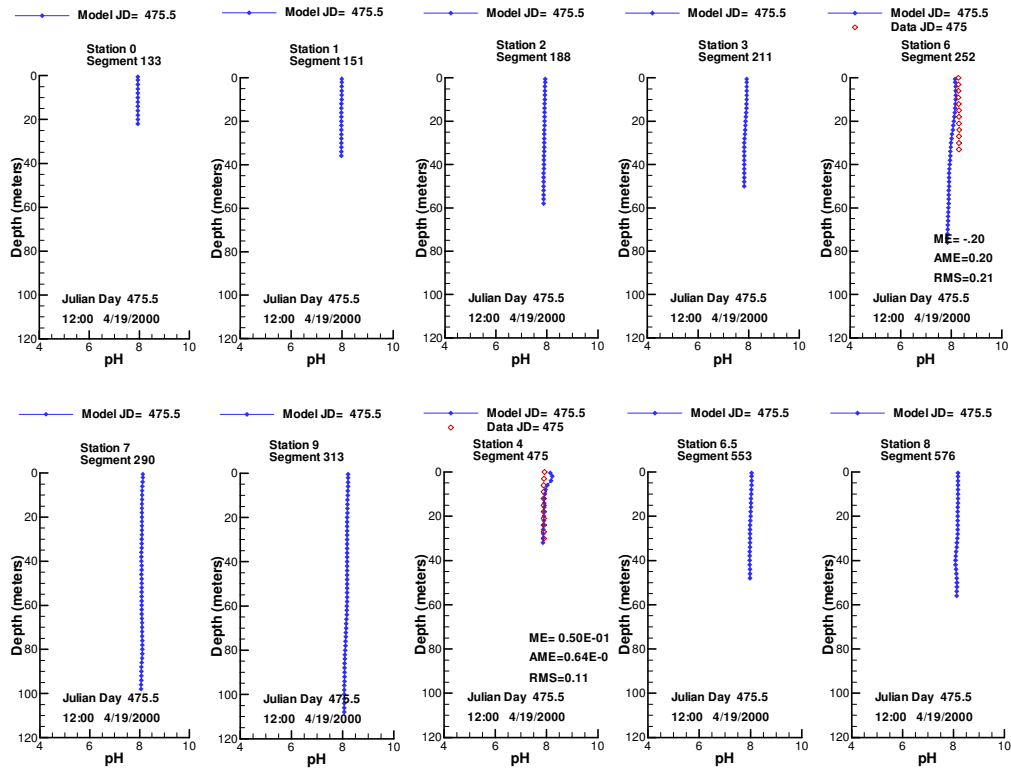


Figure 345. Vertical pH model-data comparison, J475.

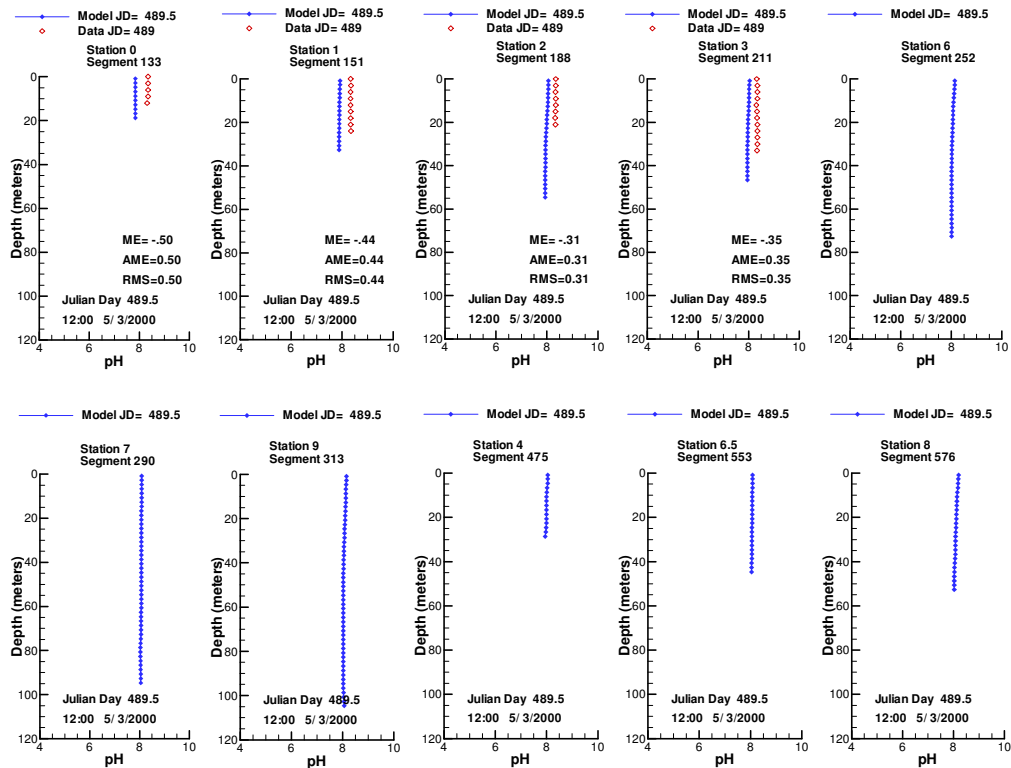


Figure 346. Vertical pH model-data comparison, J489.

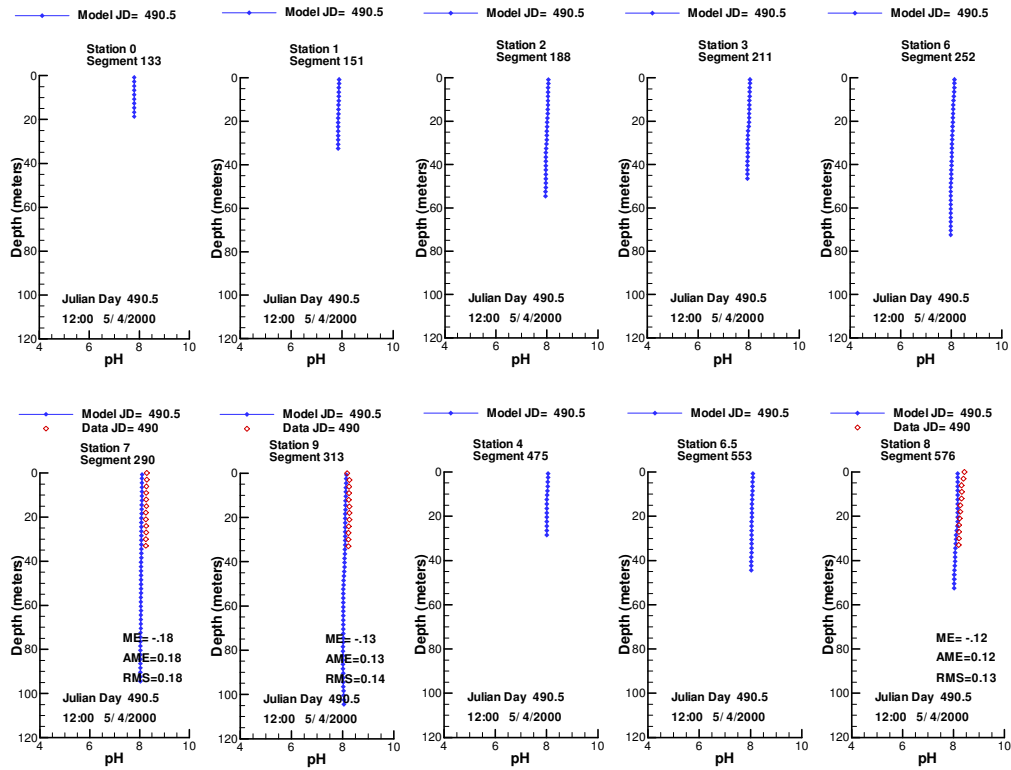


Figure 347. Vertical pH model-data comparison, J490.

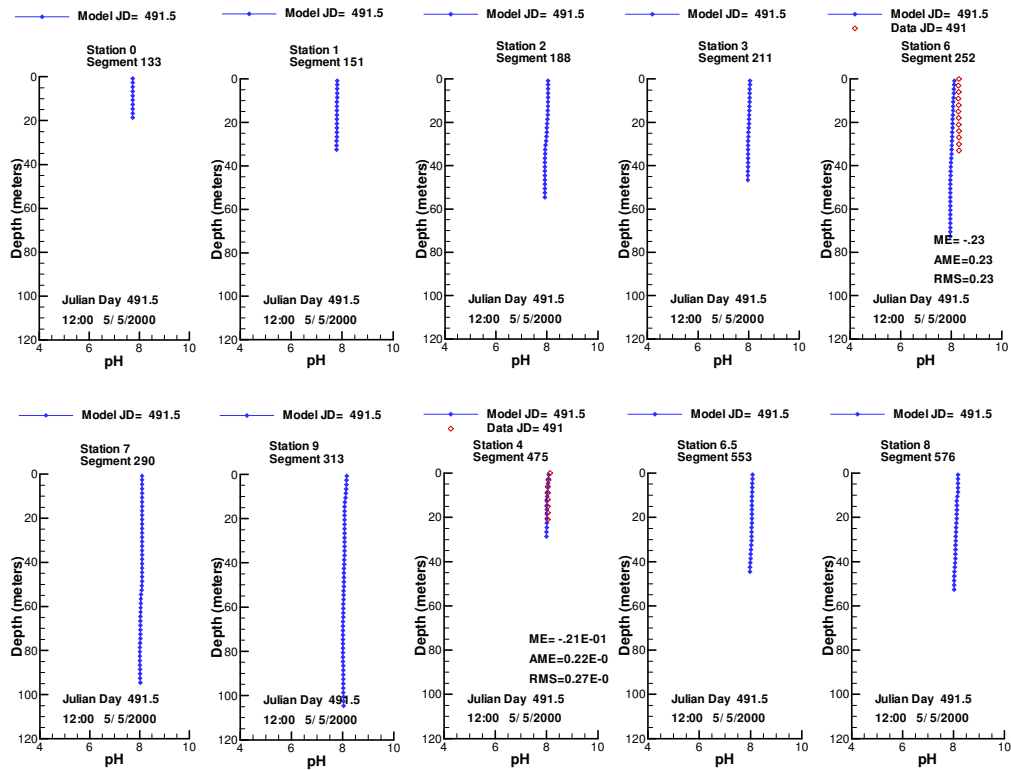


Figure 348. Vertical pH model-data comparison, J491.

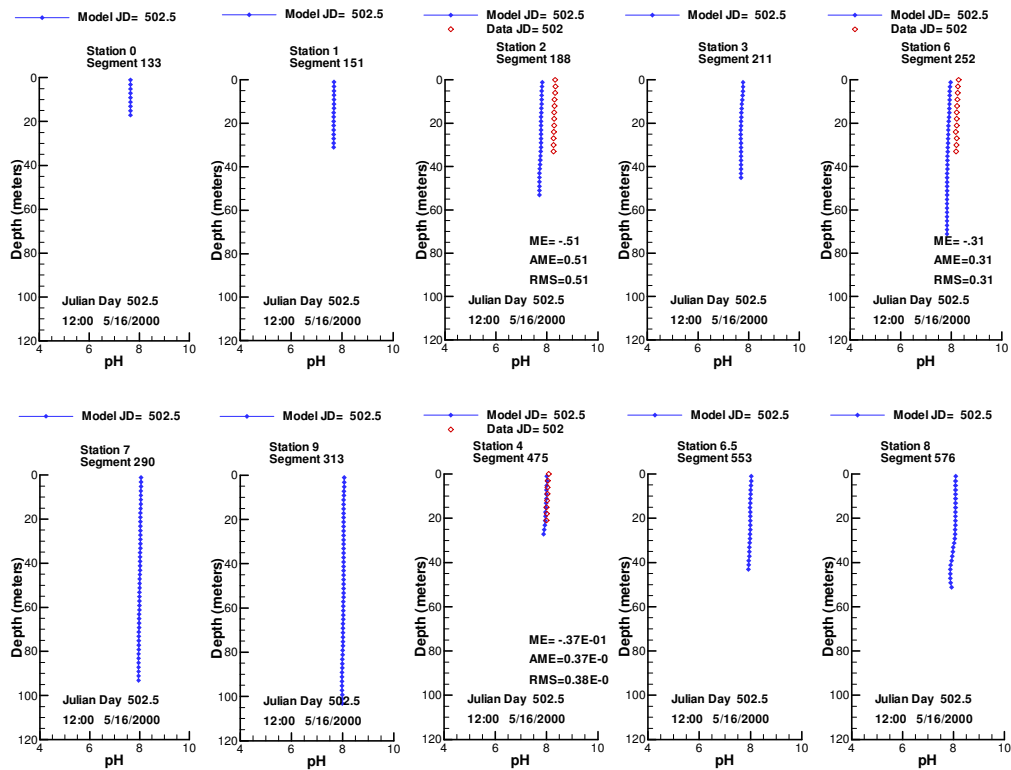


Figure 349. Vertical pH model-data comparison, J502.

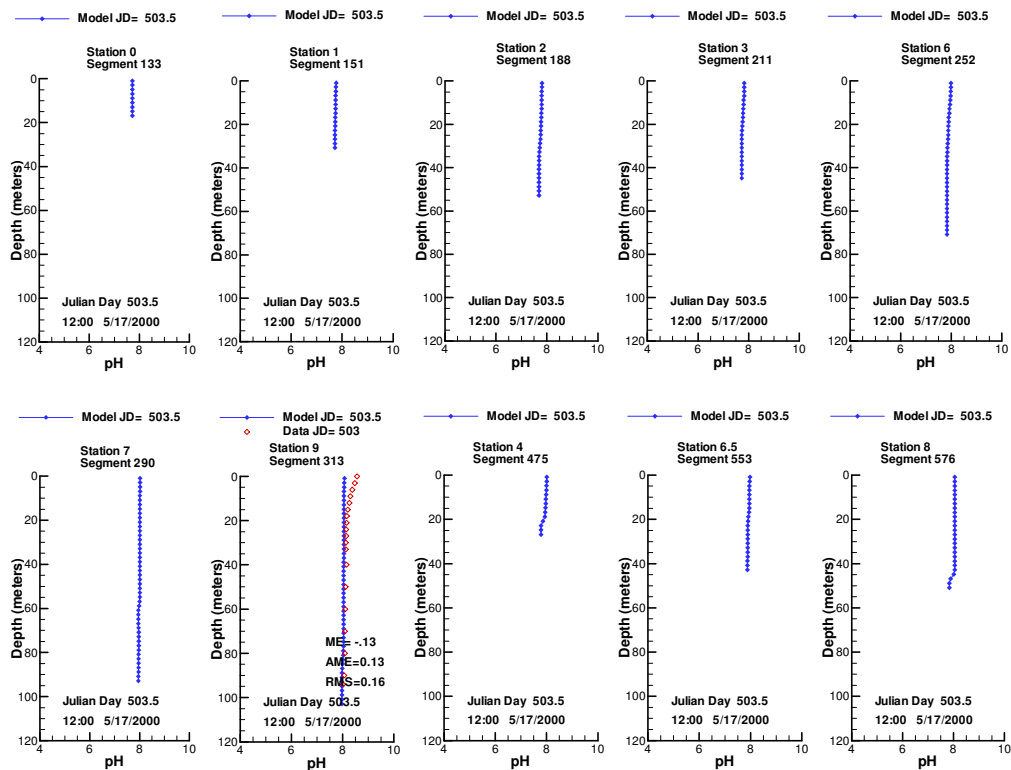


Figure 350. Vertical pH model-data comparison, J503.

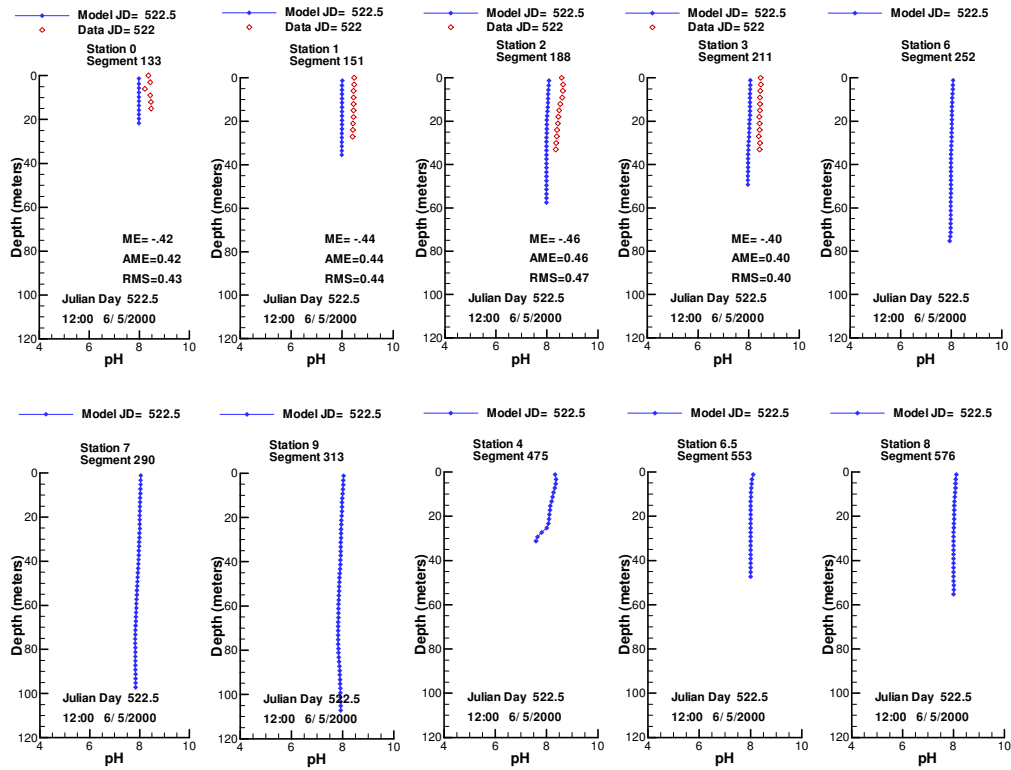


Figure 351. Vertical pH model-data comparison, J522.

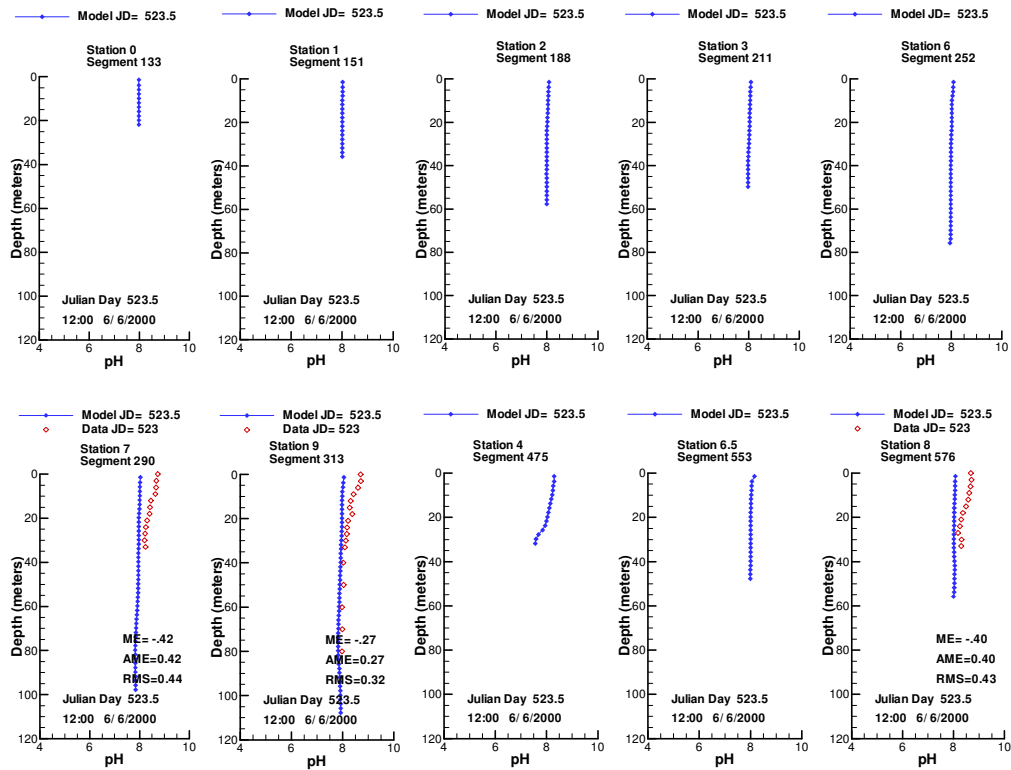


Figure 352. Vertical pH model-data comparison, J523.

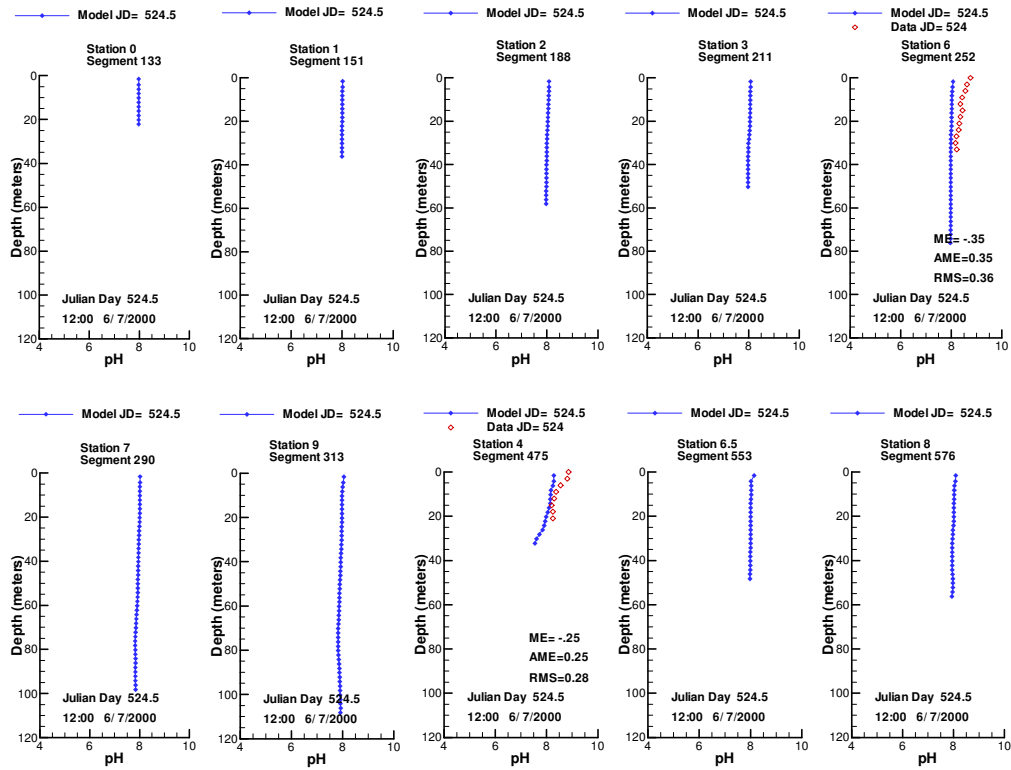


Figure 353. Vertical pH model-data comparison, J524.

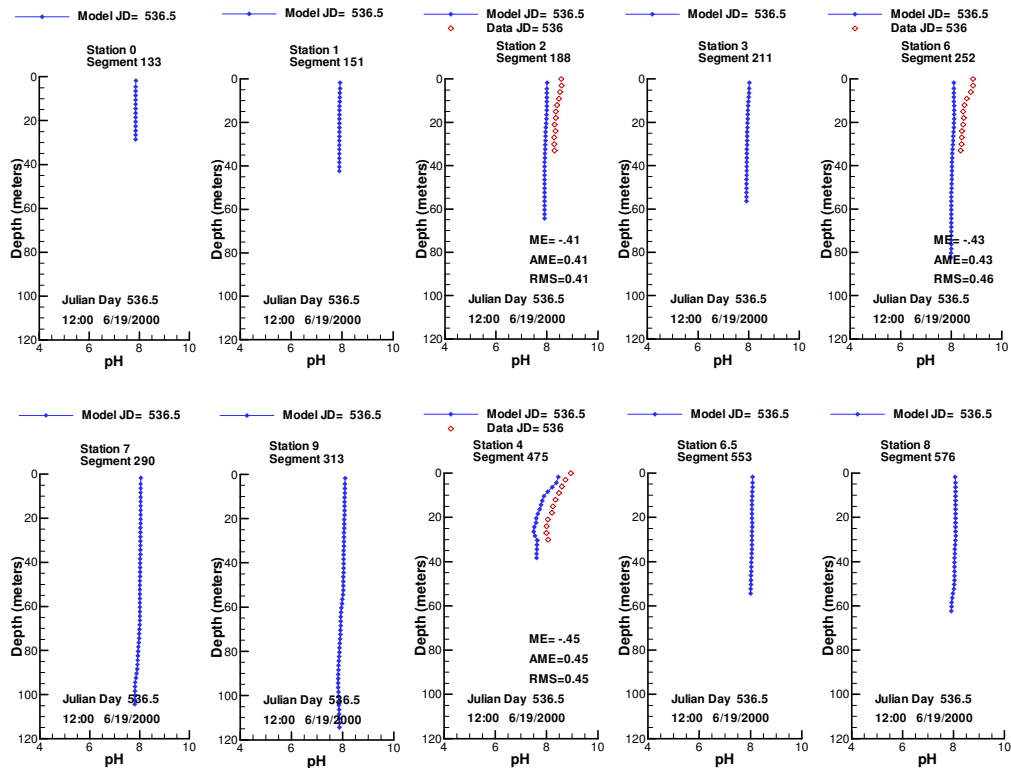


Figure 354. Vertical pH model-data comparison, J536.

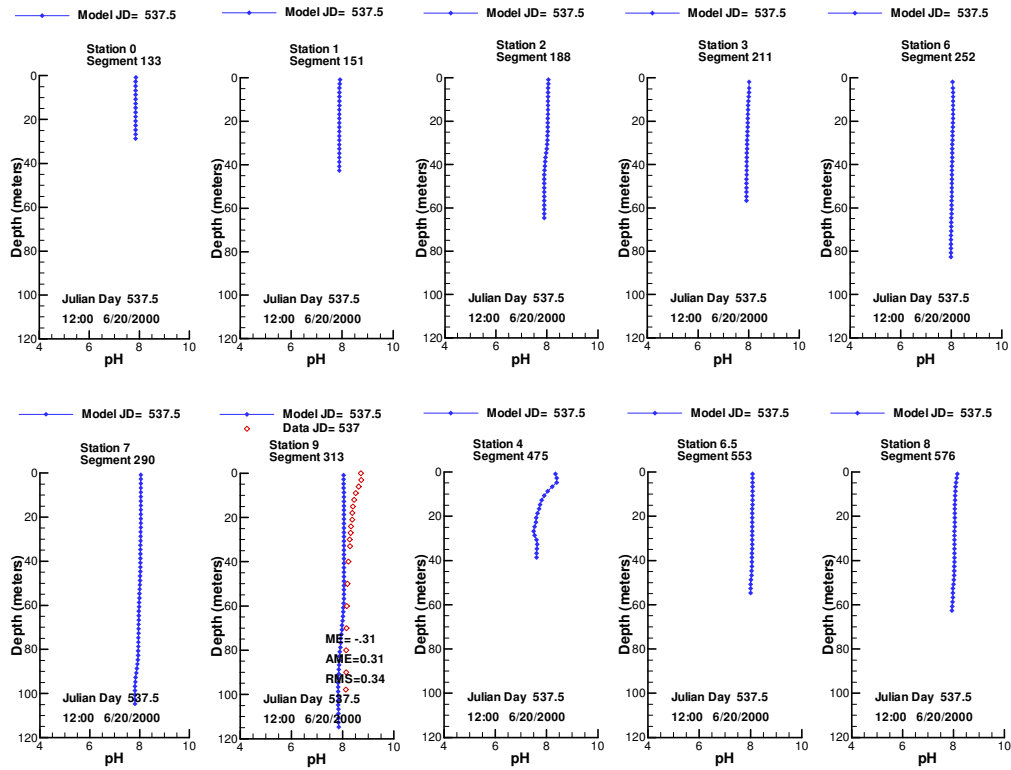


Figure 355. Vertical pH model-data comparison, J537.

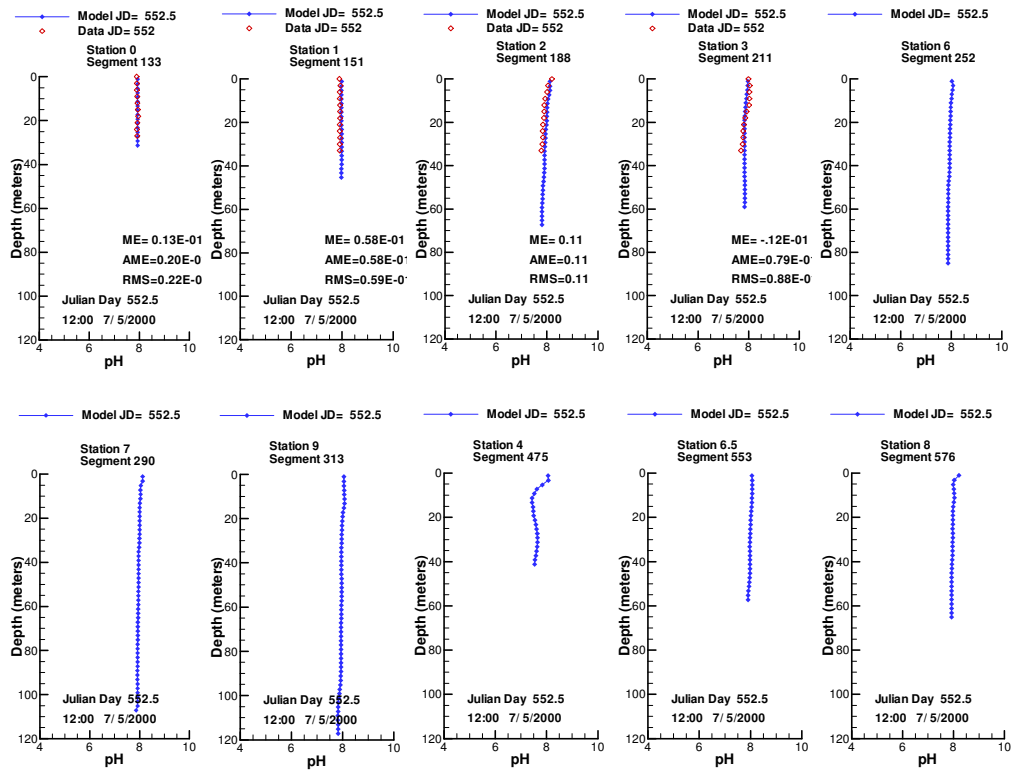


Figure 356. Vertical pH model-data comparison, J552.

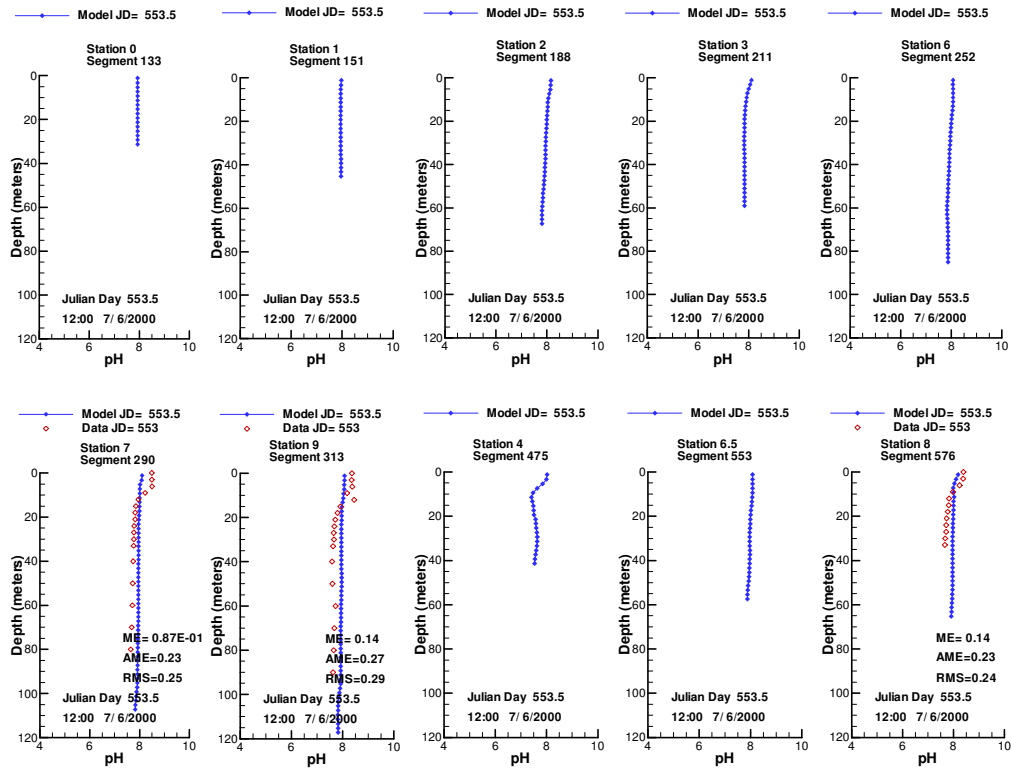


Figure 357. Vertical pH model-data comparison, J553.

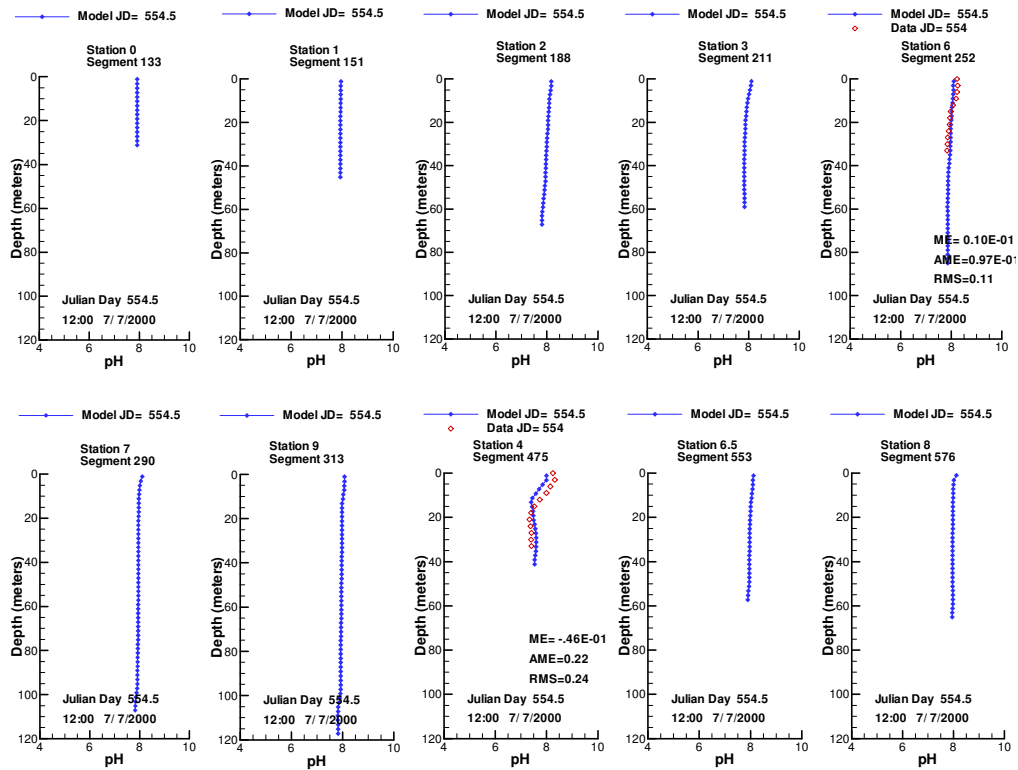


Figure 358. Vertical pH model-data comparison, J554.

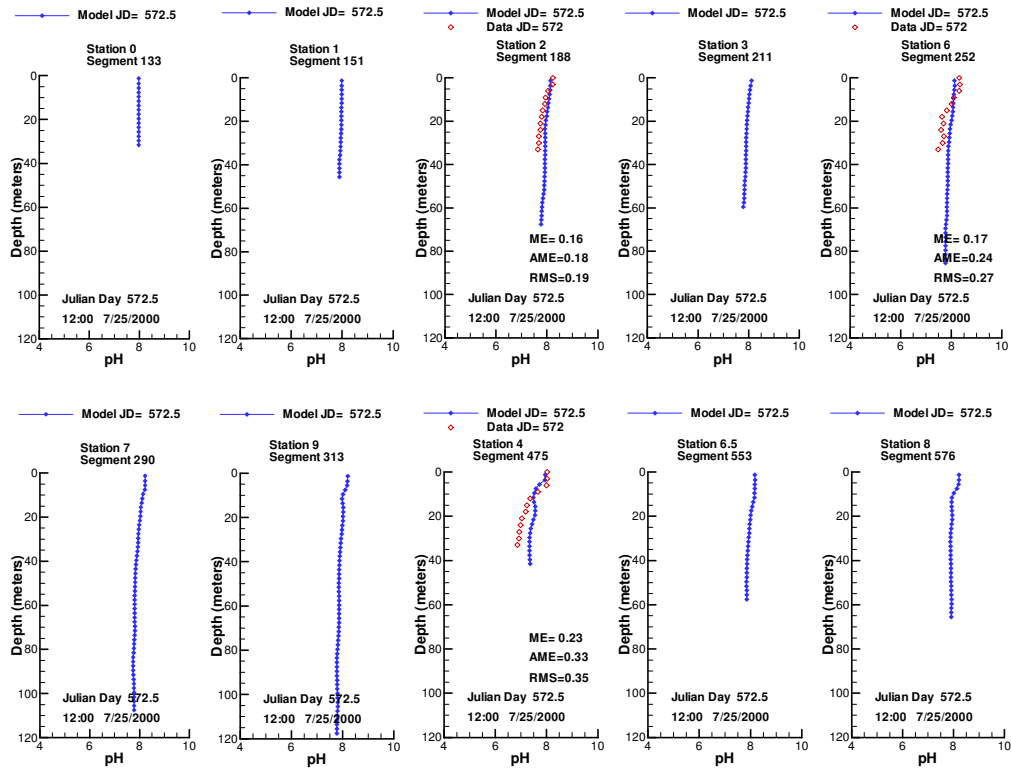


Figure 359. Vertical pH model-data comparison, J572.

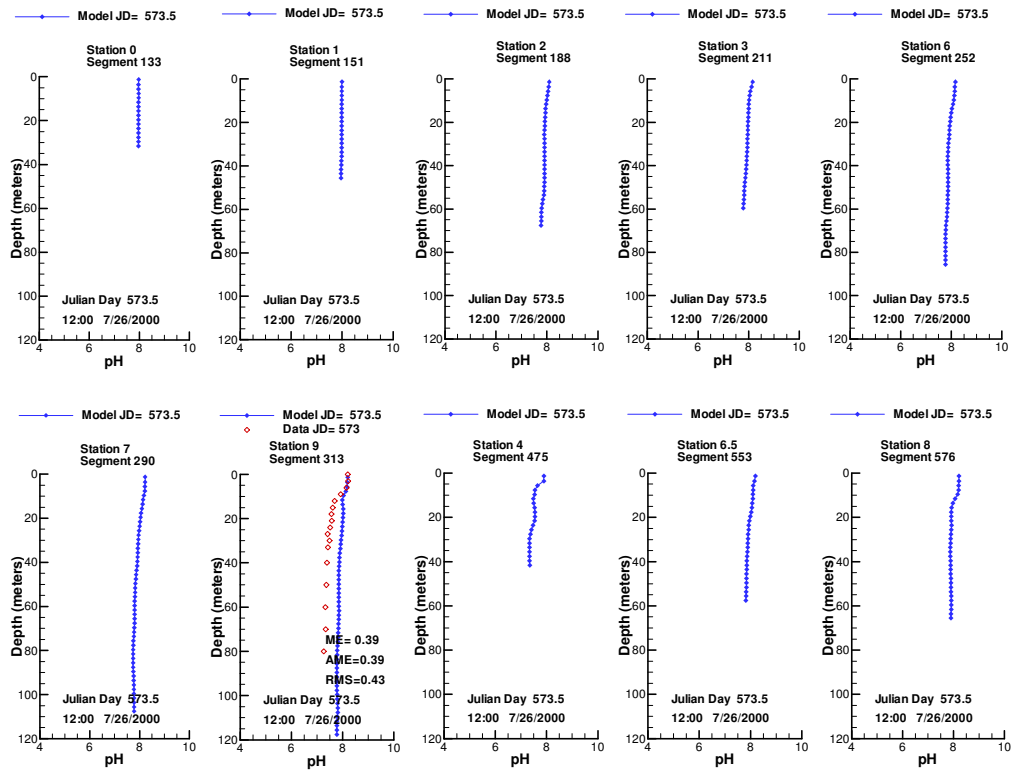


Figure 360. Vertical pH model-data comparison, J573.

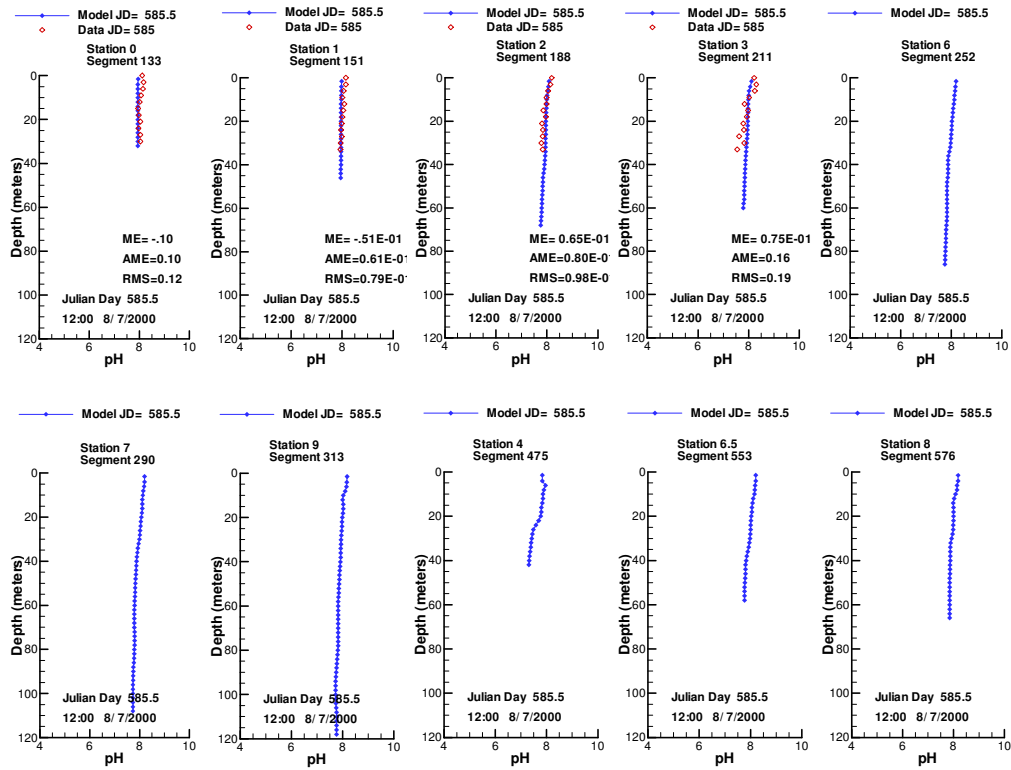


Figure 361. Vertical pH model-data comparison, J585.

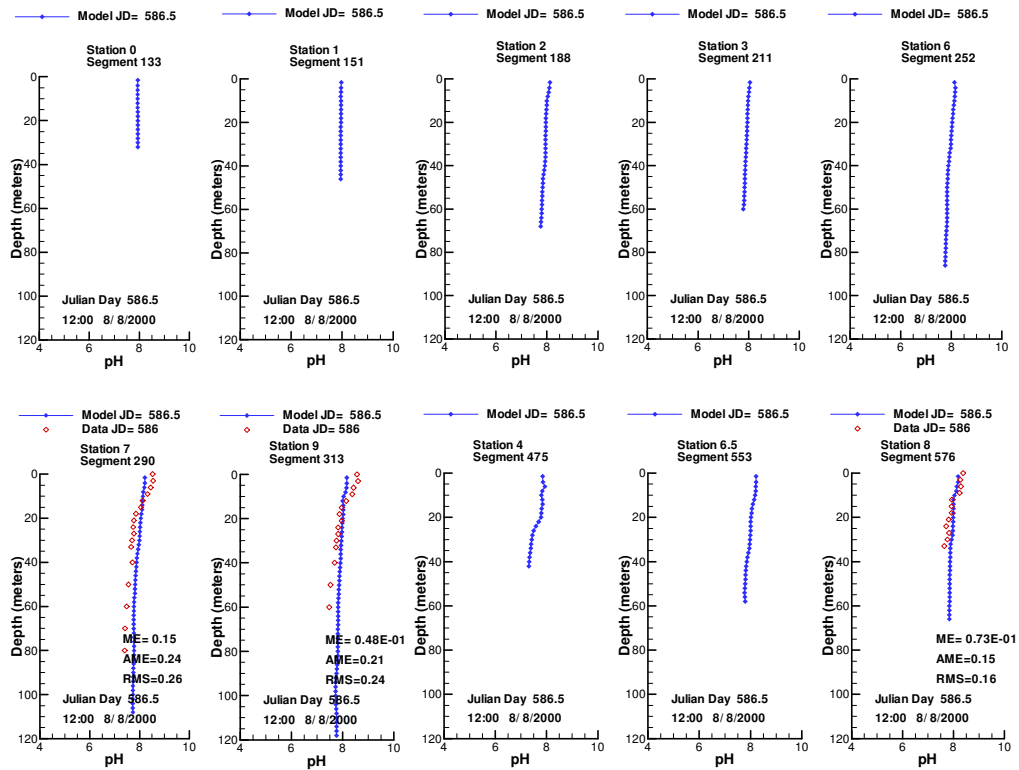


Figure 362. Vertical pH model-data comparison, J586.

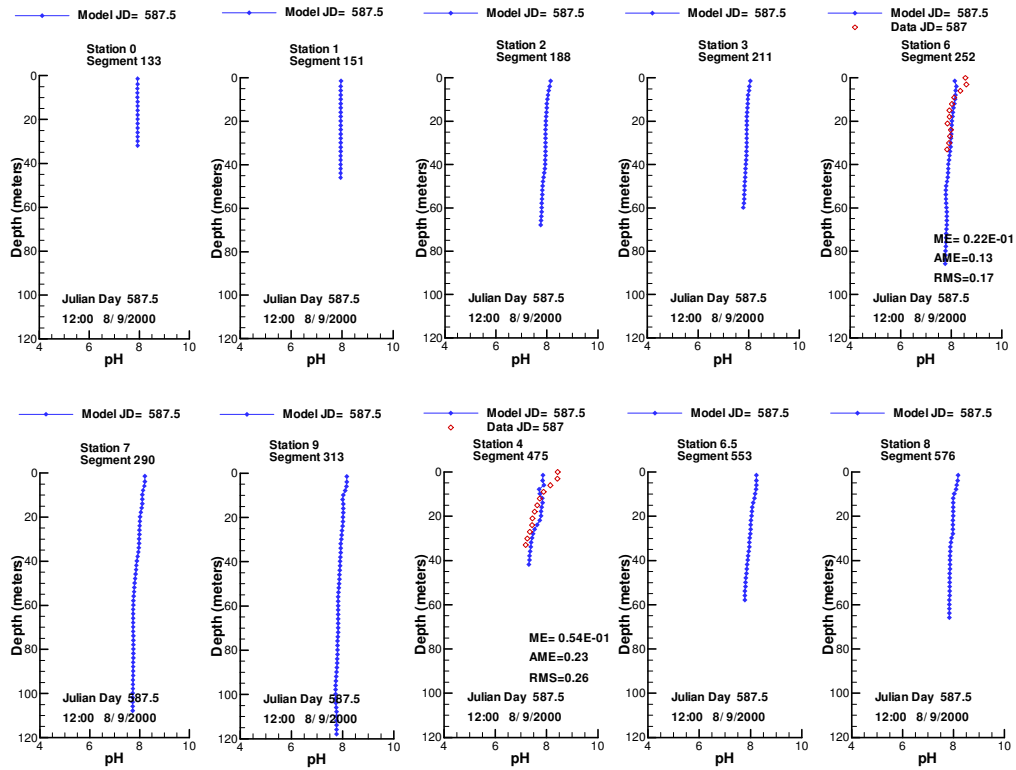


Figure 363. Vertical pH model-data comparison, J587.

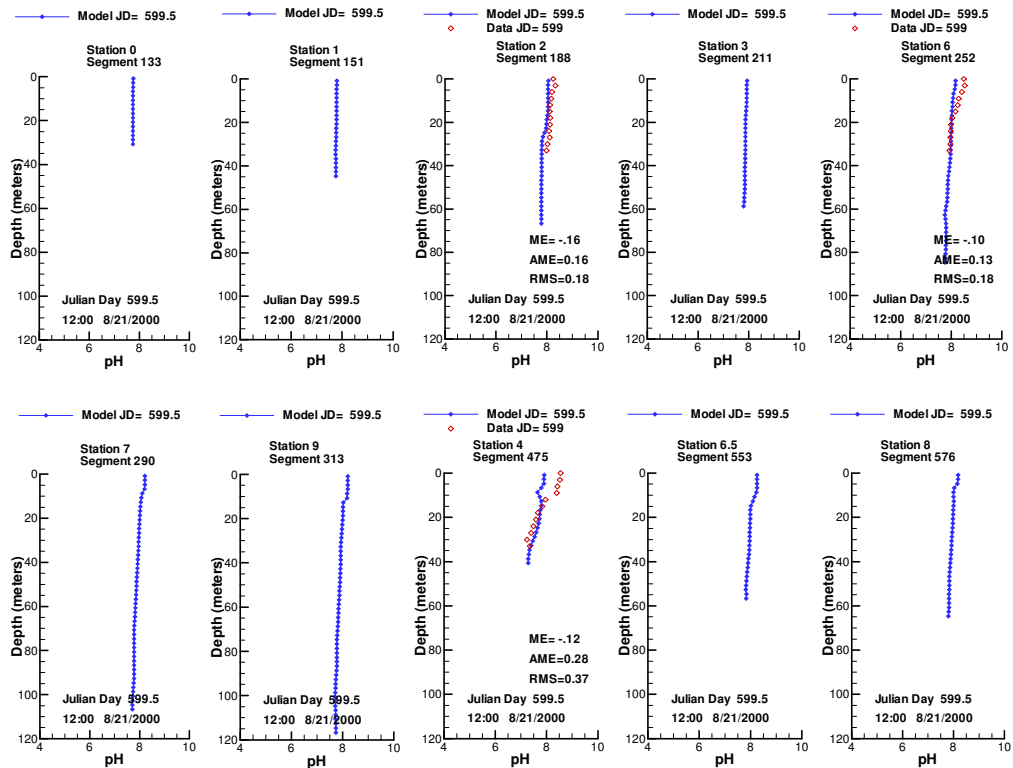


Figure 364. Vertical pH model-data comparison, J599.

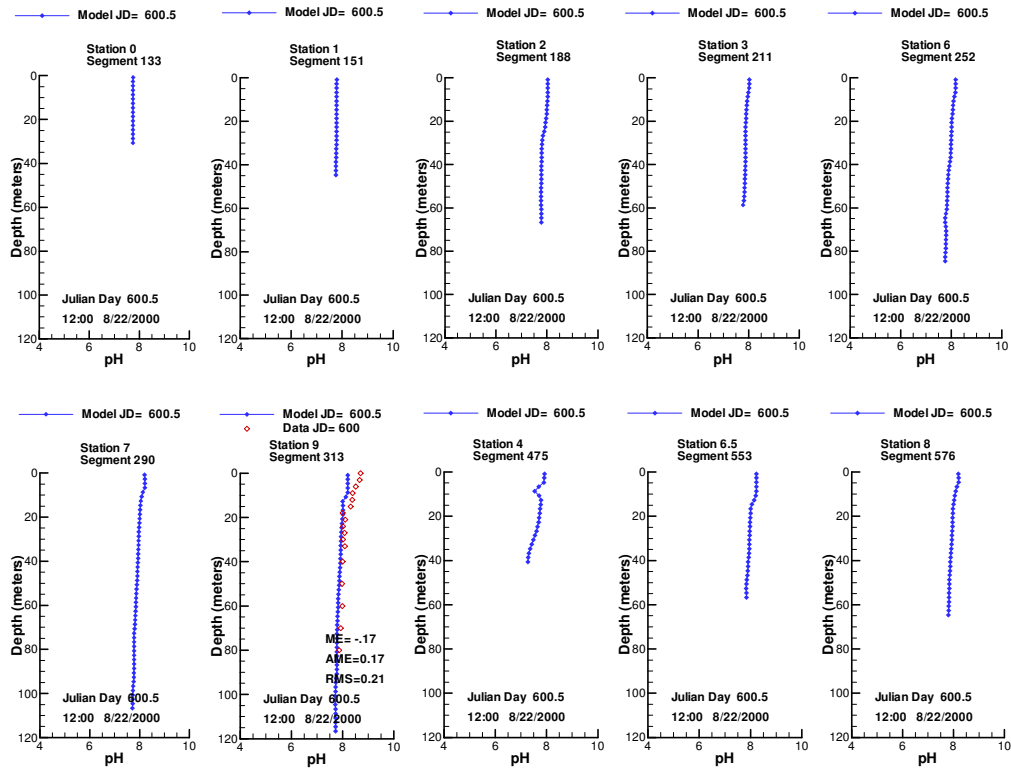


Figure 365. Vertical pH model-data comparison, J600.

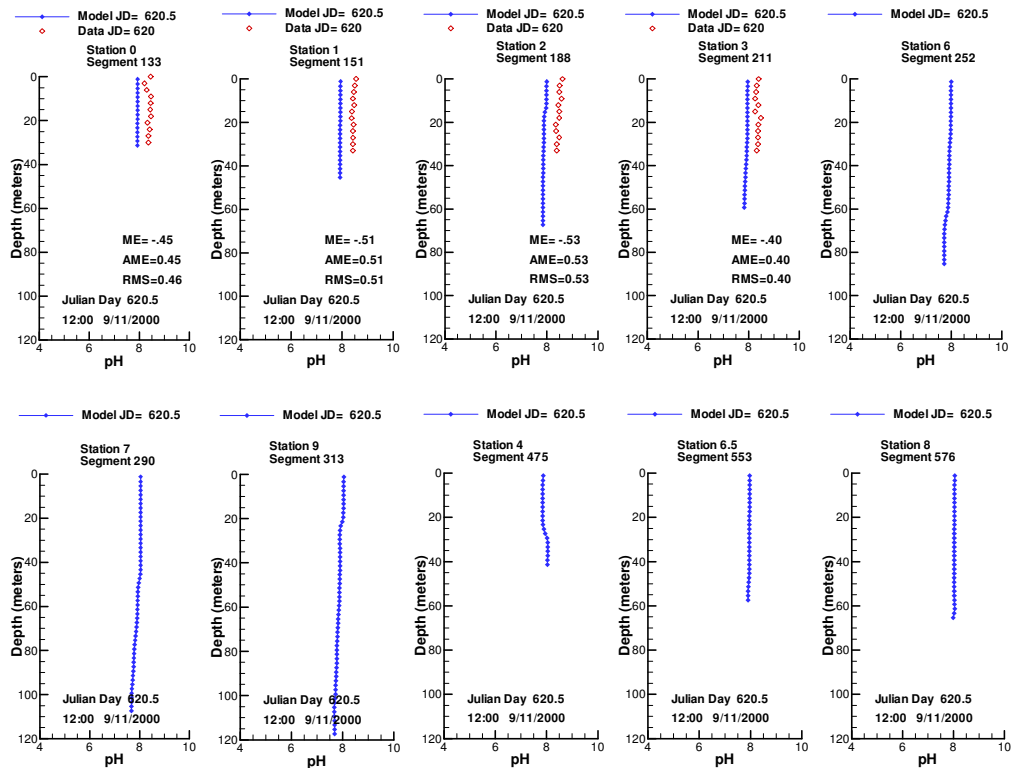


Figure 366. Vertical pH model-data comparison, J620.

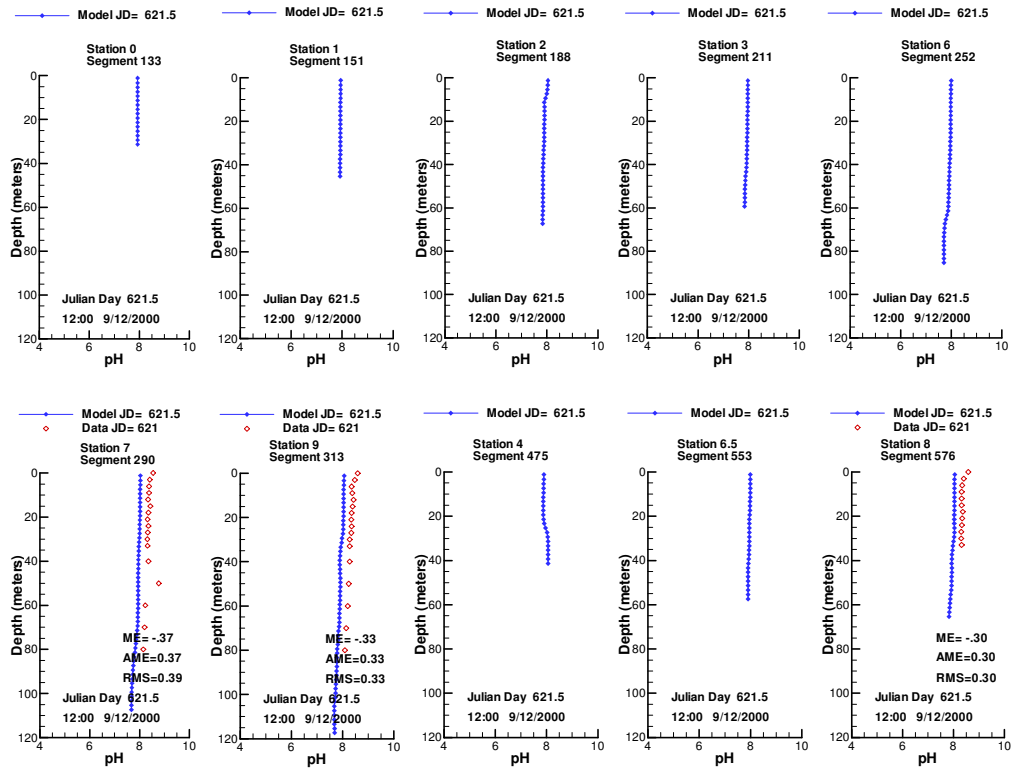


Figure 367. Vertical pH model-data comparison, J621.

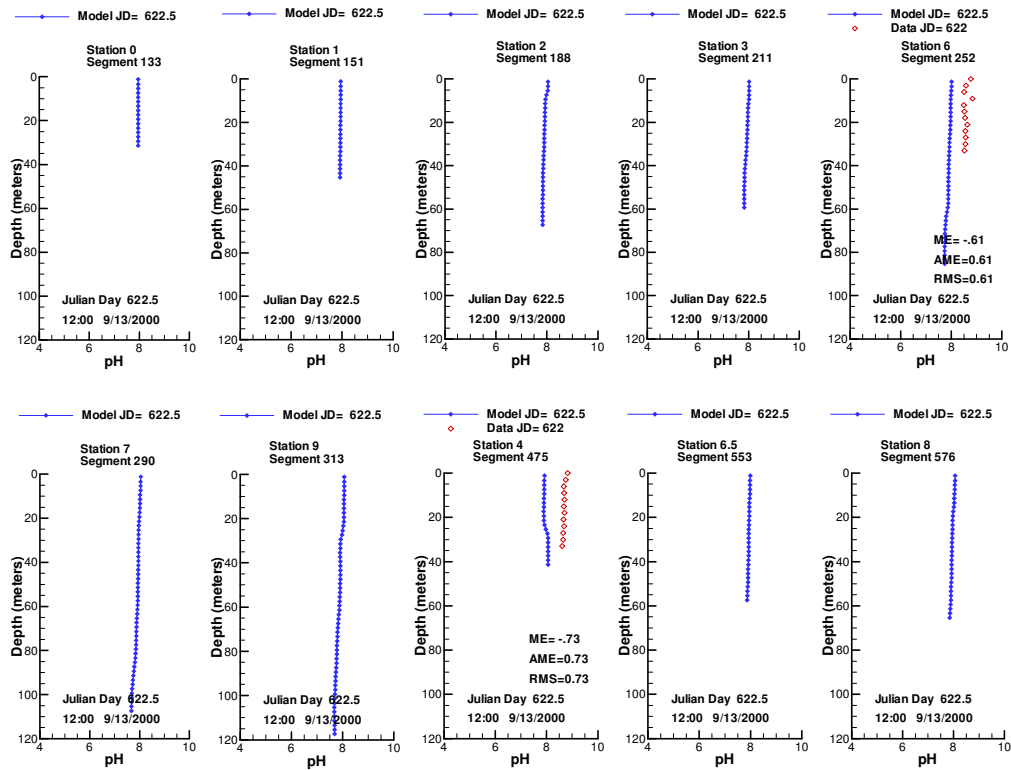


Figure 368. Vertical pH model-data comparison, J622.

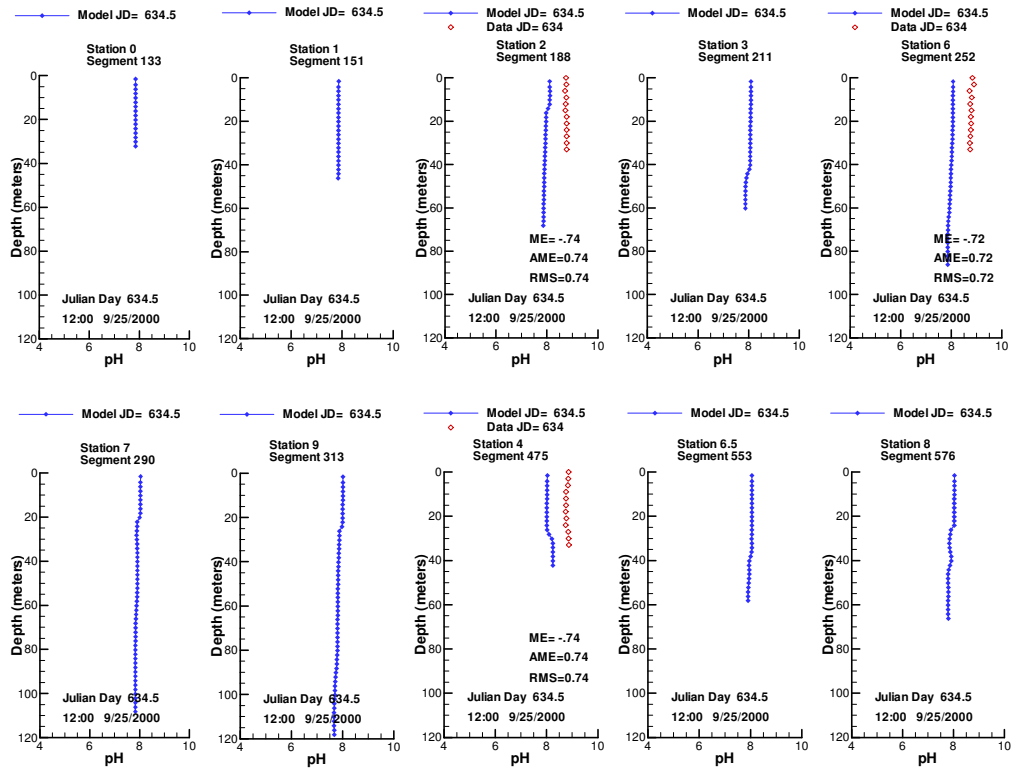


Figure 369. Vertical pH model-data comparison, J634.

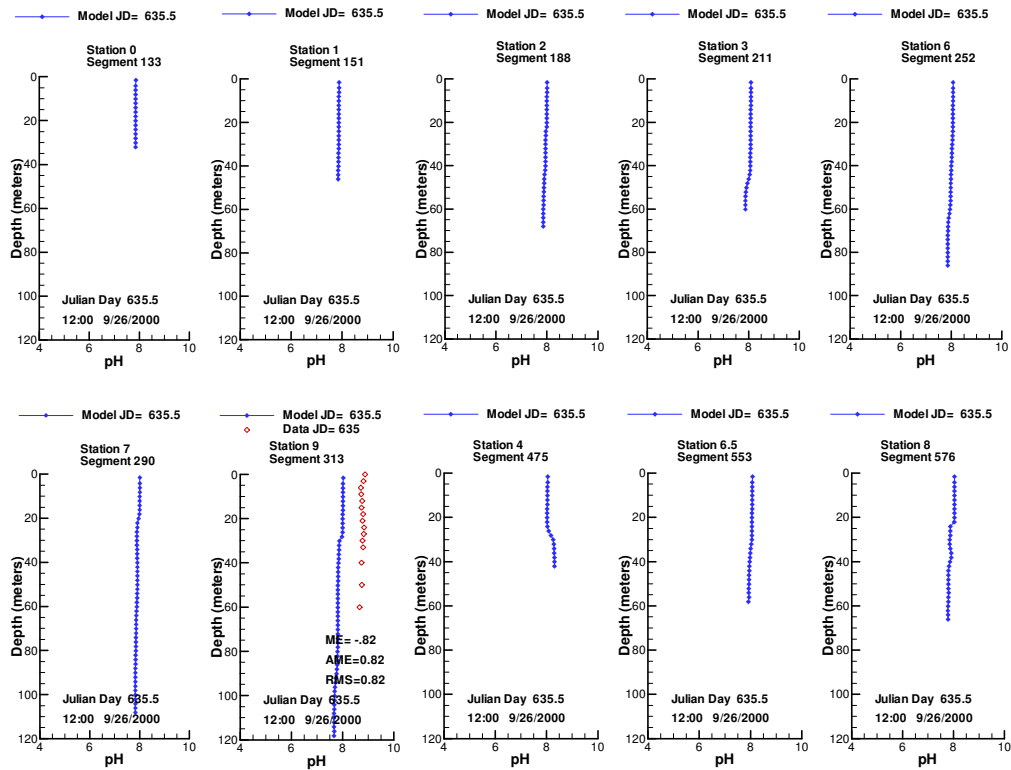


Figure 370. Vertical pH model-data comparison, J635.

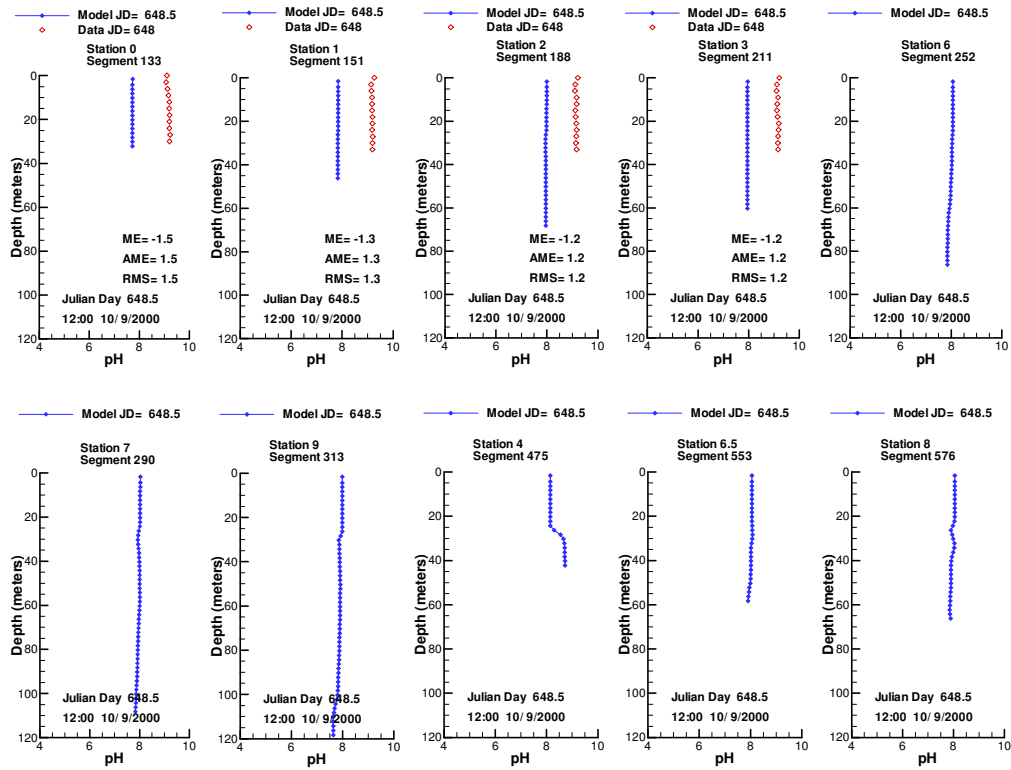


Figure 371. Vertical pH model-data comparison, J648.

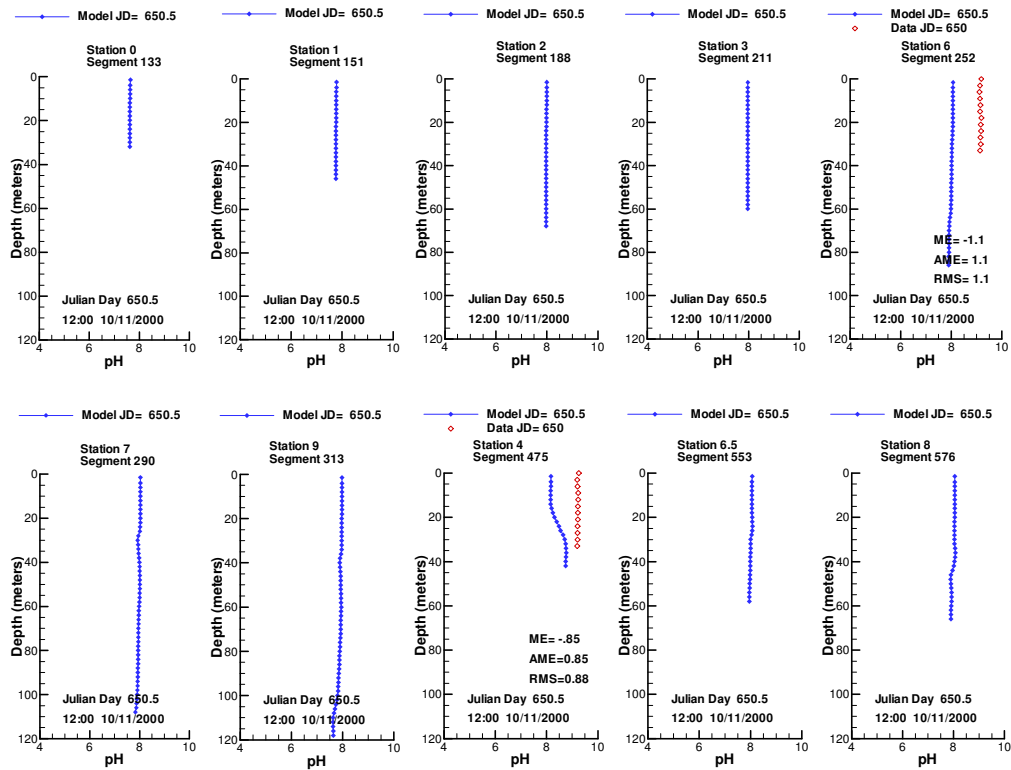


Figure 372. Vertical pH model-data comparison, J650.

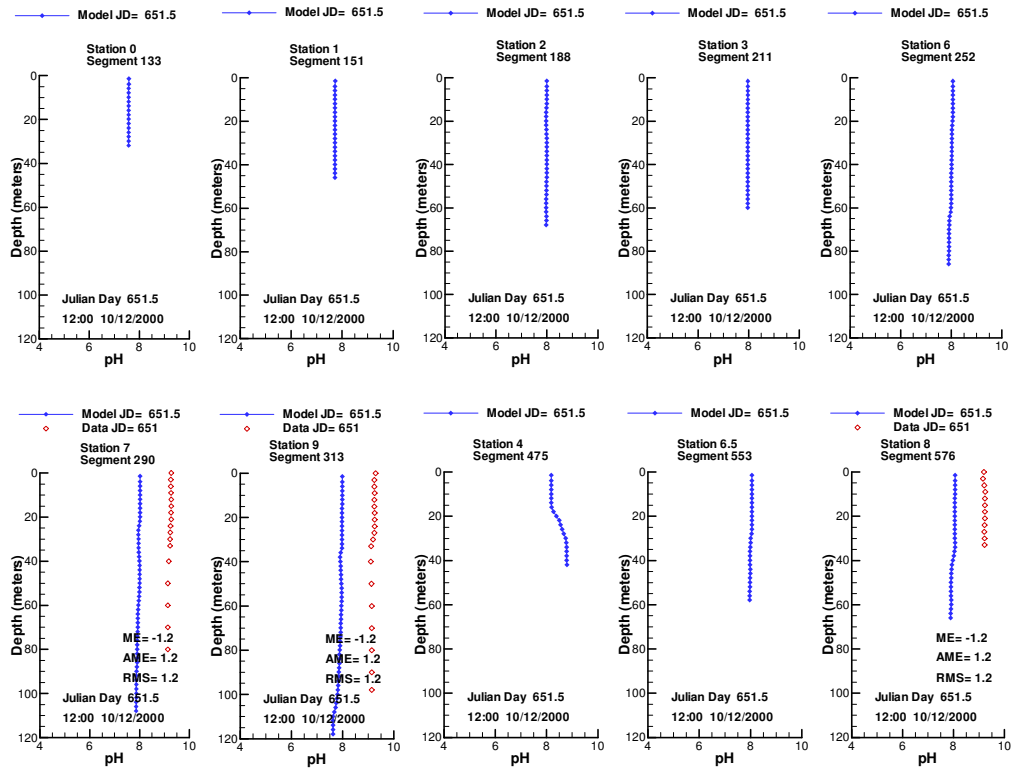


Figure 373. Vertical pH model-data comparison, J651.

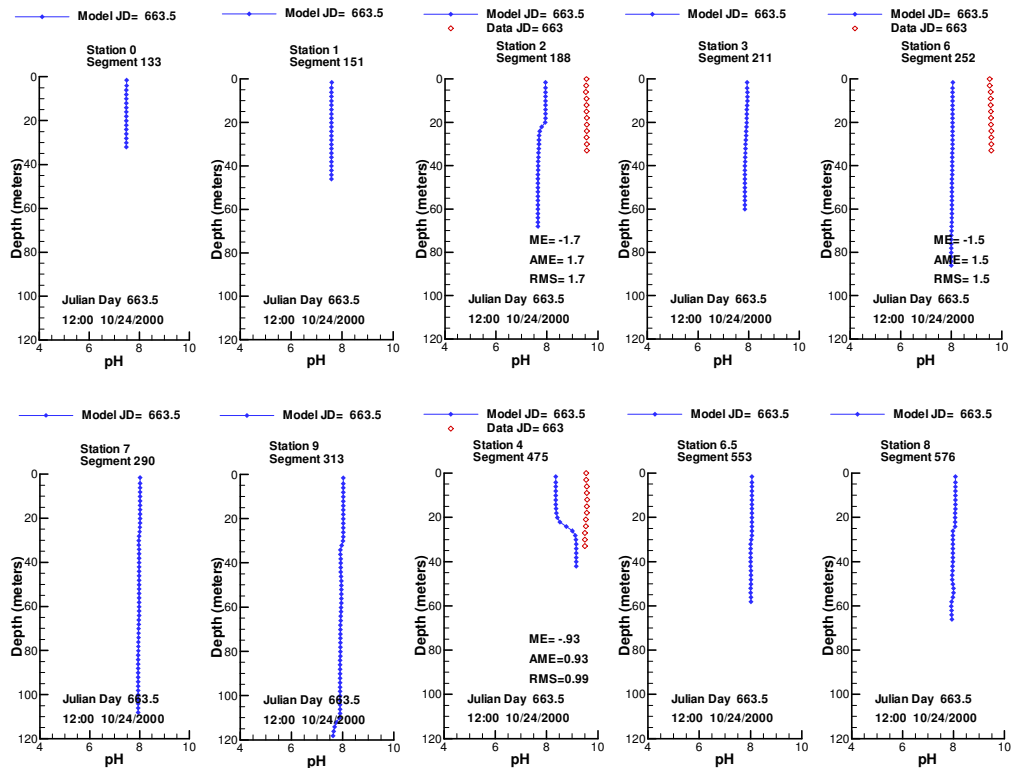


Figure 374. Vertical pH model-data comparison, J663.

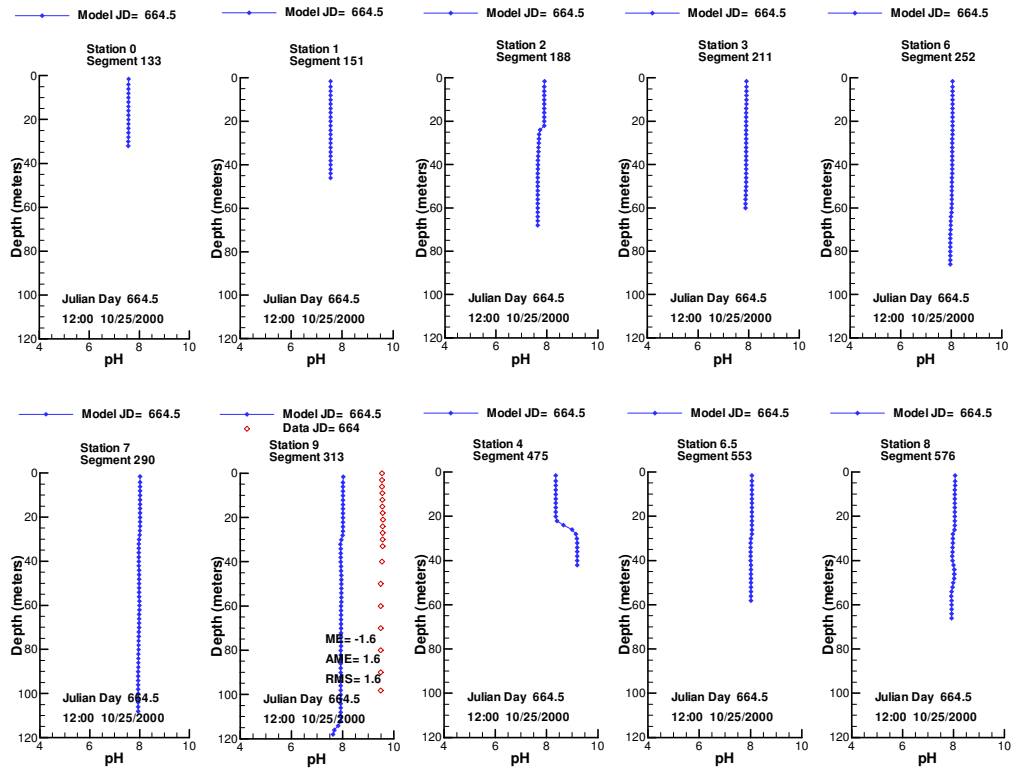


Figure 375. Vertical pH model-data comparison, J664.

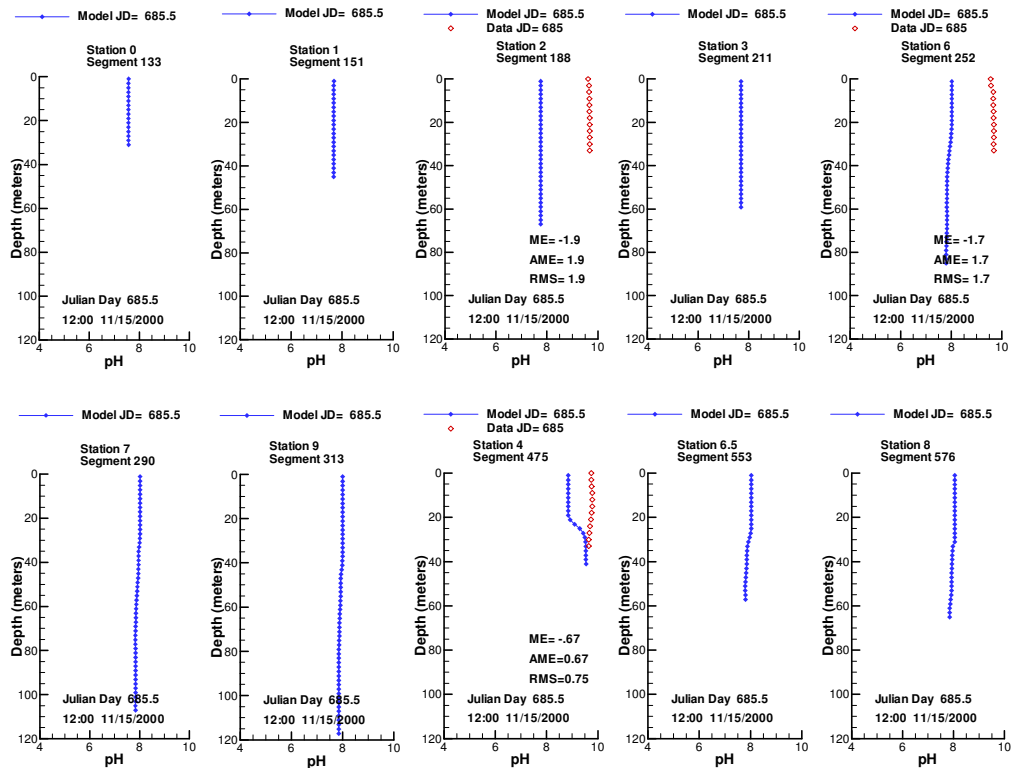


Figure 376. Vertical pH model-data comparison, J685.

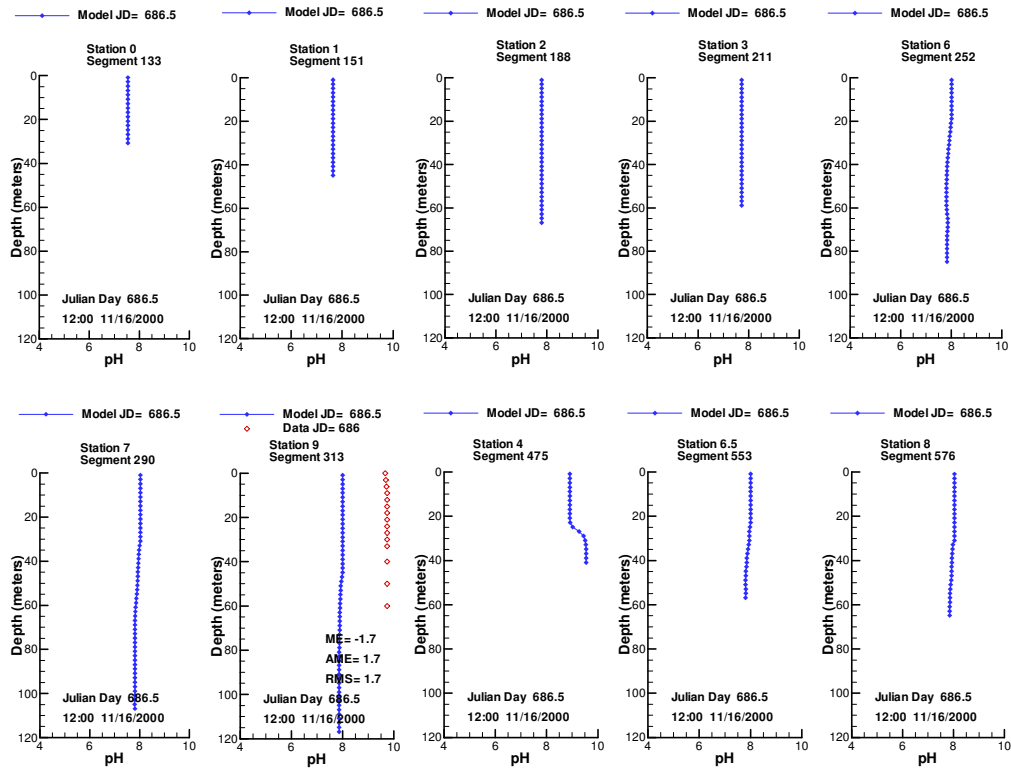


Figure 377. Vertical pH model-data comparison, J686.

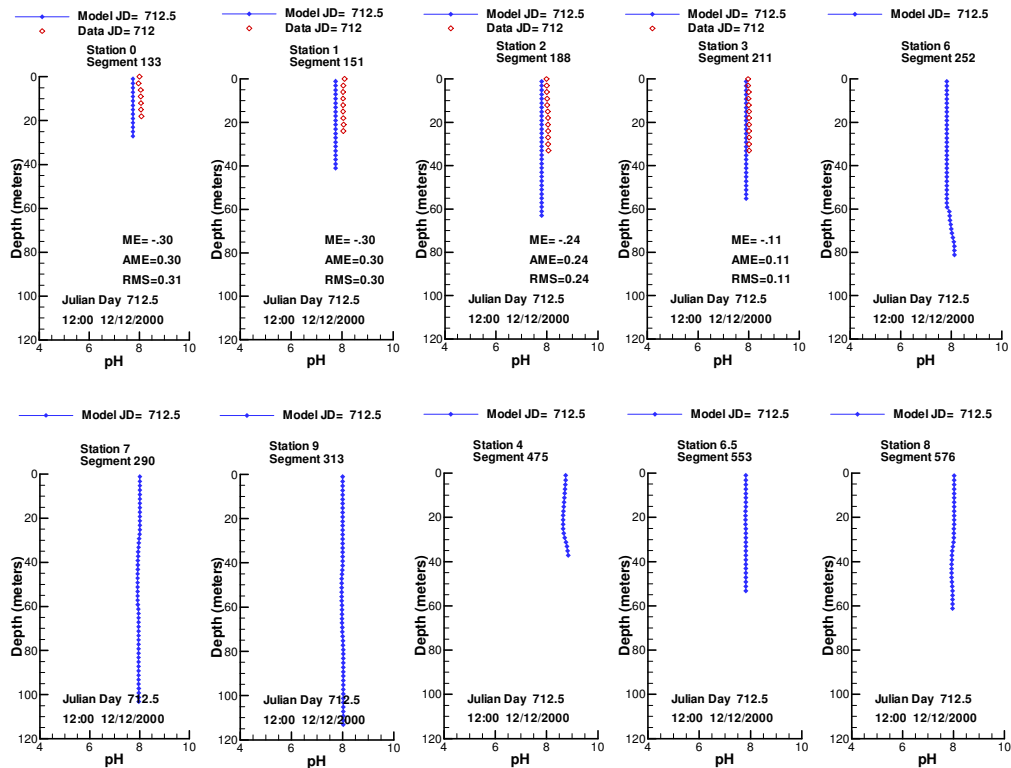


Figure 378. Vertical pH model-data comparison, J712.

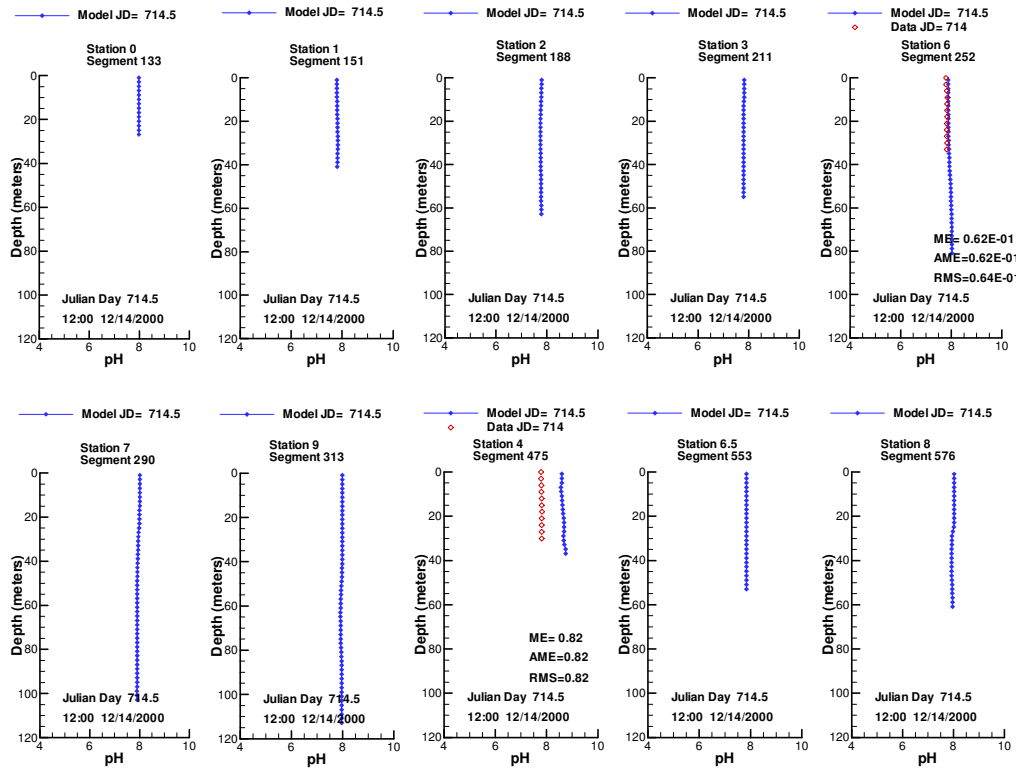


Figure 379. Vertical pH model-data comparison, J714.

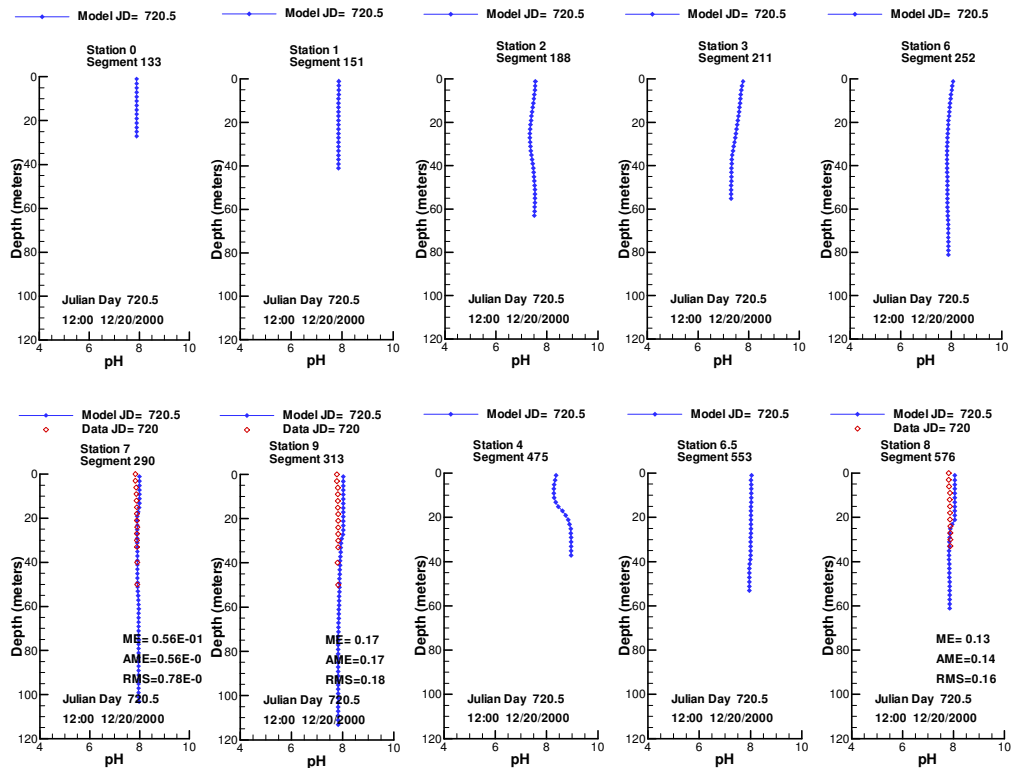


Figure 380. Vertical pH model-data comparison, J720.

Appendix G: Statistics Calculations

Model-data error statistics were computed using formulas for the mean error, absolute mean error, and root mean square error. The mean error (ME) was calculated as

$$ME = \frac{\sum_1^n (\text{model} - \text{data})}{n} \quad (19)$$

The absolute mean error was calculated as

$$AME = \frac{\sum_1^n \text{abs}(\text{model} - \text{data})}{n} \quad (20)$$

The root mean square error was calculated as

$$RMS = \sqrt{\frac{\sum_1^n (\text{model} - \text{data})^2}{n}} \quad (21)$$

where n is the number of observations, model is the model predicted state variable and data is the field data variable.

Appendix H: W2 Model Water Quality Parameters

Table 32 summarizes the model water quality parameters used. When applicable, typical literature values are included. For variables with multiple entries, such as EXA1, the last number always refers to the constituent group. Thus EXA1 is the extinction due to organic algal type 1. Similarly, PREF A23 is the feeding preference of zooplankton group 3 on algal group 2.

For clarity,

- Algal group 1 represents mixed diatoms, dominated by *asterionella formosa*.
- Algal group 2 represents mixed green algae.
- Algal group 3 represents mixed cyanobacteria.
- Zooplankton group 1 represents herbivorous copepoda
- Zooplankton group 2 represents omnivorous copepoda
- Zooplankton group 3 represents cladocera, dominated by *Daphnia*

Table 32. W2 Model Water Quality Parameters Summary

| Variable | Description | Units | Typical values* | Calibration Values |
|---|---|----------------------|-----------------|--------------------|
| Hydrodynamics and Longitudinal Transport | | | | |
| AX | Longitudinal eddy viscosity (for momentum dispersion) | m ² /sec | 1 | 1 |
| DX | Longitudinal eddy diffusivity (for dispersion of heat and constituents) | m ² /sec | 1 | 1 |
| Temperature | | | | |
| CBHE | Coefficient of bottom heat exchange | Wm ² /sec | 0.30 | 0.30 |
| TSED | Sediment (ground) temperature | °C | 12.8 | 11.5 |
| WSC | Wind sheltering coefficient | | 1.0 | 0.4-2.0 |
| BETA | Fraction of incident solar radiation absorbed at the water surface | | 0.45 | 0.45 |
| Water Quality | | | | |
| EXH20 | Extinction for water | /m | 0.25 | 0.25 |
| EXSS | Extinction due to inorganic suspended solids | m ³ /m/g | 0.01 | 0.1 |
| EXOM | Extinction due to organic | m ³ /m/g | 0.17 | 0.1 |

| Variable | Description | Units | Typical values* | Calibration Values |
|-------------------------|--|---------------------|-----------------|----------------------|
| | suspended solids | | | |
| EXA1 EXA2 EXA3 | Extinction due to organic algal type 'n' | m ³ /m/g | 0.1 | 0.2 0.2 0.2 |
| | | | | |
| SSS | Suspended solids settling rate | m/day | 2 | 1.0 |
| AG1 AG2 AG3 | Algal growth rate for algal type 'n' | /day | 1.1 | 2.5 2.5 1.0 |
| AM1 AM2 AM3 | Algal mortality rate for algal type 'n' | /day | 0.01 | 0.08 0.10 0.08 |
| AE1 AE2 AE3 | Algal excretion rate for algal type 'n' | /day | 0.01 | 0.04 0.04 0.04 |
| AR1 AR2 AR3 | Algal dark respiration rate for algal type 'n' | /day | 0.02 | 0.04 0.04 0.10 |
| AS1 AS2 AS3 | Algal settling rate for algal type 'n' | /day | 0.14 | 0.50 0.05 0.10 |
| ASAT1 ASAT2 ASAT3 | Saturation intensity at maximum photosynthetic rate for algal type 'n' | W/m ² | 150 | 85 36 75 |
| APOM1 APOM2 APOM3 | Fraction of algal biomass lost by mortality to detritus for algal type 'n' | | 0.8 | 0.50 0.50 0.50 |
| AT11 AT12 AT13 | Lower temperature for algal growth for algal type 'n' | °C | 10 | 1.0 1.0 10.0 |
| AT21 AT22 AT23 | Lower temperature for maximum algal growth for algal type 'n' | °C | 30 | 10.0 10.0 33.0 |
| AT31 AT32 AT33 | Upper temperature for maximum algal growth for algal type 'n' | °C | 35 | 20.0 20.0 35.0 |
| AT41 AT42 AT43 | Upper temperature for algal growth for algal type 'n' | °C | 40 | 28.0 25.0 40.0 |
| AK11 AK12 | Fraction of algal growth rate at ALGT1 for algal | | 0.1 | 0.20 0.20 |

| Variable | Description | Units | Typical values* | Calibration Values |
|-------------------------------|--|-------|-----------------|-------------------------------|
| AK13 | type 'n' | | | 0.05 |
| AK21 AK22 AK23 | Fraction of maximum algal growth rate at ALGT2 for algal type 'n' | | 0.99 | 0.99 0.99 0.99 |
| AK31 AK32 AK33 | Fraction of maximum algal growth rate at ALGT3 for algal type 'n' | | 0.99 | 0.99 0.99 0.99 |
| AK41 AK42 AK43 | Fraction of algal growth rate at ALGT4 for algal type 'n' | | 0.1 | 0.1 0.1 0.1 |
| ALGP-A1 ALGP-A2 ALGP-A3 | Stoichiometric equivalent between organic matter and phosphorus for algal type 'n' | | 0.011 | 0.003 0.002 0.008 |
| ALGN-A1 ALGN-A2 ALGN-A3 | Stoichiometric equivalent between organic matter and nitrogen for algal type 'n' | | 0.08 | 0.034 0.040 0.080 |
| ALGC-A1 ALGC-A2 ALGC-A3 | Stoichiometric equivalent between organic matter and carbon for algal type 'n' | | 0.45 | 0.450 0.535 0.475 |
| LDOMDK | Labile DOM decay rate | /day | 0.12 | Spokane: 0.1 Others: 0.01 |
| LRDDK | Labile to refractory DOM decay rate | /day | 0.001 | 0.01 |
| RDOMDK | Maximum refractory decay rate | /day | 0.001 | 0.001 |
| LPOMDK | Labile Detritus decay rate | /day | 0.06 | Spokane: 0.25 Others: 0.05 |
| LRPDK | Labile to refractory POM decay rate | /day | 0.01 | 0.01 |
| POMS | Detritus settling rate | m/day | 0.35 | 0.1 |
| RPOMDK | Refractory Detritus decay rate | /day | | 0.005 |
| OMT1 | Lower temperature for organic matter decay | °C | 4 | 4 |
| OMT2 | Lower temperature for maximum organic matter decay | °C | 20 | 30 |
| OMK1 | Fraction of organic matter decay rate at OMT1 | | 0.1 | 0.1 |
| OMK2 | Fraction of organic matter | | 0.99 | 0.99 |

| Variable | Description | Units | Typical values* | Calibration Values |
|-------------------------|---|------------------|------------------------|---------------------------|
| | decay rate at OMT2 | | | |
| SEDK | Sediment decay rate | /day | 0.06 | 0.1 |
| PARTP | Phosphorous partitioning coefficient for suspended solids | | 1.2 | 0.2 |
| AHSP1 AHSP2 AHSP3 | Algal half-saturation constant for phosphorous – algae 1 | g/m | 0.009 | 0.002 0.003 0.003 |
| NH4DK | Ammonia decay rate (nitrification rate) | /day | 0.12 | 0.12 |
| AHSN | Algal half-saturation constant for ammonia | g/m ³ | 0.014 | 0.014 |
| NH4T1 | Lower temperature for ammonia decay | °C | 5 | 5 |
| NH4T2 | Lower temperature for maximum ammonia decay | °C | 20 | 25 |
| NH4K1 | Fraction of nitrification rate at NH4T1 | | 0.1 | 0.1 |
| NH4K2 | Fraction of nitrification rate at NH4T2 | | 0.99 | 0.99 |
| NO3DK | Nitrate decay rate (denitrification rate) | /day | 0.102 | 0.03 |
| NO3S | Denitrification rate from sediments | m/day | 1.0 | 1.0 |
| NO3T1 | Lower temperature for nitrate decay | °C | 5 | 5 |
| NO3T2 | Lower temperature for maximum nitrate decay | °C | 20 | 25 |
| NO3K1 | Fraction of denitrification rate at NO3T1 | | 0.1 | 0.1 |
| NO3K2 | Fraction of denitrification rate at NO3T2 | | 0.99 | 0.99 |
| O2NH4 | Oxygen stoichiometric equivalent for ammonia decay | | 4.57 | 4.57 |
| O2OM | Oxygen stoichiometric equivalent for organic matter decay | | 1.4 | 1.4 |
| O2AR | Oxygen stoichiometric equivalent for dark respiration | | 1.4 | 1.1 |
| O2AG | Oxygen stoichiometric | | 1.4 | 1.4 |

| Variable | Description | Units | Typical values* | Calibration Values |
|-------------------------------|---|---------------------|-----------------|-------------------------|
| | equivalent for algal growth | | | |
| ORGP | Stoichiometric equivalent between organic matter and phosphorus | | 0.011 | 0.05 |
| ORGN | Stoichiometric equivalent between organic matter and nitrogen | | 0.08 | 0.05 |
| ORGC | Stoichiometric equivalent between organic matter and carbon | | 0.45 | 0.45 |
| O2LIM | Dissolved oxygen concentration at which anaerobic processes begin | g/m ³ | 0.1 | 0.1 |
| EXZ1 EXZ2 EXZ3 | Extinction due to organic zooplankton type 'n' | m ³ /m/g | 0.1 | 0.2 0.2 0.2 |
| ZMAX1 ZMAX2 ZMAX3 | Maximum zooplankton growth rate for zooplankton type 'n' | /day | 1.5 | 2.5 2.5 2.5 |
| ZRESP1 ZRESP2 ZRESP3 | Maximum zooplankton respiration rate for zooplankton type 'n' | /day | 0.1 | 0.1 0.1 0.1 |
| ZMORT1 ZMORT2 ZMORT3 | Maximum zooplankton mortality rate for zooplankton type 'n' | /day | 0.01 | 0.1 0.1 0.1 |
| ZEFFIC1 ZEFFIC2 ZEFFIC3 | Zooplankton assimilation efficiency or the proportion of food assimilated to food consumed for zooplankton type 'n' | | 0.5 | 0.6 0.6 0.6 |
| PREFP1 PREFP2 PREFP3 | Preference factor of zooplankton for detritus or LPOM for zooplankton type 'n' | | 0.5 | 0.1 0.1 0.1 |
| ZOONIN1 ZOONIN2 ZOONIN3 | Threshold food concentration at which zooplankton feeding begins for zooplankton type 'n' | g/m ³ | 0.01 | 0.001 0.001 0.001 |
| ZS2P1 ZS2P2 ZS2P3 | Zooplankton half-saturation constant for food (includes LPOM, algae, and zooplankton) ingestion for | g/m ³ | 0.3 | 0.3 0.5 0.3 |

| Variable | Description | Units | Typical values* | Calibration Values |
|-------------------------------|---|-------|-----------------|----------------------|
| | zooplankton type 'n' | | | |
| PREFA11 PREFA12 PREFA13 | Preference factor of zooplankton type 'n' for algae type 1 | | | 0.35 0.30 0.35 |
| PREFA21 PREFA22 PREFA23 | Preference factor of zooplankton type 'n' for algae type 2 | | | 0.55 0.40 0.55 |
| PREFA31 PREFA32 PREFA33 | Preference factor of zooplankton type 'n' for algae type 3 | | | 0.00 0.00 0.00 |
| PREFZ11 PREFZ12 PREFZ13 | Preference factor of zooplankton type 'n' for zooplankton 1 | | | 0.00 0.10 0.00 |
| PREFZ21 PREFZ22 PREFZ23 | Preference factor of zooplankton type 'n' for zooplankton 2 | | | 0.00 0.00 0.00 |
| PREFZ31 PREFZ32 PREFZ33 | Preference factor of zooplankton type 'n' for zooplankton 3 | | | 0.00 0.10 0.00 |
| ZT11 ZT12 ZT13 | Lower temperature for algal growth for zooplankton type 'n' | °C | 5.0 | 1.0 1.0 1.0 |
| ZT21 ZT22 ZT23 | Lower temperature for maximum zooplankton growth for zooplankton type 'n' | °C | 25.0 | 10.0 10.0 10.0 |
| ZT31 ZT32 ZT33 | Upper temperature for maximum zooplankton growth for zooplankton type 'n' | °C | 35.0 | 25.0 25.0 25.0 |
| ZT41 ZT42 ZT43 | Upper temperature for zooplankton growth for zooplankton type 'n' | °C | 40.0 | 36.0 36.0 36.0 |
| ZK11 ZK12 ZK13 | Fraction of zooplankton growth rate at ZT1 for zooplankton type 'n' | | 0.1 | 0.30 0.30 0.30 |
| ZK21 ZK22 ZK23 | Fraction of maximum zooplankton growth rate at ZT2 for zooplankton type 'n' | | 0.99 | 0.90 0.90 0.90 |
| ZK31 ZK32 | Fraction of maximum zooplankton growth rate at | | 0.99 | 0.98 0.98 |

| Variable | Description | Units | Typical values* | Calibration Values |
|-------------------------|--|--------------|------------------------|---------------------------|
| ZK33 | ZT3 for zooplankton type 'n' | | | 0.98 |
| ZK41 ZK42 ZK43 | Fraction of zooplankton growth rate at ZT4 for zooplankton type 'n' | | 0.1 | 0.20 0.20 0.20 |
| ZP1 ZP2 ZP3 | Stoichiometric equivalent between organic matter and phosphorus for zooplankton type 'n' | | 0.005 | 0.007 0.007 0.007 |
| ZN1 ZN2 ZN3 | Stoichiometric equivalent between organic matter and nitrogen for zooplankton type 'n' | | 0.08 | 0.080 0.080 0.080 |
| ZC1 ZC2 ZC3 | Stoichiometric equivalent between organic matter and carbon for zooplankton type 'n' | | 0.45 | 0.045 0.045 0.045 |
| * Cole and Wells (2006) | | | | |